

# Legacy of Intensive Agricultural History in the Health of (Sub)Tropical Landscapes

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## Article

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**Title:** Legacy of intensive agricultural history in the health of (sub)tropical landscapes

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**Author Contributions:** An authorship matrix tabulated across four areas of contribution, each of which were split into three additional compartments and weighted. 1) Ideas: original idea, development, interpretation/intellectual input, 2) Work: field, sample analysis, data analysis, 3) Writing: preliminary thesis and draft, final manuscript, editing, and 4) Stewardship: funding, database, continued responsibility to research group. SEC (28.6% total contribution), in decreasing order across writing, ideas, stewardship, and work. HH (26.6%), across work, ideas, and preliminary thesis and draft. JLD (20.1%) ideas, stewardship, writing, and work. TMM (11.7%), across writing, work, stewardship, and ideas. CTG (7.3%), primarily in work and stewardship. EV (3.6%) in work and stewardship. JRZ (2.1%) in stewardship and ideas.

## 1 **Abstract**

2 Soil health conceptualized as a measurable ecosystem property provides a powerful tool for  
3 monitoring progress in restoration projects or implementation of best management practices to  
4 promote sustainable agroecosystems. We surveyed soils collected from a range of land uses (i.e.,  
5 protected native and non-native forest, managed pasture, unmanaged previously intensive  
6 agricultural lands, organic cropland, and conventional cropland) across a range of soil orders  
7 (Oxisol, Mollisol, Andisol, Inceptisol, and Vertisol) on three Hawaiian Islands. Forty-six metrics  
8 associated with soil health and encompassing biological, chemical, and physical properties were  
9 measured. In this multivariate survey, the most distinct group was the unmanaged, previously  
10 intensive agriculture lands, which was significantly different from all other land uses regardless  
11 of mineralogy. Importantly, the soil health of well-managed pastures in Hawai‘i was not different  
12 from protected forests, suggesting that well-managed grazing lands may be as healthy and  
13 resilient as protected forests. A suite of 11 readily measured parameters emerged out of a first-  
14 principle approach to determining a holistic indication of soil health across a range of soils and  
15 systems in Hawai‘i encompassing much of the diversity in the tropics and subtropics. Every land  
16 use may improve its soil health status within a reasonable range of expectations for a soil’s land  
17 use history, current land use, and mineralogy. Key drivers of the measures for soil health,  
18 including intensive land use history, current land use practices, and mineralogy, must be  
19 interwoven into the soil health index, which should set minimum and maximum benchmarks and  
20 weight parameters according to equitable standards.

## 21 **Significance Statement**

22 The heterogeneity of natural and working lands may be harnessed to help broadly define soil  
23 health for an expansive role in describing diverse, multi-functional and sustainable landscapes  
24 beyond the current focus on agriculture. Hawai‘i is geographically tropical and climatically  
25 subtropical; it encompasses 83% of global soil diversity and recent patterns of deforestation and  
26 extractive, intensive agriculture mirror histories worldwide. Intensive and extensive  
27 monocropping cultivation leaves a legacy of poor soil health that challenges efforts to aggrade it  
28 with improved practices or restoration. Now, with large-scale agriculture abandoned and  
29 ambitious state-level carbon and energy mandates, Hawai‘i is emerging as a model system for  
30 land-based action that simultaneously meets food, energy, and water needs while rebuilding  
31 healthy relationships between humans and managed or unmanaged landscapes.

## 32 **Introduction**

33 The degradation of tropical and subtropical soils is widespread, and degraded lands increasingly  
34 are targeted for land-based climate action and other efforts to restore ecosystem function and  
35 resilience <sup>1</sup>. Globally, the soil carbon (C) debt driven by the expansion and intensification of  
36 agriculture is substantial <sup>2</sup> and recent studies of biodiversity <sup>3</sup> and reforestation to potentially  
37 sequester C <sup>4</sup> highlight the challenges of regaining what is lost following deforestation.  
38 Conceptually, climate smart soils <sup>5</sup>, the 4 per mille initiative <sup>6,7</sup>, and other efforts aimed at  
39 building soil C as a natural climate solution (e.g., <sup>8</sup>) focus on C and climate change mitigation;  
40 but more broadly encompass soil health, which is a holistic concept with multiple co-benefits to  
41 the environment, economy, and society <sup>9,10</sup>.

42 Soil organic matter is the critical link between C sequestration and soil health. Soil organic matter  
43 is approximately 58% C and central to soil functions that are supported by biological, chemical,  
44 and physical properties and affect the balance and flow of water, nutrients, and energy through  
45 the soil ecosystem <sup>11</sup>. Measures of soil health, which include key biological, chemical, and  
46 physical properties, are connected to ecosystem services through their functional roles (such as  
47  
48  
49

50 erosion control, C storage, nutrient transformation, water filtration, and essential food, indigenous  
51 crop, forest, etc. production). Through this lens, the well-being of humans improves as a result of  
52 enhanced soil health and function, thus directly supporting a number of sustainability goals<sup>12</sup>  
53 such as UN Sustainable Development goals 2 (Zero Hunger) and 3 (Good Health and Well-  
54 Being)<sup>13</sup>. Thus, soil organic matter and healthy soils increasingly are linked to healthy societies  
55<sup>14</sup>.

56  
57 To this point, the technical discussion surrounding soil health centers primarily on agronomic  
58 systems to target improving crop yields and economic return and biological properties of the soil  
59 microbiome using innovative technology not readily accessible<sup>10,15,16</sup>. No study has yet embraced  
60 the heterogeneity of natural and working lands and multiple land use needs to broadly define soil  
61 health for a more expansive role in diverse, multi-functional landscapes. Complex, competing  
62 demands on natural and working landscapes for food, fiber, fuel, and urbanization will continue  
63 to drive sustainable development plans. Land use and management options that reconcile  
64 productivity with maintenance and enhancement of biodiversity, soil health, and associated  
65 ecosystem services in human-dominated landscapes are critical<sup>17</sup>.

66  
67 The unique diversity of tropical/subtropical soils and ecosystems (including natural and working  
68 lands, or agroecosystems) in the small geographic space of Hawai‘i is an opportunity to explore  
69 complex relationships between land use, land use history, soil type, and soil health. In Hawai‘i,  
70 the reconciliation of potentially competing issues of development, food production, and  
71 biodiversity, together with the added pressure of climate change<sup>18</sup>, is urgent. In the last few  
72 decades, large-scale plantation agriculture declined drastically, leaving large areas of abandoned  
73 agricultural lands across the islands. Current state law mandates improvements in soil health, C  
74 sequestration, and yields across agricultural sectors and forested land while in pursuit of  
75 achieving at least state-level C neutrality by 2045. But, like other regions across the tropics and  
76 subtropics, there are not yet science-based programs in place to support this outcome due, in part,  
77 to insufficient science specific to Hawai‘i’s soils and systems. Knowledge addressing this gap in  
78 Hawai‘i, serving as a model system, can be transferred to other tropical and subtropical regions.

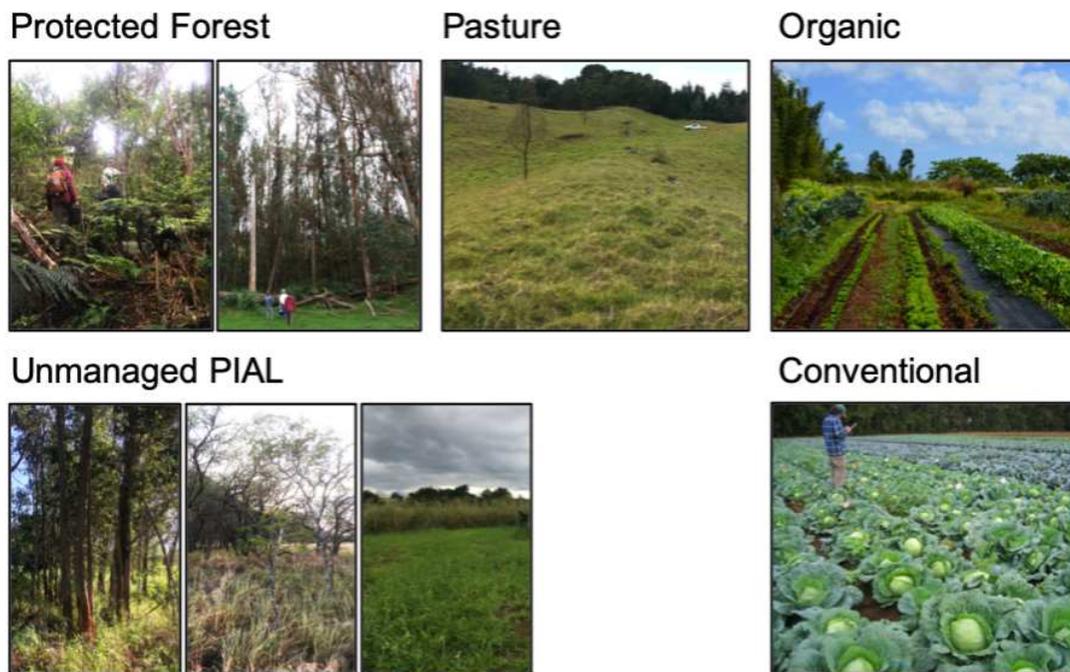
79  
80 In this context, we asked: as a dynamic ecosystem property not limited to agricultural systems,  
81 what were the predominant drivers of healthy soils and how is soil health most effectively  
82 assessed across tropical and subtropical regions and volcanic islands? We hypothesized that,  
83 within the land uses and soils studied, volcanic ash-derived soil would exhibit fundamentally  
84 different soil health characteristics than the others and that current land uses would affect soil  
85 health parameters secondarily to inherent soil differences.

## 86 87 **Results**

### 88 89 **What comprises soil health?**

90 Soil was collected from a range of land uses (identified *a priori* as protected native and non-  
91 native forest, managed pasture, unmanaged previously intensive agricultural lands (UPIAL),  
92 organic cropland, and conventional cropland) (**Fig. 1**) across a range of soil orders and clay  
93 mineralogy (Oxisol, Mollisol, Andisol, Inceptisol, and Vertisol) on three islands. When possible,  
94 pairs or triplets of sites were obtained on the same, or related, soil series but different land use.  
95 Forty-six parameters across biological, chemical, and physical soil properties were measured and  
96 analyzed with a multivariate approach to 1) test for the predominant drivers of soil health on a  
97 heterogenous landscape and 2) deduce a key set of indicators that represent soil health as an  
98 ecosystem property. Four significant principle components analysis (PCA) axes cumulatively

99 explained 71.7 % of the variance within the soil health dataset. The two dominant axes explained  
100 43.0 and 12.3 % of variance, followed by the next two that explained a further 9.0 and 7.4 %.  
101 Many of the parameters across biological, chemical and physical soil properties, strongly  
102 correlated (i.e.,  $r > 0.5$ ) to the positive or negative side of axis 1 (**Table S1**).  
103

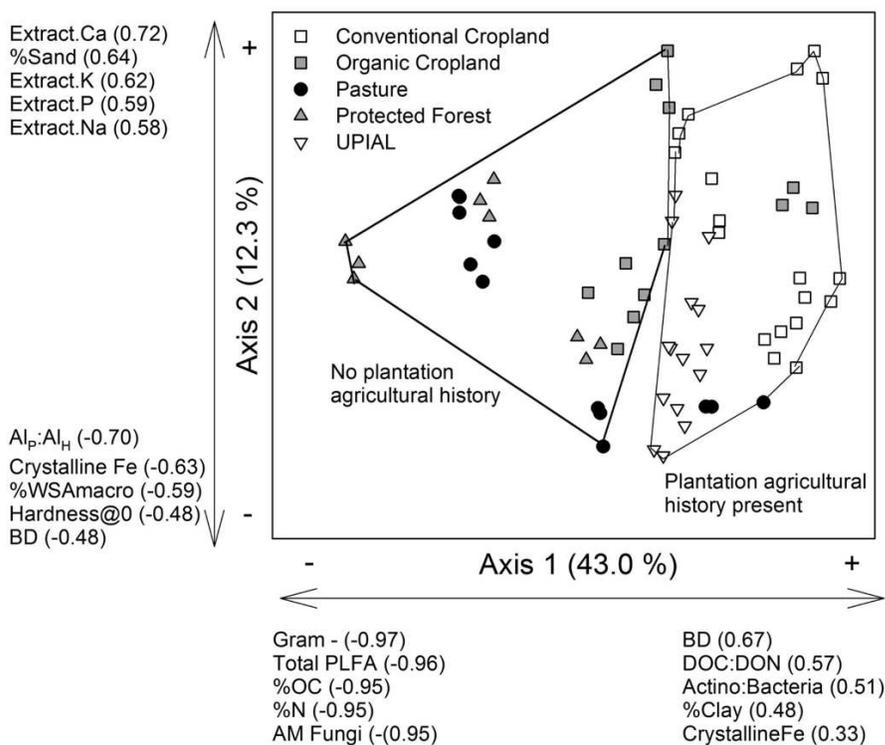


104  
105 **Figure 1.** Images from field sites in each category of current land use. Protected forest included  
106 both native (left) and non-native (right) stands. Unmanaged, previous intensive agriculture lands  
107 (PIAL) included forest stands (left), shrub lands (center), and grasslands (right). Pasture sites  
108 were managed grazing lands; croplands included organic and conventional managements.  
109

110  
111 Land use, specifically the legacy of intensive cultivation, predominated over soil type to influence  
112 soil health. Visualization of axis 1 and 2 of the PCA showed that regardless of current land use  
113 and soil type, sites with a history of long-term intensive cultivation clustered independently from  
114 other land uses within forest and pasture classifications (**Fig. 2**). Sites without a plantation  
115 agricultural history (i.e., 80+ years of sugarcane or pineapple) separated out from those sites with  
116 intensive land use history along axis one. The top five strongest (i.e.,  $r > 0.95$ ) parameters driving  
117 the sites with no intensive agricultural history toward the negative side of axis 1 were high gram  
118 negative bacteria, total phospholipid fatty acids (PLFA), organic carbon (OC) concentration, total  
119 N concentration, and actinomycetes. The negative side of axis 1 was also driven by high values of  
120 many additional biological parameters not listed as well (**Table S1**). In contrast, those sites with a  
121 plantation history had low concentration of those parameters negatively related to axis one, and  
122 high bulk density (BD), dissolved OC (DOC) to dissolved organic N (DON) ratio, actinomycetes  
123 to bacteria ratio, clay concentration, and crystalline Fe oxides. Axis 2 did not provide a clear  
124 separation among the past or current land use classifications.  
125

126 To a certain degree, some current land uses correspond to areas with an intensive agricultural past  
127 and others to areas without due to land availability and suitability. For example, the sampling

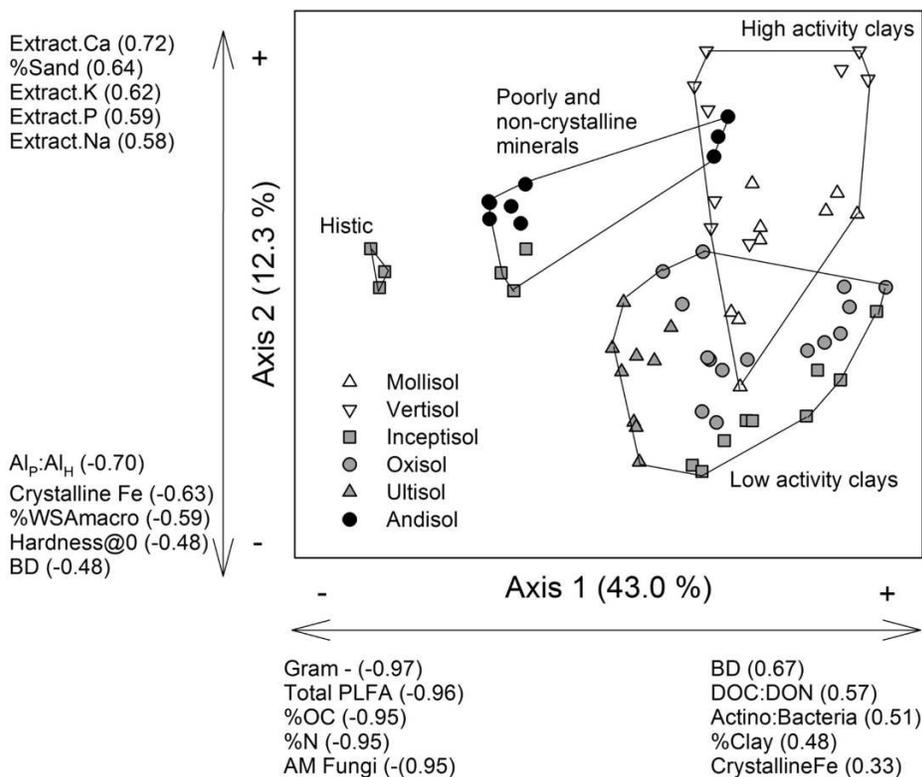
128 sites classified as conventional cropland and UPIAL (by its own definition) reside exclusively  
 129 within the cluster defined by an intensive agricultural history. Likewise, all the sampled protected  
 130 forests occurred in areas without the agricultural past. However, organic croplands and pastures  
 131 resided within both types of areas with and without the intensive agricultural past. Visualization  
 132 of axes two and three showed no separation among land use history or current status (**Fig. S1**).  
 133



134 **Figure 2.** Principal components analysis of all potential soil health indicators with the top five  
 135 most negatively and positively correlated variables to axis 1 and axis 2. Correlation values for  
 136 each parameter to the axis follow it in parentheses. Groups are delineated based on whether an  
 137 intensive plantation agricultural history is present or absent at that site. The current land use of  
 138 each site is also indicated: conventional cropland (white square), organic cropland (grey square),  
 139 pasture (black circle), protected forest (grey triangle), or unmanaged previously intensive  
 140 agriculture (UPIAL, upside down white triangle). The amount of variability explained by each  
 141 axis is in parentheses.  
 142  
 143  
 144

145 Soil type, as defined by soil order and more broadly by mineralogical class (i.e., high activity  
 146 clays, low activity clays, poorly and non-crystalline minerals, and histic) was not strongly  
 147 associated with axis 1 (**Fig. 3**). Both high and low activity clays aligned with the positive side of  
 148 axis 1 associated with high BD, DOC:DON, clay and crystalline minerals. However, high activity  
 149 clays included Vertisols and Mollisols, and associated with the positive side of axis two, driven  
 150 by high extractable Ca<sup>2+</sup>, K<sup>+</sup>, P, and Na<sup>2+</sup>, and sand concentration. Low activity clays included  
 151 Ultisols and Oxisols, and associated more with the negative side of axis 2, driven by high ratio of

152 pyrophosphate to hydroxylamine extractable Al, crystalline Fe oxides, concentration of mega size  
 153 class water stable aggregates, soil hardness at the surface layers, and BD. Histic soils fell on the  
 154 negative side of Axis 1 which correlated highly with total OC, total N, Actino:Bacteria, and AM  
 155 Fungi. Visualization of axis 2 and 3 helped further separate out the histic and poorly and non-  
 156 crystalline mineral (PNCM) groups (**Fig. S1**). Particularly with respect to the Andisols and andic  
 157 Inceptisols having high concentration of poorly and non-crystalline minerals ( $Al_H+0.5Fe_H$ ),  
 158 PNCM separation from other soils was driven by silt and sand concentration, fungi to bacteria  
 159 ratio, and concentration of “mega” size class water stable aggregates.  
 160



161  
 162  
 163 **Figure 3.** Principal components analysis of all potential soil health indicators with the top five  
 164 most negatively and positively correlated variables to axis 1 and axis 2. Correlation values for  
 165 each parameter to the axis follow it in parentheses. Groups are delineated based on broad  
 166 mineralogical categorization at that site. The soil order of each site is indicated as Mollisol  
 167 (white triangle), Vertisol (upside down white triangle), Inceptisol (grey square), Oxisol (grey  
 168 circle), Ultisol (grey triangle), and Andisol (black circle). The amount of variability explained by  
 169 each axis is in parentheses.  
 170  
 171

### 172 **Predominant drivers of the measures for soil health**

173 Multiple multi-response permutation procedure (MRPP), a nonparametric multivariate test of  
 174 differences between groups<sup>19</sup>, models were run among the categorical classifications and selected  
 175 combinations for hypothesis testing to determine the drivers of soil health parameters (**Table S2**).

176 The combination of agricultural history and disturbance level was the second most significant  
177 contrast (adjusted  $p = 0.008$ ,  $A$  value = 0.2636). The significant pairwise comparisons indicate  
178 that those sites classified as PIAL-medium are different than those PIAL-high and none-high;  
179 none-low is different than PIAL-medium and none-high. These results indicate first that sites with  
180 a plantation history and medium disturbance classification were more like one another than to  
181 those with a high disturbance level (i.e., currently in intensive cultivation), regardless of whether  
182 there was plantation past land use or not. Second, undisturbed sites with no plantation history  
183 were more like one another than to those with a plantation history and medium current  
184 disturbance classification or those with no plantation history but currently under intensive  
185 practices. The combination of agricultural history and current land use was also among the  
186 significant contrasts tested and yielded similar results to the interaction with plantation history  
187 and disturbance level. In the pairwise comparisons, those sites classified as PIAL and currently  
188 unmanaged (UPIAL) are different from PIAL and currently in conventional cropland, and from  
189 sites with no plantation history and currently in organic cropland, pasture, or protected forest. The  
190 simpler current land use model had the third highest  $A$  value (0.2583) and 10 pairwise contrasts  
191 that separated from one another along the dominant PCA axis.

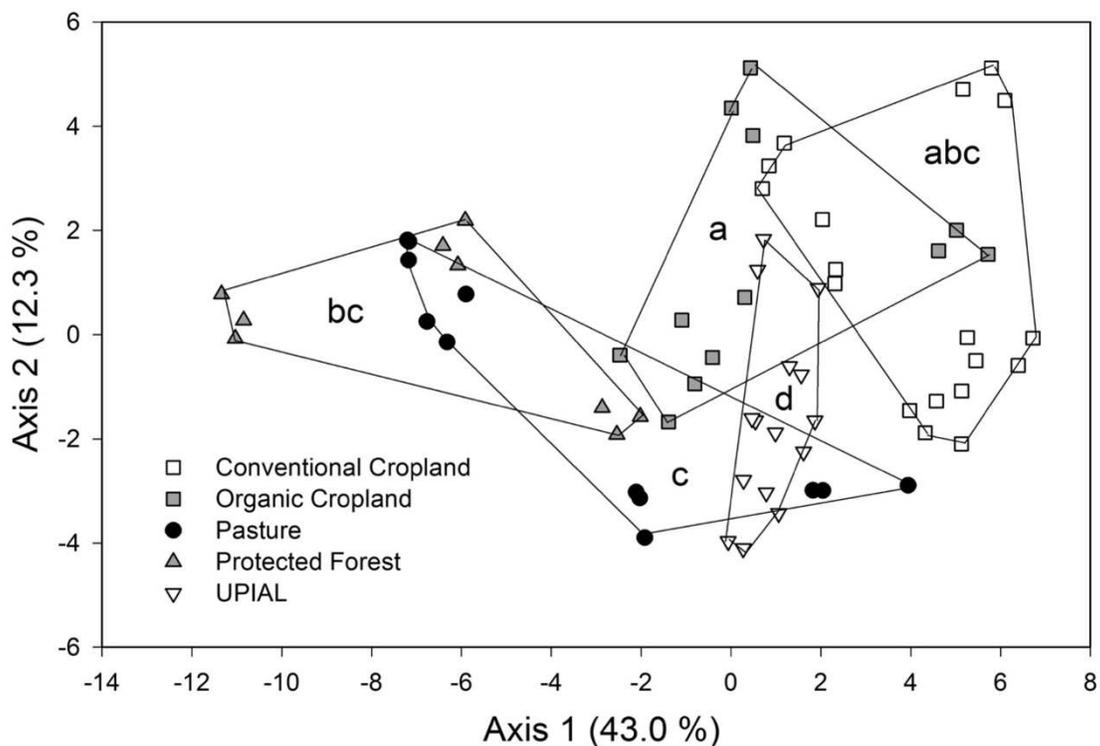
192  
193 The most distinct group was the UPIAL land use class, which was significantly different from all  
194 other land uses (**Fig. 4**). Pasture sample units spanned axis 1 yet were significantly different than  
195 UPIAL and organic croplands in multivariate space indicating a latent interaction with  
196 mineralogy unable to be explored further within the constraints of our dataset. Pastures were not  
197 different from protected forests and the two groups showed a lot of overlap in the 2-dimensional  
198 visualization of axis 1 and 2, also signaling a potential influence of similar mineralogy. In the  
199 case of conventional agriculture, the sample units were so dispersed within the group, that no  
200 differences emerged between it and the other current land uses, except for UPIAL. However,  
201 organic croplands were similar enough to one another to be significantly different than UPIAL,  
202 pastures, and protected forests.

203  
204 The combination of cropland and mineral classifications also was among the significant contrasts  
205 tested (**Table S2**) and provided additional insight into the nature of the interaction of minerals  
206 with land use by sub-setting the dataset to reduce the complexity of the five current land use  
207 classes to simply cropland (organic and conventional) and not cropland (UPIAL, pasture and  
208 protected forest). The non-cropland PNCM sites were different from non-cropland HAC and  
209 LAC. For both HAC and LAC, those sites in croplands were different from those not in cropland.  
210 Within the constraints of the dataset, which has greater coverage of HAC and LAC across the  
211 cropland/not cropland designations, being in cropland affected soil health for both HAC and  
212 LAC. Additionally, HAC in cropland was different from LAC not in cropland, and vice versa.

### 213 214 **Key soil health indicators for tropical/subtropical soils**

215 Eleven dynamic soil parameters emerged from a multi-step dimension reduction process as  
216 indicators of ecosystem health across diverse land uses, histories, and soil types. First, at its  
217 foundation PCA is a dimension reduction approach, and 26 parameters correlated strongly ( $\geq 0.5$   
218 or  $\leq -0.5$ ) with axis one. A correlation matrix of the untransformed values showed covariance  
219 among many of those 26 parameters (**Fig. S2**). From this covarying block, consideration of the  
220 practicality of the parameter's inclusion in a rapid, accessible soil health index (i.e., cost and  
221 difficulty) and coverage of biological, chemical, and physical parameters further reduced the list  
222 to CO<sub>2</sub> burst, HWEC, PMN, total OC %, and WHC. Among the parameters not included in the  
223 block:  $\beta$ -glucosidase,  $\beta$ -glucosaminidase, and water stable mega aggregates (mega WSA), the

224 DOC to DON ratio, actinomycetes to bacteria ratio, and BD remained. Because actinomycetes to  
 225 bacteria ratio is not feasible in a rapid soil health test, it was removed from the final list.  
 226



227  
 228 **Figure 4.** Axes 1 and 2 of the principal components analysis for all potential soil health indicators  
 229 including the multi-response permutation procedures results comparing the multivariate within  
 230 and between group testing among the current land uses. Conventional cropland (white square),  
 231 organic cropland (grey square), pasture (black circle), protected forest (grey triangle), or  
 232 unmanaged previously intensive agriculture (UPIAL, upside down white triangle). The amount  
 233 of variability explained by each axis is in parentheses. Groups with different letters have  
 234 statistically greater similarity within the group than to others in multivariate space.  
 235  
 236

237 With the criteria of strong relationship to PCA axis 1, non-covariance, practicality, and inclusion  
 238 of biological, chemical, and physical parameters, 11 parameters emerged as potential indicators  
 239 of a soil health gradient across the soils and ecosystems in Hawai'i (**Table 1**). Summary values  
 240 show the range, mean, and median of each parameter across the dataset (**Table 2**). For  
 241 contextualization, all parameters except pH and BD are greater in soils without an intensive  
 242 plantation history than in those with. For several of these parameters (total OC %, CO<sub>2</sub> burst, β-  
 243 glucosidase, β-glucosaminidase, PMN, and HWEC), protected forests are greater than  
 244 conventional croplands, while the DOC to DON ratio was lower. Similar differences were also  
 245 present for pasture compared to conventional cropping except for a few (PMN and HWEC).  
 246 Among those parameters with significant contrasts, total OC %, CO<sub>2</sub> burst, PMN, and HWEC are  
 247 the same for UPIAL versus conventional cropland, but β-glucosidase and β-glucosaminidase are  
 248 greater while the DOC to DON ratio was lower for UPIAL than conventional cropland. There

249 were fewer differences between organic and conventional cropping, but organic cropping had  
 250 greater total OC%, and  $\beta$ -glucosaminidase than conventional.  
 251

**Table 1.** After assessing the sensitivity, interpretation value, and feasibility (i.e., resources required for field collection and laboratory assays), the recommended indicators to use in a routine soil health test for Hawai‘i and potentially other tropical-subtropical and volcanic regions, were reduced to 11 parameters.

<u>Proposed Hawai‘i Soil Health Indicators</u>	
Parameter	Function and interpretation
Total organic carbon (%)	As the backbone of soil organic matter, a proxy measurement of the amount of soil organic matter; higher value typically relates to benefits of multiple biological, chemical, and physical aspects of soil function
<u>Biological Properties</u>	
24 hr CO <sub>2</sub> burst ( $\mu\text{g g}^{-1}$ )	Soil respiration in response to readily available substrate; higher value indicates high microbial activity and high quality organic matter pools
$\beta$ -glucosidase (mg p-nitrophenol kg <sup>-1</sup> soil h <sup>-1</sup> )	Proximate microbial metabolism of amino-containing substrate; higher value indicates nutrient, predominantly N, mineralization
$\beta$ -glucosaminidase (mg p-nitrophenol kg <sup>-1</sup> soil h <sup>-1</sup> )	Potential N supply; higher value indicates bioavailable N forms to support soil productivity
Mineralizable nitrogen ( $\mu\text{g g}^{-1}$ )	Potential N supply; higher value indicates bioavailable N forms to support soil productivity
<u>Chemical Properties</u>	
pH	Biological and nutrient availability; 6.0—7.0 is ideal, this is the pH range where plant essential elements are most available, and toxicities are negligible
DOC:DON ratio	Integrated indicator of the balance of organic carbon and organic nitrogen pools; lower is better; higher value indicates disturbance - high DOC indicates available microbial substrate but also potential runoff, priming, and loss if too high, DON is readily broken down by soil microbes into inorganic forms, but low values are associated with N-deposition or poor nutrient management in disturbed systems
Hot water extractable carbon ( $\mu\text{g g}^{-1}$ )	Readily available metabolic substrate; higher value indicates soluble organic matter and lysed microbial cells that support microbial activity
<u>Physical Properties</u>	
Water holding capacity (%)	Plant-water relations; higher values indicate improved water storage
Water stable mega-aggregates (%)	Water infiltration, porosity, aeration; higher values improve retention/transport water, promote root growth, provide habitat for microbes, reduce bulk density, and resist erosion
Bulk density (g cm <sup>-3</sup> )	Infiltration, porosity, and rooting environment; lower values indicate soils that are light, aerated, porous, promote root growth, and more workable

Table 2. Data summary for the proposed key indicators of health for subtropical/tropical and volcanic soils. Significant differences are indicated by \* for the past land use (PIAL versus None) comparison and letters for the current land use comparisons (Tukey-adjusted comparison, p-value <0.05).

	Min	Max	Mean	Median	None n=27	PIAL n=39	Protected Forests n= 9	Pasture n=12	Unmgd PIAL n=15	Organic Cropland n=12	Conv Cropland n=18
%OC	0.83	32.5	5.73	2.33	11.0 ± 3.34*	2.10 ± 0.39*	18.5 ± 7.84 <sup>a</sup>	8.58 ± 3.55 <sup>a</sup>	2.20 ± 0.15 <sup>ab</sup>	3.09 ± 0.91 <sup>a</sup>	2.12 ± 1.25 <sup>b</sup>
CO <sub>2</sub> Burst	13.3	527.1	102.6	51.4	195.2 ± 50.79*	37.2 ± 5.22*	274.5 ± 126.7 <sup>a</sup>	177.8 ± 69.5 <sup>a</sup>	52.1 ± 8.87 <sup>ab</sup>	69.5 ± 19.0 <sup>ab</sup>	30.5 ± 8.67 <sup>b</sup>
β-Gluc	20.7	230.5	92.1	83.8	119.5 ± 18.24*	72.3 ± 12.4*	117.3 ± 32.2 <sup>a</sup>	131.1 ± 37.6 <sup>a</sup>	113.6 ± 18.9 <sup>a</sup>	78.2 ± 9.86 <sup>ab</sup>	44.9 ± 14.6 <sup>b</sup>
β-Glucmin	7.71	134.1	47.6	39.0	77.5 ± 12.8*	27.7 ± 4.72*	81.3 ± 26.4 <sup>a</sup>	90.9 ± 19.9 <sup>a</sup>	44.52 ± 5.31 <sup>a</sup>	33.6 ± 7.15 <sup>a</sup>	13.9 ± 2.13 <sup>b</sup>
PMN	0.00	304.8	41.1	20.3	83.7 ± 28.7*	11.6 ± 2.88*	152.8 ± 76.2 <sup>a</sup>	54.3 ± 15.1 <sup>ab</sup>	21.7 ± 3.48 <sup>bc</sup>	23.2 ± 8.67 <sup>bc</sup>	4.70 ± 2.79 <sup>c</sup>
pH	3.71	7.86	6.44	6.7	6.43 ± 0.39	6.42 ± 0.26	6.04 ± 1.17	6.51 ± 0.32	6.21 ± 0.39	7.13 ± 0.27	6.32 ± 0.57
DOC:DON	2.03	808.9	169.1	38.0	94.1 ± 89.4*	203.0 ± 51.6*	2.68 ± 0.45 <sup>b</sup>	17.0 ± 13.9 <sup>b</sup>	169.5 ± 110.8 <sup>b</sup>	313.7 ± 191.9 <sup>ab</sup>	257.0 ± 73.3 <sup>a</sup>
HWEC	48.4	13,400	1096.5	331.6	2378.1 ± 1390.7*	197.3 ± 29.8*	5245.0 ± 4085.6 <sup>a</sup>	1001.1 ± 444.0 <sup>ab</sup>	297.2 ± 27.1 <sup>ab</sup>	466.3 ± 143.0 <sup>ab</sup>	172.0 ± 59.1 <sup>b</sup>
WHC	56.7	208.5	85.2	69.2	108.5 ± 16.7*	69.7 ± 2.14*	136.7 ± 40.3	97.9 ± 23.8	67.4 ± 2.49	76.0 ± 5.37	72.0 ± 5.6
%WSA <sub>mega</sub>	0.00	96.9	96.9	47.1	67.4 ± 11.0*	29.5 ± 8.54*	73.2 ± 15.0	79.9 ± 12.7	46.4 ± 14.9	29.6 ± 15.7	24.4 ± 18.2
BD	0.22	1.19	0.84	0.91	0.69 ± 0.10	0.94 ± 0.05	0.54 ± 0.20	0.80 ± 0.20	1.01 ± 0.04	0.86 ± 0.03	0.85 ± 0.11

PIAL = previously intensive agricultural lands; None = no plantation history; %OC = Total organic carbon; β-Gluc = β-glucosidase; β-Glucmin = β-glucosaminidase; PMN = Potentially mineralizable nitrogen; DOC:DON = DOC to DON ratio ; HWEC = Hot water extractable carbon; WHC = water holding capacity; %WSA<sub>mega</sub> = Water stable mega-aggregates; BD = bulk density; Unmgd = Unmanaged; Conv = Conventional

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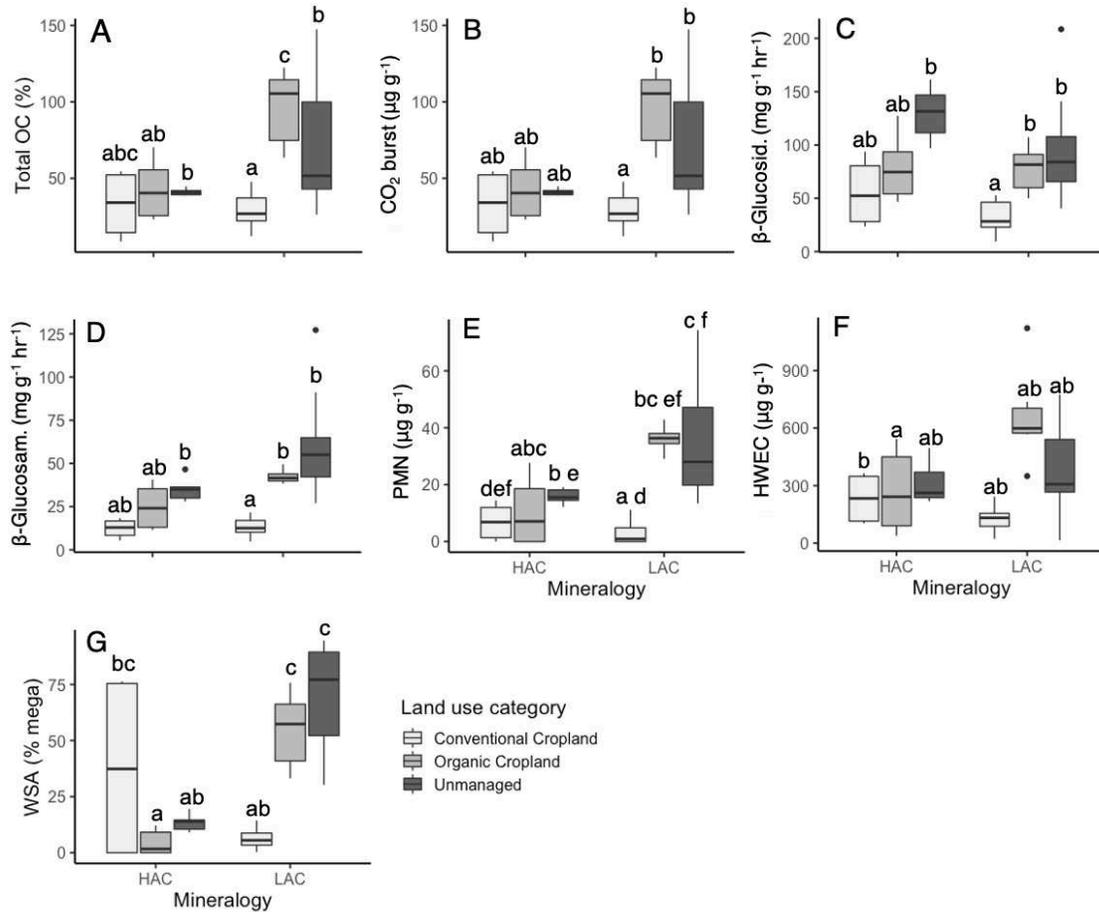
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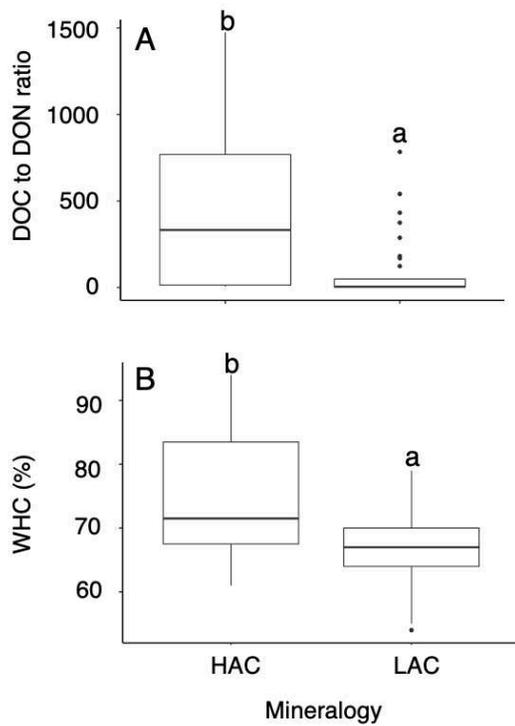
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For the subset of soils with adequate representation across mineralogy (HAC and LAC) and current land use (UPIAL, organic cropland, and conventional cropland) the interactive effects of these parameters on several soil health indicators, total OC %, CO<sub>2</sub> burst, β-glucosidase and β-glucosaminidase, PMN, HWEC, and mega-WSA, were significant (**Fig. 5**). For LAC soils, many parameters were consistently lower for conventional than organic croplands, including OC %, CO<sub>2</sub> burst, β-glucosidase, β-glucosaminidase, PMN, and mega-WSA. In general, UPIAL tended to have lower values than organic management, but greater than conventional cropland, which was especially apparent for total OC %. In contrast to LAC soils, very few significant effects of land use on HAC soils were detected, and only PMN and HWEC was less in conventional than in organic croplands while mega-WSA were greater. Of those parameters without significant interactions, mineralogy was significant for the DOC to DON ratio and WHC, and in both cases HAC was greater than LAC (**Fig. 6**). Current land use was significant for the DOC to DON ratio

266 (conventional > UPIAL), WHC (organic > UPIAL and conventional croplands), and BD (UPIAL  
 267 > organic croplands) (**Fig. 7**). Mineralogy and land use had no detectable effect on soil pH.  
 268

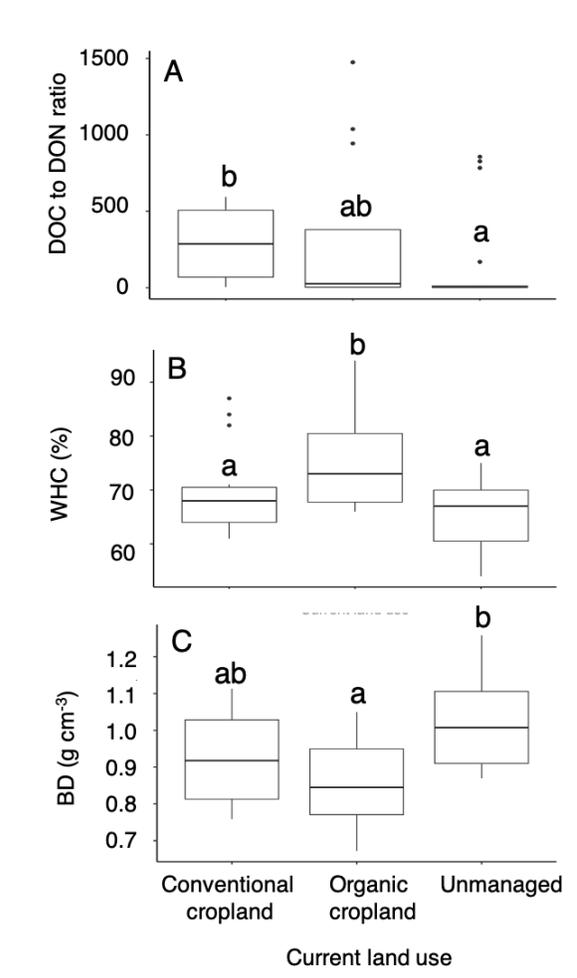


269  
 270 **Figure 5.** Boxplots for the proposed key indicators of health for a subset of high (HAC) and low  
 271 activity clay (LAC) soils under land use classes for conventional cropland, organic cropland, and  
 272 unmanaged previously intensive agriculture. Means sharing a letter are not significantly different  
 273 (Tukey-adjusted comparison of least square means at p-value < 0.05.) n=51  
 274  
 275



276

277 **Figure 6.** Boxplots for the proposed key indicators of health for a subset of high (HAC) and low  
 278 activity clay (LAC) soils averaged across land use classes. Means sharing a letter are not  
 279 significantly different (Tukey-adjusted comparison of least square means at p-value < 0.05.) n=51  
 280



281

282 **Figure 7.** Boxplots for the proposed key indicators of health for a subset of soils across land use  
 283 classes for conventional cropland, organic cropland, and unmanaged previously intensive  
 284 agriculture averaged across high (HAC) and low activity clay (LAC) soils . Means sharing a letter  
 285 are not significantly different (Tukey-adjusted comparison of least square means at p-value <  
 286 0.05.) n=51  
 287

288 **Discussion**

289

290 **Soil health is an ecosystem property**

291 Soil health conceptualized as a measurable ecosystem property provides a powerful tool for  
292 monitoring progress in restoration projects or implementation of best management practices to  
293 promote sustainable agroecosystems. A new paradigm of soil organic matter dynamics, which is  
294 central to soil health, is driving the development of new compartmental models tied to  
295 measurable soil parameters<sup>20</sup>. Soil organisms, particularly microbial communities that are  
296 proximately responsible for the flow of nutrients, C, and energy in the soil ecosystem, rely on  
297 accessible organic matter for metabolic substrate. Many of the emergent process-based ecosystem  
298 models are microbial models (e.g.,<sup>21,22</sup>) and hold promise to improve both decision support tools  
299 and earth system projections.

300

301 In policy and programs intended to incentivize maintaining and aggrading soil health across  
302 multifunctional landscapes and diverse stakeholders, expectations must be gauged accordingly.  
303 Long-term, intensive monocrop agriculture, which in Hawai‘i was predominantly pineapple and  
304 sugarcane plantations established post-Western contact, leaves a detrimental legacy on soil  
305 health. The adverse effects on soil biological properties and microbial communities persists  
306 following both abandonment and land use/management change to practices consistent with soil  
307 health management principles (e.g., perennial grasses or crops, organic matter inputs, and no or  
308 reduced tillage). Especially because the legacy of intensive cultivation history may carry-over  
309 into the success of land-based initiatives now and into the future, it is important to understand the  
310 resultant differences in baseline conditions as well as the limitations to improvements in soil  
311 health when building decision support tools and programs.

312

313 Soil health metrics interwoven with process-based ecosystem models that underpin decision  
314 support tools used by policy makers may also assist in accessing aid and improved economic  
315 outcomes that are critical to success in overcoming adoption barriers. Ecosystem functions such  
316 as greenhouse gas emission, C storage, nutrient transformation, biomass production, and  
317 regulation of hazards and extreme events link directly to key services contributing to human well-  
318 being<sup>12,23</sup>. Conventional soil organic matter models (e.g., RothC, DNDC, EPIC, and  
319 CENTURY) embedded within established decision support tools that assist land managers and  
320 policy makers alike contain some of these ecosystem functions. However, their ability to  
321 simulate (sub)tropical or volcanic soils which have very different properties from the temperate  
322 or continental soils for which they were developed currently limits their usefulness.  
323 Alternatively, microbial models such as (MEMS v1.0) developed from the Microbial Efficiency-  
324 Matrix Stabilization framework incorporates measurable parameters that constrain C pools sizes  
325 and modulate fluxes<sup>22</sup>. Measurable pools and rate modifiers in MEMS v1.0 include microbial  
326 biomass and turnover, dissolved organic matter, sorption/desorption dynamics, and exoenzyme  
327 activity. The overlap between these parameters and the measures of soil health suggest that  
328 compartmental models designed to simulate soil organic SOM matter dynamics and nutrient and  
329 GHG fluxes may benefit from the integration of soil health into their initialization and projections  
330 for (sub)tropical soils.

331

332 In the process of aggrading soil health, landscapes regain resilience through improved soil  
333 functions. For Hawai‘i and other (sub)tropical and volcanic regions, land-based management  
334 relating to conservation, biocultural restoration, climate action for mitigation/adaptation, and  
335 increased local food production strive for their specific goals. But, also contribute more broadly  
336 to sustainability when viewed through the lens of improved soil health and the associated

337 expansive network of co-benefits and regulation services. Understanding, representing, and  
338 projecting outcomes associated with soil health is critical to incorporating their full value into  
339 complex watershed-based management and interdisciplinary social-ecological forecasting that  
340 link directly to building resilient, climate ready landscapes and communities.

341

342 **Land use history, current land use practices, and mineralogy are predominant drivers of**  
343 **ecosystem soil health**

344 Key drivers of the measures for soil health, including land use history, current land use practices,  
345 and mineralogy, must be understood and integrated into the development of a soil health index.  
346 Any index should set minimum and maximum benchmarks and weigh parameters according to  
347 equitable standards. Therefore, the state of each driver (e.g., timeline of intensive use history,  
348 time since implementation of current land use and management, and predominant mineralogy)  
349 must be ascertained and recorded in databases designed for syntheses of soil health into the  
350 future.

351

352 The legacy of a plantation history is a strong driver of soil health, but greater complexity  
353 associated with current land use, management, and soil type also is present and important to  
354 understand while developing a robust soil health index for the (sub) tropics. For example, the  
355 level of disturbance in current management practices and, outside of croplands, mineralogy both  
356 also affected soil health. Results suggest that soil health may differ inherently for high versus low  
357 activity clays and whether a system is cultivated intensively for food production (cropland) under  
358 conventional versus organic management affects soil health regardless of mineralogy.

359

360 Importantly, the soil health of pastures was not different from protected forests, suggesting that  
361 well-managed grazing lands may be as healthy and resilient as protected forests. However, the  
362 most distinct group was the unmanaged, previously intensive agriculture lands (UPIAL land use  
363 class), which was significantly different from all other land uses. Unmanaged abandoned  
364 agricultural lands are more similar to each other than to sites that remain in intensive cultivation.  
365 But, they are also more similar to each other than to sites without plantation history and currently  
366 are in organic croplands, pasture, or protected forest. Upon further inspection, the univariate  
367 analysis suggests that, while the relationship between organic versus conventional cropland was  
368 largely consistent across soil type for most soil health parameters, abandoned cropland was more  
369 variable. For UPIAL, in some cases, soil health indicators fell in between organic and  
370 conventional croplands, while sometimes aligning more closely to conventional or organic for  
371 other indicators. This finding further highlights the imprint that intensive agriculture may have on  
372 the health of a soil and demonstrates the constraints to rebuilding soil health upon the cessation of  
373 soil disturbance without proactive management strategies.

374

375 **Proposed “Hawai‘i Soil Health Indicators” for (sub)tropical and volcanic soils and systems**

376 A suite of readily measured parameters emerged out of a first-principle approach to determining a  
377 holistic indication of soil health across a range of soils and systems in Hawai‘i encompassing  
378 much of the diversity in the tropics and subtropics. These parameters integrate biological,  
379 chemical, and physical properties with key functions associated with soil C and nutrient cycling,  
380 water relations, and generally, the provisions of a soil environment conducive to a diverse soil  
381 organismal community. These parameters are consistent with current measures of soil health, but  
382 developed with a more organic and equitable process, without carry over of ingrained bias. In the  
383 development of a soil health index, parameters may be weighted differentially for systems.  
384 Further, cropping systems should be paired with additional fertility testing and nutrient  
385 management for optimal environmental and yield outcome. Every land use may improve its soil

386 health status within a reasonable range of expectations when considering land use history, current  
387 land use, and mineralogy.

388

389 The measurement of soil health as a dynamic ecosystem property is only possible by properly  
390 identifying the right suite of parameters specific to a region and metering that measurement to  
391 appropriate benchmarks for a system defined by past land use, current land use, and mineralogy.  
392 Moving forward, providing a soil health index of (sub)tropical and volcanic soils will help to  
393 assist currently underserved producers and land managers improve the health and productivity of  
394 their lands and simultaneously reap co-benefits of a healthier environment and society. Within  
395 this framework, fair and equitable programs can be established to improve economic outcomes as  
396 well as C neutrality goals.

397

### 398 **Conclusion**

399 Land use, particularly where a legacy of intensive cultivation existed, predominated soil health  
400 metrics, which supports continued policies and programs that help incentivize producers and land  
401 managers to implement best practices. Because of the close association of soil health and C  
402 cycling, climate change mitigation is a powerful co-benefit of improving soil health in degraded  
403 systems. As Amundson and Biardeau<sup>24</sup> put forward, “soil carbon sequestration is an elusive  
404 climate mitigation tool.” However, soil health is a more inclusive measure of the holistic value of  
405 improving the state of a natural resource key to achieving multiple sustainability goals  
406 worldwide.

407

408 Competing demands for food, fiber, fuel, and urbanization will continue. In Hawai‘i, especially,  
409 competing land uses associated with development, food production, and biodiversity under  
410 climate change is a pressing issue. Improved land use projections are critical for reducing  
411 uncertainties in indicators for ecosystem services in a changing environment<sup>25</sup>. Land use and  
412 management options that reconcile production with maintenance and enhancement of  
413 biodiversity, soil health and associated ecosystem services in human dominated landscapes now  
414 and into the future are critical. We conclude that soil health is a measurable ecosystem property  
415 and that land use history, current land use practices, and mineralogy are all predominant drivers  
416 of soil health in landscapes. Our proposed “Hawai‘i Soil Health Indicators” may be further  
417 validated for (sub)tropical and volcanic soils and systems and are critical to developing regionally  
418 appropriate incentives programs and policy.

419

### 420 **Materials and Methods**

421 Site selection and general approach Twenty-two sites were selected across three islands (Oahu,  
422 Maui, Molokai) within the main Hawaiian archipelago to cover a diversity of soil management,  
423 fertility, and taxonomy to maximize the variance in parameters associated with soil health.  
424 Preliminary assessments helped categorize sites into 1) current land management/use, 2) land use  
425 history, 3) disturbance level, 4) soil order, and 5) predominant mineralogy. Current land  
426 management/uses included protected forests (managed to preserve long-term non-native or native  
427 forest, greater than 100 years no disturbance from feral ungulates), unmanaged previously  
428 intensive agricultural lands (UPIAL, previously monocrop plantation with no current  
429 management system, grasses, shrubs, or forest as dominate cover, and less than 100 years no  
430 disturbance), pasture (managed with pasture grasses for rearing livestock), organic croplands (no  
431 use of chemical pesticides), and conventional croplands (use of chemical pesticides). Land use  
432 history indicated simply whether an intensive plantation history (for Hawai‘i, this is typically  
433 sugarcane or pineapple) was present or absent. Disturbance level was defined categorically as  
434 low, medium, and high (**Table 3**). Soil order was according to final GPS coordinates of sample

435 location and NRCS NCSS taxonomic classification (Web Soil Survey). Predominant mineralogy  
 436 was assigned using taxonomic classification and a diagnostic key (**Table S3**).  
 437

**Table 3.** Assessment of level of disturbance for each study site is categorized based on the time frames described since the most recent soil disturbance, based on available history of land use. Disturbance was considered to be land that has undergone man-made change to soil's surface layer by physical disruption of the soil structure and ecosystem, such as tillage or compaction.

Level of Disturbance	Description
Low	At least 50 years no disturbance
Medium	Disturbed in the last 50 years
High	Disturbed in the last 10 years

438  
 439 The final compilation included an integration of sites across soil types, land use history, and  
 440 natural versus agricultural landscapes that is representative of Hawai'i (**Table 4**). However, as is  
 441 reflective of reality, some soil types are more represented in some land uses and some current  
 442 land uses are more likely to be represented in one past land use history or another. Therefore, we  
 443 purposefully designed this study as a multivariate approach to identifying parameters indicative  
 444 of soil function, specifically healthy soil function, for the ecosystems of Hawai'i (and other  
 445 similar tropical/subtropical, and volcanic regions). Then, we narrowed down to key parameters  
 446 that can be linked to drivers to facilitate the next steps of developing an index of soil health and  
 447 refining the parameters for specific systems with the goal of assisting landowners, managers, and  
 448 farmers to improve the health and resilience of their lands.  
 449

**Table 4.** Summary of sample numbers for each mineralogy class and current land use.

Mineralogy	Protected		Unmanaged	Organic	Conventional
	Forest	Pasture	PIAL	Cropland	Cropland
HAC	0	0	6	6	6
LAC	3	6	9	6	9
PNCM	3	6	0	0	3
Histic	1	0	0	0	0

HAC = High activity clays; LAC = Low activity clays; PNCM = Poorly and non-crystalline minerals

450  
 451 Soil collection Three replicate samples were collected from each of the 22 sites. Each sample  
 452 was comprised of five soil cores taken from the 0-15 cm depth of mineral soil using established  
 453 sampling methods from the Cornell Soil Health Manual<sup>26,27</sup>. Briefly, the organic horizon was  
 454 removed prior to soil coring 0-15 cm of mineral soil at five locations within a 1m<sup>2</sup> quadrant. The  
 455 soil from five cores was homogenized into one sample in a bucket in the field. Three quadrants  
 456 that each produced one soil sample for a site were located at least five m apart within the site.  
 457 Thus, 66 soil samples were packaged in a cooler with ice and transported to the lab for analysis.  
 458 Samples were transported to processing and storage facilities at UH Mānoa and subsets were  
 459 frozen at -20 °C and air-dried (<10 % moisture). Additionally, a subset for phospholipid fatty acid  
 460 testing (see below) was kept chilled, not frozen, and shipped immediately under refrigeration to  
 461 the analysis facility.  
 462

463 Soil health parameters Forty-six parameters classified as biological, chemical, or physical and  
 464 tied to soil function or health were measured for each of the 66 samples (**Table 5**).

**Table 5.** Methods and functional interpretation of forty-six parameters classified as biological, chemical, or physical and tied to soil function or health were measured for each of the 66 samples.

Protocol	Methods reference(s)	Associated parameter	Function	Key functional reference(s)
Biological				
PLFA	(31)	Total PLFA	A key component of microbial cell membranes and analysis of PLFA provides a snapshot of microbial community structure; estimate of total microbial biomass	(32–34)
		Actinomycetes	Sometimes known as actinobacteria, are active in organic matter decomposition, similar filamentous growth form to fungal hyphae, bridge gaps between water films and withstand water stress	(35)
		Gram + bacteria	Dispersed in soil profile, decompose partially decayed organic matter, resistant to water stress	(34, 36, 37)
		Gram – bacteria	Plant rhizosphere member, improve plant growth by increasing solubility of nutrients, include Rhizobium which are N-fixing species	(34, 36, 37)
		Eukaryotes	Includes fungi, algae, nematodes, earthworms, insects, arthropods, and protozoa soil community members; consume organic matter, bacteria, plants, and each other	(38)
		Arbuscular mycorrhizae	Colonize roots in symbiotic relationship, grow as long, thin hyphae that extend the reach of roots into the rhizosphere, increase access to nutrients (e.g., N, P, S, Zn, and Cu) and water during drought	(34, 36, 37)
		Anaerobic bacteria	Low oxygen environments of wet or deep soils, sediments, or interior of aggregates	(34, 36, 37)
		Fungi	Particularly important lignin degraders, wide variety of forms, saprophytic fungi breakdown organic matter and release nutrients, prefer more acidic environments than bacteria	(38)
		Actino:bacteria ratio	Indicates actinomycetes taking up the role of fungi when fungal abundance is low, e.g., application of fungicide	(39, 40)
		Fungi:bacteria ratio	Significant role of fungi in litter decomposition and potentially higher C storage potential	(32, 40)
Enzyme activity	(41)	$\beta$ -glucosidase	Recycling C compounds into energy for microbes, a reliable predictor of organic matter decomposition	(42, 43)
		$\beta$ -glucosaminidase	Soil enzymes to decompose organic matter and release nutrients into plant available forms of N	(42–44)
		Acid phosphatase	Recycling of phosphorus, both important cycles related to nutrient uptake by plants	(42–44)
Potentially mineralizable N	(46)	Potentially mineralizable N	Estimate of the capacity of the soil microbes to recycle nitrogen into plant available forms	(26)
CO <sub>2</sub> burst	(47)	CO <sub>2</sub> burst	Rapid soil quality indicator of microbial activity, highly related to soil fertility	(48, 49)

Chemical				
Elemental analysis	(26)	OC concentration (%)	Associated with organic matter content, which correlates with various critical soil functions.	(26)
		N concentration (%)	Total available N resource pool	(26)
		OC to N ratio	Indicates microbial substrate quality, plant nutrient availability.	(26)
Water-extractable C	(50)	Hot water extractable carbon (HWEC)	Associated with biological activity as the hot water lyses microbe cells and releases biomass components	(49)
		% of total OC that was HWEC	Extractable carbon pool is associated with aggregate formation as well as a reserve of nutrients and energy for plants and microbes	(49)
		Total water extractable C pool (Cold water extractable organic carbon + HWEC)	Soluble and hot water lysed organic carbon	(49)
Mineralizable C pool	(51)	C pool respired in 4 months	Labile C pool readily accessible by soil microbes during laboratory incubation	(51, 52)
pH		pH	Influences essential nutrient availability, plant toxicity, and microbial community	(32, 53)
Extractable nutrients	(54)	Extractable Ca <sup>2+</sup>	Soil fertility	(54)
		Extractable K <sup>+</sup>	Soil fertility	(54)
		Extractable Na <sup>2+</sup>	Soil fertility and salinity	(54)
		Extractable P	Soil fertility	(54)
Dissolved C and N	(47, 49, 55)	Dissolved organic C	Cold water extractable organic carbon (WEOC) provides energy source for soil microbes as an active soil C pool	(26, 49, 56)
		Total dissolved N	Cold water extractable organic and inorganic forms of N are available for microbial and plant uptake	(26, 49)
		Ammonium	Microbe and plant-available nutrient	(26, 49)
		Nitrate	Microbe and plant-available nutrient	(26, 49)
		Dissolved inorganic N	Total microbe and plant-available nutrients	(26, 49)
		Dissolved organic N	Organic forms of soluble microbe and plant-available nutrients under certain environmental conditions and plant communities. Highly related to WEOC pool and contains potentially mineralizable N.	(26, 49, 56)
		Ratio of DOC to DON	Balance of C and N forms in solution, imbalance of inorganic N forms and/or DOC indicates disturbance, N deposition, or inefficient nutrient cycle	(56)
Extractable Fe and Al	(57)	Crystalline Fe-oxides	The amount of “free” secondary crystalline Fe oxides	(58)
		Poorly and non-crystalline minerals	Sorption of C to mineral surfaces, water retention	(59)
		Ratio of Al <sub>p</sub> to Al <sub>H</sub>	Controls aluminum in surface horizons of mineral soils and O-horizons of organic matter-rich soils	(60–62)

Physical				
Bulk density	(63)	Bulk density	Degree of soil compaction and potential root growth restriction and pore space.	(26)
Hardness	(63)	Hardness measured at surface	Relates to the compactness of a soil as well as the cementing features of its mineral structure.	(26)
		Hardness measured at 15 cm	Presence and penetrability of plow layer.	(26)
Water holding capacity	(64)	Water holding capacity	Vital for sustained plant growth and supporting microbial life	(26, 49)
Aggregate stability	(65)	% WSA in the mega size class	Increased water infiltration, water storage, water and gas exchange, and resistance to erosion	(66)
		% WSA in the macro size class	Store and protect organic carbon from being lost from the soil as a physical protection mechanism from microbes as well as restricting the diffusion of oxygen and enzymes	(67, 68)
Soil texture	(69)	% Sand	Water infiltration, available pore space, poor water retention	(69)
		% Silt	Water infiltration, available pore space, plant available water, soil fertility.	(69)
		% Clay	Water infiltration, plant available water, soil fertility.	(69)

465

466

467 Principal components analysis The soil health of each sample location was summarized using  
468 principal components analysis (PCA in *PC-ORD* v.7.0) The response matrix included 46  
469 parameters in 66 soil samples. Uneven distribution of data (i.e., a slight horseshoe shape with  
470 clear outliers) in the 2D output suggested the need to transform data, which was confirmed by  
471 assessing distribution tables of each variable for non-normality<sup>19</sup>. Transformations of log, cube  
472 root, and square root were tested on each highly skewed variable since all were either positively  
473 or negatively skewed with single peaks. Transformed variables were rerun for skewness and the  
474 transformation with the lowest skewness value was selected as the best possible transformation. A  
475 new PCA using variables transformed for normality showed an improved graphical display  
476 regarding spatial distribution of plots and outlier assessment.

477

478 Values from the 46 measured parameters served as the main dataset for PCA ordination, and  
479 overlays of supplemental environmental data operated as the second matrix including current land  
480 management, historical disturbance, soil order, and soil series. Potential variables were  
481 individually removed and re-added to determine their influence on the PCA and assess any  
482 significant impact on results, while using the second matrix overlays to identify issues with the  
483 distributions of data balance or flag potential errors in data manipulation. To avoid potential bias,  
484 variables problematic to overall balance were removed. Supplemental environmental data  
485 operated as the second matrix including current land management/use, land use history,  
486 disturbance level, soil order, predominant mineralogy, and combinations relevant to hypothesis  
487 testing.

488

489 The normally distributed soil health matrix was analyzed using a standardized PCA with a  
490 correlation cross-products matrix, which produces correlation coefficients among the variables  
491 and further standardizes non-comparable response variables. This method provides a broken stick  
492 eigenvalue; the broken stick eigenvalue was less than the actual eigenvalue for the first four axes,  
493 therefore these are all presented and interpreted to some degree<sup>19</sup>. Rnd-Lambda randomization  
494 results agreed with the broken stick method, the last useful axis is four with  $p = 0.001$  and  
495 cumulative variance explained at 71.7 %.

496

497 Multi-response permutation procedure (MRPP) is a nonparametric multivariate test of differences  
498 between groups<sup>19</sup>. The  $A$  statistic describes effect size with respect to how similar within-group  
499 samples are compared to outside the group samples. When  $A = 1$ , sample units within each group  
500 are identical, when  $A = 0$ , groups are no more different than expected by chance. We tested all  
501 possible models based on the site attributes - agricultural history, current land use, soil order,  
502 mineralogical class, disturbance level, and cropland versus non-cropland. The MRPP was run  
503 with Euclidean distance measure on the transformed data. Any contrast with less than two sites (6  
504 sample points) was excluded. In the final model, an adjusted  $p$  value was calculated by dividing  
505 the model  $p$  value by the number of pairwise contrasts, the adjusted  $p$  value was used to  
506 determine whether a pairwise contrast was significant or not.

507

508 Dimension reduction The list of 46 potential indicators of soil health was reduced to a short list of  
509 key indicators that meet multiple criteria for capturing the breadth of soil health as an ecosystem  
510 property, reducing multicollinearity with other variables, and practicality for inclusion in a  
511 routine soil test. First, potential parameters were removed if  $r < 0.50$  or  $> -0.50$  with axis 1 in the  
512 PCA, which left 26 selected for further assessment. A hierarchical ordering correlogram of the  
513 untransformed values for remaining 26 selected parameters was performed in R. Highly  
514 correlated parameters were reduced further on the basis of practicality (i.e., combination of cost  
515 and difficulty). The final list was cross checked to maintain balanced coverage across biological,

516 chemical, and physical properties. Within the constraints of the original sample design, the key  
517 parameters of soil health were compared across a subset of mineral and land use classes to assess  
518 their utility as indicators of soil health.

519

520 Univariate analyses were conducted to first determine the effect of past land use (PIAL versus  
521 none) and then assess the effect of current land use (protected forest, pasture, UPIAL, organic  
522 cropland, versus conventional cropland) on each of the 11 soil health indicators. A mixed model  
523 ANOVA approach was used to assess the main effect of past or present land use with soil  
524 mineralogical classes as the random effect (lmer function in the lme4 package <sup>28</sup>). General linear  
525 hypothesis testing (glht function in multcomp package <sup>29</sup>) was performed to compare group  
526 means (Tukey-adjusted). For a subset of the data (including UPIAL, organic cropland, and  
527 conventional cropland in LAC and HAC soils), mixed factorial ANOVA was performed to  
528 examine the interactive effect of soil mineralogical class and current land use on each soil health  
529 indicator with farm/location as the random effect (lmer function in the lme4 package). Group  
530 differences were assessed by Tukey multiple comparisons of least square means (lsmeans  
531 function in emmeans package). For all post-hoc multiple comparisons with the Tukey test, letter  
532 groupings were assigned with the cld function (multcomp package <sup>30</sup>).

533

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535

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544 **References**

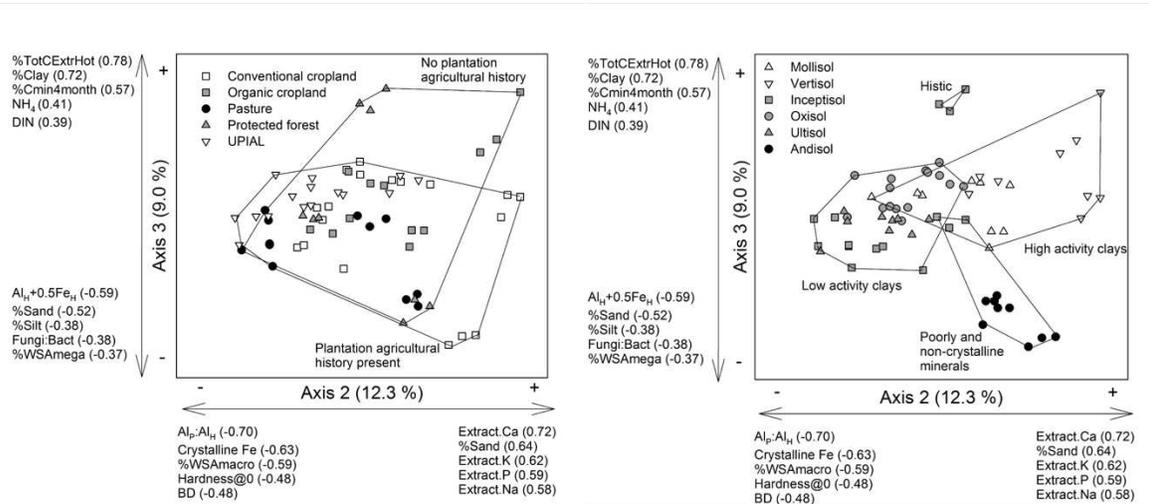
545

- 546 1. Olsson, L. et al. Land Degradation. in *Climate Change and Land: an IPCC special report on*  
547 *climate change, desertification, land degradation, sustainable land management, food*  
548 *security, and greenhouse gas fluxes in terrestrial ecosystems* (2019).
- 549 2. Sanderman, J., Hengl, T. & Fiske, G. J. Soil carbon debt of 12,000 years of human land use.  
550 *Proc Natl Acad Sci USA* **114**, 9575–9580 (2017).
- 551 3. Isbell, F., Tilman, D., Reich, P. B. & Clark, A. T. Deficits of biodiversity and productivity linger  
552 a century after agricultural abandonment. *Nat Ecol Evol* **3**, 1533–1538 (2019).
- 553 4. Zhu, K., Zhang, J., Niu, S., Chu, C. & Luo, Y. Limits to growth of forest biomass carbon sink  
554 under climate change. *Nat Commun* **9**, 2709 (2018).
- 555 5. Paustian, K. et al. Climate-smart soils. *Nature* **532**, 49–57 (2016).
- 556 6. Minasny, B. et al. Soil carbon 4 per mille. *Geoderma* **292**, 59–86 (2017).
- 557 7. Rumpel, C. et al. Response to “The ‘4p1000’ initiative: A new name should be adopted” by  
558 Baveye and White (2019). *Ambio* **49**, 363–364 (2020).
- 559 8. Bossio, D. A. et al. The role of soil carbon in natural climate solutions. *Nat Sustain* **3**, 391–  
560 398 (2020).
- 561 9. Bradford, M. A. et al. Soil carbon science for policy and practice. *Nat Sustain* **2**, 1070–1072  
562 (2019).
- 563 10. Lehmann, J., Bossio, D. A., Kögel-Knabner, I. & Rillig, M. C. The concept and future prospects  
564 of soil health. *Nat Rev Earth Environ* **1**, 544–553 (2020).
- 565 11. Kibblewhite, M. G., Ritz, K. & Swift, M. J. Soil health in agricultural systems. *Philosophical*  
566 *Transactions of the Royal Society B: Biological Sciences* **363**, 685–701 (2008).
- 567 12. Adhikari, K. & Hartemink, A. E. Linking soils to ecosystem services — A global review.  
568 *Geoderma* **262**, 101–111 (2016).
- 569 13. Bach, E. M., Ramirez, K. S., Fraser, T. D. & Wall, D. H. Soil Biodiversity Integrates Solutions  
570 for a Sustainable Future. *Sustainability* **12**, 2662 (2020).
- 571 14. Amundson, R. et al. Soil and human security in the 21st century. *Science* **348**, 1261071–  
572 1261071 (2015).
- 573 15. Lal, R. Soil health and carbon management. *Food Energy Secur* **5**, 212–222 (2016).
- 574 16. Kibblewhite, M. G. Soil and soil health: an overview. in *Burleigh Dodds Series in Agricultural*  
575 *Science* (ed. Reicosky, D.) vol. 1 3–15 (Burleigh Dodds Science Publishing, 2018).
- 576 17. Verburg, P. H. et al. Land system science and sustainable development of the earth system:  
577 A global land project perspective. *Anthropocene* **12**, 29–41 (2015).
- 578 18. Kurashima, N., Fortini, L. & Ticktin, T. The potential of indigenous agricultural food  
579 production under climate change in Hawai‘i. *Nat Sustain* **2**, 191–199 (2019).
- 580 19. Peck, J. E. *Multivariate analysis for ecologists: step-by-step.* (2016).
- 581 20. Blankinship, J. C. et al. Improving understanding of soil organic matter dynamics by  
582 triangulating theories, measurements, and models. *Biogeochemistry* **140**, 1–13 (2018).
- 583 21. Wieder, W. R., Grandy, A. S., Kallenbach, C. M., Taylor, P. G. & Bonan, G. B. Representing life  
584 in the Earth system with soil microbial functional traits in the MIMICS model. *Geoscientific*  
585 *Model Development* **8**, 1789–1808 (2015).
- 586 22. Robertson, A. D. et al. Unifying soil organic matter formation and persistence frameworks:  
587 the MEMS model. *Biogeosciences Discussions* 1–36 (2018) doi:10.5194/bg-2018-430.

- 588 23. Saco, P. M., McDonough, K. R., Rodriguez, J. F., Rivera-Zayas, J. & Sandi, S. G. The role of  
589 soils in the regulation of hazards and extreme events. *Phil. Trans. R. Soc. B* **376**, 20200178  
590 (2021).
- 591 24. Amundson, R. & Biardeau, L. Opinion: Soil carbon sequestration is an elusive climate  
592 mitigation tool. *Proc Natl Acad Sci USA* **115**, 11652–11656 (2018).
- 593 25. Bayer, A. D. et al. Diverging land-use projections cause large variability in their impacts on  
594 ecosystems and related indicators for ecosystem services. *Earth Syst. Dynam.* **12**, 327–351  
595 (2021).
- 596 26. Moebius-Clune, B. N. Comprehensive assessment of soil health: the Cornell framework  
597 manual. (2016).
- 598 27. Methods for assessing soil quality. (Soil Science Society of America, 1996).
- 599 28. Bates, D., Mächler, M., Bolker, B. & Walker, S. Fitting Linear Mixed-Effects Models Using  
600 **lme4**. *Journal of Statistical Software* **67**, (2015).
- 601 29. Lenth, R. Estimated Marginal Means, aka Least-Squares Means. (2020).
- 602 30. Piepho, H.-P. An Algorithm for a Letter-Based Representation of All-Pairwise Comparisons.  
603 *Journal of Computational and Graphical Statistics* **13**, 456–466 (2004).
- 604 31. Buyer, J. S. & Sasser, M. High throughput phospholipid fatty acid analysis of soils. *Applied*  
605 *Soil Ecology* **61**, 127–130 (2012).
- 606 32. Frostegård, A. & Bååth, E. The use of phospholipid fatty acid analysis to estimate bacterial  
607 and fungal biomass in soil. *Biol Fert Soils* **22**, 59–65 (1996).
- 608 33. Zelles, L. Fatty acid patterns of phospholipids and lipopolysaccharides in the  
609 characterisation of microbial communities in soil: a review. *Biology and Fertility of Soils* **29**,  
610 111–129 (1999).
- 611 34. Frostegård, Å., Tunlid, A. & Bååth, E. Use and misuse of PLFA measurements in soils. *Soil*  
612 *Biology and Biochemistry* **43**, 1621–1625 (2011).
- 613 35. Moore-Kucera, J. & Dick, R. P. PLFA Profiling of Microbial Community Structure and Seasonal  
614 Shifts in Soils of a Douglas-fir Chronosequence. *Microb Ecol* **55**, 500–511 (2008).
- 615 36. Bossio, D. A. & Scow, K. M. Impacts of Carbon and Flooding on Soil Microbial Communities:  
616 Phospholipid Fatty Acid Profiles and Substrate Utilization Patterns. *Microbial Ecology* **35**,  
617 265–278 (1998).
- 618 37. Bossio, D. A., Scow, K. M., Gunapala, N. & Graham, K. J. Determinants of Soil Microbial  
619 Communities: Effects of Agricultural Management, Season, and Soil Type on Phospholipid  
620 Fatty Acid Profiles. *Microbial Ecology* **36**, 1–12 (1998).
- 621 38. Ruess, L., Häggblom, M. M., García Zapata, E. J. & Dighton, J. Fatty acids of fungi and  
622 nematodes—possible biomarkers in the soil food chain? *Soil Biology and Biochemistry* **34**,  
623 745–756 (2002).
- 624 39. Helfrich, M., Ludwig, B., Thoms, C., Gleixner, G. & Flessa, H. The role of soil fungi and  
625 bacteria in plant litter decomposition and macroaggregate formation determined using  
626 phospholipid fatty acids. *Applied Soil Ecology* **96**, 261–264 (2015).
- 627 40. Malik, A. A. et al. Soil Fungal:Bacterial Ratios Are Linked to Altered Carbon Cycling. *Front.*  
628 *Microbiol.* **7**, (2016).
- 629 41. Tabatabai, M. A. Soil Enzymes. in *SSSA Book Series* (eds. Weaver, R. W. et al.) 775–833 (Soil  
630 Science Society of America, 2018). doi:10.2136/sssabookser5.2.c37.
- 631 42. Alkorta, I. et al. Soil Enzyme Activities as Biological Indicators of Soil Health. *Reviews on*  
632 *Environmental Health* **18**, (2003).

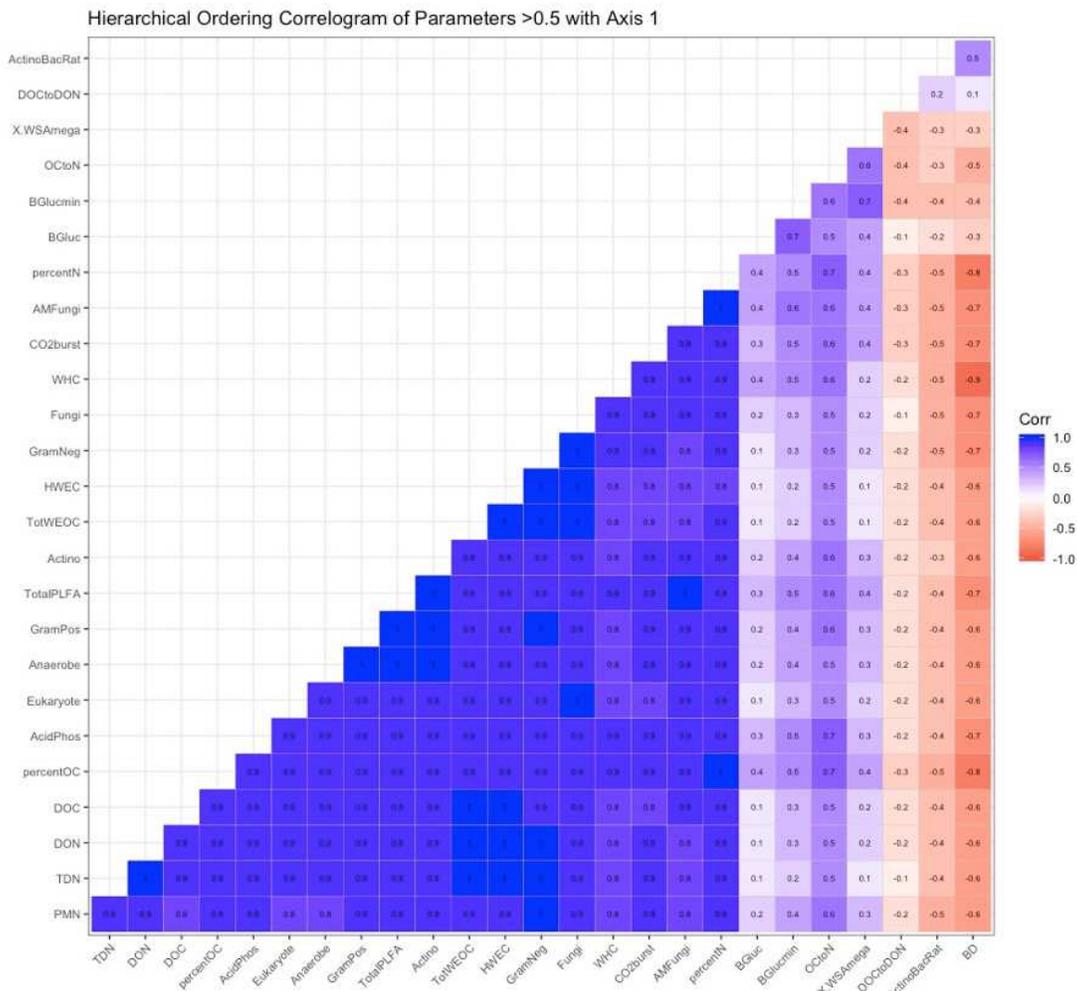
- 633 43. Acosta-Martinez, V., Cano, A. & Johnson, J. Simultaneous determination of multiple soil  
634 enzyme activities for soil health-biogeochemical indices. *Applied Soil Ecology* **126**, 121–128  
635 (2018).
- 636 44. Parham, J. A. & Deng, S. P. Detection, quantification and characterization of  $\beta$ -  
637 glucosaminidase activity in soil. *Soil Biology and Biochemistry* **32**, 1183–1190 (2000).
- 638 45. Acosta-Martínez, V. & Ali Tabatabai, M. Phosphorus Cycle Enzymes. in *SSSA Book Series* (ed.  
639 Dick, R. P.) 161–183 (American Society of Agronomy, Crop Science Society of America, and  
640 Soil Science Society of America, 2015). doi:10.2136/sssabookser9.c8.
- 641 46. Drinkwater, L. E., Cambardella, C. A., Reeder, J. D. & Rice, C. W. Potentially Mineralizable  
642 Nitrogen as an Indicator of Biologically Active Soil Nitrogen. in *SSSA Special Publications*  
643 (eds. Doran, J. W. & Jones, A. J.) 217–229 (Soil Science Society of America, 2015).  
644 doi:10.2136/sssaspecpub49.c13.
- 645 47. Haney, R. L. & Haney, E. B. Simple and Rapid Laboratory Method for Rewetting Dry Soil for  
646 Incubations. *Communications in Soil Science and Plant Analysis* **41**, 1493–1501 (2010).
- 647 48. Haney, R. L. & Franzluebbers, A. J. Soil CO<sub>2</sub> evolution: Response from arginine additions.  
648 *Applied Soil Ecology* **42**, 324–327 (2009).
- 649 49. Haney, R. L., Haney, E. B., Smith, D. R., Harmel, R. D. & White, M. J. The soil health tool—  
650 Theory and initial broad-scale application. *Applied Soil Ecology* **125**, 162–168 (2018).
- 651 50. Ghani, A., Dexter, M. & Perrott, K. W. Hot-water extractable carbon in soils: a sensitive  
652 measurement for determining impacts of fertilisation, grazing and cultivation. *Soil Biology*  
653 *and Biochemistry* **35**, 1231–1243 (2003).
- 654 51. Schädel, C. et al. Decomposability of soil organic matter over time: the Soil Incubation  
655 Database (SIDb, version 1.0) and guidance for incubation procedures. *Earth Syst. Sci. Data*  
656 **12**, 1511–1524 (2020).
- 657 52. Awale, R., Emeson, M. A. & Machado, S. Soil Organic Carbon Pools as Early Indicators for Soil  
658 Organic Matter Stock Changes under Different Tillage Practices in Inland Pacific Northwest.  
659 *Front. Ecol. Evol.* **5**, 96 (2017).
- 660 53. Karlen, D. L., Ditzler, C. A. & Andrews, S. S. Soil quality: why and how? *Geoderma* **114**, 145–  
661 156 (2003).
- 662 54. Chapman, H. D. Total Exchangeable Bases. in *Agronomy Monographs* (ed. Norman, A. G.)  
663 902–904 (American Society of Agronomy, Soil Science Society of America, 2016).  
664 doi:10.2134/agronmonogr9.2.c7.
- 665 55. Haney, R. L., Brinton, W. H. & Evans, E. Estimating Soil Carbon, Nitrogen, and Phosphorus  
666 Mineralization from Short-Term Carbon Dioxide Respiration. *Communications in Soil Science*  
667 *and Plant Analysis* **39**, 2706–2720 (2008).
- 668 56. Jansen, B., Kalbitz, K. & McDowell, W. H. Dissolved Organic Matter: Linking Soils and Aquatic  
669 Systems. *Vadose Zone Journal* **13**, vzj2014.05.0051 (2014).
- 670 57. Soil sampling and methods of analysis. (Canadian Society of Soil Science ; CRC Press, 2008).
- 671 58. Loeppert, R. L. & Inskeep, W. P. Colorimetric Determination of Ferrous Iron and Ferric Iron  
672 by the 1,10-Phenanthroline Method. in *Methods of Soil Analysis: Part 3, Chemical Methods*  
673 659–661 (SSSA, 1996).
- 674 59. Chao, T. T. & Zhou, L. Extraction Techniques for Selective Dissolution of Amorphous Iron  
675 Oxides from Soils and Sediments. *Soil Science Society of America Journal* **47**, 225–232  
676 (1983).

- 677 60. Ross, G. J., Wang, C. & Schuppli, P. A. Hydroxylamine and Ammonium Oxalate Solutions as  
678 Extractants for Iron and Aluminum from Soils. *Soil Science Society of America Journal* **49**,  
679 783–785 (1985).
- 680 61. McKeague, J. A. AN EVALUATION OF 0.1 M PYROPHOSPHATE AND PYROPHOSPHATE-  
681 DITHIONITE IN COMPARISON WITH OXALATE AS EXTRACTANTS OF THE ACCUMULATION  
682 PRODUCTS IN PODZOLS AND SOME OTHER SOILS. *Can. J. Soil. Sci.* **47**, 95–99 (1967).
- 683 62. Shoji, S., Dahlgren, R. & Nanzyo, M. Chapter 2 Morphology of Volcanic Ash Soils. in  
684 *Developments in Soil Science* (eds. Shoji, S., Nanzyo, M. & Dahlgren, R.) vol. 21 7–35  
685 (Elsevier, 1993).
- 686 63. Arshad, M. A. & Coen, G. M. Characterization of soil quality: Physical and chemical criteria.  
687 *American Journal of Alternative Agriculture* **7**, 25 (1992).
- 688 64. Awale, R. & Chatterjee, A. Soil Moisture Controls the Denitrification Loss of Urea Nitrogen  
689 from Silty Clay Soil. *Communications in Soil Science and Plant Analysis* **46**, 2100–2110  
690 (2015).
- 691 65. Kemper, W. D. & Rosenau, R. C. Aggregate Stability and Size Distribution. 425–442 (1986).
- 692 66. Arai, M. et al. Changes in water stable aggregate and soil carbon accumulation in a no-tillage  
693 with weed mulch management site after conversion from conventional management  
694 practices. *Geoderma* **221–222**, 50–60 (2014).
- 695 67. Blankinship, J. C., Fonte, S. J., Six, J. & Schimel, J. P. Plant versus microbial controls on soil  
696 aggregate stability in a seasonally dry ecosystem. *Geoderma* **272**, 39–50 (2016).
- 697 68. Berhe, A. A. et al. Persistence of soil organic matter in eroding versus depositional landform  
698 positions: EROSION AND SOIL ORGANIC MATTER DYNAMICS. *J. Geophys. Res.* **117**, n/a-n/a  
699 (2012).
- 700 69. Soil Survey Staff, Natural Resources Conservation Service, United States Department of  
701 Agriculture. (2020). Web Soil Survey. Available online at  
702 <https://websoilsurvey.nrcs.usda.gov/>.  
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**Figure S1.** Principal components analysis of all potential soil health indicators with the top five most negatively and positively correlated variables to each axis. Correlation values for each parameter to the axis follow it in parentheses. Left - Groups are delineated based on whether an intensive plantation agricultural history is present or absent at that site. The current land use of each site is also indicated: conventional cropland (white square), organic cropland (grey square), pasture (black circle), protected forest (grey triangle), or unmanaged previously intensive agriculture (UPIAL, upside down white triangle). Right - Groups are delineated based on broad mineralogical categorization at that site. The soil order of each site is indicated as Mollisol (white triangle), Vertisol (upside down white triangle), Inceptisol (grey square), Oxisol (grey circle), Ultisol (grey triangle), and Andisol (black circle). The amount of variability explained by each axis is in parentheses.



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**Figure S2.** A correlation matrix between the 26 untransformed parameters with > 0.50 correlation to axis 1 of the principal components analysis that could be eliminated from a soil health index due to multicollinearity. Values range from -1 to 1 and are also indicated by color. In this figure, values are hierarchically clustered to clearly show those that can be considered for reduction. A large block of highly positively correlated parameters with  $r \geq 0.8$  includes 1) biological - Total PLFA (TotalPLFA), actinomycetes (Actino), gram + bacteria (GramPos), gram - bacteria (GramNeg), eukaryotes, arbuscular mycorrhizae (AMfungi), anaerobic bacteria (anaerobe), fungi, acid phosphatase, potentially mineralizable N (PMN), CO<sub>2</sub> burst (CO<sub>2</sub>burst), 2) chemical – total OC %, total N %, hot water extractable C (HWEC), total water extractable C pool (TotWEOC), dissolved organic C (DOC), total dissolved N (TDN), dissolved organic N (DON), and 3) physical - water holding capacity (WHC). Soil OC to N ratio (OCtoN) also was positively correlated ( $r \geq 0.5$ ) to the parameters in this block.

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**Table S1.** Correlation values of indicators to all four significant principal components analysis axes. The five parameters most positively and negatively correlated to each axis are in **bold italics**. Parameters with correlation values to axes  $\pm 0.75$ -1.0 are dark grey,  $\pm 0.50$ -0.75 are medium grey, and  $\pm 0.25$ -0.50 are light grey. The biological factors and total organic C (OC) and TN concentrations were strongly negatively correlated with the axis 1. One exception was the actinomycetes to bacteria ratio, which was one of the strongest correlates with the positive side of axis 1. Other strong positive correlates to the first axis included the ratio of dissolved organic carbon (DOC) to dissolved organic nitrogen (DON) and bulk density (BD). Biological parameters did not contribute to the spread across axis 2, whereas chemical and physical parameters correlated strongly with both sides (positive and negative). The strongest correlates with axis 3 were fewer than for the first two axes, but largely derived from the chemical and physical categories. Interestingly, biological parameters emerged again as strong negative correlates with axis 4, and chemical and physical parameters were associated with both sides of that axis.

	Response (abbreviated)	Full Response Variable Name	Axis 1 (43.0 %)	Axis 2 (12.3 %)	Axis 3 (9.0 %)	Axis 4 (7.4 %)	
B	Total PLFA	Total PLFA	<b>-0.9611</b>	-0.0056	0.1262	-0.1651	
	Actino	Actinomycetes	-0.908	0.0299	0.1767	<b>-0.2706</b>	
	Gram+	Gram + bacteria	-0.9492	-0.0527	0.174	-0.1579	
	Gram-	Gram - bacteria	<b>-0.9707</b>	0.0040	0.1316	-0.0520	
	Eukaryote	Eukaryotes	-0.8176	-0.1146	-0.0352	-0.1549	
	AM Fungi	Arbuscular mycorrhizal fungi	<b>-0.9492</b>	0.0428	0.0107	-0.1549	
	Anaerobe	Anaerobic bacteria	-0.7087	0.1248	0.0966	<b>-0.4195</b>	
	Fungi	Fungi	-0.9094	-0.0076	0.0995	0.1364	
	Actino:Bact	Actinomycetes to bacteria ratio	<b>0.5116</b>	0.1913	0.061	<b>-0.5679</b>	
	Fungi:Bact	Fungi to bacteria ratio	-0.4173	0.0256	<b>-0.3756</b>	0.2604	
	$\beta$ -Gluc	$\beta$ -glucosidase	-0.5852	0.0055	-0.0975	<b>-0.5143</b>	
	$\beta$ -Glucmin	$\beta$ -glucosaminidase	-0.8102	-0.2495	-0.1007	-0.3571	
	AcidPhos	Acid phosphatase	-0.8671	-0.1148	-0.0128	-0.0296	
	PMN	Potentially mineralizable N	-0.8944	-0.1944	0.1554	-0.0483	
	CO <sub>2</sub> burst	CO <sub>2</sub> burst	-0.9277	-0.0536	0.1401	-0.0422	
	C	%OC	OC concentration (%)	<b>-0.9529</b>	0.0876	-0.2058	0.0844
		%N	N concentration (%)	<b>-0.9522</b>	0.1432	-0.1757	0.1013
		OC:N	OC to N ratio	-0.7622	-0.1112	-0.2812	0.0077
		HWEC	Hot water extractable carbon	-0.9196	0.0779	0.2539	-0.0423
%TotCEXtrHot		% of total OC that was HWEC	-0.4575	-0.0135	<b>0.7813</b>	-0.1075	
TotWEOC		Total water extractable C pool	-0.9321	0.0791	0.2392	-0.0443	
%Cmin4mon		C pool respired in 4 months	-0.0225	-0.3394	<b>0.5742</b>	-0.2223	
pH		pH	0.3022	0.2938	-0.2281	<b>-0.7024</b>	
Extract.Ca		Extractable Ca <sup>2+</sup>	-0.0853	<b>0.7181</b>	-0.1962	<b>-0.4073</b>	
Extract.K		Extractable K <sup>+</sup>	0.0014	<b>0.6168</b>	0.2059	-0.3681	
Extract.Na		Extractable Na <sup>2+</sup>	0.3222	<b>0.5818</b>	0.282	-0.1768	
Extract.P		Extractable P	0.2578	<b>0.5947</b>	0.2913	0.0933	
DOC		Dissolved organic C	-0.8177	0.0882	0.1608	-0.0374	
TDN		Total dissolved N	-0.8202	0.2732	0.2589	0.0508	
NH <sub>4</sub>		Ammonium	0.1177	0.4561	<b>0.4144</b>	-0.0299	
NO <sub>3</sub>		Nitrate	-0.4452	0.1096	0.2782	<b>0.5894</b>	
DIN		Dissolved inorganic N	0.0298	0.4531	<b>0.3907</b>	<b>0.3026</b>	
DON		Dissolved organic N	-0.7385	0.0281	0.3527	<b>0.3867</b>	
DOC:DON		Ratio of DOC to DON	<b>0.5715</b>	0.3174	0.1873	0.1719	
CrystalFe		Crystalline Fe-oxides	<b>0.3346</b>	<b>-0.6349</b>	0.1083	-0.2700	
Al <sub>H</sub> +0.5Fe <sub>H</sub>		Poorly and non-crystalline minerals	-0.4269	0.4558	<b>-0.5928</b>	0.0952	
Al <sub>P</sub> :Al <sub>H</sub>		Ratio of Al <sub>P</sub> to Al <sub>H</sub>	-0.1610	<b>-0.7001</b>	0.2242	0.2471	
P		BD	Bulk density	<b>0.6703</b>	<b>-0.4751</b>	0.1398	-0.3749
	Hardness@0	Hardness measured at surface	-0.1323	<b>-0.5626</b>	0.0964	<b>0.2647</b>	
	Hardness@15	Hardness measured at 15 cm	0.1545	-0.3770	0.1843	-0.3087	
	WHC	Water holding capacity	-0.7694	0.4398	0.0272	<b>0.3018</b>	
	%WSAmega	% WSA in the mega size class	-0.6471	-0.3222	<b>-0.3692</b>	-0.2533	
	%WSAmacro	% WSA in the macro size class	-0.3920	<b>-0.5985</b>	-0.3143	-0.2202	
	%Sand	% sand	-0.2938	<b>0.6362</b>	<b>-0.5157</b>	-0.0281	
	%Silt	% silt	-0.3697	-0.3906	<b>-0.3809</b>	0.1959	
	%Clay	% clay	<b>0.4800</b>	-0.1171	<b>0.7205</b>	-0.1778	

B = Biological; C = Chemical; P = Physical

**Table S2.** Results of a multi-response permutation procedure to test significance of multidimensional spatial differences between the proposed varying management groups. Higher A value indicates stronger model, adjusted p-value is 0.05 divided by the number of pairwise contrasts and was used to determine the significance of pairwise comparisons in the final model. Groups with fewer than one site (i.e., three within-site sample units) were excluded from the model. These classifications include agricultural history (previously intensive agricultural lands “PIAL” or “none”), current land use (conventional cropland, organic cropland, pasture, protected forest, or unmanaged previously intensive agriculture), soil order (Mollisol, Vertisol, Inceptisol, Oxisol, Ultisol, and Andisol) mineralogical class (high activity clays “HAC”, low activity clays “LAC”, poorly and non-crystalline minerals “PNCM”, and histic “HIS”), disturbance (“low” at least 50 yr no tillage, “medium” tilled in the last 10-50 yr, and “high” tilled in the last 10 yr), and cropland (“not cropland” or “cropland”). The categorical combination of current land use and minerals was the most significant model tested ( $A = 0.2926$ ,  $p < 0.0001$ ). However, due to the high numbers of pairwise contrasts and excluded groups, this model was not accepted as valid.

Classification or combination	A	p-value	# contrasts	Adj. p-value	Excluded groups
Current land use x Minerals	0.2926	< 0.0001	28	0.0018	ConvPNCM, ProForPMCN, ProForHIS, ProFor LAC
Agricultural history x Disturbance	0.2636	< 0.0001	6	0.0083	
Current land use	0.2583	< 0.0001	10	0.0050	
Disturbance	0.2559	< 0.0001	3	0.0167	
Agricultural history x Current land use	0.2544	< 0.0001	10	0.0050	PIALOrg and PIALPas
Cropland x Minerals	0.2203	< 0.0001	10	0.0050	NotCropHIS, CropPNCM
Agricultural history x Minerals	0.1396	0.001651	6	0.0083	NoneHIS, NoneHAC, PIALPNCM
Agricultural history	0.0944	0.000300	1	0.0500	
Cropland	0.0890	0.000453	1	0.0500	
Order	0.0879	n.s.	15	0.0033	
Minerals	0.0416	n.s.	3	0.01667	HIS

ConvPNCM = Conventional cropland; poorly and non-crystalline minerals; ProFor = Protected forest; HIS = Histic; LAC = low activity clays; PIAL = Previously intensive agricultural lands (i.e. plantation agricultural history present); Org = organic croplands; Pas = Pasture; NotCrop = A non-cropland land use designation; Crop = Cropland land use designation; None = No plantation history; HAC = high activity clays.

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**Table S3.** Broad mineralogical classification into high activity clay (HAC), low activity clay (LAC), poorly and non-crystalline minerals (PNM), and histic (HIS) was made using a key system with criteria derived from standard NRCS taxonomy.

Key	Criteria		
Q1	High poorly and non-crystalline mineral concentration?	$Al_H + 0.5Fe_H$ concentration over 2 %	If yes, then PNCM* If no, then Q2
Q2	High concentration of total oxides?	CrystalAl+CrystalFe over 50 g /kg soil	If yes, then LAC* If no, then Q3
Q3	High extractable bases?	Ca + Na + K cmol c over 10	If yes, then HAC* If no, then Q4
Q4	At least one classification?		If yes, then proceed to final decision If no, then proceed to HISTIC
HISTIC	Does the site belong as Histic?	high OC and low BD	If yes, then HIS If no, then Additional information needed
<b>Final decision</b>			
Q5	If LAC + HAC, is it currently cropland?		If yes, then LAC, but double check taxonomy and management If no, then proceed to Q6
Q6	If LAC + HAC, does it have a recent agricultural history?		If yes, then LAC, but double check taxonomy and management If no, then proceed to Q7
Q7	If LAC + HAC, is it an undisturbed, or protected system?		If yes, then HAC, but double check taxonomy and management If no, then Additional information needed

\* all samples proceed

Classification notes for tropical soils

- Inceptisols, in particular, need additional information to determine which class they belong to.
- Other than Inceptisols, classes usually align with taxonomy.