

# Aminophosphonate mineralisation is a major step in the global oceanic phosphorus redox cycle

**Andrew Murphy**

University of Warwick

**David Scanlan**

University of Warwick <https://orcid.org/0000-0003-3093-4245>

**Yin Chen**

University of Warwick <https://orcid.org/0000-0002-0367-4276>

**Andrew Bottrill**

Medical Research Council Toxicology Unit, University of Leicester <https://orcid.org/0000-0002-5182-3643>

**Gary Bending**

University of Warwick

**John Hammond**

University of Reading <https://orcid.org/0000-0002-6241-3551>

**Elizabeth Wellington**

Warwick University

**Ian Lidbury** (✉ [I.lidbury@sheffield.ac.uk](mailto:I.lidbury@sheffield.ac.uk))

University of Sheffield <https://orcid.org/0000-0001-7190-315X>

---

## Article

**Keywords:** Planktonic Synthesis, Organophosphorus, Aminophosphonate Uptake, Marine Heterotrophs, Mesopelagic and Epipelagic Waters, Regenerated Phosphate

**Posted Date:** December 18th, 2020

**DOI:** <https://doi.org/10.21203/rs.3.rs-114998/v1>

**License:**  This work is licensed under a Creative Commons Attribution 4.0 International License.

[Read Full License](#)

---

**Version of Record:** A version of this preprint was published at Nature Communications on July 27th, 2021. See the published version at <https://doi.org/10.1038/s41467-021-24646-z>.

1 **Aminophosphonate mineralisation is a major step in the global oceanic phosphorus redox**  
2 **cycle**

3 Andrew R. J. Murphy<sup>1</sup>, David J. Scanlan<sup>1</sup>, Yin Chen<sup>1</sup>, Andrew Bottrill<sup>1</sup>, Gary Bending<sup>1</sup>, John P.  
4 Hammond<sup>2</sup>, Elizabeth M. H. Wellington<sup>1</sup>, Ian D.E.A. Lidbury<sup>3\*</sup>

5 <sup>1</sup> School of Life Sciences, University of Warwick, Gibbet Hill Road, Coventry, UK

6 <sup>2</sup> School of Agriculture, Policy, and Development, University of Reading, Earley Gate,  
7 Whiteknights, Reading, UK

8 <sup>3</sup>Department of Animal and Plant Sciences, University of Sheffield, Sheffield, UK

9 **\*Corresponding author: [i.lidbury@sheffield.ac.uk](mailto:i.lidbury@sheffield.ac.uk)**

10

11

12

13

14

15

16

17

18

19

20

## 21 **Abstract**

22 The planktonic synthesis of reduced organophosphorus molecules, such as  
23 alkylphosphonates and aminophosphonates, represents one half of a vast global oceanic  
24 phosphorus redox cycle. Whilst alkylphosphonates tend to accumulate in recalcitrant  
25 dissolved organic matter, aminophosphonates do not. Thus, we hypothesised unknown  
26 pathways for the uptake of aminophosphonates must exist in seawater. Here, we identify  
27 three novel bacterial 2-aminoethylphosphonate (2AEP) transporters, named AepXVW, AepP  
28 and AepSTU, whose expression is independent of phosphate concentrations (phosphate-  
29 insensitive). AepXVW, is found in diverse marine heterotrophs and is ubiquitously distributed  
30 in mesopelagic and epipelagic waters. Unlike the archetypal phosphate-regulated  
31 phosphonate binding protein, PhnD, the newly identified AepX is heavily transcribed (~100-  
32 fold>PhnD) in the global ocean independently of phosphate and nitrogen concentrations.  
33 Collectively, our data identifies a mechanism responsible for the oxidative step in the marine  
34 phosphorus redox cycle and suggests 2AEP may be an important source of regenerated  
35 phosphate, which is required for oceanic primary production.

36

## 37 **Introduction**

38 Phosphonates are reduced organic phosphorus (P) molecules with a carbon (C)-P  
39 bond, as opposed to the more common C-oxygen (O)-P ester bonds found in many other  
40 organic P molecules<sup>1</sup>. Phosphonates are synthesised as both primary and secondary  
41 metabolites in various bacterial, archaeal and eukaryotic organisms<sup>1-7</sup> where they are  
42 incorporated into lipids (phosphonolipids) and glycans (phosphonoglycans)<sup>4,8</sup>. A significant  
43 proportion can also be released from the cell to facilitate favourable biotic interactions<sup>9</sup>. Thus,  
44 they are ubiquitous in terrestrial and aquatic ecosystems<sup>10-14</sup>. Phosphonates also represent a

45 major fraction of the marine organic phosphorus pool<sup>11,15-17</sup> and recent studies have now  
46 identified several cosmopolitan marine microorganisms capable of synthesising significant  
47 quantities of these compounds<sup>6,7,9,18,19</sup>. Collectively, this synthesis drives a vast global oceanic  
48 phosphorus redox cycle with reduced phosphorus input in the surface ocean estimated to be  
49 an order of magnitude greater than pre-anthropogenic riverine phosphorus input<sup>9</sup>. Whilst  
50 much attention has focused on the degradation of alkylphosphonates, such as (hydroxy-)  
51 methylphosphonate (MPn) and 2-hydroxyethylphosphonate (HEP), as a source of P<sup>16,17,20,21</sup>,  
52 the potential for aminophosphonates such as 2-aminoethylphosphonate (2AEP) to serve as  
53 sources of C and N (Pi-insensitive) has been neglected. However, emerging evidence suggests  
54 that Pi-insensitive 2AEP catabolism occurs in nature<sup>22,23</sup>. Notably, the absence of 2AEP from  
55 otherwise phosphonate rich high molecular weight (HMW) DOM<sup>16,17</sup>, despite its supposed  
56 ubiquitous production<sup>9,19,24,25</sup>, suggests preferential catabolism of this molecule in  
57 comparison to alkylphosphonates.

58         The C-P bond is resistant to the activity of standard inorganic phosphate (Pi) liberating  
59 enzymes such as alkaline phosphatases and requires specific mechanisms to break, such as  
60 the C-P lyase<sup>26,27</sup>. Several 2AEP-specific phosphonate degradation systems (phosphonatases)  
61 have been characterised (Fig 1A). The C-P lyase, which is a non-specific promiscuous  
62 phosphonatase, is only induced in response to Pi-starvation, being regulated by the two-  
63 component master regulator of the Pi-stress response regulon, PhoBR<sup>24</sup>. In marine surface  
64 waters, genes encoding the C-P lyase are enriched in bacterial genomes found in regions  
65 typified by low Pi concentrations<sup>20</sup> where they are also heavily expressed<sup>28</sup>. Recent data has  
66 shown Pi-insensitive regulation of 2AEP degradation, facilitated by the 2AEP-specific  
67 phosphonatase systems, occurs in a few strains related to marine *Alphaproteobacteria*<sup>22</sup> and  
68 a terrestrial *Gammaproteobacteria*<sup>23</sup>. In both cases, a major consequence of Pi-insensitive

69 2AEP degradation was the remineralisation and release of labile  $\text{Pi}^{22}$ , due to the greater  
70 cellular demand for N over P and the 1:1 N:P stoichiometry of 2AEP.

71 To date, only two 2-AEP transport systems have been identified. Both are ATP-binding  
72 cassette (ABC) transporters which consist of a periplasmic substrate-binding protein (SBP), an  
73 ATP-binding domain protein, and a transmembrane permease. The first is located within the  
74 C-P lyase operon (*phnCDEFGHIJKLMNOP*)<sup>29,30</sup> with genes encoding the SBP, ATP-binding  
75 domain and transmembrane domains designated *phnD*, *phnC*, and *phnE*, respectively. The  
76 second is another ABC-transporter *phnSTUV* shown by Jiang *et al.* to complement a C-P lyase  
77 knockout mutant of *E. coli* together with *phnWX*<sup>31,32</sup>. PhnD has a restricted distribution in  
78 seawater whose abundance is highly correlated with regions of Pi-limitation<sup>20</sup>. However, no  
79 2AEP transporter has been identified in the majority of bacteria possessing PhnWX and  
80 PhnWAY phosphonatasases, which is surprising, given the fact 2AEP is a charged molecule and  
81 ubiquitous in marine and terrestrial ecosystems<sup>1,20</sup>.

82 Here, we sought to identify transporters, which provide superb molecular tools for  
83 investigating the *in situ* cycling of specific environmental metabolites<sup>28,33-35</sup>, required for 2AEP  
84 catabolism in environmental bacteria. Through combining laboratory-based molecular and  
85 genetic analyses with environmental meta-omics, we identified three novel transporters  
86 which have a role in 2AEP uptake and revealed Pi-insensitive 2AEP catabolism is widespread  
87 in the global ocean, likely representing a major step in the marine phosphorus redox cycle.

88

## 89 Results

### 90 *Pseudomonas putida* BIRD-1 possesses a novel Pi-sensitive 2-aminoethylphosphonate ABC 91 transporter, AepXVW

92 We recently identified several candidate 2-AEP transporters (ABC-type) in  
93 *Pseudomonas* rhizobacteria that contain the PhnWX phosphonatase but lack both the  
94 archetypal PhnCDE transporter and PhnSTUV<sup>36</sup> (Fig 1A). In *Pseudomonas putida* BIRD-1  
95 (hereafter BIRD-1), a periplasmic substrate binding protein associated with one of these  
96 putative transporters (PPUBIRD1\_4925), which we hereafter refer to as AepX, was induced  
97 under Pi-deplete growth conditions in a PhoBR-dependent manner<sup>36</sup>. AepX belongs to the  
98 same family (pfam13343) as PhnS, iron and sulphate SBPs but is clearly distinct (Coverage =  
99 40%, Identity = 25.09%,  $1.1e^{-05}$ ) (Fig 1B).

100 BIRD-1 was capable of growth on 1.5 mM 2AEP as either a sole N, P, or N and P source,  
101 the latter resulting in mineralisation of Pi which was subsequently exported from the cell (Fig  
102 1C, Fig S1). Mutagenesis of *phnWX* confirmed this phosphonatase was essential for 2AEP  
103 catabolism under both growth conditions in this bacterium (Fig S2A and S2B). Next, we  
104 investigated if AepX and its corresponding ABC transporter components, the ATP binding  
105 domain protein (AepV), and the permease domain protein (AepW) were essential for its  
106 growth on 2AEP. Surprisingly, deletion of *aepXVW*<sup>BIRD</sup> had no effect on growth as a sole N  
107 source (Fig 1C). However, the mutant ( $\Delta aepXVW$ <sup>BIRD</sup>::Gm) had significantly (P <0.0001)  
108 attenuated growth on 2AEP as a sole P source (Fig 1C). The growth defect observed during  
109 growth on 2AEP as a sole P source was largely restored by complementing the mutant with a  
110 plasmid-encoded native homolog (Fig. 1C). These data suggest that whilst the *aepXVW*  
111 transporter is not essential, it is involved in 2AEP uptake as a sole P source but is not involved

112 in its growth as an N source. Therefore, another 2AEP transport system must also exist in this  
113 bacterium.

## 114 **Identification of two Pi-insensitive 2-aminoethylphosphonate transporters in *P. putida***

### 115 **BIRD-1**

116 Next, by subjecting the BIRD-1  $\Delta aepXVW^{BIRD}::gm$  mutant to comparative proteomics,  
117 we identified a major facilitator-type (MFS) transporter, (PPUBIRD1\_3129), hereafter referred  
118 to as AepP for 2-aminoethylphosphate permease, whose expression was significantly  
119 increased during growth on 2 AEP as a sole N source (Pi-insensitive) (Fig S3). AepP is related  
120 to the glycerol-3-phosphate (G3P): Pi antiporter, GlpT<sup>37,38</sup> (identity to GlpT = 27.75%,  $9e^{-37}$ ), a  
121 member of the organophosphate: phosphate antiporter (OPA) family of MFS transporters.  
122 AepP shares conserved residues essential for binding Pi and the phosphate moiety of G3P<sup>39-</sup>  
123 <sup>42</sup> with GlpT, whereas residues that impact the binding affinity to the glycerol moiety of G3P  
124 but not Pi are not conserved<sup>41</sup> (Fig S4). Mutation of *aepP* in either the wild type parental strain  
125 ( $\Delta aepP$ , Fig1D) or the *aepXVW* mutant ( $\Delta aepXVW:Gm:aepP$ , Fig 1E) led to an inability to grow  
126 on 2AEP as sole N source. Subsequent complementation of  $\Delta aepP$  with its native homolog  
127 ( $\Delta aepP + pBB:aepP$ ) restored growth on 2AEP as a sole N source (Fig 1D). Interestingly, delayed  
128 but significant growth on 2AEP as a P source still occurred in the  $\Delta aepXVW:Gm:aepP$  double  
129 mutant, revealing the presence of a third 2AEP transporter in this bacterium.

130 To identify the unknown 2AEP transporter, we reanalysed our proteomics data (Fig  
131 S3). Another substrate binding protein (PPUBIRD1\_3891) containing the same pfam domain  
132 (pfam13343) as AepX (Fig 1B), hereafter named AepS, was constitutively expressed in all  
133 growth conditions. In order to uncover the role of AepS in the utilisation of 2AEP as a sole P  
134 source, a triple mutant  $\Delta aepXVW:Gm;aepP:aepSTU$  was generated in BIRD-1 (Fig 1E). This

135 triple knockout mutant was unable to grow on 2AEP as a sole P source (Fig 1E), suggesting  
136 AepSTU is a functional 2AEP transporter. However, generation of a single *aepSTU* knockout  
137 mutant did not affect Pi-sensitive growth compared to the wild type (Fig S2C), suggesting  
138 AepXVW is the major transporter involved in Pi-sensitive 2AEP uptake and AepSTU only has  
139 an auxiliary role in 2AEP uptake. Despite expression of AepS during Pi-insensitive growth (Fig  
140 S3), this transporter was unable to facilitate growth on 2AEP as a sole N source in the absence  
141 of AepP. Together, these data reveal the presence of three novel and differentially regulated  
142 2AEP transporters in BIRD-1. AepXVW is the primary Pi-sensitive 2AEP transporter with  
143 AepSTU having an auxiliary role, whereas AepP is essential for Pi-insensitive 2AEP growth.

144

145 **AepXVW is found in several marine bacteria capable of Pi-insensitive mineralisation and is**  
146 **a functional 2-aminoethylphosphonate transporter**

147 Using PPUBIRD1\_4925 (AepX) as the query, we scrutinised the genomes of several  
148 isolates related to marine *Rhodobacteraceae* (*Stappia* spp., *Terasakiella* spp.,  
149 *Falsirhodobacter* spp.) capable of Pi-insensitive phosphonate catabolism<sup>22</sup>. ORFs encoding  
150 orthologs of AepX were identified in the genomes of *Stappia stellulata* DSM 5886 and  
151 *Terasakiella pusilla* DSM 6293, in addition to several other marine roseobacter strains:  
152 *Aliiroseovarius crassostreae* (DSM 16950), *Aliiroseovarius sediminilitoris* (DSM 29439), *Shimia*  
153 *marina* (DSM 26895), and *Thalassobius aestuarii* (DSM 15283) (Fig 2A). We also found an  
154 orthologous ORF in the model rhizosphere alphaproteobacterium *Sinorhizobium meliloti*  
155 strain 1021 that is capable of 2AEP catabolism via a phosphonatase<sup>43</sup> (Fig 2A). In all cases,  
156 ORFs encoding AepXVW were located adjacent to ORFs encoding the phosphonatases,  
157 PhnWAY or PhnWX, strongly suggesting a role in 2AEP transport. *S. stellulata* DSM 5886, A.

158 *crassostreae* DSM 16950, *A. sediminilitoris* DSM 29439, *S. marina* DSM 26895, and *T. aestuarii*  
159 DSM 15283 were all capable of growth on 2AEP as either the sole N or P source (Table S1).  
160 Indeed, both *Aliiroseovarius* strains lack other characterised 2AEP transport and degradation  
161 systems (Table S2). As previously reported, Pi was exported from cells and accumulated in the  
162 medium during growth on 2AEP as a sole N source (Fig S5A).

163 To confirm that *Stappia* AepXVW can take up 2AEP, we complemented the BIRD-1 null  
164 2AEP transporter mutant ( $\Delta aepXVW:aepP:aepSTU:gm$ ) with this transporter fused with the  
165 *aepXVW*<sup>BIRD-1</sup> promoter (Fig 2B). This duly restored growth of the triple mutant confirming *S.*  
166 *stellulata* AepXVW is also a functional 2AEP transporter.

#### 167 **AepXVW is highly expressed in the marine bacterium *Stappia stellulata* during Pi-sensitive** 168 **and Pi-insensitive growth on 2-aminoethylphosphonate**

169 In BIRD-1, AepXVW was only involved in Pi-sensitive growth whilst AepP was induced  
170 during Pi-insensitive metabolism (Fig 1). However, *S. stellulata* lacks AepP but is still capable  
171 of Pi-insensitive growth and Pi export (Fig S5A). In addition to *aepXVW* and genes encoding  
172 the 2AEP phosphonate (*phnWAY*), *S. stellulata* also possesses genes (*phnCDEFGHIJKLMN*)  
173 encoding the P-regulated C-P lyase operon and we also confirmed this strain grew on several  
174 other alkylphosphonates as sole P source (Fig S5B). Therefore, to determine which transport  
175 and degradation systems were upregulated during growth on 2AEP as either a sole N or P  
176 source, we performed comparative proteomics. Unlike BIRD-1, AepX was abundantly  
177 expressed during growth on 2AEP as either a sole N or P source, as was the phosphonate  
178 (*PhnWAY*), whilst the C-P lyase operon was not (Fig 3). Importantly, whilst we detected  
179 several general nitrogen stress-response proteins induced under Pi-insensitive growth, we

180 didn't identify any other potential 2AEP transporters (Table S3). Therefore, unlike in BIRD-1,  
181 AepXVW likely represents the major route for 2AEP uptake in this bacterium.

## 182 ***aepX* and *aepP* are found in distantly related and cosmopolitan bacterial taxa**

183 Using the Integrated Microbial Genomes/Microbiomes from the Joint Genome  
184 Institute (IMG/M/JGI) database, we identified ORFs encoding AepX and AepP (but not AepS)  
185 in genomes retrieved from both taxonomically divergent isolates as well as single amplified  
186 genomes (SAGs) and metagenome assembled genomes (MAGs), which revealed an  
187 unexpected diversity for these substrate binding proteins (Fig 4). For AepX, this included  
188 cosmopolitan marine *Alphaproteobacteria* other than *Rhodobacteraceae*, marine  
189 *Deltaproteobacteria*, as well as marine *Vibrio* spp. AepX was also found in terrestrial  
190 *Betaproteobacteria*, *Firmicutes*, and other gram-positive bacteria (Fig 4). AepX was  
191 partitioned into several subclades, with AepX<sup>Stappia</sup> and AepX<sup>BIRD</sup> well separated (Fig 4). Many  
192 taxonomically divergent AepX ORFs were co-localised with ORFs encoding the various  
193 phosphonate systems or the C-P lyase, supporting a role in 2AEP transport (Fig 4).

194 AepP was also found in a wide range of phylogenetically divergent taxa, such as  
195 *Acidobacteria* (*Granuliella mallensis*) and *Bacteroidetes* (*Kriegella aquimaris*), *Actinobacteria*  
196 (*Streptomyces albulus* CCRC 11814) and *Verrucomicrobia* (*Haloferula* sp. BvORR071 and  
197 *Verrucomicrobia* sp. SGGC AC-337 J20) (Fig S6). However unlike AepX, AepP was not found in  
198 cosmopolitan marine bacteria. Again, for all of these strains ORFs encoding AepP were co-  
199 localised with ORFs encoding phosphonates (Fig S6). Notably, AepP was found in fewer  
200 marine isolates compared to AepX.

201

202 ***aepX* gene and transcript abundance is far greater than the archetypal *phnD/phnS* in the**  
203 **global ocean**

204 Using the TARA oceans OM-RGCv2+G metagenome (MG) and OM-RGCv2+T  
205 metatranscriptome (MT) datasets<sup>44</sup>, we calculated the abundance of our newly-identified  
206 *aepX*, *aepS* and *aepP* transporters and compared this with *phnS* and the archetypal  
207 phosphonate transporter *phnD*, whose gene abundance in seawater was recently  
208 calculated<sup>20</sup>. We analysed data from both the epipelagic and mesopelagic zones where  
209 phosphonate mineralisation is believed to occur<sup>15</sup>. Across all oceanic sampling sites in both  
210 the epipelagic and mesopelagic, *aepX* gene and transcript abundance was significantly greater  
211 (MG; post-hoc Dunn's test  $z = 10.4$ ,  $p < 0.001$  and  $z = 4.8$ ,  $p < 0.001$ , respectively) than *phnD* (Fig  
212 5A and B). On average, in the mesopelagic almost 10% of bacterial cells possess *aepX* whilst  
213 only ~0.3% and 0.5% possess *phnD* and *aepP*, respectively (Fig 5A). *aepX* transcription was 40-  
214 fold and 350-fold greater than *phnD* in the epipelagic and mesopelagic, respectively (Fig 5B).  
215 The majority of *aepX* sequences were related to the cosmopolitan *Alphaproteobacteria* and  
216 *Deltaproteobacteria* (Fig 4). We confirmed that these abundant environmental sequences  
217 were also co-localised with phosphonate degradation genes (Fig 4). In broad agreement with  
218 *aepX*, the cumulative transcription of the two phosphonatase markers *phnA* and *phnX* is  
219 significantly greater than the C-P lyase marker *phnJ* (Kruskal-wallis  $X^2 = 206.6$ ,  $p < 0.001$ )  
220 strengthening this observation that 2AEP mineralisation is a major oceanic process.

221 The gene abundances of *aepP*, *aepS* and *phnS* were all significantly lower than both  
222 *aepX* and *phnD* (post-hoc Dunn's test  $z = 13.1$  and  $9.0$ ,  $p < 0.001$ ,  $z = 14.5$  and  $12.6$ ,  $p < 0.001$ ,  
223 and  $z = 9.9$  and  $8.9$ ,  $p < 0.001$  respectively) in the epipelagic, whilst only *aepS* and *phnS* were

224 significantly lower in the mesopelagic (post-hoc Dunn's test  $z = 10.1$  and  $10.8$ ,  $p < 0.001$ , and  $z$   
225  $= 4.6$  and  $5.5$ ,  $p < 0.001$  respectively) (Fig 5A and S7).

226 Unlike *phnD* and *aepP*, *aepX* abundance was comparable across all oceanic regions  
227 within both MG and MT at each depth suggesting 2AEP mineralisation is a ubiquitous process  
228 in seawater (Fig 5C & D). For all sites at each depth, the relative abundance of *aepX* transcripts  
229 was always significantly greater (Wilcoxon rank sum  $W=3771$ ,  $p < 0.001$ , estimated log2  
230 difference =  $2.24$  (95CI  $1.97-2.52$ )) than its own gene abundance. For *phnD*, we observed  
231 significantly greater transcript abundance compared to its own gene abundance only in the  
232 Mediterranean Sea, a region typified by Pi-depletion (Wilcoxon rank sum  $W=0$ ,  $p < 0.001$ ,  
233 estimated log2 difference =  $3.11$  (95CI  $1.90-4.42$ )). Finally, in agreement with previous work<sup>20</sup>,  
234 *phnD* gene abundance was inversely correlated ( $R^2 = 0.340$ ,  $p < 0.001$ ) with standing stock  
235 concentrations of Pi (Fig 5E) as was *phnD* transcript abundance (Fig 5F). In contrast, *aepX* and  
236 *aepP* gene abundance were positively correlated ( $R^2 = 0.098$ ,  $p < 0.001$  and  $R^2 = 0.291$ ,  $p < 0.001$ ,  
237 respectively) with Pi concentration (Fig 5E), whilst no significant relationship between Pi  
238 concentration and *aepX/aepP* transcription was found (Fig 5F) suggesting their activation is  
239 independent of Pi in seawater globally.

240 To better understand the parameters controlling 2AEP catabolism in the ocean, we  
241 compared both gene abundance and transcript in relation to  $R^*$ , a measure of N vs P limitation  
242 calculated as  $[\text{NO}_2] + [\text{NO}_3] - 16[\text{PO}_4]$  (adapted from Smith *et al.*<sup>45</sup>). As expected, *phnD* gene and  
243 transcript abundance were positively correlated with  $R^*$  ( $R^2 = 0.168$ ,  $p < 0.001$  and  $R^2 = 0.197$ ,  
244  $p < 0.001$ , respectively (Fig 5G and 5H), i.e. regions typified with Pi depletion, though notably  
245 Pi concentration alone was a better predictor of *phnD* gene abundance (Fig 5E). However,  
246 *aepX* and *aepP* gene abundance was (weakly) inversely correlated with  $R^*$  ( $R^2 = 0.029$ ,  $p < 0.05$

247 and  $R^2 = 0.108$ ,  $p < 0.001$ , respectively) (Fig 5G) and no significant relationship was found  
248 between  $R^*$  and *aepX/aepP* transcript abundance (Fig 5H). These data are consistent with the  
249 proteomic response of *S. stellulata* under laboratory conditions and suggests AepX is induced  
250 in the presence of 2AEP (substrate-inducible) and not in response to nutrient limitation.

251

## 252 Discussion

253 Both phosphonate biosynthesis<sup>1,3,5,46</sup> and catabolic<sup>1,23,25,43,47-52</sup> genes are ubiquitous in  
254 marine, soil and gut microbiomes, suggesting phosphonate cycling is widespread in nature. In  
255 contrast, the uptake of these molecules is comparatively understudied, with only two  
256 characterised ABC transport systems confirmed, both of which are linked solely to P-  
257 acquisition<sup>29-31</sup>. The abundance of these Pi-sensitive transporters in marine systems is not  
258 equivalent to the abundance of catabolic genes<sup>20</sup>, especially those (*phnWAY*) recently shown  
259 to be involved in Pi-insensitive catabolism<sup>22</sup>. This would suggest our knowledge of the  
260 microbial uptake of phosphonates, particularly 2AEP, is incomplete. Using transporter  
261 expression as a proxy for the cycling of specific nutrients has significantly advanced our  
262 understanding of *in situ* biogeochemical cycling<sup>33,34,53</sup>. This molecular approach helped  
263 resolve the biogenesis of the climate-active gas methane in oxygenated surface waters, driven  
264 through the uptake and degradation of methylphosphonate<sup>21,28</sup>. Thus, a gap in mechanistic  
265 knowledge on 2AEP metabolism impairs our ability to survey the *in situ* cycling of reduced  
266 organophosphorus compounds, especially when high resolution separation of such  
267 compounds is difficult<sup>9,54</sup>. Here, identification of novel Pi-insensitive 2AEP transporters  
268 allowed us to develop molecular markers to investigate the cycling of 2AEP on a global scale  
269 and compare these with previously characterised Pi-sensitive markers.

270 Genomic and biochemical analyses have revealed 2AEP, MPn and HEP are ubiquitously  
271 synthesised in the marine environment in relatively large quantities<sup>1,3,9,25,54</sup>. However,  
272 several studies have shown 2AEP is absent in 'semi-labile' DOM whilst alkylphosphonates,  
273 such as MPn and HEP, tend to accumulate<sup>16,17,55</sup>. Collectively, this would suggest 2AEP is more  
274 susceptible to microbial mineralisation<sup>22</sup> and thus shorter residence times. Here, we reveal  
275 pathways for the Pi-insensitive uptake (*aepX*) and catabolism (*phnA*) are expressed at  
276 significantly higher levels across the global ocean than the Pi-repressible *phnD* and *phnJ*,  
277 providing a clear mechanism for this phenomenon. Our data also suggests 2AEP is  
278 preferentially mineralised independently of both N and P status explaining why phosphonates  
279 are metabolised in regions where phosphate concentrations are high enough to repress C-P  
280 lyase-expression<sup>20,21,56</sup>. Together, this adds further weight to the notion that phosphonates  
281 are rapidly cycled between reduced and oxidised forms<sup>9,22</sup>

282 Substrate inducible expression of 2AEP catabolic genes irrespective of nutrient status  
283 has previously been shown to mineralise organic N to ammonium which can cross feed into  
284 another bacterium<sup>57</sup>. Assuming the link between Pi-insensitive expression of *AepXVW*,  
285 *PhnWAY* and subsequent cellular export of mineralised Pi observed in our laboratory cultures  
286 and as demonstrated by Chen et al.<sup>22</sup> is comparable to *in situ* metabolism, we provide a clear  
287 mechanism to support the hypothesis that phosphonates are a source of regenerated Pi  
288 throughout the water column<sup>15,22</sup>, especially at depths where *AepP* is also found in higher  
289 abundances. Thus, aminophosphonates likely represent another source of regenerated Pi  
290 from DOP, a mechanism which is important for maintaining biological production in Pi-  
291 deplete regions of the ocean<sup>58,59</sup>.

292 In summary, this study identified three novel 2AEP transporters in marine and  
293 terrestrial bacteria capable of Pi-insensitive 2AEP catabolism. One of these, AepXVW, is the  
294 most abundant phosphonate transporter in seawater and is ubiquitously transcribed at high  
295 levels suggesting 2AEP mineralisation is a major process in the marine organic C, N and P  
296 cycles and may present a significant source of regenerated Pi available for oceanic production.  
297 This conclusion is strengthened by two key observations: 1) AepX transcription is not  
298 repressed by standing stock concentrations of Pi, and 2) extracellular Pi export occurs during  
299 Pi-insensitive 2AEP metabolism. Thus, we provide further evidence for the role of low  
300 molecular weight phosphonates acting as a phosphorus currency between autotrophic and  
301 heterotrophic microbes<sup>9</sup>.

302

303

## 304 **Materials and Methods**

### 305 **Bacterial strains and growth conditions**

306 *Pseudomonas* strains used in this work were maintained on Luria Bertani (LB) agar  
307 (1.5% w/v) medium at 30°C. *Stappia stellulata* and the *Roseobacter* strains were maintained  
308 on Marine Broth agar (1.5% w/v) medium at 30°C. *Pseudomonas* mutants and complemented  
309 mutants were maintained on similar plates containing the appropriate antibiotic. For all  
310 growth and proteomics experiments cultures were grown in an adapted Minimal A medium<sup>36</sup>  
311 using Na-Succinate (20 mM) as the sole carbon source and, where applicable, 10mM NH<sub>4</sub>Cl  
312 was added as the sole N source. 2AEP and KH<sub>2</sub>PO<sub>4</sub> were added to a final concentration of 100  
313 μM or 1.5mM as specified in the text and figure legends. *Pseudomonas* strains were pre-  
314 cultured in minimal medium A containing 100 μM Pi and 1.5 mM NH<sub>4</sub>Cl to ensure adequate  
315 growth while minimising the potential for carryover of residual nutrients into experimental  
316 cultures. Co-culture experiments were carried out according to the protocol described in<sup>60</sup>.  
317 Culture experiments were performed using a FLUOStar Omega 96-well plate reader using  
318 Sarstedt 96-well plates incubated at 30°C, shaking at 200 rpm. For 2AEP as N source  
319 proteomics, *S. stellulata* was grown in modified defined marine ammonium mineral salt  
320 (MAMS) media, lacking ammonium, and where HEPES replaced the phosphate buffer<sup>61</sup>. For  
321 all other growth and proteomics experiments these marine bacteria were grown in Sea Salts  
322 media<sup>62</sup>, where HEPES replaced the phosphate buffer.

### 323 **Generation and complementation of *Pseudomonas* mutants**

324 Mutants were generated and complemented via the protocols outlined in<sup>36,60</sup>. A full  
325 list of strains, plasmids, and primers used in this study is presented in Table S4. Briefly, regions  
326 of genomic DNA at the 5' and 3' end of each deletion target were amplified, along with in  
327 some cases the gentamicin resistance cassette from p34S-Gm<sup>63</sup>. Other mutants were

328 constructed to be marker-less, and so lacked this resistance cassette. DNA fragments were  
329 ligated into linearised pk18mobsacB<sup>64</sup> using the HiFi DNA Assembly Kit (New England Biolabs,  
330 Hitchin, UK) according to the manufacturer's instructions. For the construction of plasmids  
331 for complementation, genes and their corresponding promoters were cloned into linearised  
332 pBBR1McS-km again using the HiFi DNA Assembly Kit.

333 Plasmids were subsequently electroporated into *Escherichia coli* S17.1 and mobilized  
334 into *Pseudomonas* by conjugation. Transconjugants were selected via gentamicin (50 µg ml<sup>-1</sup>)  
335 or kanamycin (50 µg ml<sup>-1</sup>) selection and chloramphenicol (10 µg ml<sup>-1</sup>) counter-selection. Single  
336 crossover transconjugants were identified by PCR, and double crossover mutants selected via  
337 plating on LB containing either gentamicin or no antibiotic, and 10% (w/v) sucrose.  
338 Homologous recombination was confirmed by PCR and Sanger sequencing. Complemented  
339 mutants were selected using kanamycin with chloramphenicol counter-selection.

#### 340 **Proteomics preparation and analysis**

341 To identify proteins involved in 2AEP uptake and catabolism in *S. stellulata*, total  
342 protein (n=3 for each treatment), was retrieved by sampling cell cultures (OD<sub>540</sub> 0.8-1.0) and  
343 pelleting cells (centrifugation at 16200 x g for 5 mins). Cell lysis was achieved via boiling in  
344 100µl LDS buffer (Expedeon) prior to loading 20µl onto a 4-20% Bis-Tris SDS precast gel  
345 (Expedeon). For enrichment of the membrane protein fraction of the *P. putida* BIRD-1  
346 *aepXVW::gm* mutant, we adapted the methods outlined in<sup>36</sup>. Briefly, 40ml cells were grown  
347 to an OD<sub>600</sub> of 0.8-1.0, a volume equivalent to 5 OD<sub>600</sub> units was centrifuged at 2000 x g for 10  
348 mins, resuspended in 1ml 50mM Tris-HCl, pH 7.6, and centrifuged at 2000 x g for 10 mins.  
349 Pellets were resuspended in 500µl 200mM MgCl<sub>2</sub>, 50mM Tris-HCl, pH 7.6 and incubated for  
350 30 mins at 30°C with gentle shaking. Cells were cooled on ice for 5 mins and incubated at

351 room temperature for 15 mins, then centrifuged at 8000 x g for 10 mins at 4°C. Pellets were  
352 washed in 1ml 50mM Tris-HCl, pH 7.6, and again centrifuged at 8000 x g for 10 mins at 4°C.  
353 Pellets were resuspended in 0.5ml 50mM Tris-HCl, pH 7.6, and sonicated for 30 s twice, on  
354 ice. The solution was centrifuged at 2000 x g for 15 mins at 4°C, and supernatants were then  
355 ultracentrifuged at 120000 x g for 45 mins at 4°C. Following this, the supernatant was  
356 discarded and the pellet resuspended in 50µl LDS buffer (Expedeon) prior to loading 20µl onto  
357 a 4-20% Bis-Tris SDS precast gel (Expedeon). Gel sections were de-stained with 50mM  
358 ammonium bicarbonate in 50% ethanol, dehydrated with 100% ethanol, reduced and  
359 alkylated with Tris-2-carboxyethylphosphine (TCEP) and iodoacetamide (IAA), washed with  
360 50mM ammonium bicarbonate in 50% ethanol and dehydrated with 100% ethanol prior to  
361 overnight digestion with trypsin. Peptides were extracted and analysed using an Orbitrap  
362 Fusion Ultimate 3000 RSLCNano System (Thermo Scientific) in electrospray ionization mode  
363 at the Warwick Proteomics Research Technology Platform.

364 Resulting MS/MS files were searched against the relevant protein sequence database  
365 (*P. putida* BIRD-1 , NC\_017530.1, *S. stellulata*, GCF\_000423715.1) using MaxQuant<sup>65</sup> with  
366 default settings and quantification was achieved using Label Free Quantification (LFQ). The  
367 proteomics analysis software Perseus (1.6.12)<sup>66</sup> was used to identify differentially expressed  
368 proteins based on LFQ values, using a False Discovery Rate (FDR) of 0.01. Identified proteins  
369 were retained if they were present in at least two biological replicates within a treatment.  
370 Missing (N/A) values were imputed from a normal distribution using the default parameters.  
371 Differential expression was identified by two-sample Student's *t*-test, using an  $s_0$  constant of  
372 0.1, or ANOVA, where appropriate.

373 **Bioinformatics analyses of *aepXVW/aepP***

374 AepX homologs were identified using the IMG/JGI (<https://img.jgi.doe.gov/>)  
375 'Customized Homolog Display' search tool. Strains containing homologs were identified (cut-  
376 off values:  $e^{-70}$ , min. identity 40%), preferentially from type strains. In addition, representative  
377 strains from soil and marine environments were added to this list (see Table S2). Protein  
378 sequences were aligned using ClustalOmega<sup>67</sup> and profile Hidden Markov Models (pHMMs)  
379 were constructed from these sequences using the hmmbuild function of hmmer 3.3  
380 (<http://hmmer.org>)<sup>68</sup>. The previously characterised *Escherichia coli* K-12 PhnD<sup>69</sup> and the  
381 SAR11 clade isolate *Pelagibacter* sp. HTCC7211 PhnD<sup>21</sup> showed surprisingly low sequence  
382 homology (BLAST %ID 28.46%, query coverage 76%, e-value 2e-25). We therefore developed  
383 two pHMMs for PhnD to reflect this. There was no overlap in environmental sequences  
384 retrieved from each search using either hmm model. Therefore, abundance counts for each  
385 PhnD form were combined together as a collective PhnD group. These pHMMs were used to  
386 search the TARA ocean metagenome (OM-RGC\_v2\_metaG) and metatranscriptome (OM-  
387 RGC\_v2\_metaT) via the Ocean Gene Atlas web interface<sup>44</sup>, using a stringency of  $1E^{-80}$ .  
388 Sequence abundances were expressed as average percentage of genomes containing a copy  
389 by dividing the percentage of total mapped reads by the median abundance (as a percentage  
390 of total mapped reads) of 10 single-copy marker genes<sup>70</sup> for both MG and MT. The pHMMs  
391 were used to search the soil MG via hmmsearch<sup>68</sup> using the same stringency as above.  
392 Similarly, abundances were calculated as average percentage of genomes containing a copy  
393 as above.

394 Phylogenetic analyses were performed using IQ-TREE 2<sup>71</sup> using the following  
395 parameters: -m TEST -bb 1000 -alrt 1000. Evolutionary relationships were inferred by  
396 maximum-likelihood analysis, and visualised using the Interactive Tree of Life (iTOL) v5.6.3  
397 online platform (<https://itol.embl.de/>)<sup>72</sup>

## 398 **Statistical analysis**

399 Unless specified above, all statistical analysis was performed using R (version 4.02)<sup>73</sup>, within  
400 the RStudio programme (version 1.3)<sup>74</sup>.

## 401 **Acknowledgements**

402 We thank the Warwick Proteomics Research Facility, namely Dr. Cleidiane Zampronio for her  
403 assistance in generating and processing the mass-spectrometry data. This study was funded  
404 by the Biotechnology and Biological Sciences Research Council (BBSRC) under project codes  
405 BB/L026074/1 and BB/T009152/1 linked to The Soil and Rhizosphere Interactions for  
406 Sustainable Agri-ecosystems (SARISA) programme and a Discovery Fellowship (IL),  
407 respectively.

408

## 409 **References**

- 410 1 Villarreal-Chiu, J. F., Quinn, J. P. & McGrath, J. W. The genes and enzymes of phosphonate  
411 metabolism by bacteria, and their distribution in the marine environment. *Front Microbiol* **3**,  
412 19, doi:10.3389/fmicb.2012.00019 (2012).
- 413 2 Mukhamedova, K. S. & Glushenkova, A. I. Natural Phosphonolipids. *Chem Natural*  
414 *Compounds* **36**, 329-341, doi:10.1023/A:1002804409503 (2000).
- 415 3 Yu, X. *et al.* Diversity and abundance of phosphonate biosynthetic genes in nature. *Proc Natl*  
416 *Acad Sci U S A* **110**, 20759-20764, doi:10.1073/pnas.1315107110 (2013).
- 417 4 Ju, K.-S., Doroghazi, J. R. & Metcalf, W. W. Genomics-enabled discovery of phosphonate  
418 natural products and their biosynthetic pathways. *J Industrial Microbiol & Biotechnol* **41**,  
419 345-356, doi:10.1007/s10295-013-1375-2 (2014).
- 420 5 Peck, S. C. & van der Donk, W. A. Phosphonate biosynthesis and catabolism: a treasure trove  
421 of unusual enzymology. *Curr Opin Chem Biol* **17**, 580-588,  
422 doi:<https://doi.org/10.1016/j.cbpa.2013.06.018> (2013).
- 423 6 Metcalf, W. W. *et al.* Synthesis of methylphosphonic acid by marine microbes: a source for  
424 methane in the aerobic ocean. *Science* **337**, 1104-1107, doi:10.1126/science.1219875  
425 (2012).
- 426 7 Born, D. A. *et al.* Structural basis for methylphosphonate biosynthesis. *Science* **358**, 1336,  
427 doi:10.1126/science.aao3435 (2017).
- 428 8 Hildebrand, R. *The Role of Phosphonates in Living Systems*. (CRC Press, 1983).
- 429 9 Van Mooy, B. A. S. *et al.* Major role of planktonic phosphate reduction in the marine  
430 phosphorus redox cycle. *Science* **348**, 783, doi:10.1126/science.aaa8181 (2015).

431 10 Kolowith, L. C., Ingall, E. D. & Benner, R. Composition and cycling of marine organic  
432 phosphorus. *Limnol Oceanogr* **46**, 309-320, doi:10.4319/lo.2001.46.2.0309 (2001).

433 11 Clark, L. L., Ingall, E. D. & Benner, R. Marine phosphorus is selectively remineralized. *Nature*  
434 **393**, 426-426, doi:10.1038/30881 (1998).

435 12 Turner, B. L., Baxter, R., Mahieu, N., Sjögersten, S. & Whitton, B. A. Phosphorus compounds  
436 in subarctic Fennoscandian soils at the mountain birch (*Betula pubescens*)—tundra ecotone.  
437 *Soil Biol Biochem* **36**, 815-823, doi:<https://doi.org/10.1016/j.soilbio.2004.01.011> (2004).

438 13 Tate, K. R. & Newman, R. H. Phosphorus fractions of a climosequence of soils in New Zealand  
439 tussock grassland. *Soil Biol Biochem* **14**, 191-196, doi:[https://doi.org/10.1016/0038-](https://doi.org/10.1016/0038-0717(82)90022-0)  
440 [0717\(82\)90022-0](https://doi.org/10.1016/0038-0717(82)90022-0) (1982).

441 14 Cade-Menun, B. J., Navaratnam, J. A. & Walbridge, M. R. Characterizing dissolved and  
442 particulate phosphorus in water with 31P nuclear magnetic resonance spectroscopy. *Environ*  
443 *Sci Technol* **40**, 7874-7880, doi:10.1021/es061843e (2006).

444 15 Clark, L. L., Ingall, E. D. & Benner, R. Marine organic phosphorus cycling; novel insights from  
445 nuclear magnetic resonance. *American J Sci* **299**, 724-737 (1999).

446 16 Repeta, D. J. *et al.* Marine methane paradox explained by bacterial degradation of dissolved  
447 organic matter. *Nature Geosci* **9**, 884-887, doi:10.1038/ngeo2837 (2016).

448 17 Sosa, O. A. *et al.* Phosphonate cycling supports methane and ethylene supersaturation in the  
449 phosphate-depleted western North Atlantic Ocean. *Limnol Oceanogr* **65**, 2443-2459,  
450 doi:10.1002/lno.11463 (2020).

451 18 Dyhrman, S. T., Benitez-Nelson, C. R., Orchard, E. D., Haley, S. T. & Pellechia, P. J. A microbial  
452 source of phosphonates in oligotrophic marine systems. *Nature Geoscience* **2**, 696-699,  
453 doi:10.1038/ngeo639 (2009).

454 19 Acker, M. *et al.* Phosphonate production by marine microbes: exploring new sources and  
455 potential function. *bioRxiv*, 2020.2011.2004.368217, doi:10.1101/2020.11.04.368217 (2020).

456 20 Sosa, O. A., Repeta, D. J., DeLong, E. F., Ashkezari, M. D. & Karl, D. M. Phosphate-limited  
457 ocean regions select for bacterial populations enriched in the carbon-phosphorus lyase  
458 pathway for phosphonate degradation. *Environ Microbiol* **21**, 2402-2414, doi:10.1111/1462-  
459 2920.14628 (2019).

460 21 Carini, P., White, A. E., Campbell, E. O. & Giovannoni, S. J. Methane production by  
461 phosphate-starved SAR11 chemoheterotrophic marine bacteria. *Nature Comms* **5**, 4346,  
462 doi:10.1038/ncomms5346 (2014).

463 22 Chin, J. P., Quinn, J. P. & McGrath, J. W. Phosphate insensitive aminophosphonate  
464 mineralisation within oceanic nutrient cycles. *ISME J* **12**, 973-980, doi:10.1038/s41396-017-  
465 0031-7 (2018).

466 23 Ternan, N. G. & Quinn, J. P. Phosphate starvation-independent 2-aminoethylphosphonic acid  
467 biodegradation in a newly isolated strain of *Pseudomonas putida*, NG2. *Syst Appl Microbiol*  
468 **21**, 346-352, doi:10.1016/S0723-2020(98)80043-X (1998).

469 24 White, A. K. & Metcalf, W. W. Microbial metabolism of reduced phosphorus compounds.  
470 *Annu Rev Microbiol* **61**, 379-400, doi:10.1146/annurev.micro.61.080706.093357 (2007).

471 25 Martinez, A., Tyson, G. W. & DeLong, E. F. Widespread known and novel phosphonate  
472 utilization pathways in marine bacteria revealed by functional screening and metagenomic  
473 analyses. *Environ Microbiol* **12**, 222-238, doi:10.1111/j.1462-2920.2009.02062.x (2010).

474 26 Wanner, B. L. Molecular genetics of carbon-phosphorus bond cleavage in bacteria.  
475 *Biodegradation* **5**, 175-184 (1994).

476 27 Kononova, S. V. & Nesmeyanova, M. A. Phosphonates and their degradation by  
477 microorganisms. *Biochemistry (Mosc)* **67**, 184-195 (2002).

478 28 Sowell, S. M. *et al.* Transport functions dominate the SAR11 metaproteome at low-nutrient  
479 extremes in the Sargasso Sea. *ISME J* **3**, 93-105,  
480 doi:<http://www.nature.com/ismej/journal/v3/n1/supinfo/ismej200883s1.html> (2008).

481 29 Alicea, I. *et al.* Structure of the *Escherichia coli* phosphonate binding protein PhnD and  
482 rationally optimized phosphonate biosensors. *J Mol Biol* **414**, 356-369,  
483 doi:10.1016/j.jmb.2011.09.047 (2011).

484 30 Rizk, S. S., Cuneo, M. J. & Hellinga, H. W. Identification of cognate ligands for the *Escherichia*  
485 *coli phnD* protein product and engineering of a reagentless fluorescent biosensor for  
486 phosphonates. *Protein Sci* **15**, 1745-1751, doi:10.1110/ps.062135206 (2006).

487 31 Jiang, W., Metcalf, W. W., Lee, K. S. & Wanner, B. L. Molecular cloning, mapping, and  
488 regulation of Pho regulon genes for phosphonate breakdown by the phosphonate  
489 pathway of *Salmonella typhimurium* LT2. *J Bacteriol* **177**, 6411, doi:10.1128/jb.177.22.6411-  
490 6421.1995 (1995).

491 32 Kim, A. D. *et al.* The 2-aminoethylphosphonate-specific transaminase of the 2-  
492 aminoethylphosphonate degradation pathway. *J Bacteriol* **184**, 4134,  
493 doi:10.1128/JB.184.15.4134-4140.2002 (2002).

494 33 Ottesen, E. A. *et al.* Pattern and synchrony of gene expression among sympatric marine  
495 microbial populations. *Proc Natl Acad Sci USA* **110**, E488-E497,  
496 doi:10.1073/pnas.1222099110 (2013).

497 34 Lidbury, I., Murrell, J. C. & Chen, Y. Trimethylamine N-oxide metabolism by abundant marine  
498 heterotrophic bacteria. *Proc Natl Acad Sci USA* **111**, 2710-2715,  
499 doi:10.1073/pnas.1317834111 (2014).

500 35 Mauchline, T. H. *et al.* Mapping the *Sinorhizobium meliloti* 1021 solute-binding protein-  
501 dependent transportome. *Proc Natl Acad Sci* **103**, 17933-17938,  
502 doi:10.1073/pnas.0606673103 (2006).

503 36 Lidbury, I. D. *et al.* Comparative genomic, proteomic and exoproteomic analyses of three  
504 *Pseudomonas* strains reveals novel insights into the phosphorus scavenging capabilities of  
505 soil bacteria. *Environ Microbiol* **18**, 3535-3549, doi:10.1111/1462-2920.13390 (2016).

506 37 Lemieux, M. J., Huang, Y. & Wang, D. N. Crystal structure and mechanism of GlpT, the  
507 glycerol-3-phosphate transporter from *E. coli*. *J Electron Microsc* **54 Suppl 1**, i43-46,  
508 doi:10.1093/jmicro/54.suppl\_1.i43 (2005).

509 38 Elvin, C. M., Hardy, C. M. & Rosenberg, H. Pi exchange mediated by the GlpT-dependent sn-  
510 glycerol-3-phosphate transport system in *Escherichia coli*. *J Bacteriol* **161**, 1054-1058 (1985).

511 39 Enkavi, G. & Tajkhorshid, E. Simulation of spontaneous substrate binding revealing the  
512 binding pathway and mechanism and initial conformational response of GlpT. *Biochemistry*  
513 **49**, 1105-1114, doi:10.1021/bi901412a (2010).

514 40 Law, C. J. *et al.* Salt-bridge dynamics control substrate-induced conformational change in the  
515 membrane transporter GlpT. *J Mol Biol* **378**, 828-839,  
516 doi:<https://doi.org/10.1016/j.jmb.2008.03.029> (2008).

517 41 Law, C. J., Enkavi, G., Wang, D.-N. & Tajkhorshid, E. Structural basis of substrate selectivity in  
518 the glycerol-3-phosphate: phosphate antiporter GlpT. *Biophysical J* **97**, 1346-1353,  
519 doi:10.1016/j.bpj.2009.06.026 (2009).

520 42 Moradi, M., Enkavi, G. & Tajkhorshid, E. Atomic-level characterization of transport cycle  
521 thermodynamics in the glycerol-3-phosphate:phosphate antiporter. *Nature Comms* **6**, 8393,  
522 doi:10.1038/ncomms9393 (2015).

523 43 Borisova, S. A. *et al.* Genetic and biochemical characterization of a pathway for the  
524 degradation of 2-aminoethylphosphonate in *Sinorhizobium meliloti* 1021. *J Biol Chem* **286**,  
525 22283-22290, doi:10.1074/jbc.M111.237735 (2011).

526 44 Villar, E. *et al.* The Ocean Gene Atlas: exploring the biogeography of plankton genes online.  
527 *Nucleic Acids Res* **46**, W289-W295, doi:10.1093/nar/gky376 (2018).

528 45 Smith, A. F. *et al.* Elucidation of glutamine lipid biosynthesis in marine bacteria reveals its  
529 importance under phosphorus deplete growth in *Rhodobacteraceae*. *ISME J* **13**, 39-49,  
530 doi:10.1038/s41396-018-0249-z (2019).

531 46 Chin, J. P., McGrath, J. W. & Quinn, J. P. Microbial transformations in phosphonate  
532 biosynthesis and catabolism, and their importance in nutrient cycling. *Curr Opin Chem Biol*  
533 **31**, 50-57, doi:10.1016/j.cbpa.2016.01.010 (2016).

534 47 White, A. K. & Metcalf, W. W. Two C-P lyase operons in *Pseudomonas stutzeri* and their roles  
535 in the oxidation of phosphonates, phosphite, and hypophosphite. *J Bacteriol* **186**, 4730-  
536 4739, doi:10.1128/JB.186.14.4730-4739.2004 (2004).

537 48 Ermakova, I. T. *et al.* Organophosphonates utilization by soil strains of *Ochrobactrum*  
538 *anthropi* and *Achromobacter* sp. *Arch Microbiol* **199**, 665-675, doi:10.1007/s00203-017-  
539 1343-8 (2017).

540 49 Hartley, L. E., Kaakoush, N. O., Ford, J. L., Korolik, V. & Mendz, G. L. Characterisation of  
541 *Campylobacter jejuni* genes potentially involved in phosphonate degradation. *Gut Pathog* **1**,  
542 13, doi:10.1186/1757-4749-1-13 (2009).

543 50 Imazu, K. *et al.* Enhanced utilization of phosphonate and phosphite by *Klebsiella aerogenes*.  
544 *Appl Environ Microbiol* **64**, 3754-3758 (1998).

545 51 Mendz, G. L., Mégraud, F. & Korolik, V. Phosphonate catabolism by *Campylobacter* spp. *Arch*  
546 *Microbiol* **183**, 113-120, doi:10.1007/s00203-004-0752-7 (2005).

547 52 Metcalf, W. W. & Wanner, B. L. Evidence for a fourteen-gene, *phnC* to *phnP* locus for  
548 phosphonate metabolism in *Escherichia coli*. *Gene* **129**, 27-32, doi:10.1016/0378-  
549 1119(93)90692-v (1993).

550 53 Hultman, J. *et al.* Multi-omics of permafrost, active layer and thermokarst bog soil  
551 microbiomes. *Nature* **521**, 208-212, doi:10.1038/nature14238 (2015).

552 54 Frischkorn, K. R. *et al.* *Trichodesmium* physiological ecology and phosphate reduction in the  
553 western tropical South Pacific. *Biogeosciences* **15**, 5761-5778, doi:10.5194/bg-15-5761-2018  
554 (2018).

555 55 Young, C. L. & Ingall, E. D. Marine dissolved organic phosphorus composition: insights from  
556 samples recovered using combined electro dialysis/ reverse osmosis. *Aquatic Geochem* **16**,  
557 563-574, doi:10.1007/s10498-009-9087-y (2010).

558 56 Benitez-Nelson, C. R., O'Neill, L., Kolowith, L. C., Pellechia, P. & Thunell, R. Phosphonates and  
559 particulate organic phosphorus cycling in an anoxic marine basin. *Limnol Oceanogr* **49**, 1593-  
560 1604, doi:10.4319/lo.2004.49.5.1593 (2004).

561 57 Lidbury, I. D. E. A., Murrell, J. C. & Chen, Y. Trimethylamine and trimethylamine N-oxide are  
562 supplementary energy sources for a marine heterotrophic bacterium: implications for  
563 marine carbon and nitrogen cycling. *ISME J* **9**, 760-769, doi:10.1038/ismej.2014.149 (2015).

564 58 Dyhrman, S. T. *et al.* Phosphonate utilization by the globally important marine diazotroph  
565 *Trichodesmium*. *Nature* **439**, 68-71, doi:10.1038/nature04203 (2006).

566 59 Karl, D. M. *et al.* Aerobic production of methane in the sea. *Nature Geosci* **1**, 473-478,  
567 doi:10.1038/ngeo234 (2008).

568 60 Lidbury, I. D. E. A. *et al.* Identification of extracellular glycerophosphodiesterases in  
569 *Pseudomonas* and their role in soil organic phosphorus remineralisation. *Sci Rep* **7**, 2179,  
570 doi:10.1038/s41598-017-02327-6 (2017).

571 61 Lidbury, I., Kimberley, G., Scanlan, D. J., Murrell, J. C. & Chen, Y. Comparative genomics and  
572 mutagenesis analyses of choline metabolism in the marine *Roseobacter* clade. *Environ*  
573 *Microbiol* **17**, 5048-5062 (2015).

574 62 Chen, Y. Comparative genomics of methylated amine utilization by marine *Roseobacter* clade  
575 bacteria and development of functional gene markers (*tmm*, *gmaS*). *Environ Microbiol* **14**,  
576 2308-2322, doi:10.1111/j.1462-2920.2012.02765.x (2012).

577 63 Dennis, J. J. & Zylstra, G. J. Plasposons: modular self-cloning minitransposon derivatives for  
578 rapid genetic analysis of gram-negative bacterial genomes. *Appl Environ Microbiol* **64**, 2710-  
579 2715 (1998).

580 64 Schäfer, A. *et al.* Small mobilizable multi-purpose cloning vectors derived from the  
581 *Escherichia coli* plasmids pK18 and pK19: selection of defined deletions in the chromosome

582 of *Corynebacterium glutamicum*. *Gene* **145**, 69-73, doi:10.1016/0378-1119(94)90324-7  
583 (1994).

584 65 Tyanova, S., Temu, T. & Cox, J. The MaxQuant computational platform for mass  
585 spectrometry-based shotgun proteomics. *Nature Protocols* **11**, 2301-2319,  
586 doi:10.1038/nprot.2016.136 (2016).

587 66 Tyanova, S. *et al.* The Perseus computational platform for comprehensive analysis of  
588 (prote)omics data. *Nature Methods* **13**, 731-740, doi:10.1038/nmeth.3901 (2016).

589 67 Madeira, F. *et al.* The EMBL-EBI search and sequence analysis tools APIs in 2019. *Nucleic  
590 Acids Res* **47**, W636-W641, doi:10.1093/nar/gkz268 (2019).

591 68 Finn, R. D., Clements, J. & Eddy, S. R. HMMER web server: interactive sequence similarity  
592 searching. *Nucleic Acids Res* **39**, W29-W37, doi:10.1093/nar/gkr367 (2011).

593 69 Wackett, L. P., Wanner, B. L., Venditti, C. P. & Walsh, C. T. Involvement of the phosphate  
594 regulon and the *psiD* locus in carbon-phosphorus lyase activity of *Escherichia coli* K-12. *J  
595 Bacteriol* **169**, 1753-1756, doi:10.1128/jb.169.4.1753-1756.1987 (1987).

596 70 Milanese, A. *et al.* Microbial abundance, activity and population genomic profiling with  
597 mOTUs2. *Nature Comms* **10**, 1014, doi:10.1038/s41467-019-08844-4 (2019).

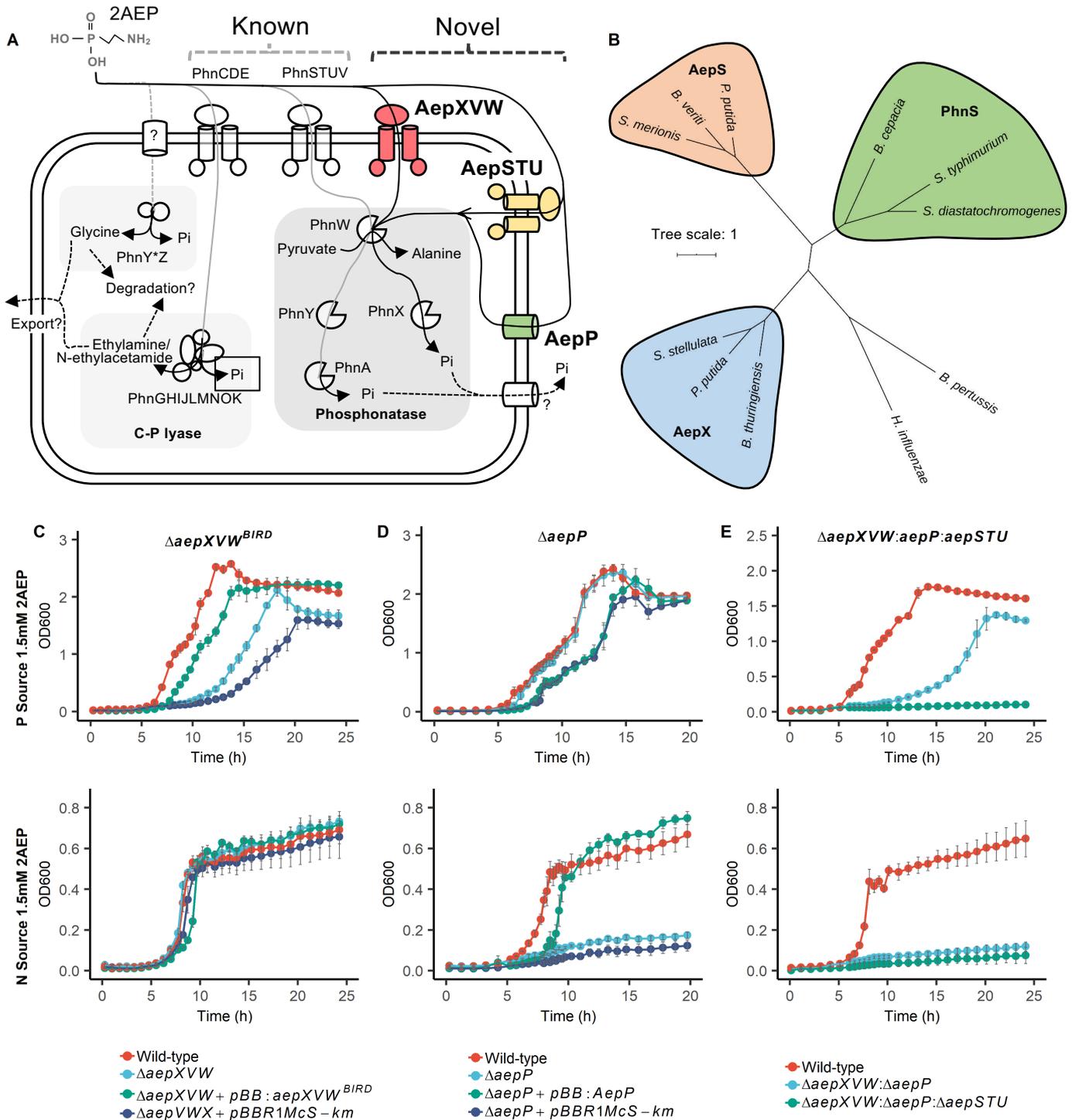
598 71 Minh, B. Q. *et al.* IQ-TREE 2: New models and efficient methods for phylogenetic inference in  
599 the genomic era. *Mol Biol Evol* **37**, 1530-1534, doi:10.1093/molbev/msaa015 (2020).

600 72 Letunic, I. & Bork, P. Interactive Tree Of Life (iTOL) v4: recent updates and new  
601 developments. *Nucleic Acids Res* **47**, W256-W259, doi:10.1093/nar/gkz239 (2019).

602 73 (2020)., R. C. T. *R: A language and environment for statistical computing*. R Foundation for  
603 Statistical Computing, Vienna, Austria., <<https://www.R-project.org/>> (2020).

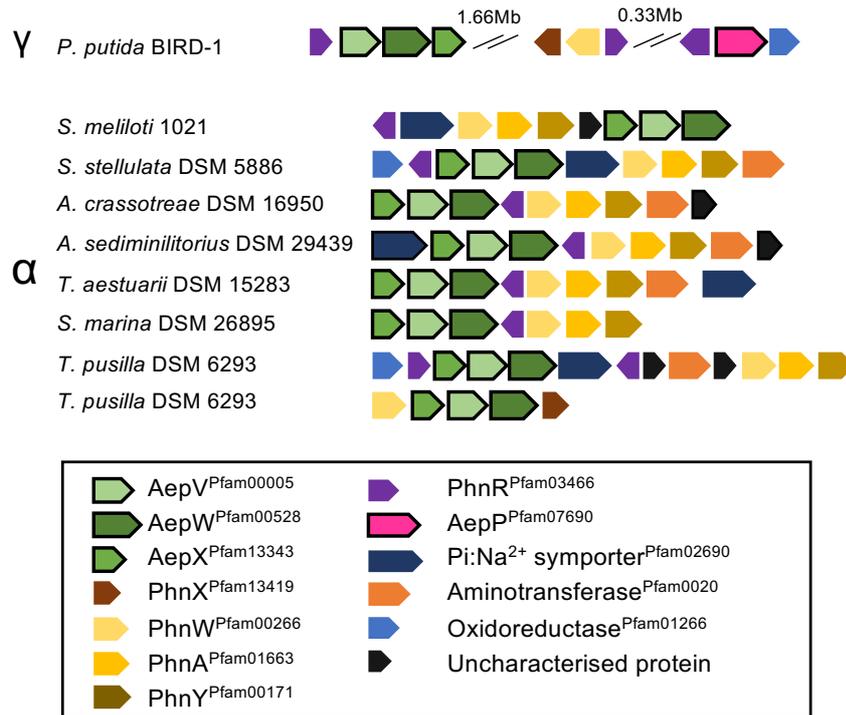
604 74 (2020)., R. T. *RStudio: Integrated Development for R.*, <<http://www.rstudio.com/>> (2020).

605

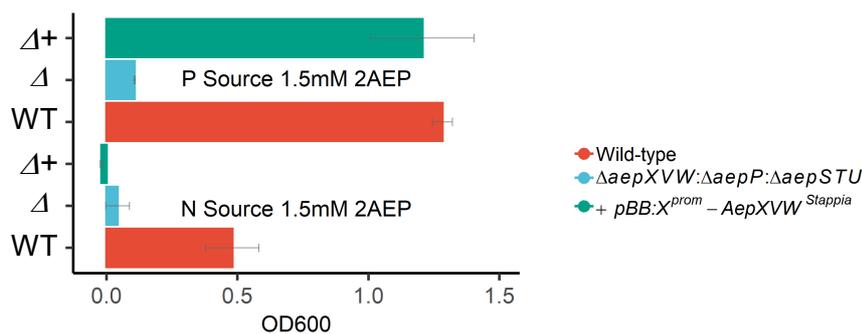


**Figure 1. 2AEP transport and catabolism in *P. putida* BIRD-1.** (A) Schematic representation of novel (highlighted in bold and coloured) routes for 2AEP transport, together with existing characterised and putative 2AEP transport routes. Each catabolic system for degradation is highlighted and includes i) the phosphonate system comprising a 2AEP-pyruvate transaminase (PhnW) and either a phosphonoacetaldehyde hydrolase (PhnX)<sup>74,75</sup> or a NAD<sup>+</sup>-dependent phosphonoacetaldehyde dehydrogenase (PhnY) and a phosphonoacetate hydrolase (PhnA)<sup>43</sup>, and ii) the PhnY\*Z system comprising phosphohydrolase (PhnZ)<sup>21,76</sup> and a 2-oxoglutarate dioxygenase (PhnY\*)<sup>77</sup>. In addition, the promiscuous multi subunit enzyme C-P lyase (PhnGHIJKLMN) can also act on 2AEP<sup>78</sup>, as well as alkylphosphonates<sup>14,15,23</sup>. Pathways found in BIRD-1 are represented by black lines, pathways absent from BIRD-1 are shaded grey. Characterised pathways are shown with solid lines, uncharacterised pathways are shown with dashed lines. Transporters found in BIRD-1 are red if Pi-sensitive, green if Pi-insensitive, and yellow if constitutive. Unknown mechanisms are denoted by a '?'. (B) Phylogenetic tree of AepX, PhnS, and AepS, using the characterised Fe<sup>3+</sup> substrate binding protein FBPa from *Haemophilus influenzae* and *Bordetella pertussis* as an outgroup. *P. putida* = *Pseudomonas putida* BIRD-1, *S. stellulata* = *Stappia stellulata*, *B. cepacia* = *Burkholderia cepacia*, *B. vireti* = *Bacillus vireti*, *S. merionis* = *Streptococcus merionis*, *S. typhimurium* = *Salmonella typhimurium*, *S. diastatochromogenes* = *Streptomyces diastatochromogenes*, *B. thuringiensis* = *Bacillus thuringiensis*. (C) Growth (n=4) of *P. putida* BIRD-1 wild type,  $\Delta aepXVW$ , and the complemented mutant. (D),  $\Delta aepP$  and the complemented mutant, and (E), the 2AEP null mutant,  $\Delta aepXVW::aepP::aepSTU$ . All strains used 2AEP as a sole P (top panel) or N (bottom panel) source. Error bars denote standard deviation of the mean.

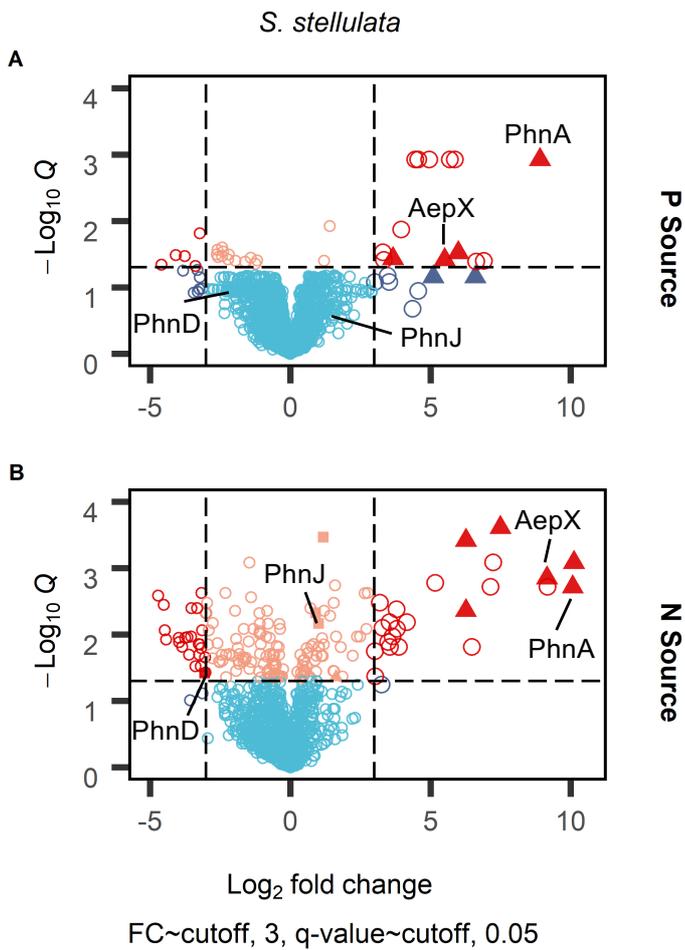
A



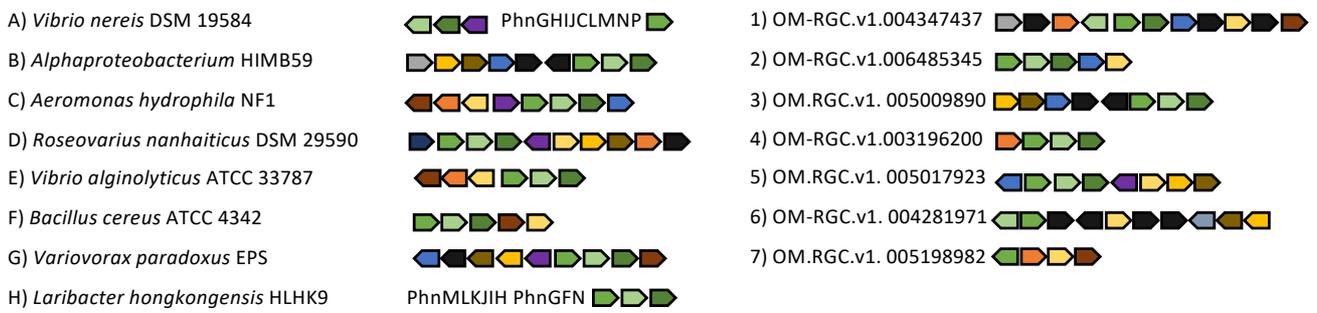
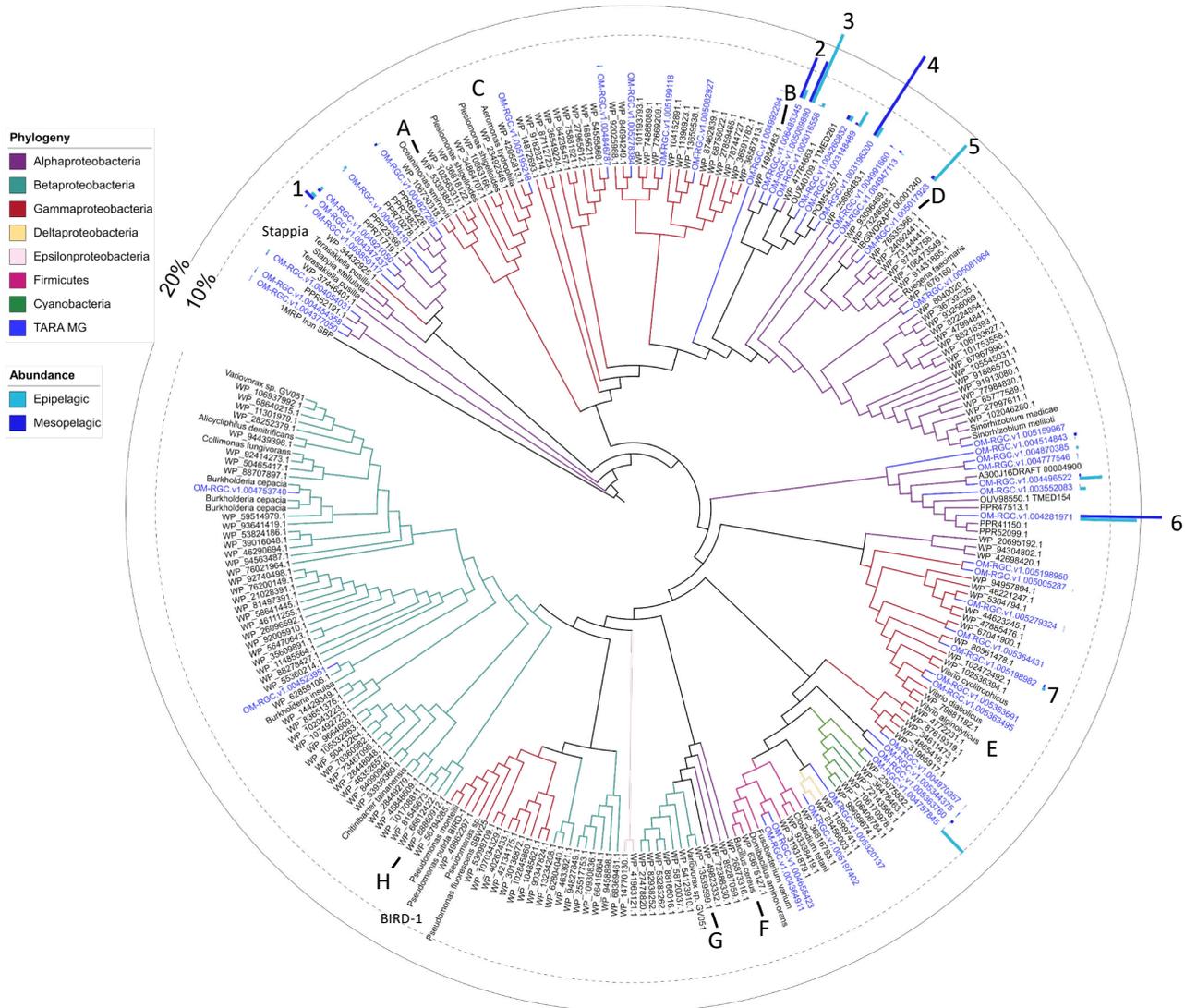
B



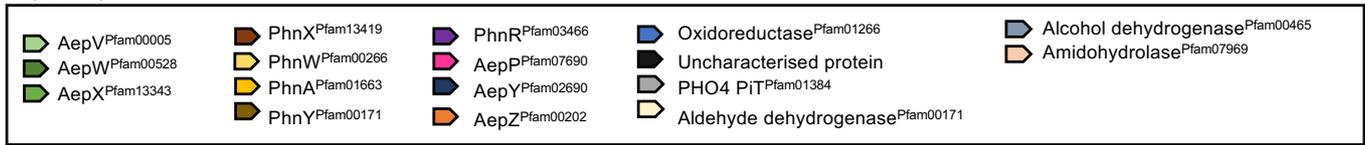
**Figure 2. Distribution and functional characterisation of AepXVW in marine bacteria. (A)** Genetic neighbourhoods of *aepXVW* within marine *Alpha*- and terrestrial *Gamma*-proteobacteria. Strains shown are *Pseudomonas putida* BIRD-1, *Sinorhizobium meliloti* 1021, *Stappia stellulata* DSM 5886, *Aliiroseovarius crassostreae* DSM 16950, *Aliiroseovarius sediminilitorius* DSM 29439, *Thalassobius aestuarii* DSM 15283, and *Shimia marina* DSM 26895. ORFs separated on the genome are indicated by breaks with the corresponding gap given in megabases (Mb). **(B)** Growth of the *P. putida* BIRD-1 triple 2AEP transporter mutant ( $\Delta aepXVW:aepP:aepSTU:gm$ ) complemented with *aepXVW*<sup>Stappia</sup> concatenated with the promoter region from *aepXVW*<sup>BIRD</sup> on 2AEP as either a sole N (60 h) or P (48 h) source. Data represents the mean of triplicates cultures. Error bars denote standard deviation.



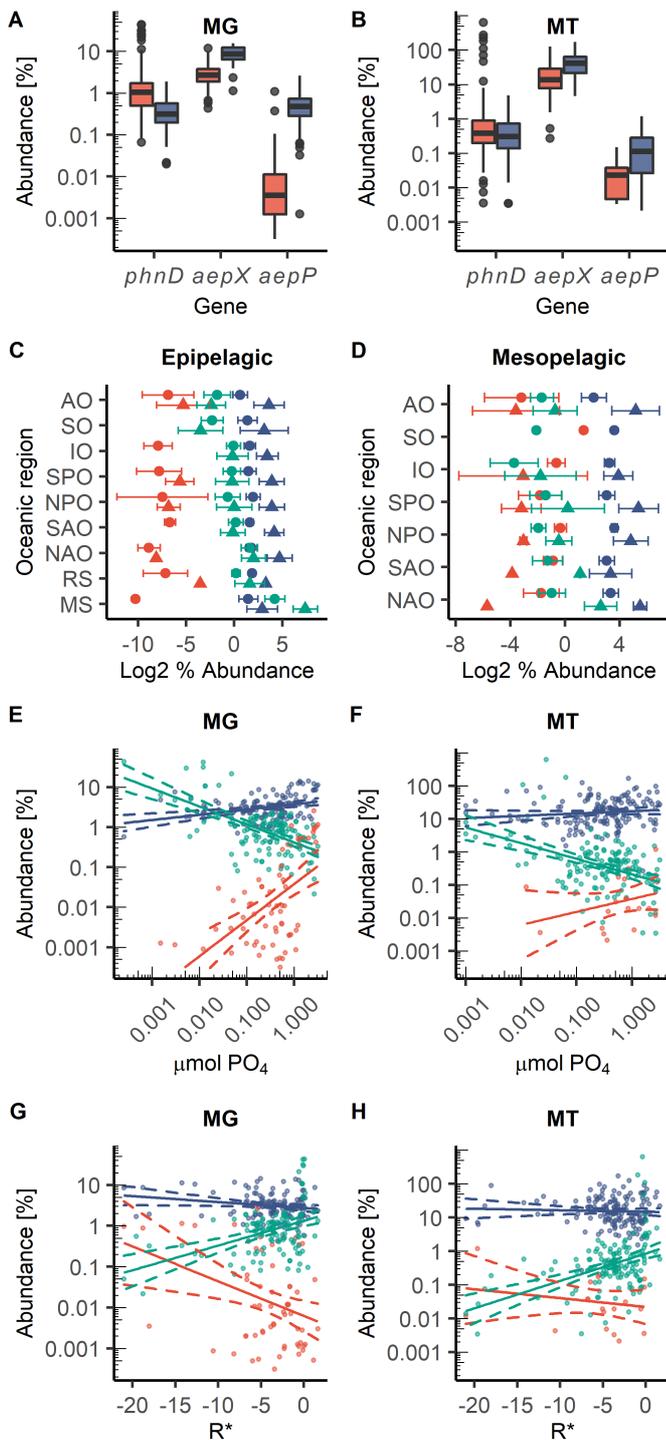
**Figure 3. Proteomic analysis of 2AEP-grown *S. stellulata* cells.** Whole-cell protein expression profiles (n=3) for *S. stellulata* grown using either Pi or 2AEP as sole P source (**A**) or NH<sub>4</sub> or 2AEP as the sole N source (**B**). Fold change represents the difference in Log<sub>2</sub> LFQ values between each treatment and the statistical value on the Y axis is generated from Q values (FDR corrected P values). Members of the *aepXVW-phnWAY* operon are shown as triangles, members of the CP lyase operon are shown as squares, all other proteins are shown as open circles. Data plotted represents the mean of triplicate cultures.



**ORFs**



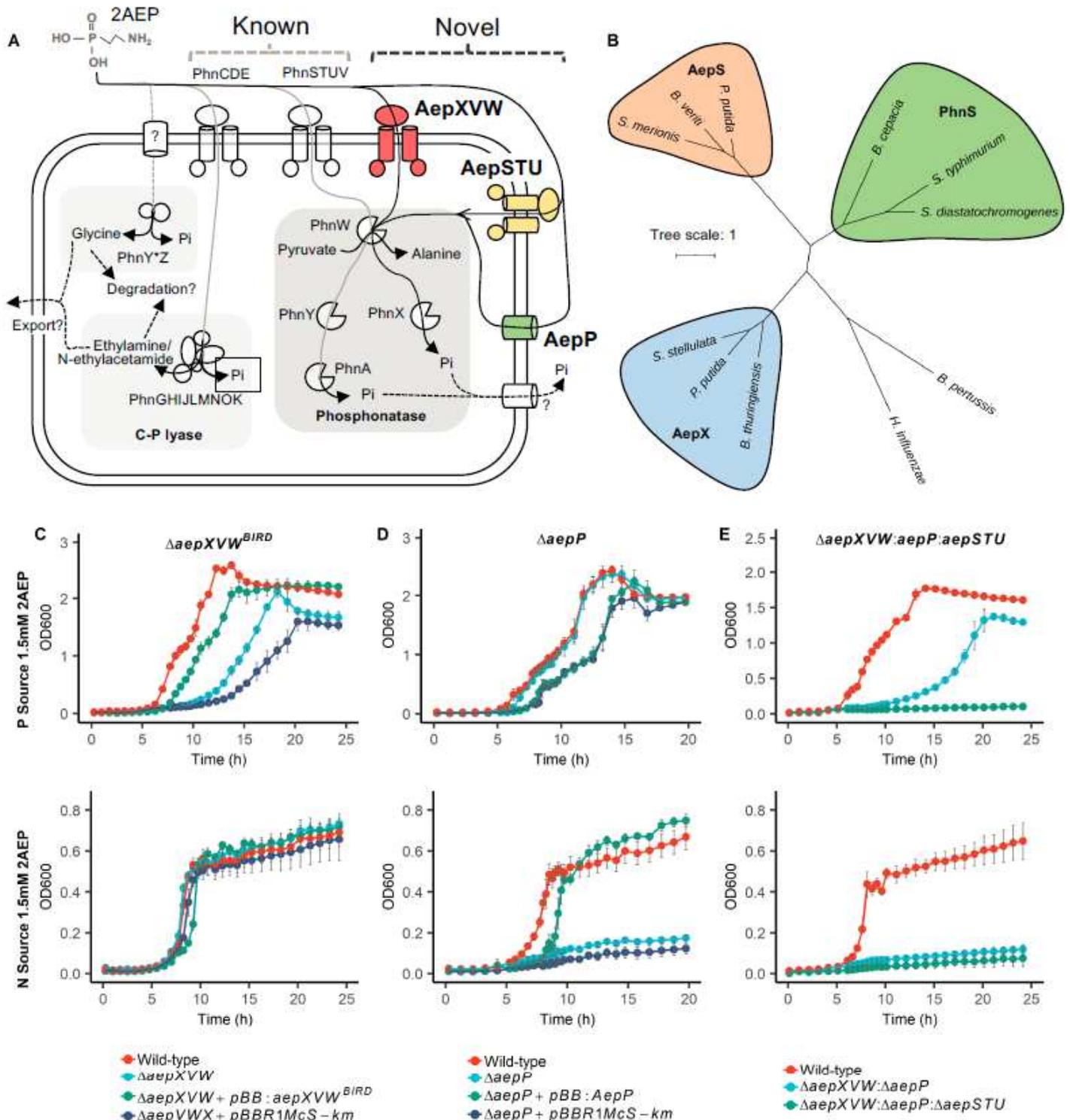
**Figure 4. Phylogenetic and genomic analyses of aepX in marine and terrestrial bacteria.** Genetic neighbourhoods for selected aepX homologs are presented adjacent to trees. Numbers indicate environmental OTUs and letters indicate isolates or MAGs/SAGs. Tree topology and branch lengths were calculated by maximum likelihood using the LG+F+I+G4 model of evolution for amino acid sequences based on 744 sites in IQ-TREE software 70. A consensus tree was generated using 1000 bootstraps. Branches representing isolates or MAGs/SAGs are colour coded based on their phylogenetic affiliation (see legends). Branches and identifiers for representative environmental OTU sequences (clustered at 0.8) retrieved from the TARA Oceans database are highlighted blue. The outer ring denotes the relative abundance of environmental AepX OTUs using the same colour scheme; 10% (dashed line) and 20% (filled line) thresholds are shown for scale. *S. stellulata* DSM 5886 and *P. putida* BIRD-1 aepX are labelled.



**Figure 5. Distribution and transcriptional regulation of phosphonate transporter genes in the global ocean.**

Abundance of *phnD*, *aepX*, *aepP* in marine epipelagic (red) and mesopelagic (blue) waters, split by metagenome (MG) (A), and metatranscriptome (MT) (B). Abundance (Log<sub>2</sub> % abundance related to single copy core genes) of *phnD*, *aepX*, *aepP* in MG (circles) and MT (triangles) in epipelagic (C) and mesopelagic (D) waters, split by oceanic region. *aepP* (red), *phnD* (green), *aepX* (blue). AO = Arctic Ocean, SO = Southern Ocean, IO = Indian Ocean, SPO = South Pacific Ocean, NPO = North Pacific Ocean, SAO = South Atlantic Ocean, NAO = North Atlantic Ocean, RS = Red Sea, MS = Mediterranean Sea. The relationship between the standing stock Pi concentration and transporter abundance in the MG (E), (*aepX*  $R^2 = 0.098^{***}$ , *phnD*  $R^2 = 0.340^{***}$ , *aepP*  $R^2 = 0.291^{***}$ ) and MT (F), (*aepX*  $R^2 = 0.007^{ns}$ , *phnD*  $R^2 = 0.203^{***}$ , *aepP*  $R^2 = 0.058^{ns}$ ). *aepP* (red), *phnD* (green), *aepX* (blue), ns = not significant, \*\*\* =  $p < 0.001$ . The relationship between  $R^*$ , a measure of N vs P limitation defined as the sum of standing stock nitrate plus nitrite concentration minus 16x standing stock Pi concentration, and transporter abundance in the MG (G) (*aepX*  $R^2 = 0.029^*$ , *phnD*  $R^2 = 0.168^{***}$ , *aepP*  $R^2 = 0.108^{***}$ ) and MT (H) (*aepX*  $R^2 = -0.005^{ns}$ , *phnD*  $R^2 = 0.197^{***}$ , *aepP*  $R^2 = -0.014^{ns}$ ). *aepP* (red), *phnD* (green), *aepX* (blue), ns = not significant, \* =  $p < 0.05$ , \*\*\* =  $p < 0.001$ .

# Figures

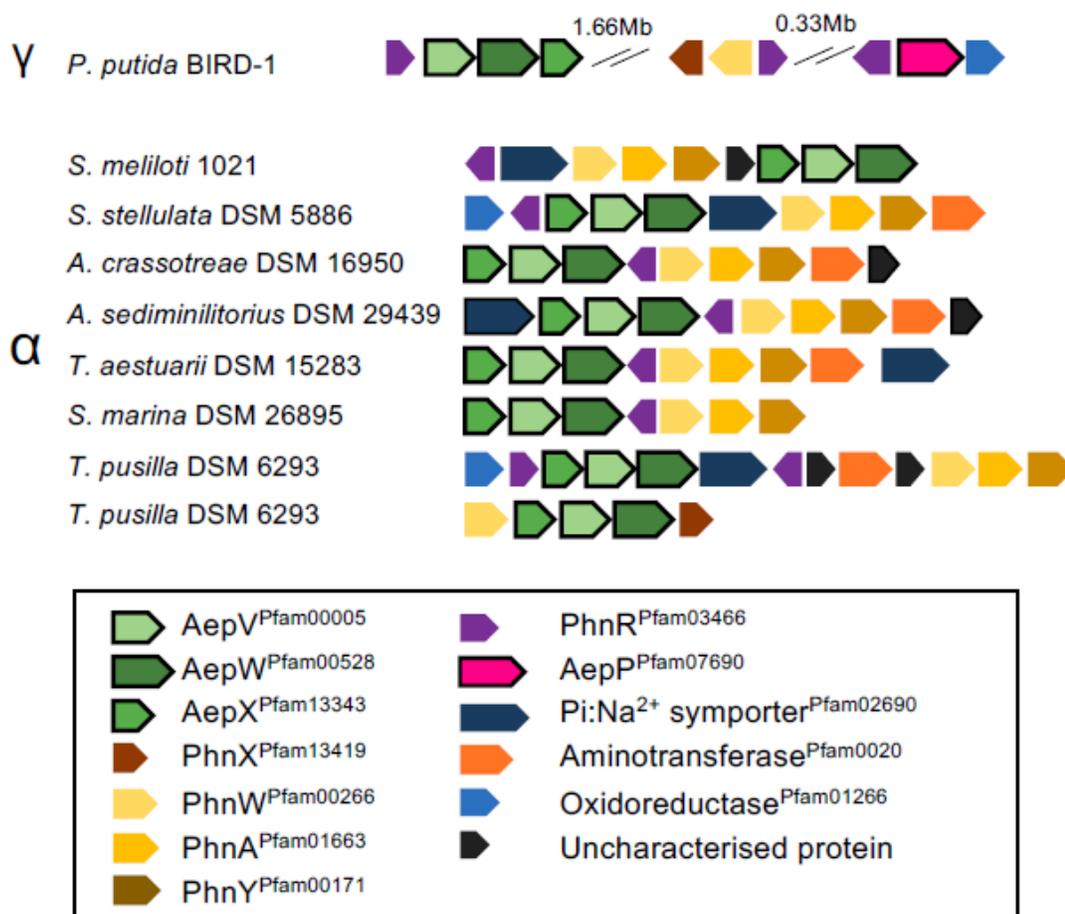


**Figure 1**

2AEP transport and catabolism in *P. putida* BIRD-1. (A) Schematic representation of novel (highlighted in bold and coloured) routes for 2AEP transport, together with existing characterised and putative 2AEP transport routes. Each catabolic system for degradation is highlighted and includes i) the phosphonate

system comprising a 2AEP-pyruvate transaminase (PhnW) and either a phosphonoacetaldehyde hydrolase (PhnX)<sup>74,75</sup> or a NAD<sup>+</sup>-dependent phosphonoacetaldehyde dehydrogenase (PhnY) and a phosphonoacetate hydrolase (PhnA)<sup>43</sup>, and ii) the PhnY\*Z system comprising phosphohydrolase (PhnZ)<sup>21,76</sup> and a 2-oxoglutarate dioxygenase (PhnY\*)<sup>77</sup>. In addition, the promiscuous multi subunit enzyme C-P lyase (PhnGHIJKLMN) can also act on 2AEP<sup>78</sup>, as well as alkylphosphonates<sup>14,15,23</sup>. Pathways found in BIRD-1 are represented by black lines, pathways absent from BIRD-1 are shaded grey. Characterised pathways are shown with solid lines, uncharacterised pathways are shown with dashed lines. Transporters found in BIRD-1 are red if Pi-sensitive, green if Pi-insensitive, and yellow if constitutive. Unknown mechanisms are denoted by a '?'. (B) Phylogenetic tree of AepX, PhnS, and AepS, using the characterised Fe<sup>3+</sup> substrate binding protein FBPa from *Haemophilus influenzae* and *Bordetella pertussis* as an outgroup. *P. putida* = *Pseudomonas putida* BIRD-1, *S. stellulata* = *Stappia stellulata*, *B. cepacia* = *Burkholderia cepacia*, *B. vireti* = *Bacillus veriti*, *S. merionis* = *Streptococcus merionis*, *S. typhimurium* = *Salmonella typhimurium*, *S. diastatochromogenes* = *Streptomyces diastatochromogenes*, *B. thuringiensis* = *Bacillus thuringiensis*. (C) Growth (n=4) of *P. putida* BIRD-1 wild type,  $\Delta$ aepXVW, and the complemented mutant. (D),  $\Delta$ aepP and the complemented mutant, and (E), the 2AEP null mutant,  $\Delta$ aepXVW::aepP::aepSTU. All strains used 2AEP as a sole P (top panel) or N (bottom panel) source. Error bars denote standard deviation of the mean.

A



B

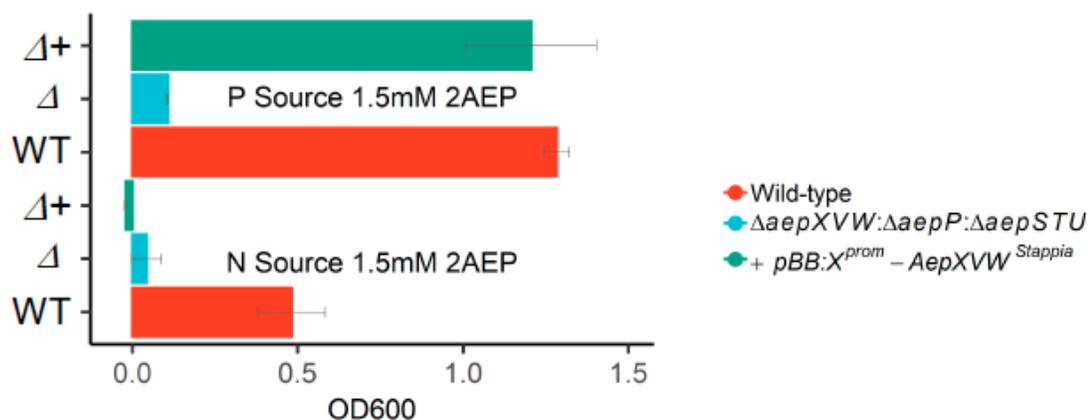


Figure 2

Distribution and functional characterisation of AepXVW in marine bacteria. (A) Genetic neighbourhoods of aepXVW within marine Alpha- and terrestrial Gamma-proteobacteria. Strains shown are *Pseudomonas putida* BIRD-1, *Sinorhizobium meliloti* 1021, *Stappia stellulata* DSM 5886, *Aliiroseovarius crassostreae* DSM 16950, *Aliiroseovarius sediminilitorius* DSM 29439, *Thalassobius aestuarii* DSM 15283, and *Shimia marina* DSM 26895. ORFs separated on the genome are indicated by breaks with the corresponding gap

given in megabases (Mb). (B) Growth of the *P. putida* BIRD-1 triple 2AEP transporter mutant ( $\Delta aepXVW:aepP:aepSTU:gm$ ) complemented with *aepXVW* from *Stappia* concatenated with the promoter region from *aepXVW* BIRD on 2AEP as either a sole N (60 h) or P (48 h) source. Data represents the mean of triplicates cultures. Error bars denote standard deviation.

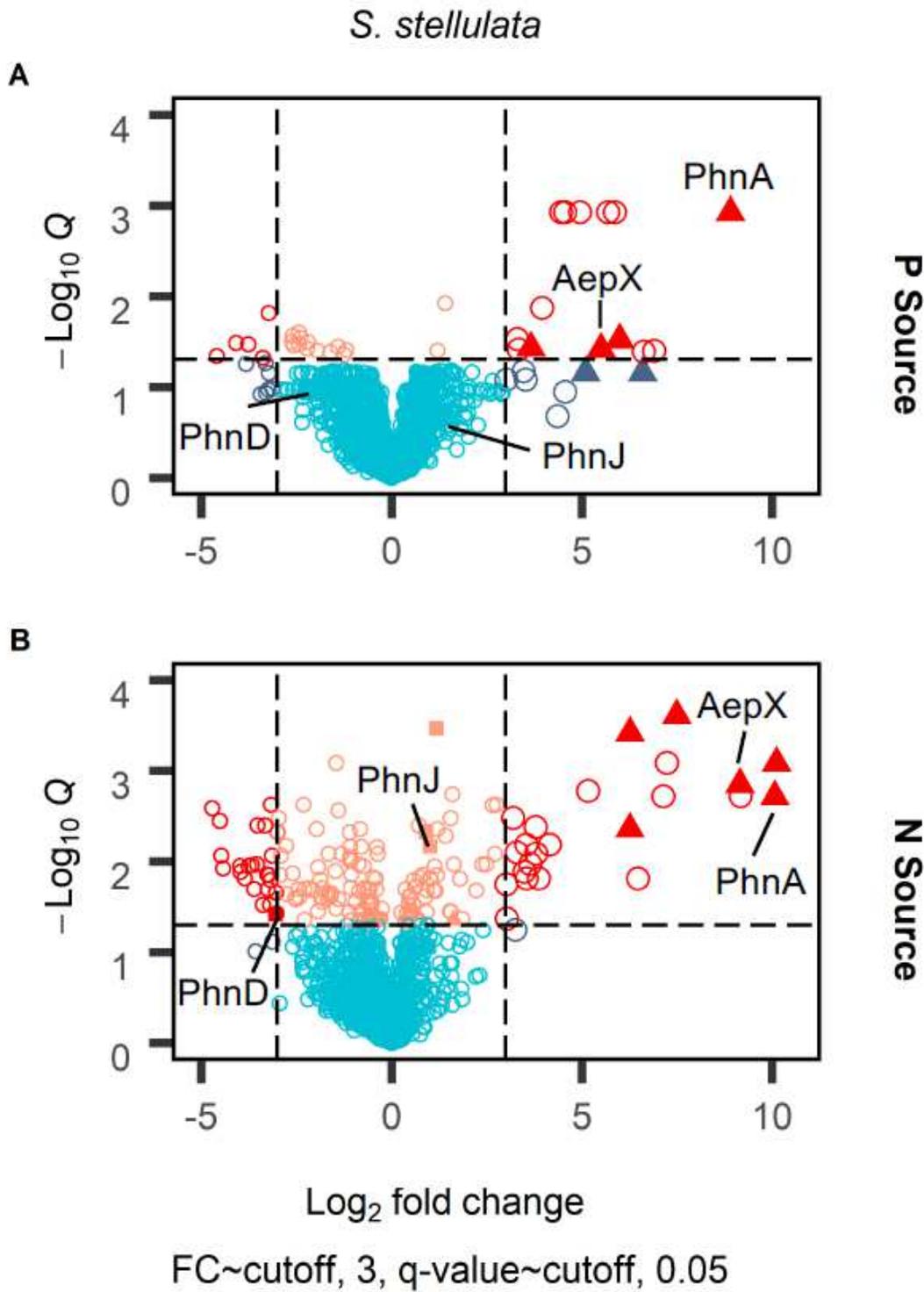


Figure 3

Proteomic analysis of 2AEP-grown *S. stellulata* cells. Whole-cell protein expression profiles (n=3) for *S. stellulata* grown using either Pi or 2AEP as sole P source (A) or NH<sub>4</sub> or 2AEP as the sole N source (B). Fold change represents the difference in Log<sub>2</sub> LFQ values between each treatment and the statistical value on the Y axis is generated from Q values (FDR corrected P values). Members of the *aepXVW-phnWAY* operon are shown as triangles, members of the CP lyase operon are shown as squares, all other proteins are shown as open circles. Data plotted represents the mean of triplicate cultures.

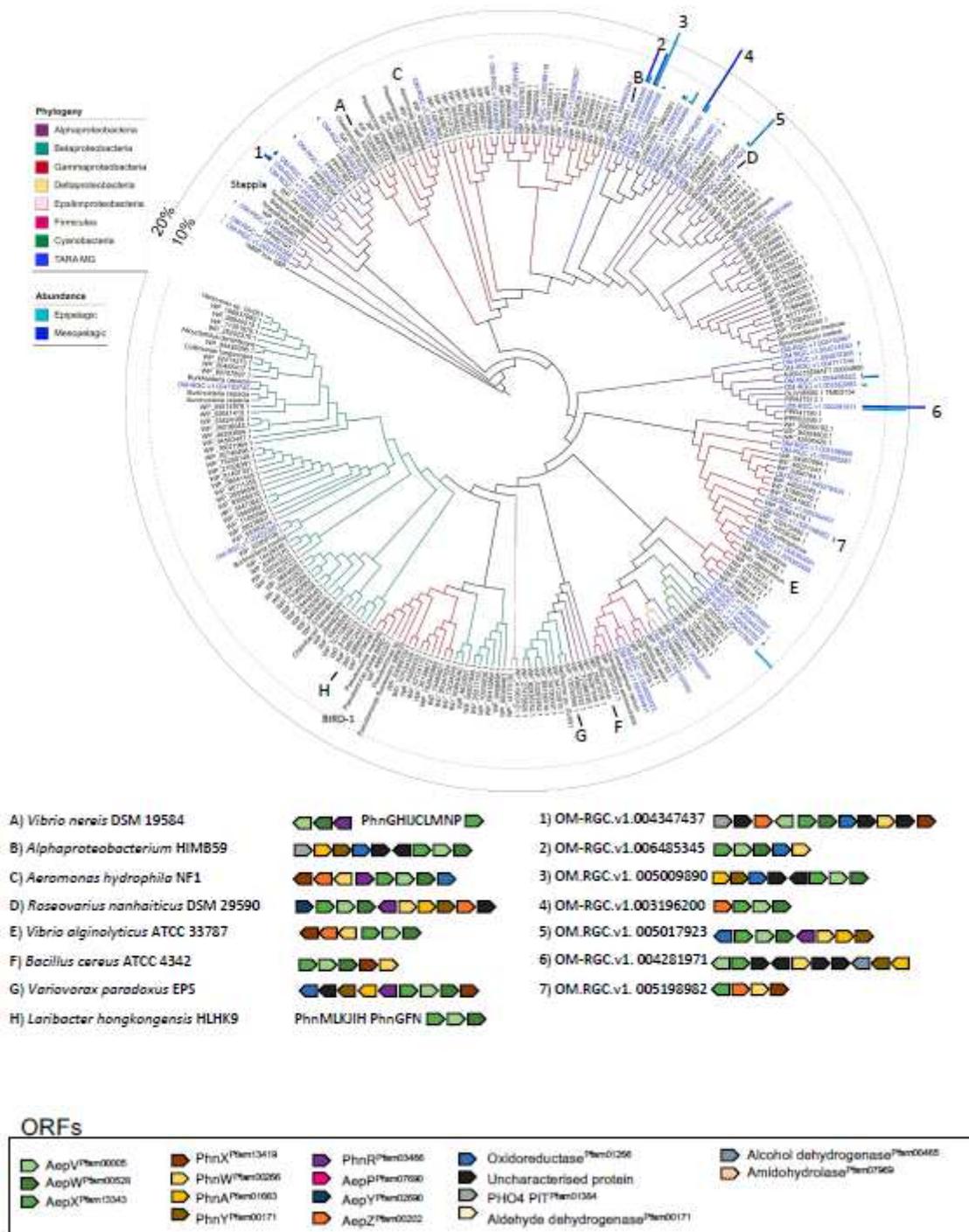
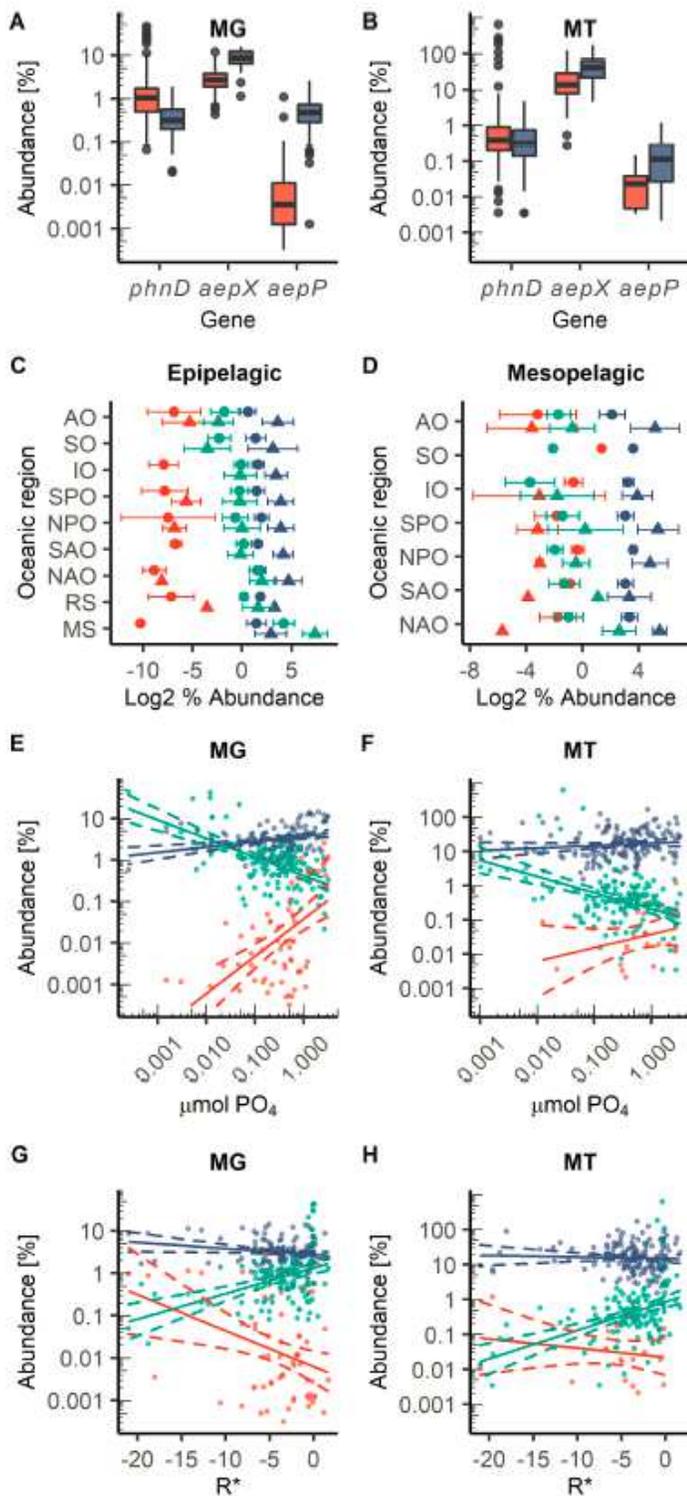


Figure 4

Phylogenetic and genomic analyses of *aepX* in marine and terrestrial bacteria. Genetic neighbourhoods for selected *aepX* homologs are presented adjacent to trees. Numbers indicate environmental OTUs and letters indicate isolates or MAGs/SAGs. Tree topology and branch lengths were calculated by maximum likelihood using the LG+F+I+G4 model of evolution for amino acid sequences based on 744 sites in IQ-TREE software 70. A consensus tree was generated using 1000 bootstraps. Branches representing isolates or MAGs/SAGs are colour coded based on their phylogenetic affiliation (see legends). Branches and identifiers for representative environmental OTU sequences (clustered at 0.8) retrieved from the TARA Oceans database are highlighted blue. The outer ring denotes the relative abundance of environmental *AepX* OTUs using the same colour scheme; 10% (dashed line) and 20% (filled line) thresholds are shown for scale. *S. stellulata* DSM 5886 and *P. putida* BIRD-1 *aepX* are labelled.



**Figure 5**

Distribution and transcriptional regulation of phosphonate transporter genes in the global ocean. Abundance of *phnD*, *aepX*, *aepP* in marine epipelagic (red) and mesopelagic (blue) waters, split by metagenome (MG) (A), and metatranscriptome (MT) (B). Abundance (Log<sub>2</sub> % abundance related to single copy core genes) of *phnD*, *aepX*, *aepP* in MG (circles) and MT (triangles) in epipelagic (C) and mesopelagic (D) waters, split by oceanic region. *aepP* (red), *phnD* (green), *aepX* (blue). AO = Arctic Ocean,

SO = Southern Ocean, IO = Indian Ocean, SPO = South Pacific Ocean, NPO = North Pacific Ocean, SAO = South Atlantic Ocean, NAO = North Atlantic Ocean, RS = Red Sea, MS = Mediterranean Sea. The relationship between the standing stock Pi concentration and transporter abundance in the MG (E), (aepX  $R^2 = 0.098^{***}$ , phnD  $R^2 = 0.340^{***}$ , aepP  $R^2 = 0.291^{***}$ ) and MT (F), (aepX  $R^2 = 0.007^{ns}$ , phnD  $R^2 = 0.203^{***}$ , aepP  $R^2 = 0.058^{ns}$ ). aepP (red), phnD (green), aepX (blue), ns = not significant,  $^{***} = p < 0.001$ . The relationship between  $R^*$ , a measure of N vs P limitation defined as the sum of standing stock nitrate plus nitrite concentration minus 16x standing stock Pi concentration, and transporter abundance in the MG (G) (aepX  $R^2 = 0.029^*$ , phnD  $R^2 = 0.168^{***}$ , aepP  $R^2 = 0.108^{***}$ ) and MT (H) (aepX  $R^2 = -0.005^{ns}$ , phnD  $R^2 = 0.197^{***}$ , aepP  $R^2 = -0.014^{ns}$ ). aepP (red), phnD (green), aepX (blue), ns = not significant,  $^* = p < 0.05$ ,  $^{***} = p < 0.001$ .

## Supplementary Files

This is a list of supplementary files associated with this preprint. Click to download.

- [MurphyPhosphonatesSuppFigsNov20Final.pptx](#)
- [SuppFigure4.docx](#)
- [MurphyPhosphonatesSuppTablesNov20Final.docx](#)
- [SupplementaryproteomicstableS3.xlsx](#)