

# City water stress and industrial water-saving potential in stringent management of China

**Zongyong Zhang**

Southern University of Science and Technology

**Yuli Shan**

University of Groningen <https://orcid.org/0000-0002-5215-8657>

**Martin Tillotson**

University of Leeds

**Philippe Ciais**

Laboratoire des Sciences du Climat et de l'Environnement <https://orcid.org/0000-0001-8560-4943>

**Hong Yang**

Eawag, Swiss Federal Institute of Aquatic Science and Technology

**Xian Li**

Southern University of Science and Technology

**Bofeng Cai**

Chinese Academy for Environmental Planning

**Dabo Guan**

Tsinghua University <https://orcid.org/0000-0003-3773-3403>

**Junguo Liu** (✉ [junguo.liu@gmail.com](mailto:junguo.liu@gmail.com))

Southern University of Science and Technology

---

## Article

**Keywords:** industrial water saving, China, city, stringent management, water scarcity

**Posted Date:** August 19th, 2021

**DOI:** <https://doi.org/10.21203/rs.3.rs-114487/v1>

**License:**   This work is licensed under a Creative Commons Attribution 4.0 International License.

[Read Full License](#)

---

1 **City water stress and industrial water-saving potential under stringent management in China**

2

3 Zongyong Zhang<sup>1,2</sup>, Yuli Shan<sup>3</sup>, Martin Tillotson<sup>4</sup>, Philippe Ciais<sup>8,9</sup>, Hong Yang<sup>10</sup>, Xian Li<sup>5,2</sup>, Bofeng  
4 Cai<sup>6</sup>, Dabo Guan<sup>7,11\*</sup>, Junguo Liu<sup>1\*</sup>

5 1. School of Environmental Science and Engineering, Southern University of Science and Technology,  
6 Shenzhen, 518055, China.

7 2. Water Security Research Centre, School of International Development, University of East Anglia,  
8 Norwich NR4 7TJ, UK.

9 3. Integrated Research on Energy, Environment and Society (IREES), Energy and Sustainability  
10 Research Institute Groningen, University of Groningen, Groningen 9747 AG, the Netherlands

11 4. Water@leeds, School of Civil Engineering, University of Leeds, Leeds, LS2 9JT, UK

12 5. Department of Statistics and Data Science, Southern University of Science and Technology,  
13 Shenzhen, 518055, China.

14 6. Centre for Climate and Environmental Policy, Chinese Academy for Environmental Planning,  
15 Beijing, 100012, China.

16 7. Department of Earth System Sciences, Tsinghua University, Beijing 100080, China.

17 8. Sino-France Institute of Earth Systems Science, Laboratory for Earth Surface Processes, College of  
18 Urban and Environmental Sciences, Peking University, Beijing 100871, China.

19 9. Laboratoire des Sciences du Climat et de l'Environnement, Commissariat à l'Énergie Atomique et  
20 aux Énergies Alternatives CNRS Université de Versailles-Saint-Quentin-en-Yvelines, 91191 Gif-sur-  
21 Yvette, France

22 10. Department of Systems Analysis, Integrated Assessment and Modelling, Swiss Federal Institute for  
23 Aquatic Science and Technology (Eawag), Dübendorf, Switzerland

24 11. The Bartlett School of Construction and Project Management, University College London, London,  
25 UK.

26 **\* Address correspondence to:**

27 Junguo Liu, School of Environmental Science and Engineering, Southern University of Science and  
28 Technology, Shenzhen, 518055, China. Email: [junguo.liu@gmail.com](mailto:junguo.liu@gmail.com) or [liujg@sustech.edu.cn](mailto:liujg@sustech.edu.cn);

29 Dabo Guan, Department of Earth System Sciences, Tsinghua University, Beijing 100080, China.  
30 Email: [guandabo@hotmail.com](mailto:guandabo@hotmail.com);

31

32

33

34 China's industrial water withdrawal soared in the last decades and remained high. Stringent water  
35 management policies were set to save water through improving industrial withdrawal efficiency by 20%  
36 between 2015 and 2020. Although China has a nation-wide water scarcity, scarcity at city-level has not  
37 been fully explored. Thus, it is meaningful to use sectoral data to investigate industrial water saving  
38 potential and implication for alleviating scarcity. Here, we account for water withdrawal and scarcity in  
39 272 prefectural cities, using a 2015 data benchmark. The top 10% of low-efficiency sectors occupied 46%  
40 water use. In scenario analysis of 41 sectors across 146 water scarce cities, we assume a convergence of  
41 below-average efficiencies to the national sector-average. Results reveal overall efficiency could be  
42 increased by 20%, with 18.9 km<sup>3</sup> ( $\pm 3.2\%$ ) water savings, equivalent to annual water demand of Australia  
43 or Hebei province in China. A minority of sectors (13%) could contribute to most (43%) water savings  
44 whilst minimizing economic perturbations. In contrast, implementing water efficiency measures in the  
45 majority of sectors would result in significant economic disruption to achieve identical savings. Water  
46 efficiency improvements should be targeted towards this minority of sectors: cloth(ing) supply-chain,  
47 chemical manufacturing, and electricity and heat supply.

48

49 Key words: industrial water saving, China, city, stringent management, water scarcity

50 Freshwater is an essential and global resource<sup>1</sup>. Over the last 50 years, China's industrial water  
51 withdrawal increased in 90% of its cities<sup>2</sup>, and has remained at a high level above 126 km<sup>3</sup>/yr from 2013  
52 to 2018<sup>3</sup> largely due to low water-use efficiency. China used to have transnationally low efficiency partly  
53 owing to mis-management<sup>4-7</sup>, specifically poor sectoral controls and water-saving initiatives<sup>8</sup>. China's  
54 response to this was to legislate for industrial water withdrawals through the so-called stringent water  
55 resources management system ("Three-Redline" regulations), introduced by the Chinese State Council  
56 in 2011<sup>9</sup>, and aimed at saving water through improving industrial withdrawal per value-added by 20%  
57 between 2015 and 2020. More recently, China established national water-saving demonstration (sponge)  
58 cities, but specific control on both industrial water withdrawal intensities and volumes still remains poor<sup>10</sup>.

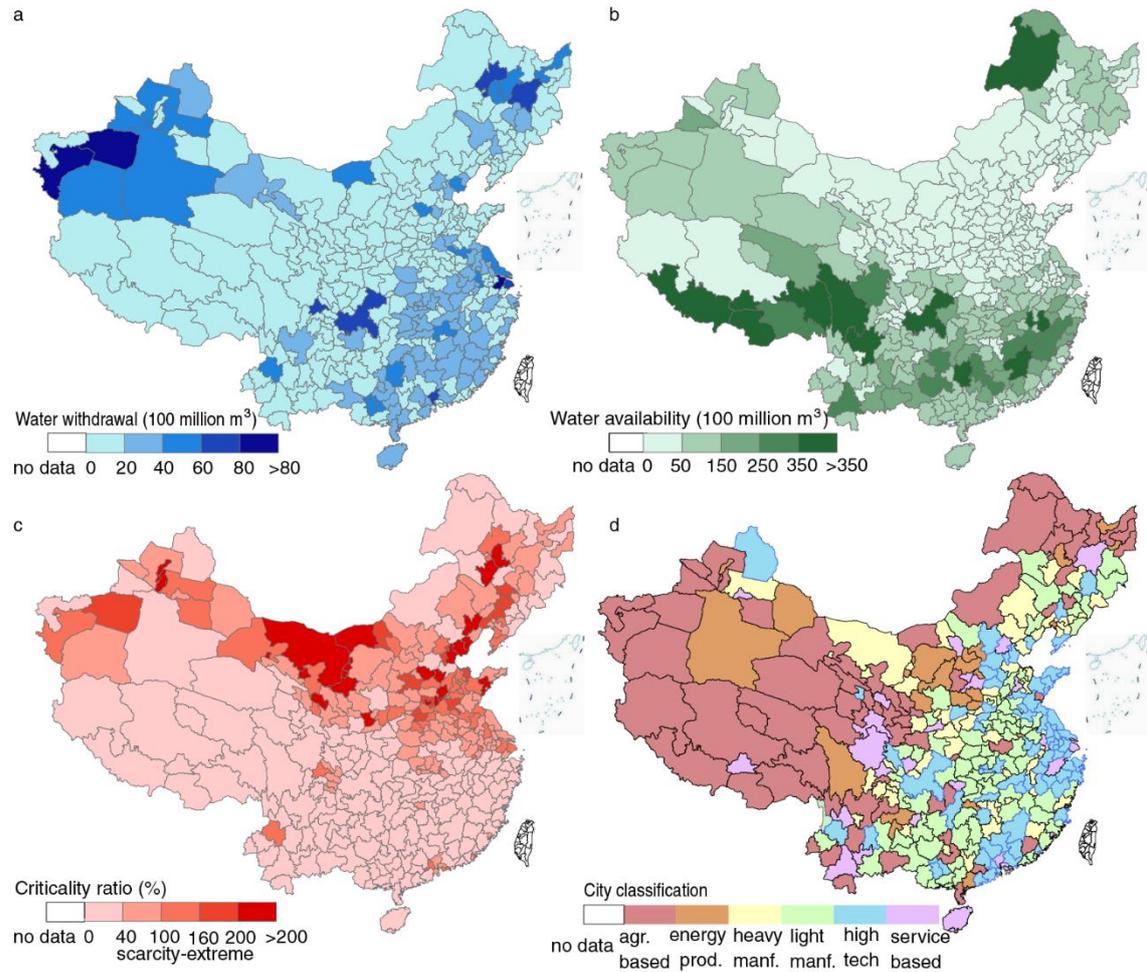
59 Although nation-wide China is deficient in water<sup>11</sup>, with a wicked problem between water demand  
60 and availability<sup>4,12</sup>, city-level water scarcity has not been fully explored<sup>13</sup>. The science of water scarcity  
61 assessment has developed for the past 30 years and, as more spatial geo-data have been available, studies  
62 have adopted more integrated and multi-faceted approaches typically based on spatial resolution in grid  
63 units at the river basin scale<sup>14,15</sup> or global levels<sup>16-18</sup>, rather than at administrative/territory based units  
64 such as the city level. There is only a single city-level based study in 2005 from the Ministry of Water  
65 Resources in China, which is not widely available to the public<sup>19</sup>. Thus far, to the best of our knowledge,  
66 an appraisal of cities and their water scarcity status is unavailable. In terms of measuring scarcity, the  
67 criticality ratio (water withdrawal to annual renewable freshwater) is a simple and classical indicator of  
68 blue water and quantitative scarcity<sup>20,21</sup>. It has thus far been applied at the provincial level<sup>16,22-24</sup>, but not  
69 at the city level due to data limitations<sup>7</sup>.

70 Water scarcity is typically exacerbated by unsustainable levels of water withdrawal; hence, society  
71 ought to be well placed to mitigate it by improving water use efficiency, especially by reducing water  
72 withdrawal intensities. Many studies have focused on agricultural intensification<sup>25,26</sup> in relation to better  
73 water management in land use<sup>27</sup> and irrigation<sup>28</sup>. However, due to lack of measured efficiency data, there  
74 remains a dearth of research especially from an industrial and sectoral perspective<sup>29</sup>, to explore water  
75 saving potential and implication on scarcity alleviation<sup>30</sup> at the city level.

76 We first accounted for datasets on water withdrawal for 41 industrial sectors in 272 prefecture-level  
77 cities (88% of China's population), and water scarcity for all cities (343) in 2015, based on a point-  
78 sourced survey in China<sup>31,32</sup>. We identified cities suffering from water scarcity, and low water efficiency  
79 sectors at the city level (compared with the national average). Second, we found the most severely

80 affected city type, and detected water scarcity and differences amongst these city-groups. Finally, in  
 81 scenario analysis we assumed a convergence of below-average efficiencies to the national sector-average,  
 82 to explore water saving potential amongst 41 industrial sectors and implication on water stress of Chinese  
 83 cities under the constraint of the 20%-intensity-reduction. For key sectors and cities, our results help to  
 84 identify priorities and optimize efforts for improving water use efficiency and facilitate more effective  
 85 water management through enabling distinctive saving strategies.

86 **Water withdrawal and water scarcity datasets at the city level**



116 **Fig. 1.** Prefecture-level cities and their water situation based on 2015 data. (a) total water withdrawal, (b)  
 117 water availability, (c) criticality ratio (%), and (d) six groups with predominant sector clustering. Average  
 118 size of cities was 2.80 million ha; average population was 4.43 million.

120 We built up datasets using a general accounting framework for Chinese cities, as developed for previous  
 121 work<sup>31,32</sup>. Drawn on the datasets, Fig. 1a represents a map of total water withdrawal at the city level.  
 122 Criticality ratio was determined by dividing total water withdrawal (1a) by water availability (1b) for  
 123 each city<sup>23,33,34</sup>. Typically an empirical threshold of 40% is regarded as water scarcity status<sup>18,35,36</sup>, and

124 over-100% as extreme water scarcity stress, signifying that annual water withdrawal exceeds renewable  
125 water resources<sup>13</sup>.

126 Overall, 146 of 272 cities (55% of population) were found to be under water scarce conditions, a  
127 result consistent with previous studies<sup>13</sup>. These cities are represented by darker colors in Fig. 1(c):  
128 Guangzhou and Shenzhen (south), Shanghai, Suzhou, and Yancheng (east), Harbin (north), and Hotan  
129 (west). Notably, in contrast to an earlier study<sup>13</sup>, we also identified some severe water-scarce areas in  
130 south China: Shenzhen (south; 108%) and Foshan (southeast; 107%). Water scarcity in China is known  
131 to already be serious, thus caution should be exercised when interpreting the south expansion of scarcity.

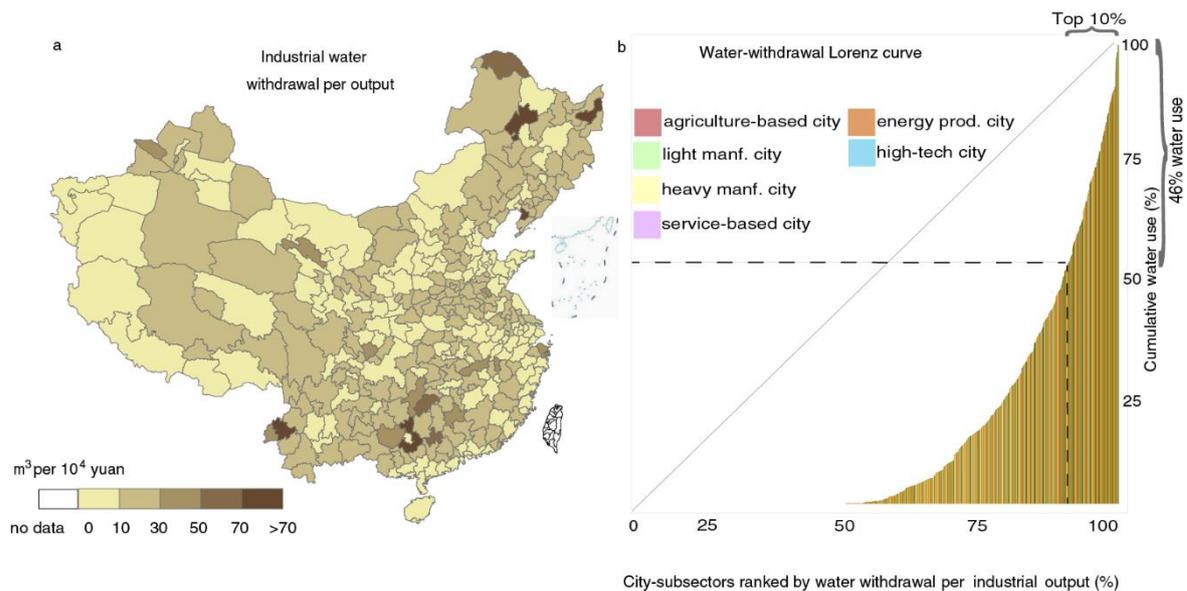
132 Sixty-nine Chinese cities (25%) were found to be under extreme water scarcity. These cities  
133 occupied 27% of the population. We identified cities in different regions experiencing extreme scarcity  
134 (Fig. 1c), for example Jiayuguan, Kelamayi and Lanzhou (northwest), Panjin (northeast), Puyang and  
135 Zhengzhou (central), and Shanghai (east). One of the adverse effects of extreme scarcity was observed  
136 in Zhengzhou, where average level of shallow groundwater decreased by 0.5 m in 2015<sup>3</sup>. Of 13  
137 metropolitan areas containing over-ten-million citizens, 11 cities were constrained by water scarcity, and  
138 6 by extreme scarcity. Median criticality ratio was 46%, varying between 0.38% in Ganzi (southwest) to  
139 over 200% in Jiayuguan (northwest). This median was six percentages exceeding the scarcity threshold  
140 of 40%.

141 Fig. 1(a), (b) and (c) show a mismatch in distribution between water use and availability at the city  
142 level. This uneven distribution results in water resources being commonly over-exploited in northern  
143 China. For example, several hotspots (with large water withdrawals) in northwest China, such as Hotan,  
144 Kuerle and Bayannur, have criticality ratios exceeding 100%. This indicates that environmental flow<sup>37-</sup>  
145 <sup>39</sup> is largely reduced for natural runoff and ecosystem survival. Fig. 1(d) shows city classifications and  
146 their intuitive spatial distribution. We classified cities into six groups, namely: agriculture-based, energy  
147 production, heavy manufacturing, light manufacturing, high-tech and service-based cities, using a  
148 clustering based methodology<sup>40</sup>.

#### 149 **Discrepancies in water withdrawal and water scarcity between cities**

150 When constrained by severe water scarcity, one might expect industries in water scarce cities to adopt  
151 water saving technologies, hence their industrial water withdrawal intensities should be lower than  
152 comparable industries in water sufficient areas. In other words, water scarcity should force local  
153 industries to be front-runners in water use efficiency. Nevertheless, a few water scarce cities (Fig. 2(a))

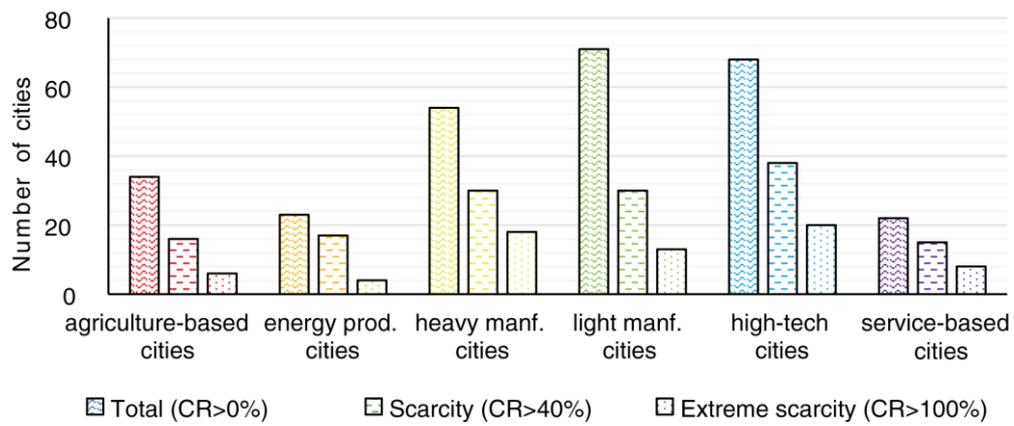
154 such as Qiqihar (north), Yingkou (east), Wuhai (west) and Puyang (central), had water intensities which  
 155 were much higher than in cities abundant in water resources. Although China has set intensity reduction  
 156 targets in stringent management since 2011, reducing intensities of sectors in water-scarce cities should  
 157 therefore be prioritized. Awareness of industrial water savings should be given greater focus in these  
 158 sectors in water scarce cities to prevent the situation to get worse. For example, cities such as Wuhai,  
 159 Hegang, Puyang, and Qitaihe, had water intensities which were still high, yet they were all included in  
 160 the 69 cities known to be over-exploiting resources, as released by the Chinese government in 2018<sup>41</sup>.



161  
 162  
 163 **Fig. 2.** Discrepancies in water withdrawal intensities across cities; (a) spatial distributions of overall  
 164 industrial intensities across cities; and (b) water-withdrawal Lorenz curve depicted by different  
 165 intensities of a total of  $41 \times 272 = 11,152$  city-sector combinations from six groups. (Different city groups  
 166 are represented by their corresponding color, as the same below.)  
 167

168 A disproportionately small fraction of sectors at the city level contributed to large industrial water  
 169 withdrawals. Thus sectors of low-efficiencies across cities should be well targeted to save water. We  
 170 ranked a total of  $41 \times 272 = 11,152$  city-sector combinations by order of water intensity from low to high  
 171 and then calculated share of cumulative water withdrawal accordingly. We depicted these shares relative  
 172 to shares of cumulative numbers of sectors and obtained a water-withdrawal Lorenz curve (Fig. 2b). The  
 173 curve indicates that the top 10% of high-intensity sectors account for 46% of water withdrawal, as a  
 174 disproportionate fraction. Such high-intensity water users were mostly found in small and developing  
 175 cities, with representative industries such as papermaking and product manufacturing in Chenzhou  
 176 (central), Lincang (southwest) and Qiqihar (northeast); liquor, beverage and tea manufacturing in  
 177 Jingdezhen (mid-east), Anqing (mid-south) and Wuzhou (southwest); and electricity and hot water  
 178 supply in Changde (mid-south).

179 We compared water scarcity occurrence amongst different city-groups. The most-severely affected  
 180 were found in the high-tech group (Fig. 3); 38 cities over the 40% criticality-ratio (water scarce) and 20  
 181 above 100% (extremely scarce). These are the highest in their corresponding tier, indicating economic  
 182 growth limitations subject to water resources constraints. Notably, population in high-tech cities  
 183 accounts for 33% of the total, and are commonly affected from severe water scarcity. Heavy- and light-  
 184 manufacturing cities were also ranked, following high-tech cities. These water scarce cities with sectors  
 185 of low water withdrawal efficiencies should be targeted.



186

187 **Fig. 3.** Statistics of city numbers in different criticality-ratio categories.

188

189 **Industrial water saving potential based on efficiency improvement**

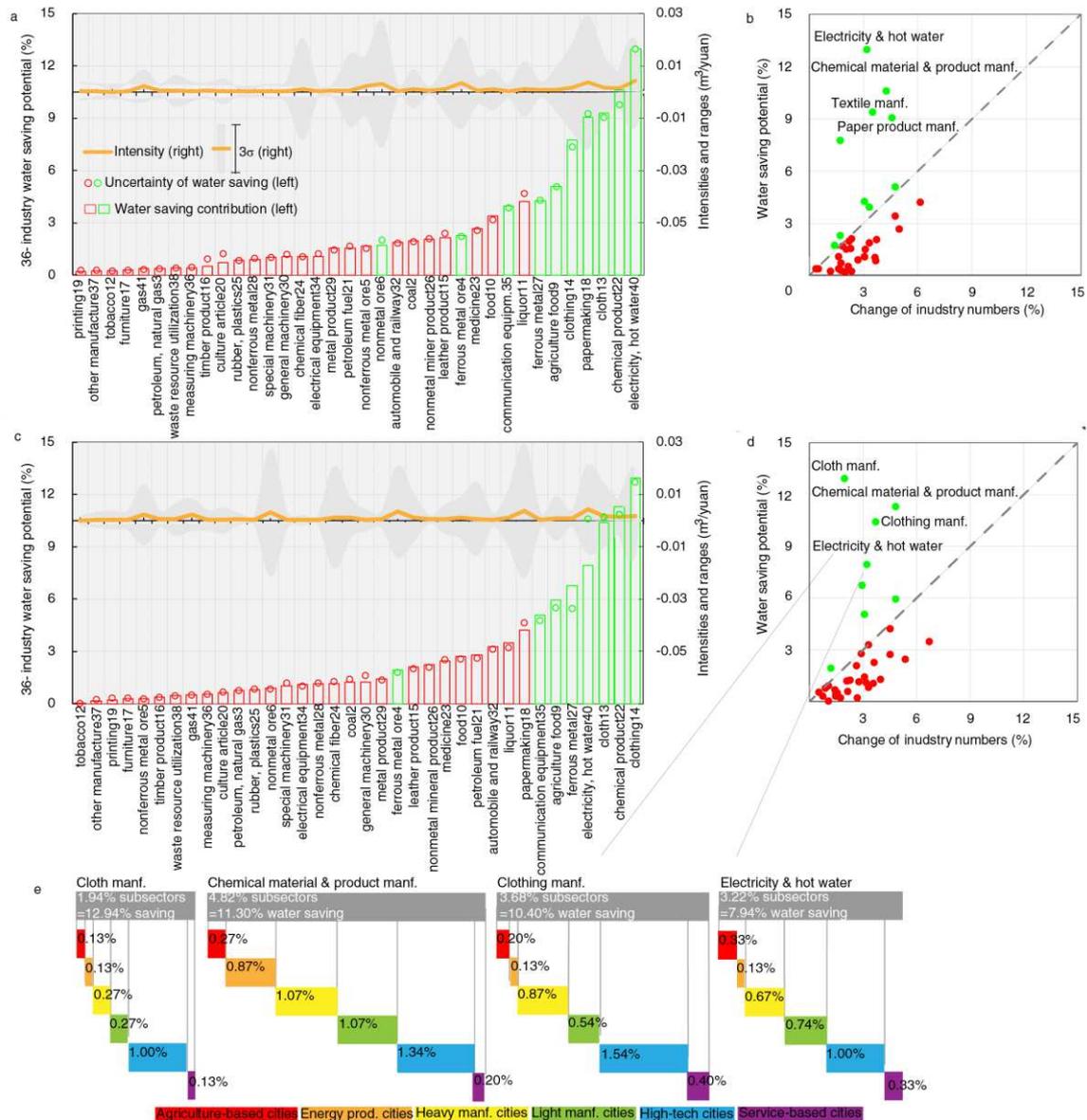
190 Industrial sectors in 272 cities were investigated for two reasons: first, there were special regulations for  
 191 industrial water withdrawal intensity in the redline policy; a number of cities were even required to  
 192 implement the most up-to-date technologies or regulatory standards for water savings during industrial  
 193 production. Second, the 41 industrial sectors we considered in total (see appendix III details) showed  
 194 high heterogeneity in water use and saving potential<sup>32</sup>.

195 For scenario analysis in individual of 41 industrial sectors, we substituted above-average water  
 196 intensities with average ones, by assuming technical progress in water use efficiency. Scenario A was  
 197 for all 272 cities and B was for the 146 water-stressed cities. Water saving strategies are more stringent  
 198 in A than B. If water withdrawal intensity of a sector in a city was lower than the national sector-average,  
 199 we left water intensity as it was. This would help maintain a stable technological and economic structure  
 200 whilst improving efficiency; If intensity of a sector was higher than the national sector-average, but it  
 201 occurred in a city with no water stress (criticality ratio less than 40%), we did not substitute it either;

202 Only for sectors that both had above-average intensities and were located in water-stressed cities, we did  
203 substitute intensities with national sector-averages. In fact, technology is a vital factor underpinning  
204 different intensities in the same sector. For example, in Suzhou, electricity and hot water supply  
205 consumed as much as 5.3 km<sup>3</sup> p.a. (64% of total water use) due to once-through cooling technology  
206 (water-intensive) accounting for 99% in thermal plants. Conversion of these plants to circulating cooling  
207 technologies, would result in large water savings. In contrast, food or general machinery manufacturing  
208 in Dongguan and Hanzhong, which stood out as high-efficiency exemplars, should be set as  
209 demonstration sites for peers in the same sector.

210 For all 272 cities, we estimated 41.91 km<sup>3</sup> ( $\pm 4.45\%$ ) water could be saved. This amount equates to  
211 7% of total water use for the whole of China, and is more than total industrial water consumption (31  
212 km<sup>3</sup>), twice the water demand of Australia or Hebei province of China in 2015<sup>42</sup>, and almost 2,000 times  
213 the water storage capacity of the West Lake in Hangzhou, China. A relatively small fraction (27%) of  
214 11,152 city-sector combinations contributed to large water savings (39%) of total industrial water  
215 withdrawals. Fig. 4(a) illustrates sectors towards right-hand side of x-axis could contribute approximately  
216 10% water savings, whilst those on the left could contribute a 0.2% reduction.

217 Furthermore, large contributors arose from fewer sectors at the city level, as shown in 4(b) (above  
218 the dotted line), whilst it was less effective to tap saving potential for sectors in the lower section (below  
219 the dotted line). Typically, there will be more than a single sector affected in most cities. Jiang (2009)<sup>43</sup>  
220 recommended exploration of cost-effective and long-term saving options by considering perturbations  
221 caused to economy. Here we hypothesized that the fewer individual sectors substituted, the less economic  
222 perturbation would result. Interestingly, a minority of sectors could save most water whilst affecting  
223 fewer cities. This seems a win-win opportunity. Instead, most sectors needed to disturb more economy  
224 to achieve the same saving. From an industrial water usage perspective, we therefore recommended water  
225 saving initiatives in five key sectors which potentially contributed half the available water savings:  
226 electricity and hot water supply (13.0%), chemical material and product manufacturing (10.6%), cloth  
227 (textile) manufacturing (9.4%), papermaking and products manufacturing (9.0%), and clothing (apparel,  
228 footwear and hats) manufacturing (7.8%). Requiring all industrial sectors to improve water efficiency  
229 does not therefore represent an optimal policy choice. This finding also applies to



230  
231  
232  
233  
234  
235  
236  
237  
238

Fig. 4. Water saving potential and withdrawal intensity in each sector; (a) and (b) for Scenario A, and (c), (d) and (e) for B. Grey shading indicates empirical distribution range of intensities in each sector. Upper and lower boundaries were calculated by the three-standard-deviation method. (e) shows the top four water-saving sectors and their structure within different city-groups. For brevity, we listed a product and a code in each sector; 2-8 represent mining and processing, 9-39 are manufacturing, and 40-42 are production and supply of electricity, gas and hot water. For full names and descriptions please refer to appendix III. We excluded sectors of small contributions.

239 water scarce cities (Scenario B, Fig. 4(c) and (d)). In 4(a) and (c), uncertainty arose from treatment of  
240 high-intensity sectors during the survey, considering considerable heterogeneity of water use technology  
241 across cities for the same sector. For Jing-Jin-Ji agglomeration 0.96 km<sup>3</sup> ( $\pm 9.8\%$ ) water could be saved.

242 In the 146 water scarce cities, reducing high water intensities in a relatively small fraction (13%) of  
243 11,152 city-sector combinations would result in large water saving (18%) of total industrial water  
244 withdrawals. A level of 18.9 km<sup>3</sup> ( $\pm 3.2\%$ ) of water would be saved. This equates to annual water demand

245 of Australia or Hebei province of China, and almost 1,000 times the West Lake capacity. For individual  
246 city-groups, water savings would reach 7.90 km<sup>3</sup> for high-tech cities, 4.17 km<sup>3</sup> for heavy manufacturing  
247 cities, 3.40 km<sup>3</sup> for service-based cities, 2.71 km<sup>3</sup> for light manufacturing cities, 0.7 km<sup>3</sup> for energy  
248 production cities, and 0.62 km<sup>3</sup> for agriculture-based cities. For individual cities, water savings ranged  
249 from 118,700 m<sup>3</sup> in Beijing, to 2.0 km<sup>3</sup> in Guangzhou. We hypothesized industrial value-added levels  
250 remained unchanged, in which case water withdrawal per value added would decrease by 20%, equating  
251 to the 2015-20 efficiency target in the stringent management. At identical water availability levels,  
252 criticality-ratio reduction ranged from 0.72% in Dongguan to 62% in Lanzhou. A small number of cities  
253 would be alleviated below the scarcity threshold (40%) and shake off water scarcity, for example Jilin  
254 city (northeast), Jincheng (northwest), Yulin and Tongchuan (west), and Xiangtan (mid-south). Heavy-  
255 manufacturing cities would be alleviated by 11% on average to sub extreme-scarcity level. At the national  
256 level, although the situation would remain severe, mean water scarcity level of 146 cities would fall by  
257 six percentage points from 95% to 89%.

258 Notably, in contrast to conventional understanding, electricity and hot water supply was not the  
259 largest contributor to water savings. Manufacturing of clothing, chemical materials and products, and  
260 textile would bring greater savings. The largest potential was in the cloth-clothing supply chain, including  
261 from cotton to intermediate products (yarn, cloth and other materials), and from yarn etc. to final clothing  
262 products such as apparels, footwears, hats, masks, and trims. This finding is supported by a previous  
263 study<sup>44</sup>, and could be useful in water saving management for relevant industrial committees.

264 We also decomposed structure of the important 13% sector fraction into different cities and groups,  
265 and identified four sectors (Fig. 4(d)) which contributed to half of total water savings; cloth(ing) supply  
266 chain, chemical material and product manufacturing, and electricity and hot water supply. Fig. 4(e) shows  
267 proportions of affected sectors from individual city-groups, respectively. For example, cloth  
268 manufacturing contributed to 12.94% (~2.37 km<sup>3</sup>) of water saving in total, yet these sectors accounted  
269 for just 1.94% overall at the city level. These subsectors and cities should be prioritized. A list is provided  
270 in appendix IV.

271 Most severely scarce city-groups were effectively pinned down, such as high-tech, heavy- and light-  
272 manufacturing cities. These city-groups basically hold the top three places for efficiency improvements.  
273 For example, proportions of affected cities (sectors) in heavy-manufacturing and high-tech cities were  
274 all highest; 78% (37%) and 56% (26%) respectively. Proportions of water-saving contributions from

275 individual city-groups were also checked and consistent (upon request). Thus, we were able to reliably  
276 and robustly validate discussion on substitution.

277 Of course, realization of water intensity reductions is likely to be different<sup>29</sup> from our rather crude  
278 scenario analyses; technologies between sectors and cities vary, and we must consider institutional as  
279 well as technical interventions. In fact, China's water saving potential in this regard is significant, with  
280 opportunities for factories and enterprises to adopt or advance efficient water-use equipment from their  
281 respective sector in the global environment. The main improvements we would recommend are in water  
282 recirculation (wet tower) in power generation, for example abstraction per kWh could be improved from  
283 168 liters to 5 liters<sup>45</sup>. Alternatively, we would encourage sectoral water abstraction and use rights, and  
284 incentives such as trade and other subsidies for water-saving sectors and cities<sup>46</sup> through water  
285 management contracts<sup>47</sup>. Regularly updated indices for leading-edge enterprises and high water  
286 efficiency manufactured products should be promoted by water efficiency labels<sup>48</sup> and national awards.  
287 Finally, online/real-time monitoring on water withdrawal of key sectors at the city level through roll-out  
288 of smart meters should be considered<sup>22</sup>.

289 In summary, we have reported water withdrawal and scarcity accounting for 272 Chinese cities,  
290 using a 2015 data benchmark. The top 10% of low-efficiency sectors made up 46% industrial water use.  
291 In scenario analysis of 41 sectors across 146 water-scarce cities, through efficiency improvements by 20%  
292 and satisfying the stringent management policy, 18.9 km<sup>3</sup> ( $\pm 3.2\%$ ) water saving would be realized.

293 Yet, here we recommend water saving potential in a handful of sectors, as these sectors identified  
294 to contribute to half of total water savings amongst 41 sectors. Focusing on these sectors makes sense in  
295 terms of producing water saving returns, whilst minimizing potential economic disruption across the  
296 industrial base. China may therefore target key sectors and cities in stringent water management, rather  
297 than requiring all industries and cities to be involved in water saving.

298

299 **Methods**

300 **City-level industrial water withdrawal data sources.** Industrial total water withdrawal and water-  
301 withdrawal per value added were compiled from water resources bulletins at provincial and city levels.  
302 Industrial water withdrawal is a newly withdrawn water amount<sup>3</sup>. This variable may depict pressure on  
303 available water resources from domestic economic activities more accurately since it excludes reused  
304 water.

305 Industrial water withdrawal intensities for individual sectors in each city were derived from the  
306 China High Resolution Emission Gridded Dataset<sup>31</sup>, in which a key survey of spot-sites covered 162,000  
307 enterprises, across 41 industrial sectors for all 343 prefecture-level cities (including leagues, regions and  
308 autonomous prefectures) in China. Sectoral industrial outputs were sourced from statistical yearbook for  
309 each city.

310 We applied the general accounting framework used in previous work, and built up city-level and  
311 territory-based industrial water withdrawal data for individual sector and city, according to IPCC  
312 administrative boundary (scope 1)<sup>49</sup>. For method validation please refer to detailed discussion in previous  
313 paper<sup>32</sup> and Turner et al. (2010)<sup>50</sup>. Of 343 cities, only 272 cities' data were available for sectoral  
314 accounting datasets, and 343 were further accounted for total blue-water withdrawal, availability and  
315 quantitative blue-water scarcity status.

316

317 **Clusters for city classification.** Cluster analysis usually refers to magnitudes of a series of pre-provision  
318 indicators (or variables) for specific datasets<sup>51</sup>. In the result, difference within a group would be  
319 significantly small, whilst relatively large between groups i.e., clusters represent variables with similar  
320 attributes<sup>52,53</sup>. Beyond administrative or provincial territories, city-level studies<sup>54,55</sup> concerning resource  
321 use across industries have utilized Shan et al. methodology<sup>40</sup> to classify Chinese cities into different  
322 groups (a k-mean cluster analysis). We used a similar treatment (employing proportions of industrial  
323 output) and supplemented with an agriculture-based grouping, to Shan et al. method. Agriculture-based  
324 cities occupied greater proportions of farming, forestry, animal husbandry, and fisheries in their GDP  
325 than other cities. We thought six groups represented different economic development stages by assuming  
326 a development time lag. For example, representatives of service-based cities were the so-called first-tier  
327 cities, including Beijing, Shanghai, Guangzhou, Shenzhen, as well as provincial capitals such as Wuhan  
328 and Nanjing. These were typified as wealthy and industrialized economies, as demonstrated by average

329 per capita GDP of 132,302 Yuan. This ranked 1<sup>st</sup> in all six groups, and was more than twice that of energy  
330 production cities. Service-based cities were assumed to take leading position for industrialization process  
331 in all Chinese cities.

332 Fig. 1 in the Appendix shows top-/bottom-ten sectors for water withdrawal efficiency and GDP  
333 statistics in six groups. Some low-efficiency and large water-users should be targeted to save water.  
334 Examples of energy production cities include Daqing, Panjin, Changzhi and Liupanshui. Although the  
335 top and bottom ten for water withdrawal intensity were amongst the smallest, this group appeared  
336 vulnerable since some cities such as Wuhai, Panjin, Hegang, Huozhou, and Qitaihe, have exhausted  
337 energy and water resources. High-tech cities followed, of which examples included Dalian, Nanchang,  
338 and Shaoguan. In heavy manufacturing cities, water withdrawal intensities were complex: these were  
339 amongst the largest, for example Panzhihua, Sanmenxia, Anshan and Handan, and most withdrawal  
340 efficiency varied across a large range. Service-based city water withdrawal intensities were not high.  
341 Furthermore, some cities were featured through cluster sectors with large water-use, such as Changchun  
342 (heavy manufacturing: special purpose machinery), Suzhou (high-tech manufacturing: communications  
343 equipment), and Yangzhou (heavy manufacturing: chemical materials and products). These sectors could  
344 learn from their peers within the same group.

345

346 **Application of criticality ratio as an indicator for water scarcity.** The criticality ratio (%) was applied  
347 to measure annual water scarcity<sup>23</sup>, i.e.:

$$348 \quad \text{Criticality Ratio}_i = \text{Water withdrawal}_i / \text{Water availability}_i \quad (1)$$

349 where i represents a city (one ratio number for one city); water withdrawal was the total amount from  
350 including farming, forestry, animal husbandry, fisheries, industry, construction, service, household, and  
351 ecosystem and environment preservation; and water availability included surface water and groundwater.

352 There are mainly three indicators in the current study: net runoff, natural streamflow, and natural  
353 streamflow minus consumptive use from upstream human activities<sup>13</sup>. We adopted the natural-  
354 streamflow measure and obtained relevant data from water resources bulletins for the cities, referring to  
355 Zhao et al. (2019)<sup>56</sup>. Basically produced from domestic precipitation, it is calculated through surface  
356 water plus groundwater minus double measurements. In 2015, China's precipitation (and water  
357 availability) was 2.8% (0.9%) more than, but close to, its average values through multiple years (1957-  
358 2000, with statistics)<sup>3</sup>. Criticality ratio takes into consideration environmental flows<sup>39,57</sup> and connects

359 with water quality and biodiversity<sup>58</sup>. The higher the ratio is, the more stress is placed on available water  
360 resources from withdrawal, and the greater the probability of water scarcity occurrence<sup>35</sup>.

361 In addition to Fig. 3, we further found there appeared to be discrepancies in criticality ratio in  
362 different city-types, indicating frequency and severity of water scarcity occurrence, referring to  
363 Veldkamp et al. (2016)<sup>59</sup>. For energy production cities (Appendix Fig. 2), frequency seemed relatively  
364 higher, but not as severe when compared to heavy manufacturing group. Trendline curve peaked at 50%,  
365 exceeding the 40% definition for water scarcity. In other words, most cities appeared to be distributed to  
366 the right of scarcity threshold. Reassuringly, there appeared to be relatively few instances of cities  
367 occurring in the extreme scarcity region (i.e. >100%).

368 In contrast, heavy manufacturing cities had lower frequencies of water scarcity occurrence, but once  
369 over the 40% threshold it tended to be more severe. The peak in the frequency trendline appeared at  
370 approximately 10-15% i.e., most cities tended to be distributed in a narrow band to the left of scarcity  
371 threshold. However, there was a greater, more even spread of samples above the extreme scarcity  
372 threshold, with a slight frequency approximately 5% for each distance, so the trendline tended to decrease  
373 gradually. Examples were Jiayuguan (3507%, northwest), Shizuishan (962%, northwest), Baiyin (489%,  
374 northwest), Tangshan (290%, north), Alashan (287%, northwest), Dongying (200%, east) and Baotou  
375 (189%, north). This small subset (approximately 13%) of cities in this group mainly influenced our  
376 findings for water scarcity in heavy manufacturing cities.

377 According to discrepancies of scarcity occurrence in different city-types, we also considered distinct  
378 water saving strategies. For heavy manufacturing cities, policy focus should therefore be on a small  
379 number of scarce cities at this stage. By comparison, for energy production cities, policy makers should  
380 focus on a greater number of cities. For agriculture-based and light-manufacturing cities, given their  
381 relatively lower GDP per capita, balance between economic development and water saving needs to be  
382 better coordinated in decision-making.

383

384 **Uncertainty analysis.** We also clustered cities based on economic shares of GDP for primary, secondary  
385 and tertiary industries, then classified cities into three groups for sensitivity analysis. We found only  
386 minor differences between ratios of cities at individual water scarcity levels, from the groups using  
387 proportions of industrial output. Specifically, for agriculture-based cities, the >40% and >100%  
388 criticality ratios accounted for 46% and 17% respectively; for industry-based cities they were 54% and

389 25%; whilst for service-based cities they were 67% and 35%. Although clusters were based on different  
390 indexes, we found no significant differences in water-scarcity distribution and status. We also verified  
391 water withdrawal per GDP of agriculture-based cities of 211m<sup>3</sup> per 10<sup>4</sup> Yuan, which was close to the  
392 magnitude of representative agriculture province such as Heilongjiang at 210 in 2015<sup>60</sup>. Finally, for  
393 individual city groups we validated median and average criticality ratios and water intensities; these  
394 results as well as significance tests for our group classification are available upon request.

395 Besides, we may over-estimate criticality ratio, considering water withdrawal statistics do include  
396 those from reservoirs and upstream rivers, while water availability data do not include these parts. We  
397 were unable to incorporate these data into water availability generally due to statistical incongruence  
398 between cities. Thus, our results could suffer from an upward bias in some cities. In future, we will  
399 supplement these data by combining hydrological simulations<sup>59,61,62</sup>. In summary, verifications suggest  
400 our city clusters are unbiased, and the results are robust and credible.

#### 401 **Limitations and future work**

402 Our study collated and accounted results for a single year and did not consider fluctuations in inter-  
403 annual precipitation and withdrawal, due to data availability. Variation of water availability for  
404 individual cities should be considered in future work since we have observed significant fluctuation, for  
405 example a decrease of approximately 60% in Qingdao, Zaozhuang, Laiwu and Linyi cities in 2016, due  
406 to reduced precipitation in dry years. This further work will not only reduce uncertainty of water scarcity  
407 status, but also explore temporal insights into understanding of water scarcity and allow for more time-  
408 series and statistical-significance testing.

409 Water quality-induced scarcity<sup>16,63,64</sup> has not been included in this paper due to lack of data for water  
410 temperature and salinity, nutrient and other pollutants. Besides, the extent to which water savings could  
411 be driven by water stress needs quantitative analysis.

412 At this stage our study is also limited by data availability for agriculture; we do not find sufficient  
413 irrigation efficiency data for subdivided crops or lands in individual cities, in order to project water saving  
414 potential for agriculture. For industrial sectors, it is better to use value-added to substitute output to assess  
415 efficiency, especially when such sectoral value-added data will be accessible in the future.

416 Finally, we only considered direct water savings for isolated sectors. It is only partially feasible to  
417 assume a smooth knowledge transfer of water efficiency experience from wealthier cities to poorer ones,

418 for example technology progress for saving water. Consumption-based water accounting considers water  
419 saving throughout the entire supply chains, which would be practical in future work.

420 **Acknowledgements**

421 This study has been supported by the National Natural Science Foundation of China (Grant No.  
422 41625001), the Strategic Priority Research Program of the Chinese Academy of Sciences (Grant No.  
423 XDA20060402), the High-level Special Funding of the Southern University of Science and Technology  
424 (Grant No. G02296302, G02296402), the Pengcheng Scholar Program of Shenzhen, the National High-  
425 level Talents Special Support Plan (“Ten Thousand Talents Plan”), and the Leading Innovative Talent  
426 Program for young and middle-aged scholars by the Ministry of Science and Technology. Special  
427 appreciations go to Dr. Zhao Zeng (Tianjin University), Dr. Dandan Zhao (Aalto University), Dr. Jiashuo  
428 Li (Shandong University) and Dr. Ya Zhou (Guangdong University of Technology) for improving the  
429 article quality. The authors claim no conflicts of interest.

430 **References**

- 431 1. Wada, Y., Wisser, D. & Bierkens, M. F. P. Global modeling of withdrawal, allocation and  
432 consumptive use of surface water and groundwater resources. *Earth Syst. Dyn.* **5**, 15–40  
433 (2014).
- 434 2. Zhou, F. *et al.* Deceleration of China’s human water use and its key drivers. *Proc. Natl. Acad.*  
435 *Sci. U. S. A.* **117**, 7702–7711 (2020).
- 436 3. China, M. of W. R. of the P. R. of. China water resources bulletin. (2015).
- 437 4. Shifflett, S. C., Turner, J. L., Dong, L., Mazzocco, I. & Yunwen, B. China’s Water-Energy-  
438 Food Roadmap: A Global Choke Point Report. *Wilson Cent. Washington, DC.* <https://www.wilsoncenter.org/publication/global-choke-point-report-chinaswater-energy-food-roadmap>  
439 (2015).  
440
- 441 5. Kong, X. *et al.* Chapter Two – Groundwater Depletion by Agricultural Intensification in  
442 China’s HHH Plains, Since 1980s. **135**, 59–106 (2016).
- 443 6. Lal, R. Research and development priorities in water security. *Agron. J.* **107**, 1567–1572  
444 (2015).
- 445 7. Wang, X. *et al.* Gainers and losers of surface and terrestrial water resources in China during  
446 1989–2016. *Nat. Commun.* **11**, 1–12 (2020).
- 447 8. Zhao, X., Tillotson, M., Yang, Z., Yang, H. & Liu, J. Reduction and reallocation of water use  
448 of products in Beijing. *Ecol. Indic.* **61**, 893–898 (2016).
- 449 9. Liu, J. *et al.* Water conservancy projects in China: Achievements, challenges and way forward.  
450 *Glob. Environ. Chang.* **23**, 633–643 (2013).
- 451 10. Jiang, Y., Zevenbergen, C. & Ma, Y. Urban pluvial flooding and stormwater management: A  
452 contemporary review of China’s challenges and “sponge cities” strategy. *Environ. Sci. Policy*  
453 **80**, 132–143 (2018).
- 454 11. Liu, J., Yang, H. & Savenije, H. H. G. China’s move to higher-meat diet hits water security.  
455 *Nature* **454**, 397 (2008).
- 456 12. Liu, J. *et al.* Environmental Sustainability of Water Footprint in Mainland China. *Geogr.*  
457 *Sustain.* **1**, 8–17 (2020).

- 458 13. Liu, X. *et al.* A Spatially Explicit Assessment of Growing Water Stress in China From the Past  
459 to the Future. *Earth's Futur.* **7**, 1027–1043 (2019).
- 460 14. Gao, X., Schlosser, C. A., Fant, C. & Strzepek, K. The impact of climate change policy on the  
461 risk of water stress in southern and eastern Asia. *Environ. Res. Lett.* **13**, (2018).
- 462 15. Wang, J., Zhong, L. & Long, Y. Technical Note Baseline Water Stress : China. *World Resour.*  
463 *Inst. Tech. Note* 1–16 (2016).
- 464 16. Liu, J. *et al.* Water scarcity assessments in the past , present , and future. *Earth's Futur.* **5**, 545–  
465 559 (2017).
- 466 17. Veldkamp, T. I. E. *et al.* Water scarcity hotspots travel downstream due to human interventions  
467 in the 20th and 21st century. *Nat. Commun.* **8**, (2017).
- 468 18. Flörke, M. *et al.* Domestic and industrial water uses of the past 60 years as a mirror of socio-  
469 economic development: A global simulation study. *Glob. Environ. Chang.* **23**, 144–156 (2013).
- 470 19. More than 400 cities out of 600 suffered from water scarcity in China\_China National Radio  
471 News. [http://www.cnr.cn/sxpd/shgl/20180322/t20180322\\_524173578.shtml](http://www.cnr.cn/sxpd/shgl/20180322/t20180322_524173578.shtml) (2018).
- 472 20. Alcamo, J. *et al.* World's Water in 2025.
- 473 21. Oki, T. Global Hydrological Cycles and World Water. *Science (80-. )*. 1068–1073 (2006).
- 474 22. Zeng, Z., Liu, J. & Savenije, H. H. G. A simple approach to assess water scarcity integrating  
475 water quantity and quality. *Ecol. Indic.* **34**, 441–449 (2013).
- 476 23. Zhao, X. *et al.* Physical and virtual water transfers for regional water stress alleviation in  
477 China. *Proc. Natl. Acad. Sci.* **112**, 1031–1035 (2015).
- 478 24. Cai, J., Varis, O. & Yin, H. China's water resources vulnerability: A spatio-temporal analysis  
479 during 2003–2013. *J. Clean. Prod.* **142**, 2901–2910 (2017).
- 480 25. Tilman, D., Balzer, C., Hill, J. & Befort, B. L. Global food demand and the sustainable  
481 intensification of agriculture. *Proc. Natl. Acad. Sci. U. S. A.* **108**, 20260–20264 (2011).
- 482 26. Cai, X. & Rosegrant, M. W. Optional water development strategies for the Yellow River Basin:  
483 Balancing agricultural and ecological water demands. *Water Resour. Res.* **40**, 1–11 (2004).
- 484 27. Lambin, E. F. & Meyfroidt, P. Global land use change, economic globalization, and the  
485 looming land scarcity. *Proc. Natl. Acad. Sci. U. S. A.* **108**, 3465–3472 (2011).
- 486 28. Matson, P. A., Parton, W. J., Power, A. G. & Swift, M. J. Agricultural intensification and  
487 ecosystem properties. *Science (80-. )*. **277**, 504–509 (1997).
- 488 29. Tillotson, M. R. *et al.* Water Footprint Symposium: where next for water footprint and water  
489 assessment methodology? *Int. J. Life Cycle Assess.* **19**, 1561–1565 (2014).
- 490 30. Wenfeng, L. *et al.* Savings and losses of global water resources in food-related virtual water  
491 trade. *Wiley Interdiscip. Rev. Water* (2018).
- 492 31. Cai, B. *et al.* China high resolution emission database (CHRED) with point emission sources,  
493 gridded emission data, and supplementary socioeconomic data. *Resour. Conserv. Recycl.* **129**,  
494 232–239 (2018).
- 495 32. Zhang, Z. *et al.* City-level water withdrawal in China: Accounting methodology and  
496 applications. *J. Ind. Ecol.* (2020) doi:10.1111/jiec.12999.
- 497 33. Vörösmarty, C. J., Green, P., Salisbury, J. & Lammers, R. B. Global water resources:  
498 vulnerability from climate change and population growth. *Science* **289**, 284–8 (2000).
- 499 34. Hanasaki, N. *et al.* A global water scarcity assessment under shared socio-economic pathways  
500 &ndash; Part 1: Water use. *Hydrol. Earth Syst. Sci. Discuss.* **9**, 13879–13932 (2012).
- 501 35. Alcamo, J. & Henrichs, T. Critical regions: A model-based estimation of world water resources  
502 sensitive to global changes. *Aquat. Sci.* **64**, 352–362 (2002).

- 503 36. Wada, Y., Gleeson, T. & Esnault, L. Wedge approach to water stress. *Nat. Geosci.* **7**, 615–617  
504 (2014).
- 505 37. Jacobsen, D., Milner, A. M., Brown, L. E. & Dangles, O. Biodiversity under threat in glacier-  
506 fed river systems. *Nat. Clim. Chang.* **2**, 361–364 (2012).
- 507 38. Van Vliet, M. T. H. *et al.* Global river discharge and water temperature under climate change.  
508 *Glob. Environ. Chang.* **23**, 450–464 (2013).
- 509 39. Liu, J., Liu, Q. & Yang, H. Assessing water scarcity by simultaneously considering  
510 environmental flow requirements, water quantity, and water quality. *Ecol. Indic.* **60**, 434–441  
511 (2016).
- 512 40. Shan, Y. *et al.* City-level climate change mitigation in China. *Sci. Adv.* **4**, eaaq0390 (2018).
- 513 41. Changjun, W., Yangyang, Q. & Xiangli, W. A Study on Economic-Population Contraction  
514 Governance in Resource-exhausted Cities: A Realistic Analysis Based on Resource-exhausted  
515 Cities in Heilongjiang Province. *Macroeconomics* (2019).
- 516 42. FAO. AQUASTAT - FAO's Information System on Water and Agriculture.  
517 <http://www.fao.org/nr/water/aquastat/main/index.stm>.
- 518 43. Jiang, Y. China's water scarcity. *J. Environ. Manage.* **90**, 3185–3196 (2009).
- 519 44. Niinimäki, K. *et al.* The environmental price of fast fashion. *Nat. Rev. Earth Environ.* **1**, 189–  
520 200 (2020).
- 521 45. Byers, E. A., Hall, J. W. & Amezaga, J. M. Electricity generation and cooling water use: UK  
522 pathways to 2050. *Glob. Environ. Chang.* **25**, 16–30 (2014).
- 523 46. Wang, T., Park, S. C. & Jin, H. Will farmers save water? A theoretical analysis of groundwater  
524 conservation policies. *Water Resour. Econ.* **12**, 27–39 (2015).
- 525 47. Zhigang, Z. & Xi, L. I. U. Current situation of water saving in Beijing universities and  
526 exploration and consideration of water-saving management contract mechanism. *Beijing Water*  
527 (2019).
- 528 48. Wang, Y. J., Xue, B. & Cai, K. Revelation for China Water Efficiency Label System based on  
529 the Practices of Domestic and Foreign Product Water-saving Systems. *China Stand.* (2015).
- 530 49. IPCC. *IPCC Guidelines for National Greenhouse Gas Inventories, Volume 2, Energy.*  
531 *Intergovernmental Panel of Climate Change* (2006). doi:10.1017/CBO9781107415324.004.
- 532 50. Turner, G. M., Baynes, T. M. & McInnis, B. C. A water accounting system for strategic water  
533 management. *Water Resour. Manag.* **24**, 513–545 (2010).
- 534 51. Ramaswami, A., Jiang, D., Tong, K. & Zhao, J. Impact of the Economic Structure of Cities on  
535 Urban Scaling Factors: Implications for Urban Material and Energy Flows in China. *J. Ind.*  
536 *Ecol.* **22**, 392–405 (2018).
- 537 52. M, E., P, S., P, B. & D, B. Correction: Cluster Analysis and Display of Genome-Wide  
538 Expression Patterns. *Proc. Natl. Acad. Sci. U. S. A.* **96**, 10943.
- 539 53. Cavalli-Sforza, A. W. F. E. L. A Method for Cluster Analysis. *Biometrics* **21**, 362–375.
- 540 54. Guan, D. *et al.* Structural decline in China's CO2 emissions through transitions in industry and  
541 energy systems. *Nat. Geosci.* **11**, 551–555 (2018).
- 542 55. Wu, Z., Ye, H., Shan, Y., Chen, B. & Li, J. A city-level inventory for atmospheric mercury  
543 emissions from coal combustion in China. *Atmos. Environ.* **223**, 117245 (2020).
- 544 56. Zhao, D., Hubacek, K., Feng, K., Sun, L. & Liu, J. Explaining virtual water trade: A spatial-  
545 temporal analysis of the comparative advantage of land, labor and water in China. *Water Res.*  
546 **153**, 304–314 (2019).
- 547 57. Vörösmarty, C. J. *et al.* Global threats to human water security and river biodiversity. *Nature*  
548 **467**, 555 (2010).

- 549 58. Kirby, J. M., Connor, J., Ahmad, M. D., Gao, L. & Mainuddin, M. Climate change and  
550 environmental water reallocation in the Murray-Darling Basin: Impacts on flows, diversions  
551 and economic returns to irrigation. *J. Hydrol.* **518**, 120–129 (2014).
- 552 59. Veldkamp, T. I. E., Wada, Y., Aerts, J. C. J. H. & Ward, P. J. Towards a global water scarcity  
553 risk assessment framework: Incorporation of probability distributions and hydro-climatic  
554 variability. *Environ. Res. Lett.* **11**, (2016).
- 555 60. Li, L. J. *et al.* Change in soil organic carbon between 1981 and 2011 in croplands of  
556 Heilongjiang Province, northeast China. *J Sci Food Agric* **96**, 1275–1283 (2016).
- 557 61. Veldkamp, T. I. E. *et al.* Changing mechanism of global water scarcity events: Impacts of  
558 socioeconomic changes and inter-annual hydro-climatic variability. *Glob. Environ. Chang.* **32**,  
559 18–29 (2015).
- 560 62. Wada, Y. *et al.* Global monthly water stress: 2. Water demand and severity of water stress.  
561 *Water Resour. Res.* **47**, 1–17 (2011).
- 562 63. Chaves, P. & Kojiri, T. Deriving reservoir operational strategies considering water quantity and  
563 quality objectives by stochastic fuzzy neural networks. *Adv. Water Resour.* **30**, 1329–1341  
564 (2007).
- 565 64. Van Vliet, M. T. H., Florke, M. & Wada, Y. Quality matters for water scarcity. *Nat. Geosci.*  
566 **10**, 800–802 (2017).
- 567
- 568
- 569

## Supplementary Files

This is a list of supplementary files associated with this preprint. Click to download.

- [supplementaryinformation.pdf](#)