

# Honeycomb Wet Scrubber For Acidic Gas Control: Modeling And Long-term Test Results

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## Research

**Keywords:** Honeycomb, Wet scrubber, Acidic gas, Long-term operation, Pressure drop

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18 **Abstract**

19 A laboratory scale, 1.0 CMM wet scrubber packed with water-absorbing honeycomb material  
20 (HWS) with a very large geometric surface area of  $480 \text{ m}^2 \text{ m}^{-3}$  and a low pressure drop developed in  
21 our previous study was shown to achieve a very high removal efficiency for gaseous acidic gases but  
22 there were no long-term test data. In this study, the HWS scaled up to operate at 100 CMM flow rate  
23 was tested for removing mixed acidic gases at a semiconductor fab for a very long period of 3.5 years.  
24 Results showed that the removal efficiency for the mixed gases emitted from the fab always maintained  
25 as high as  $> 95\%$  for HF,  $\text{CH}_3\text{COOH}$ , HCl,  $\text{HNO}_3$ ,  $\text{HNO}_2$ , and  $\text{H}_2\text{SO}_4$  with the inlet concentrations  
26 ranging from supper-ppmv to sub-ppmv, respectively, during 3.5-yr period. With water jet cleaning of  
27 the honeycomb modules once per year, the pressure drop of the HWS remained to be very low at 0.5-  
28 0.8  $\text{cmH}_2\text{O}$ , indicating minimal scaling in the HWS existed. Additionally, the predicted height and  
29 removal efficiencies of the HWS were very close to the experimental data. The excellent long-term  
30 performance of the HWS warrants its potential applications in many areas in which liquid absorption  
31 is the preferred treatment method and the theoretical equations can facilitate the design of the HWS.

32

33 **Keywords:** Honeycomb, Wet scrubber, Acidic gas, Long-term operation, Pressure drop

34

## 35 1. Introduction

36 Inorganic acids such as HF, HCl, HNO<sub>3</sub>, H<sub>2</sub>SO<sub>4</sub>, and CH<sub>3</sub>COOH are widely used in  
37 semiconductor or photoelectric manufacturing factories (or fabs) for wafer cleaning and wet-etching  
38 processes [1-5] resulting in the emission of gaseous inorganic acidic pollutants from the stacks, which  
39 need to be treated to meet the emission standard of TWEPA (Taiwan Environment Protection Agency).  
40 The standard stipulates that the total mass emission rate from all stacks must be < 0.6 kg h<sup>-1</sup> for HF,  
41 HCl, and HNO<sub>3</sub>, and < 0.1 kg h<sup>-1</sup> for H<sub>2</sub>SO<sub>4</sub>, respectively, for a semiconductor fab, and the removal  
42 efficiency for each species must be ≥ 95% [1]. Among these gaseous pollutants, HF and HCl are  
43 considered as hazardous air pollutants [6]. These gases are also precursors for secondary inorganic  
44 components in PM<sub>2.5</sub> [7] which is considered as a major health-risk concern in many countries [8]. Wet  
45 scrubbers (WSs) are widely used to control the emission of the acidic gases [9-11] due to their small  
46 footprint, simple structure, and low capital cost [12]. However, the measured removal efficiency (RE)  
47 of most of the traditional WSs can't meet the required value of greater than 95% [1]. For instance, the  
48 removal efficiency of a vertical packed WS with 1,500 CMM flow rate were 70.6-93.4% for HF, HCl,  
49 HNO<sub>2</sub>, HNO<sub>3</sub>, and CH<sub>3</sub>COOH at a high-tech factory while the removal efficiency of a lab-scale packed  
50 WS with 0.6-1.8 CMM flow rate was found to be as low as 76.8%-85.7% for H<sub>2</sub>SO<sub>4</sub> mist [13]. When  
51 the removal efficiency of the WSs can't meet the standard value, they must follow the design criteria

52 and operating conditions in which the specific surface area of packing materials must be  $\geq 90 \text{ m}^2 \text{ m}^{-3}$ ,  
53 the residence time through the scrubbing section must be  $\geq 0.5 \text{ s}$ , a wetting factor must be  $\geq 0.1 \text{ m}^3 \text{ h}^{-1}$ ,  
54 and the pH of scrubbing liquid must be  $\geq 7$  as shown in Table 1 [1].

55 Among all operating parameters, the inlet concentration was found to be the most important  
56 parameter affecting the removal efficiency of the packed WS. The removal efficiency of a conventional  
57 packed WS increased with the increasing inlet concentration at low inlet concentration ( $C_{in}$ ) and  
58 approached a constant value until  $C_{in}$  reached a certain value for HF, HCl, HNO<sub>3</sub>, HNO<sub>2</sub>, and  
59 CH<sub>3</sub>COOH [14]. For example, the removal efficiency of a multistage dual-flow sieve plate WS was  
60 93.8% as  $C_{in}$  was as high as 100-300 ppmv for HCl [15] while a packed tower was found to have a  
61 removal efficiency of less than 90% when  $C_{in}$  was lower than 1.0 ppmv for HF and HCl [9, 14]. The  
62 WSs normally have high pressure drop resulting in high power consumption and operation cost [13,  
63 16]. For example, a WS installed mesh MV filters operated at 90 CMM flow rate achieved  $> 95\%$   
64 removal efficiency for NH<sub>3</sub> (L/G of  $2.0 \text{ L m}^{-3}$ ) at a high pressure drop of  $> 1,411 \text{ Pa}$  [17]. However,  
65 there were no long-term test data provided. Therefore, it is a challenge to design a WS with a high  
66 removal efficiency and low pressure drop which lasts for a long time in an actual plant.

67 The WSs are normally packed with random packing materials to enhance the removal efficiency.  
68 The packing materials, which can be in the form of a saddle, ring or tellerette and made of ceramics,

69 glass, metal, porcelain, steel or plastics [18, 19], should have a high specific surface area to maximize  
70 the gas-liquid contact surface for mass transfer [20] and liquid holdup. The removal efficiencies of  
71 acidic gases were found to increase when the packing material size decreased [21]. For instance, the  
72 removal efficiency for HCl was increased by 2.0-3.5% when the specific surface area of packing balls  
73 was increased from 150 to 224  $\text{m}^2 \text{m}^{-3}$  [22]. The packing materials also need to resist fouling and  
74 corrosion, avoid particle clogging, and enhance water retention to minimize the pressure drop and  
75 maintenance cost for long-term use. In addition, they need to provide good air distribution for uniform  
76 absorption [13] and low pressure drop in the WSs. The pressure drop of a WS with conventional plastic  
77 packing materials is about 413-830 Pa (4.2-8.5  $\text{cmH}_2\text{O}$ ) per meter of the packing column [23-25].

78 The removal efficiency can also be improved by increasing the L/G ratio due to the increasing  
79 gas-liquid contacting surface area. The removal efficiency could increase from 82.8% to 97.5% as the  
80 L/G ratio was increased from 1.5 to 3.0  $\text{L m}^{-3}$  [26]. However, a higher L/G ratio will cause a higher  
81 pressure drop [11, 23]. Adding surfactants in the scrubbing liquid can help enhance the removal  
82 efficiency of the WSs due to the charge generation on the water surface by surfactants but the  
83 surfactants are costly [10]. The overall efficiency of a packed WS for  $\text{NH}_3$ , HF, and HCl ( $C_{\text{in}} = 0.2-3$   
84 ppmv) was reported to increase from 56% to 83% at 1.1 s residence time and 10 CMM flow rate when  
85 surfactants were added [9, 10]. The pH of the scrubbing solution is another factor that affects the

86 removal efficiency of the WSs for some weak acidic gases such as  $\text{CH}_3\text{COOH}$  but not for strong acidic  
87 gases such as  $\text{HCl}$  [11]. There is no significant effect on the removal efficiency for  $\text{CH}_3\text{COOH}$  when  
88 the pH is higher than 7.5 due to the gas dissolution limit while the removal efficiency for  $\text{HCl}$  and  
89  $\text{HNO}_3$  is nearly constant for the pH ranging from 7-9 [11, 27]. To achieve good efficiency for mixed  
90 acidic gases and save the operating cost, the pH of scrubbing liquid should be maintained  $> 7.5$ .

91 Most of the previous researches have focused on enhancing removal efficiency of WSs by  
92 determining the optimal operating conditions [16, 26, 28], attempting to improve the performance of  
93 the existing WSs operating in high-tech industries [9, 10, 22], and designing lab-scale WSs for short-  
94 term study only [13, 19, 29]. A WS packed with a multi-parallel-plate (MPP) module (PPWS)  
95 developed by our group can achieve very high removal efficiencies for  $\text{HCl}$  ( $> 99\%$ ),  $\text{HNO}_3$  ( $> 98\%$ ),  
96 and  $\text{CH}_3\text{COOH}$  ( $> 99\%$ ) and very low pressure drop (23.5 Pa) at a regulated residence time of 0.5 s, a  
97 low inlet concentration ( $< 3$  ppmv), and a high L/G ratio ( $18.5 \text{ L m}^{-3}$ ) [11]. It is because that the PPWS  
98 uses polypropylene plates coated with nano- $\text{TiO}_2$  packing material with  $326.67 \text{ m}^2 \text{ m}^{-3}$  specific surface  
99 area to enhance the hydrophilicity for scrubbing liquid to form a uniform liquid film. However, this  
100 packing material requires a time-consuming and meticulous procedure for preparation and assembly  
101 with a precise gap of 3 mm. The pilot-scale 1.0 CMM WS with the honeycomb module (HWS) was  
102 further developed with the removal efficiency as high as the PPWS and the packing module is easy to

103 fabricate [14]. The HWS showed nearly 100% for HF ( $C_{in} = 0.1-0.4$  ppmv) and 99% for HCl ( $C_{in} =$   
104 3.4-10.8 ppmv) and  $CH_3COOH$  ( $C_{in} = 0.7-6.5$  ppmv) at an optoelectronic factory. Similar to the MPP  
105 module, the honeycomb packing material made of water-absorbing polypropylene (PP) fabric ( $0.875$   
106  $kg\ m^{-1}$ ) shows the higher specific surface area ( $483.2\ m^2\ m^{-3}$ ) and lower pressure drop than most of the  
107 packing materials as shown in Table 2.

108 Thus, the honeycomb packing material with the uniform arrangement and 3-mm gap becomes a  
109 promising packing material for an efficient WS. But the removal efficiency of these small scale HWS  
110 and PPWS were obtained when these WSs were run for several days. There are no scale-up, actual  
111 HWS which runs in a fab for the long-term to control the soluble gas emission. In this study, the HWS  
112 was scaled up as an actual device in a semiconductor manufacturing fab at 100 CMM flow rate to  
113 remove acidic gases and odor. The removal efficiencies of the HWS for the mixed gases containing  
114 HF, HCl,  $HNO_3$ ,  $HNO_2$ ,  $H_2SO_4$ , and  $CH_3COOH$  were tested for 13 times over 3.5 years for the removal  
115 efficiency while the pressure drop was also measured to make sure that there was no fouling occurred  
116 in the system. The theoretical analysis was conducted to predict the height of the packing module and  
117 the removal efficiency of the HWS and the traditional WS to further justify the outstanding  
118 performance of the HWS.

## 119 2. Materials and methods

120 *2.1. Field test*

121 The experimental setup of the HWS installed at a semiconductor manufacturing fab which  
122 produces DRAM in Hsinchu, Taiwan to reduce its emission and bad odor of acidic gases is shown in  
123 Fig. 1a. The HWS, which is registered as a patent in Taiwan (No. I569867 and M466724), is an add-  
124 on central scrubber device mainly to remove high concentration acidic gas pollutants collected from  
125 about 40 local scrubbers before a traditional central packed tower which has low removal efficiency  
126 with high emitted HF and CH<sub>3</sub>COOH concentrations (e.g., typically > 10 ppmv without the present  
127 HWS) and bad odor. The HWS consists of 64 honeycomb modules placed in parallel in four rows (16  
128 modules per row) as shown in Fig. 1b, spray nozzles installed at the top of honeycomb modules, and  
129 a liquid reservoir with conductivity and pH meters. Each honeycomb elements were scrolled in a  
130 cylindrical casing (D x Z = 300 x 300 mm). The operating parameters of the HWS are shown in Table  
131 1, which meet the design criteria for a packed tower to achieve the efficiency-based emission standards  
132 for acidic gases. In this study, the HWS with the specific surface area of the packing material of 480  
133 m<sup>2</sup> m<sup>-3</sup> was designed to operate at 100 CMM flow rate with 0.4 m s<sup>-1</sup> face velocity and 0.5 s retention  
134 time. The liquid flow rate was about 2,600 L min<sup>-1</sup> which was controlled by a flow meter and the L/G  
135 ratio was maintained at 26 L m<sup>-3</sup>. NaOH solution was used as the scrubbing solution with the pH  
136 maintained at 7.5-8. The theoretical pressure drop of 59.8 Pa (0.58 cmH<sub>2</sub>O) was calculated by the

137 method in Chien et al. [11], which is higher than the pressure drop (23.5 Pa) of the lab-scale, 1.0 CMM  
138 HWS in the previous study due to the increasing L/G ratio.

139 The HWS was tested for the removal efficiency of inorganic gaseous acidic pollutants including  
140 HF, HCl, HNO<sub>2</sub>, HNO<sub>3</sub>, H<sub>2</sub>SO<sub>4</sub>, and CH<sub>3</sub>COOH from August 2016 to March 2020 (3.5 years) with 13  
141 times (11/08/2016, 12/09, 11/10, 15/11, 13/12, 13/03/2017, 15/06, 15/08, 22/12, 02/05/2018,  
142 31/01/2019, 6/12, and 19/03/2020). The testing frequency was set at once per month in the first 5  
143 months (5 tests) and then extended to once per three months in the next 12 months (4 tests), once per  
144 six months (2 tests), and finally once per year (1 test). The latest testing time (19/03/2020) was about  
145 3 months after the 12<sup>th</sup> test when the pH meter was not functioning well when the NaOH solution was  
146 not pumped into the tank to maintain the pH  $\geq$  7.5. The result of the 12<sup>th</sup> test is shown in Fig. S1 in  
147 Section S1 in the supplementary information (SI). Two porous denuder samplers (PDS) with a 2 L min<sup>-1</sup>  
148 flow rate [4, 30-32] were used to sample mixture gases at the inlet and outlet of the HWS  
149 simultaneously in 20-60 min depending on inlet concentrations. The PDS consists of a Teflon filter for  
150 sampling particles followed by two porous-metal discs coated with solutions containing 10 ml, 5%  
151 (w/v) sodium carbonate and 1% (w/v) glycerol in 1:1 (v/v) methanol/water solution for sampling  
152 inorganic acidic and basic gases. An ion chromatography (model 883, Metrohm AG, Switzerland) was  
153 used to quantify the concentrations of gases and particles. The particulate phase concentration was

154 found to be much smaller than the corresponding gaseous species and is not a target of this study. The

155 RE of the acidic gases was then calculated as:

156

$$157 \quad RE(\%) = \left[ 1 - \frac{C_{out}}{C_{in}} \right] \times 100\% \quad (1)$$

158

159 where  $C_{out}$  (ppbv) was the outlet concentrations of HF, HCl, HNO<sub>2</sub>, HNO<sub>3</sub>, H<sub>2</sub>SO<sub>4</sub>, or CH<sub>3</sub>COOH in

160 the gas phase. Four tons of makeup water per day was required to compensate for water evaporation

161 and maintain the conductivity of 0.3-0.5 mS cm<sup>-1</sup> in the water reservoir. The pressure drop of the HWS

162 was also measured at five testing times (11/08/2016, 13/12, 15/08/2017, 31/01/2019, and 19/03/2020)

163 during the 3.5-yr period.

## 164 2.2. Theoretical prediction

165 The theoretical removal efficiency of the HWS is calculated by using Eqs. (2) and (3) and Eq. (4)

166 based on the convection-diffusion theory of Gormley and Kennedy ( $RE_{GK}$ , %) [33, 34] and two film-

167 theory ( $RE_{CA}$ , %) [18], respectively, as follows:

168

$$169 \quad RE_{GK}(\%) = \left[ 1 - [0.82 \exp(-11.5\xi) + 0.097 \exp(-70.1\xi)] \right] \times 100\% \quad \text{for } \xi < 0.009 \quad (2)$$

170

171 
$$RE_{GK}(\%) = \left[ 1 - \left[ 1 - 5.50\xi^{2/3} + 3.77\xi \right] \right] \times 100\% \quad \text{for } \xi \geq 0.009 \quad (3)$$

172

173 where  $\xi$  is the dimensionless parameter which is calculated as  $\xi = D_g LZ/QW$ , where  $D_g$  is the diffusion  
 174 coefficient of the gaseous pollutant in the gas phase ( $m^2 s^{-1}$ ) and calculated by the Fuller-Schettler-  
 175 Giddings equation [35],  $L$  is the total length of the PP fabric plate (m);  $Z$  is the height of the honeycomb  
 176 module (m);  $W$  is the gap of the module (m);  $Q$  is the airflow rate ( $m^3 s^{-1}$ ). The  $RE_{CA}$  is calculated as  
 177 [18]:

178

179 
$$RE_{CA}(\%) = \left[ 1 - \exp\left(-\frac{ZK_g a_t}{G_{my}}\right) \right] \times 100\% \quad (4)$$

180

181 where  $a_t$  is the total specific surface area of the module ( $m^2 m^{-3}$ );  $K_g$  is the overall mass transfer  
 182 coefficient ( $mole s^{-1} m^{-2}$ ) which is calculated as  $1/K_g = 1/k_g + m/k_w$  in which  $k_g$  and  $k_w$  are the mass  
 183 transfer coefficient of the gaseous pollutants in the gas phase and the liquid phase, respectively ( $mole$   
 184  $s^{-1} m^{-2}$ ) and  $m$  is the dimensionless Henry's law constant (H) or the dimensionless effective Henry's  
 185 law constant ( $H^*$ );  $G_{my}$  is the molar flux of the air through the HWS and calculated as  $G_{my} = V_m/S$  in  
 186 which  $V_m$  is the molar flow rate of the air ( $mole s^{-1}$ ) and  $S$  is the empty cross-sectional area of the  
 187 honeycomb module ( $m^2$ ).

188 The HWS is the wetted wall column whose the mass transfer coefficients in the gas phase ( $k_g$ )

189 and liquid phase ( $k_w$ ) are calculated as [36]:

190

$$191 \quad k_g = \frac{0.023\pi N_{Re}^{0.83} N_{Sc}^{0.44} D_g P}{W} \quad (6)$$

$$192 \quad k_w = 0.422 \sqrt{\frac{D_w \Gamma}{\rho_w B_F^2}} \quad (7)$$

193

194 where P is the total pressure (atm);  $N_{Re}$  is the Reynolds number which is calculated as  $N_{Re} = W'G_g/\mu_g$ ,

195 where  $W'$  is the equivalent inner diameter of the honeycomb module (m) which is calculated as  $W' =$

196  $\sqrt{4S/\pi}$ ,  $G_g$  is the mass flow rate of the air ( $\text{kg s}^{-1} \text{m}^{-2}$ ), and  $\mu_g$  is the air viscosity ( $\text{kg m}^{-1} \text{s}^{-1}$ );  $N_{Sc}$  is the

197 Schmidt number and calculated as  $N_{Sc} = \mu_g/\rho_g D_g$  in which  $\rho_g$  is the air density ( $\text{kg m}^{-3}$ );  $D_w$  is the

198 diffusion coefficient of the gaseous pollutant in the liquid phase ( $\text{m}^2 \text{s}^{-1}$ ) which is calculated by the

199 Stokes-Einstein equation [37];  $\Gamma$  is the liquid mass flow rate based on the wetted perimeter ( $\text{kg s}^{-1} \text{m}^{-1}$ );

200  $\rho_w$  is the liquid density ( $\text{kg m}^{-3}$ );  $B_F = (3\mu_w \Gamma/\rho_w g)^{1/3}$  in which  $\mu_w$  is the liquid viscosity ( $\text{kg m}^{-1} \text{s}^{-1}$ )

201 and  $g$  is the gravity acceleration ( $\text{m s}^{-2}$ ).

202 The theoretical height of the honeycomb module ( $Z_{CA}$ , m) can be calculated as [18]:

203

$$204 \quad Z_{CA} = \left[ \frac{G_{my}}{K_g a_t} \right] \times \left[ \int_{C_{out}}^{C_{in}} \frac{dC}{\left(1 - m \frac{V_m}{L_m}\right) C + \left(\frac{V_m}{L_m} C_{in} - C_{w,out}\right)} \right] \quad (8)$$

205

206 where  $L_m$  is the liquid molar flow rate (mole  $s^{-1}$ );  $C_{w,out}$  is the outlet concentration of the gaseous  
207 pollutant in the liquid phase (ppbv) and it is calculated as  $C_{w,out} = C_{w,in} + (V_m/L_m) \times (C_{in} - C_{out})$  in which  
208  $C_{w,in}$  is the inlet concentration of the gaseous pollutant in the liquid phase (ppbv). Assuming that the  
209 honeycomb modules are replaced by the traditional packing materials, the removal efficiency of the  
210 wet scrubber packed with different raschig rings (equivalent diameter ( $D_e$ ) = 6, 13 and 25 mm) and  
211 berl saddle ( $D_e = 25$  mm) are studied. The small raschig rings with  $D_e$  equal to 6 mm and 13 mm have  
212 the specific surface areas 50% higher than ( $710 \text{ m}^2 \text{ m}^{-3}$ ) and similar ( $370 \text{ m}^2 \text{ m}^{-3}$ ) [36] to that of the  
213 honeycomb module, respectively. Whereas, the commonly used raschig ring ( $D_e = 25$  mm and  $a_t = 190$   
214  $\text{m}^2 \text{ m}^{-3}$ ) [36] and berl saddle ( $D_e = 25$  mm and  $a_t = 250 \text{ m}^2 \text{ m}^{-3}$ ) [36] have the specific surface areas  
215 much lower than that of the honeycomb module. The gas phase and liquid phase mass transfer  
216 coefficients of the WS packed with the raschig rings and berl saddle are typically calculated by Eqs.  
217 (S1) and (S2), respectively shown in Section S2 in the SI. The theoretical height of the raschig rings  
218 and the berl saddle is then calculated by Eq. (8) and compared with that of the honeycomb module.

219 The pressure drop of the traditional WS is determined based on the correlation of  $\frac{U_t^2 a_t \rho_g \mu_w^{0.2}}{g \epsilon^2 \rho_w}$  and  
220  $\frac{G_w}{G_g} \sqrt{\frac{\rho_g}{\rho_w}}$  and compared to that of the HWS [36].

### 221 3. Results and discussion

222 3.1. Field test results

223 Fig. 2 shows the removal efficiency of the HWS versus time with a wide range of the inlet  
224 concentrations of 31,874–180,685 ppbv for HF, 12,552–72,582 ppbv for CH<sub>3</sub>COOH; 1,384–15,203  
225 ppbv for HCl, 1,234–12,390 ppbv for HNO<sub>3</sub>, 201–6,017 ppbv for HNO<sub>2</sub>, and 240–1,583 ppbv for  
226 H<sub>2</sub>SO<sub>4</sub>. Among these gases, HF and CH<sub>3</sub>COOH are dominant and constitute  $64.3 \pm 9.95$  and  $25.4 \pm$   
227  $9.3\%$  of the total mass concentration of the mixed gases, respectively while HCl, HNO<sub>3</sub>, HNO<sub>2</sub>, and  
228 H<sub>2</sub>SO<sub>4</sub> account for only  $4.0 \pm 1.7\%$ ,  $3.5 \pm 1.7\%$ ,  $2.1 \pm 1.4\%$ , and  $0.6 \pm 0.3\%$ , respectively. The figure  
229 shows that the removal efficiencies of each species are all very high, which are 95.6-98.3% ( $RE_{ave} =$   
230  $97.2 \pm 0.8\%$ ) for HF (Fig. 2a), 95.0-98.3% ( $RE_{ave} = 96.9 \pm 1.1\%$ ) for CH<sub>3</sub>COOH (Fig. 2a), 96.1-98.5%  
231 ( $RE_{ave} = 97.7 \pm 0.7\%$ ) for HCl (Fig. 2b), 95.1-98.8% ( $RE_{ave} = 97.5 \pm 1.2\%$ ) for HNO<sub>3</sub> (Fig. 2b), 95.0-  
232  $97.9\%$  ( $RE_{ave} = 96.3 \pm 0.9\%$ ) for HNO<sub>2</sub> (Fig. 2c), and 95.2-97.6% ( $RE_{ave} = 96.5 \pm 0.7\%$ ) for H<sub>2</sub>SO<sub>4</sub>  
233 (Fig. 2c), respectively. That is, the HWS shows a very good performance for mixed gas control with a  
234 wide inlet concentration range during the 3.5-yr long period.

235 The removal efficiency of the acidic gases follows the order as  
236 HCl>HNO<sub>3</sub>>HF>HNO<sub>2</sub>>CH<sub>3</sub>COOH>H<sub>2</sub>SO<sub>4</sub>. Among these gases, HCl has the highest removal  
237 efficiency during 3.5 years although HCl has the lowest H ( $H_{HCl} = 1.1 < H_{HNO_2} = 49 < H_{CH_3COOH} = 8.8$   
238  $\times 10^3 < H_{HF} = 1.3 \times 10^4 < H_{HNO_3} = 2.1 \times 10^5 < H_{H_2SO_4} = 2.9 \times 10^9 \text{ M atm}^{-1}$ ) [38, 39]. On the other hand,

239 the dissociation constants of HCl ( $K_{\text{HCl}} = 1.7 \times 10^6 \text{ M atm}^{-1}$ ) is several orders of magnitude higher than  
240 other acidic gases ( $K_{\text{H}_2\text{SO}_4} = 10^3$ ,  $K_{\text{HNO}_3} = 15.4$ ;  $K_{\text{HF}} = 6.3 \times 10^{-4}$ ;  $K_{\text{HNO}_2} = 5.1 \times 10^{-4}$ ;  $K_{\text{CH}_3\text{COOH}} = 1.7$   
241  $\times 10^{-5} \text{ M atm}^{-1}$ ) [11, 38]. It indicates that H just reflects the physical solubility of the gases while  
242 additional dissociation and chemical reaction constants also play role in the absorption process by WSs  
243 [16, 38], which is explained in detail in Section 3.2.

244 It is seen that the high removal efficiency of the HWS does not change from the first operating  
245 day (11/08/2016) until the latest testing day (19/03/2020) for  $\text{CH}_3\text{COOH}$  (1<sup>st</sup>: 96.3%; 13<sup>th</sup>: 97.4%) and  
246 HCl (1<sup>st</sup>: 97.8%; 13<sup>th</sup>: 97.8%), respectively, while it is slightly decreases for 1.3% for HF (1<sup>st</sup>: 98.0%;  
247 13<sup>th</sup>: 96.7%), 2.8% for  $\text{HNO}_3$  (1<sup>st</sup>: 97.9%; 13<sup>th</sup>: 95.1%), 2.0% for  $\text{HNO}_2$  (1<sup>st</sup>: 97.9%; 13<sup>th</sup>: 95.9%), and  
248 1.2% for  $\text{H}_2\text{SO}_4$  (1<sup>st</sup>: 96.9%; 13<sup>th</sup>: 95.7%), respectively. It may be due to the slight increasing pressure  
249 drop which is to be discussed later. It is to be noted that the HWS was cleaned manually only once  
250 every year by spraying water jet from the top of the modules during the test period (three cleaning  
251 times in total, in July 2017, April 2018 and July 2019, respectively), which helped maintain  $\geq 95\%$   
252 high removal efficiency during this long-term test. In comparison, the packing materials of the  
253 traditional packed tower need to be washed once every quarter to reduce particle clogging. After each  
254 yearly maintenance denoted as “M” in Fig. 2, the removal efficiency is increased slightly for HF (1<sup>st</sup>  
255 M: +1.4%; 2<sup>nd</sup> M: +1.5%) and is not affected for other gases which all maintain to be very high.

256 Unlike other traditional WSs, the removal efficiency of the HWS is not affected by the inlet  
257 concentration, which is always higher than 95% for all acidic gases when  $C_{in}$  varies from sub-ppmv  
258 ( $709 \pm 324$  ppbv in average) to super-ppmv ( $105,505 \pm 41,039$  ppbv in average) as shown in Fig. 3.  
259 However, it is seen that the removal efficiencies of the HWS increase slightly by 3.5%, 2.2%, and  
260 1.5% when the inlet concentration increases by 8.2, 5.6, and 3.9 times for  $HNO_3$ ,  $HCl$ , and  $H_2SO_4$ ,  
261 respectively as shown in Fig. 3a. The slightly positive correlation of the removal efficiency and the  
262 inlet concentration of  $HNO_3$  and  $HCl$  was found with  $R^2$  of 0.37 (p-value =  $0.03 < 0.05$ ) and  $R^2$  of 0.53  
263 (p-value =  $0.02 < 0.05$ ), respectively. It is due to the variation of the inlet concentrations of  $HNO_3$  and  
264  $HCl$  from low ( $< 3,000$  ppbv) to high concentrations ( $> 3,000$  ppbv). In Fig. 3a, it is found that the  
265 removal efficiency of  $H_2SO_4$  is lower than those of  $HNO_3$  and  $HCl$  since the inlet concentration of  
266  $H_2SO_4$  is lower than 1,600 ppbv and 9.2 times lower than  $HNO_3$  and 11.2 times lower than  $HCl$  in  
267 average. In comparison, the inlet concentrations of  $HF$  and  $CH_3COOH$  are always higher than 12,500  
268 ppbv. Therefore, the removal efficiency also remains high and constant without an increasing trend, as  
269 shown in Fig. 3b. The removal efficiency for  $HNO_2$  also shows no dependence on the inlet  
270 concentration which also varies from low ( $< 3,000$  ppbv) to high concentrations ( $> 3000$  ppbv) but  
271 does not range widely ( $C_{in-ave} = 2,899 \pm 1,596$  ppbv) as compared to that of  $HCl$  ( $C_{in-ave} = 8,493 \pm$   
272  $3,566$  ppbv) and  $HNO_3$  ( $C_{in-ave} = 6,969 \pm 3,353$  ppbv) with a wide concentration range (Fig. 3b).

273 During the 3.5-yr period, the pressure drop is found to be very small and increases only slightly  
274 from 49.0 to 78.4 Pa (0.5 to 0.8 cmH<sub>2</sub>O), which is also close to the theoretical value of 59.8 Pa (0.58  
275 cmH<sub>2</sub>O). The slight increase in the pressure drop during 3.5 years implies that only slight fouling of  
276 the honeycomb material might have occurred which could not be cleaned thoroughly by water jet. In  
277 Fig. 4a, it is found that the removal efficiency is decreased by 2.8% for HNO<sub>3</sub>, 2.0% for HNO<sub>2</sub>, 1.6%  
278 for HF, and 1.2% for H<sub>2</sub>SO<sub>4</sub> as the pressure drop is increased by 29.4 Pa (0.3 cmH<sub>2</sub>O). The removal  
279 efficiency shows a good negative correlation with the pressure drop for HNO<sub>3</sub> ( $R^2 = 0.88$ ; p-value =  
280  $0.02 < 0.05$ ) with 95% confident interval and H<sub>2</sub>SO<sub>4</sub> ( $R^2 = 0.73$ ; p-value =  $0.06 < 0.1$ ), HF ( $R^2 = 0.70$ ;  
281 p-value =  $0.08 < 0.1$ ), and HNO<sub>2</sub> ( $R^2 = 0.67$ ; p-value =  $0.09 < 0.1$ ) with 90% confident interval showing  
282 a slight influence of the pressure drop on the removal efficiency for these gases. On the other hand,  
283 the removal efficiency of HCl and CH<sub>3</sub>COOH does not show the dependence on the pressure drop as  
284 shown in Fig. 4b since HCl has high dissociation constant and CH<sub>3</sub>COOH has a very high inlet  
285 concentration.

286 The lab-scale HWS operated at a flow rate of 1.0 CMM and an L/G ratio of 18 L m<sup>-3</sup> in the  
287 previous study [14] while the present scale-up, actual HWS device operated at a higher flow rate of  
288 100 CMM, a higher L/G ratio of 26 L m<sup>-3</sup>, and the same retention time of 0.5 s. Although the removal  
289 efficiencies of HF, HCl, and CH<sub>3</sub>COOH for the present HWS are not as high as those in the previous

290 study, which are about 99% for HF and HCl and nearly 100% for CH<sub>3</sub>COOH, the long-term test results  
291 show that the HWS always achieved removal efficiency higher than 95% stipulated in the emission  
292 standard [1]. The inlet concentrations in the previous study ( $C_{in} = 100\text{-}450$  ppbv for HF, 3,000-11,000  
293 ppbv for HCl, and 600-6500 ppbv for CH<sub>3</sub>COOH) were lower than the current HWS. Since there are  
294 64 modules placed in parallel inside the HWS with some unused cross-section area (or dead space),  
295 the required L/G ratio in the present study needs to be 1.4 times higher, which causes the pressure drop  
296 to increase from 23.5 Pa to 49.0 Pa but it is still much lower than that of most traditional WSs.

297 The emission rates of each species in 12 sampling times were calculated from the outlet  
298 concentrations and the air flow rate, and found to be less than 0.03, 0.04,  $2.4 \times 10^{-3}$ ,  $4.0 \times 10^{-3}$ ,  $3.2 \times$   
299  $10^{-3}$ , and  $1.8 \times 10^{-3}$  kg h<sup>-1</sup> for HF, CH<sub>3</sub>COOH, HCl, HNO<sub>3</sub>, HNO<sub>2</sub>, and H<sub>2</sub>SO<sub>4</sub>, respectively. It indicates  
300 that the exhausted gaseous acidic pollutants from the manufacturing process of the fab are well  
301 controlled with only 2.5%, 3.0%, 0.3%, 0.4%, 0.2% and 0.6%, respectively, of the total regulated mass  
302 emission rates (0.1 kg h<sup>-1</sup> for H<sub>2</sub>SO<sub>4</sub> and 0.6 kg h<sup>-1</sup> for other gases) for a semiconductor manufacturing  
303 fab. In addition, the bad odor due to high HF and CH<sub>3</sub>COOH discharge concentrations disappears after  
304 the installation of the present HWS.

### 305 3.2. Results of theoretical prediction

306 Fig. 5 shows the theoretical removal efficiencies of the HWS with the  $D \times Z = 0.3 \times 0.3$  m for

307 different acidic gases. It is seen that the  $RE_{GK}$  is about 100% for all the acidic gases indicating that the  
308 pollutants diffuse completely from the gas phase into the liquid phase fast enough without limiting the  
309 ensuring dissociation and reaction processes of the pollutants in the water film. Whereas, the  $RE_{CA}$   
310 based on H ( $RE_{CA-H}$ ) is 99.7% for HF, 98.0% for  $CH_3COOH$ , 12.6% for HCl, 99% for  $HNO_3$ , 91.9%  
311 for  $HNO_2$ , and 97.8% for  $H_2SO_4$ , respectively indicating that the mass transfer resistance across the  
312 gas-liquid interface is high for HCl, moderate for  $HNO_2$ , and very low for other acidic gases. It is  
313 because that the solubility of the HCl and  $HNO_2$  in the liquid is several orders of magnitude lower than  
314 other acidic gases. The  $RE_{CA}$  based on  $H^*$  ( $RE_{CA-H^*}$ ) of HCl and  $HNO_2$  is increased to 99.3% and  
315 99.2%, respectively when  $H^*$  is used. This implies that the dissociation and reaction are the rate  
316 limiting steps in the HWS and the chemical reaction enhances the absorption rate and increases the  
317 dissolved capacity of the liquid film in the HWS. The  $RE_{CA-H^*}$  results are close to the experimental  
318 data with less than +3% biases.

319 To achieve the removal efficiency as high as the experimental data, the theoretical heights based  
320 on  $H^*$  ( $Z_{CA-H^*}$ ) are 0.19 m for HF, 0.27 m for  $CH_3COOH$ , 0.23 m for HCl, 0.25 m for  $HNO_3$ , 0.21 m  
321 for  $HNO_2$ , and 0.16 m for  $H_2SO_4$ , respectively, as shown in Fig. S2 in the SI. It indicates that the  
322 theoretical values are very similar to the designed value of 0.3 m for mixed acidic gas control. The  
323 theoretical heights based on H ( $Z_{CA-H}$ ) are also shown in Fig. S2 in the SI, which demonstrate that it is

324 required longer module to achieve the removal efficiency  $\geq 95\%$  for HCl ( $Z_{CA-H} = 8.5$  m) and HNO<sub>2</sub>  
325 ( $Z_{CA-H} = 0.4$  m) when the scrubbing NaOH solution is not added. In summary, the equations based on  
326 the two-film theory can be used to predict the theoretical removal efficiency of the HWS and  
327 theoretical height of the honeycomb module with good accuracy.

328 Fig. 6 shows the theoretical removal efficiency ( $RE_{CA-H^*}$ ) and the theoretical height ( $Z_{CA-H^*}$ ) of  
329 the WS packed with the traditional packing materials including raschig rings ( $D_e = 6, 13,$  and  $25$  mm)  
330 and berl saddle ( $D_e = 25$  mm). As compared with the honeycomb module with the similar specific  
331 surface area, theoretical removal efficiencies of the WS packed with 13-mm raschig ring are high for  
332 HF, HCl and H<sub>2</sub>SO<sub>4</sub> ( $RE_{CA-H^*} \geq 99.9\%$ ) but low for CH<sub>3</sub>COOH ( $RE_{CA-H^*} = 62.3\%$ ) and HNO<sub>2</sub> ( $RE_{CA-H^*}$   
333  $= 17.2\%$ ). It is because CH<sub>3</sub>COOH has the lowest dissociation constant among these acidic gases  
334 and both Henry's law and dissociation constants of HNO<sub>2</sub> are several orders of magnitude lower than  
335 other acidic gases. When the 6-mm raschig ring with a higher specific surface area than the honeycomb  
336 module is used, the removal efficiencies can be achieved 95% for CH<sub>3</sub>COOH and 52.6% for HNO<sub>2</sub>,  
337 respectively. However, the removal efficiencies of the traditional WS packed with 6-mm raschig rings  
338 are still lower than those of the HWS for CH<sub>3</sub>COOH ( $Z_{CA-H^*} = 98.1\%$ ) and HNO<sub>2</sub> ( $Z_{CA-H^*} = 99.2\%$ ). It  
339 implies that HWS enhances the mass transfer rate of CH<sub>3</sub>COOH and HNO<sub>2</sub> which have low solubility  
340 and dissociation ability due to its multiple small gaps with uniform water-absorbing surface.

341 Additionally, the theoretical pressure drop of the WS packed with the 6-mm and 13-mm raschig rings  
342 are about 490 Pa (5.0 cmH<sub>2</sub>O) and 245 Pa (2.5 cmH<sub>2</sub>O), respectively, which are much higher than that  
343 of the HWS (59.8 Pa or 0.58 cmH<sub>2</sub>O). The theoretical heights of the WS packed with the 6-mm and  
344 13-mm raschig rings need to be as large as 1.34 m and 5.3 m, respectively to achieve the removal  
345 efficiency higher than 95% for all the acidic gases for mixed gas control as shown in Fig. 6b. That is,  
346 the HWS is an economical control device for mixed gas control as compared with the traditional WS.

347 To reduce the pressure drop, the WS is usually packed with bigger packing materials with lower  
348 specific surface areas such as 25-mm raschig ring or 25-mm berl saddle. However, the removal  
349 efficiency of the traditional WS packed with 25-mm perl saddle is not only as low as 29.1% for  
350 CH<sub>3</sub>COOH and 6.42% for HNO<sub>2</sub> but just 92% for H<sub>2</sub>SO<sub>4</sub>, which does not meet the removal efficiency  
351 requirement. It is because that the inlet concentrations of H<sub>2</sub>SO<sub>4</sub> are very low ( $C_{in} = 0.24-1.58$  ppmv)  
352 as compared to those of the acidic gases. Although the removal efficiency of the WS packed with 25-  
353 mm raschig ring is also as high as 95.3% for H<sub>2</sub>SO<sub>4</sub>, it is still lower than that of the HWS and the WS  
354 packed with 6-mm and 13-mm raschig rings, which is due to its lower specific surface area. This  
355 indicates that the HWS outperforms the traditional WS with the similar or even larger specific surface  
356 areas for the removal efficiency of the mixed gas and pressure drop.

#### 357 4. Conclusion

358 This work evaluated the long-term performance of the honeycomb wet scrubber for removing  
359 gaseous acidic gases and odor at a semiconductor manufacturing fab in 3.5 years from August 2016 to  
360 March 2020. The results of 12 removal efficiency tests indicated that the HWS is capable of removing  
361 the gaseous acidic gases and odor with the inlet concentrations varying from low concentrations of  
362 240–1,583 ppbv for H<sub>2</sub>SO<sub>4</sub> and 201–6,017 ppbv for HNO<sub>2</sub> to moderate concentrations of 1,234–12,390  
363 ppbv for HNO<sub>3</sub> and 1,384–15,203 ppbv for HCl, and high concentrations of 12,552–72,582 ppbv for  
364 CH<sub>3</sub>COOH and 31,874–180,685 ppbv for HF. All removal efficiency data, which are 95.6%–98.3%  
365 for HF, 95.0%–98.3% for CH<sub>3</sub>COOH, 96.1%–98.5% for HCl, 95.1%–98.8% for HNO<sub>3</sub>, 95.0%–97.9%  
366 for HNO<sub>2</sub>, and 95.2%–97.6 for H<sub>2</sub>SO<sub>4</sub>, respectively, met the regulation for the control devices in  
367 Taiwan. The pressure drop of the HWS maintained as low as 49.0–78.4 Pa (0.5–0.8 cmH<sub>2</sub>O) and  
368 increased only slightly during the 3.5-yr period, which is much lower than that of the traditional WSs.  
369 The theoretical removal efficiency and the theoretical height of the HWS module were close to the  
370 experimental data indicating that the theoretical equations can facilitate the design of the HWS. The  
371 HWS showed higher theoretical removal efficiency and lower theoretical height for the mixed gas  
372 control, especially for the acidic gases with low solubility and dissociation ability as compared with  
373 the WS packed with the traditional packing materials. That is, the honeycomb wet scrubber can be  
374 used as an efficient device for mixed acidic gas and odor control with high removal efficiency and low

375 pressure drop for the long-term.

376 **Declarations**

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382 **Authors' contributions**

383 YLL and BTW designed the study and carried out the field tests. GYL and ZL contributed to data  
384 analysis. GHH drafted and edited the manuscript. LTC contributed to data analysis, modeling, and  
385 manuscript preparation. DYHP and CJT critically reviewed and edited the final manuscript. All the  
386 authors read and approved the final manuscript.

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390 **Availability of data and materials**

391 All data generated or analyzed during this study are included in this article and its supplementary

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### 393 **Competing interests**

394 The authors declare they have no competing interests.

### 395 **Nomenclature**

$a_t$	Specific surface area ( $\text{m}^2 \text{m}^{-3}$ )	$N_{RE}$	Reynolds number
$C_{in}$	Inlet concentration (ppbv)	$N_{Sc}$	Schmidt number
$C_{out}$	Outlet concentration (ppbv)	$Q$	Airflow rate ( $\text{m}^3 \text{s}^{-1}$ )
$C_{w,in}$	Inlet concentration of the gaseous pollutant in the liquid phase (ppbv)	$R$	Gas constant ( $\text{m}^3 \text{Pa K}^{-1} \text{mole}^{-1}$ )
$C_{w,out}$	Out concentration of the gaseous pollutant in the liquid phase (ppbv)	$RE$	Experimental removal efficiency (%)
$D$	The inner diameter of the packing material (m)	$RE_{CA}$	Theoretical removal efficiency based on two film theory (%)
$D_e$	The equivalent diameter of the raschig ring (m)	$RE_{GK}$	Theoretical removal efficiency based on diffusion theory (%)
$D_g$	Gas-phase diffusion coefficient ( $\text{m}^2 \text{s}^{-1}$ )	$S$	Empty cross-section surface area ( $\text{m}^2$ )
$D_w$	Liquid-phase diffusion coefficient ( $\text{m}^2 \text{s}^{-1}$ )	$T$	Temperature (K)
$g$	Gravity acceleration ( $\text{m s}^{-2}$ )	$U_t$	Superficial air velocity ( $\text{m s}^{-1}$ )
$G_g$	The mass flow rate of the air ( $\text{kg s}^{-1} \text{m}^{-2}$ )	$V_m$	The molar flow rate of the air ( $\text{mole s}^{-1}$ )

$G_w$	Mass flow rate of the liquid ( $\text{kg s}^{-1} \text{m}^{-2}$ )	$W$	Gap distance of the honeycomb module (m)
$G_{my}$	The molar flux of the air ( $\text{mole s}^{-1} \text{m}^{-2}$ )	$W'$	The equivalent diameter of the honeycomb module (m)
$H$	Henry's low constant ( $\text{atm M}^{-1}$ )	$Z$	Height of the honeycomb module (m)
$H^*$	Effective Henry's low constant ( $\text{atm M}^{-1}$ )	$\epsilon$	Fractional voids in dry packing
	<sup>1)</sup>		
$K$	Dissociation constant ( $\text{M atm}^{-1}$ )	$\rho_g$	Air density ( $\text{kg m}^{-3}$ )
$K_g$	Overall mass transfer coefficient ( $\text{mole s}^{-1} \text{m}^{-2}$ )	$\rho_w$	Liquid density ( $\text{kg m}^{-3}$ )
$k_g$	Gas phase mass transfer coefficient ( $\text{mole s}^{-1} \text{m}^{-2}$ )	$\mu_g$	Air viscosity ( $\text{kg s m}^{-1}$ )
$k_w$	Liquid phase mass transfer coefficient ( $\text{mole s}^{-1} \text{m}^{-2}$ )	$\mu_w$	Liquid viscosity ( $\text{kg s m}^{-1}$ )
$L$	The total length of the PP fabric plate (m)	$\xi$	Dimensionless parameter
$L_m$	Liquid molar flow rate ( $\text{mole s}^{-1}$ )	$\Gamma$	Liquid mass flow rate based on the wetted perimeter ( $\text{kg s}^{-1} \text{m}^{-1}$ )

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- 485

486 **Table 1** Operating parameters of the HWS at a semiconductor fab and the design criteria of TWEPA.

<b>Operating parameter</b>	<b>Value</b>	<b>Criteria</b>	<b>Unit</b>
<sup>a</sup> Wetting factor	0.1	$\geq 0.1$	$\text{m}^2 \text{h}^{-1}$
Retention time	0.5	$\geq 0.5$	sec
Pressure drop	49~78.4	—	Pa
Airflow rate	100	—	$\text{m}^3 \text{min}^{-1}$
Liquid flow rate	2,600	—	$\text{L min}^{-1}$
pH	7.5~9.0	$\geq 7$	—
Packing material	483	$\geq 90$	$\text{m}^2 \text{m}^{-3}$

487 <sup>a</sup> Wetting factor = scrubbing water flow rate divided by (the specific surface area of the packing materials times the  
 488 horizontal cross-sectional area of the tower).

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500 **Table 2** Some typical packing materials.

<b>Packing materials</b>	<sup>a</sup> <b>D×Z or L×W<sub>T</sub>×Z (cm)</b>	<b>Bulk density (kg m<sup>-3</sup>)</b>	<b>Specific surface area (m<sup>2</sup> m<sup>-3</sup>)</b>	<b>L/G ratio (L m<sup>-3</sup>)</b>	<b>Pressure drop (Pa)</b>	<b>Reference</b>
<b>Honeycomb</b>	<b>30*30</b>	<b>160</b>	<b>483</b>	<b>26</b>	<sup>b</sup> <b>59.8</b>	<b>Present</b>
Honeycomb	30*30	160	483	18.5	23.5	[14]
MPPM	30*30*30	920	327	18.5	23.5	[11]
Polypropylene ball (2.5 cm)	47*40*40	139	379	1.5-3	<196	[21]
Ceramic intalox saddle (2.5 cm)	30*61	610	250	1.4-2.1	265-1932	[29]
Raschig ring (1.3 cm)	30*61	-	350	1.4-2.1	294-1058	[29]
Polyethylene knit packing	14*10	-	150	3.0	~2000	[27]
SS wire mesh packing	14*10	-	500	3.0	~2000	[27]

501 <sup>a</sup> Packing dimensions: diameter × height (D×Z) or length × width × height (L×W<sub>T</sub>×Z); <sup>b</sup> Theoretical value

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509 **Figure captions**

510 **Fig. 1 a** Experimental setup of the HWS at a semiconductor fab, and **b** pictorial view of honeycomb  
511 packing installed in the HWS.

512 **Fig. 2** Removal efficiency versus test date for **a** HF and CH<sub>3</sub>COOH, **b** HCl and HNO<sub>3</sub>, and **c** HNO<sub>2</sub>  
513 and H<sub>2</sub>SO<sub>4</sub> (Note: **(M)**: maintenance day).

514 **Fig. 3** Removal efficiency versus inlet concentration for **a** HCl, HNO<sub>3</sub>, and H<sub>2</sub>SO<sub>4</sub>, and **b** HF,  
515 CH<sub>3</sub>COOH, and HNO<sub>2</sub>.

516 **Fig. 4** Removal efficiency versus pressure drop for **a** HF, HNO<sub>2</sub>, HNO<sub>3</sub>, and H<sub>2</sub>SO<sub>4</sub>, and **b** HCl and  
517 CH<sub>3</sub>COOH.

518 **Fig. 5** Comparison of the experimental and theoretical removal efficiencies for different acidic gases  
519 (Note: the data are slightly shifted to avoid overlapping).

520 **Fig. 6** Comparison of **a** the theoretical removal efficiency, and **b** the theoretical height of the traditional  
521 WS and the HWS (Note: the data are slightly shifted to avoid overlapping).

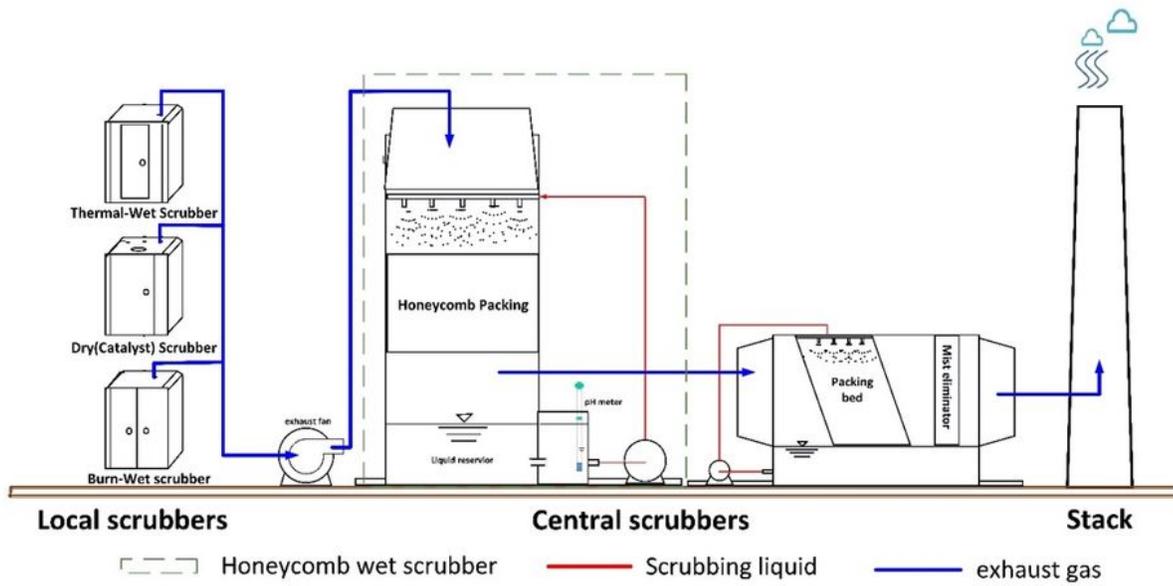
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# Figures



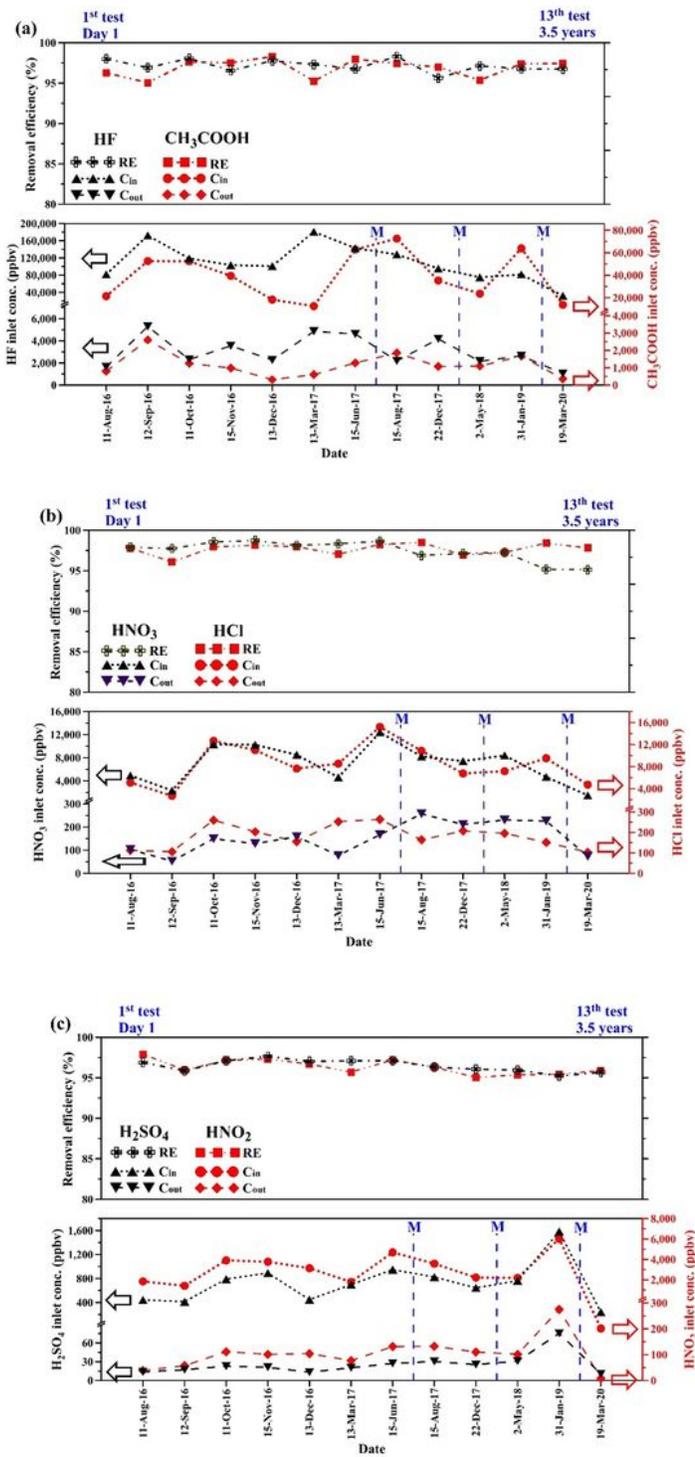
(a)



(b)

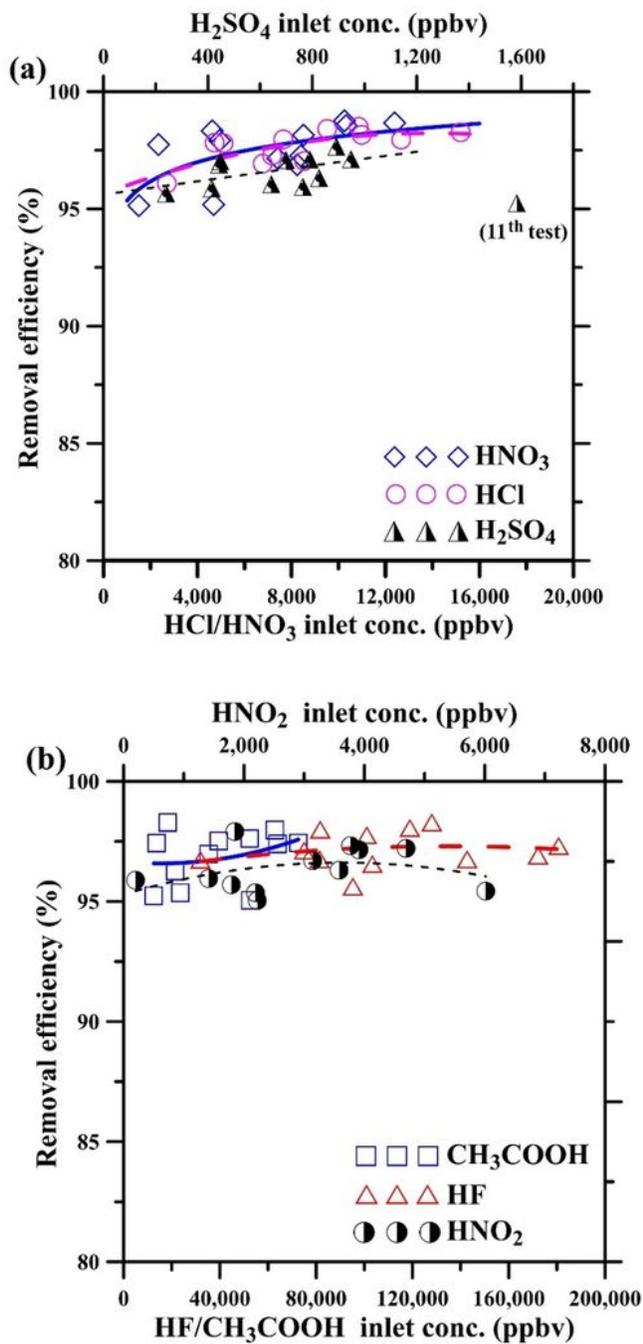
**Figure 1**

a Experimental setup of the HWS at a semiconductor fab, and b pictorial view of honeycomb packing installed in the HWS.



**Figure 2**

Removal efficiency versus test date for a HF and CH<sub>3</sub>COOH, b HCl and HNO<sub>3</sub>, and c HNO<sub>2</sub> and H<sub>2</sub>SO<sub>4</sub> (Note: (M): maintenance day).



**Figure 3**

Removal efficiency versus inlet concentration for a HCl, HNO<sub>3</sub>, and H<sub>2</sub>SO<sub>4</sub>, and b HF, CH<sub>3</sub>COOH, and HNO<sub>2</sub>.

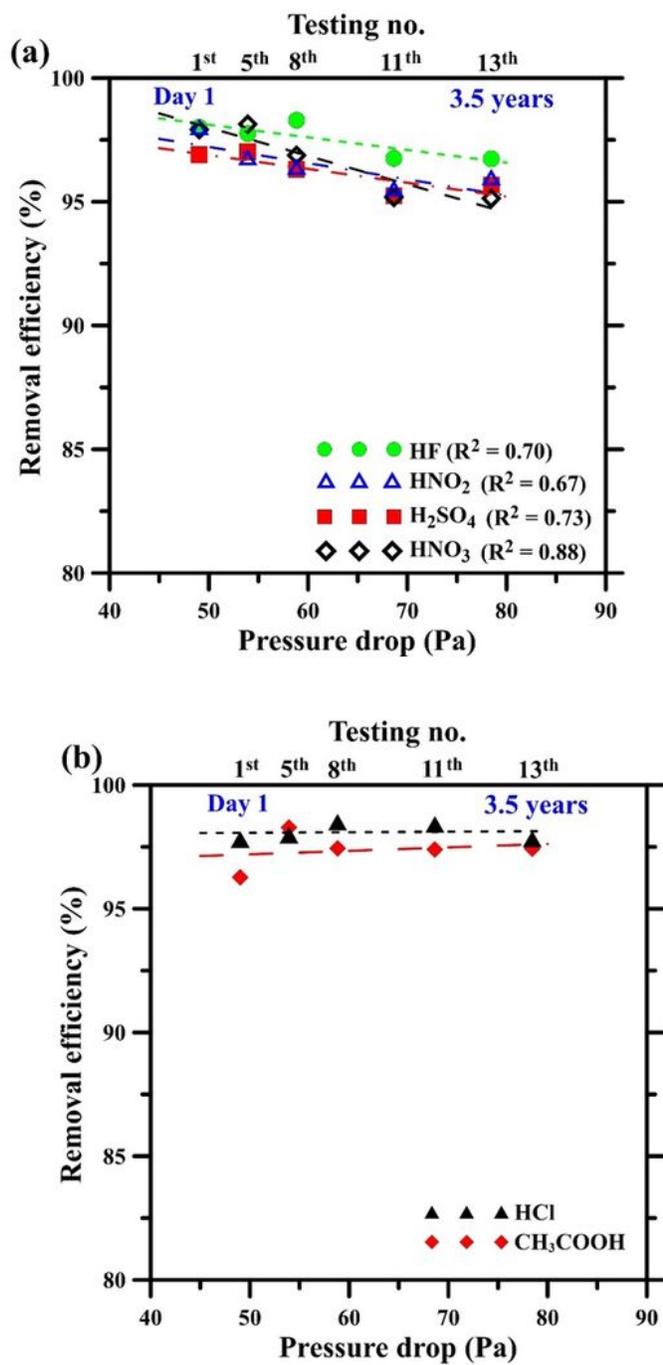


Figure 4

Removal efficiency versus pressure drop for a HF, HNO<sub>2</sub>, HNO<sub>3</sub>, and H<sub>2</sub>SO<sub>4</sub>, and b HCl and CH<sub>3</sub>COOH.

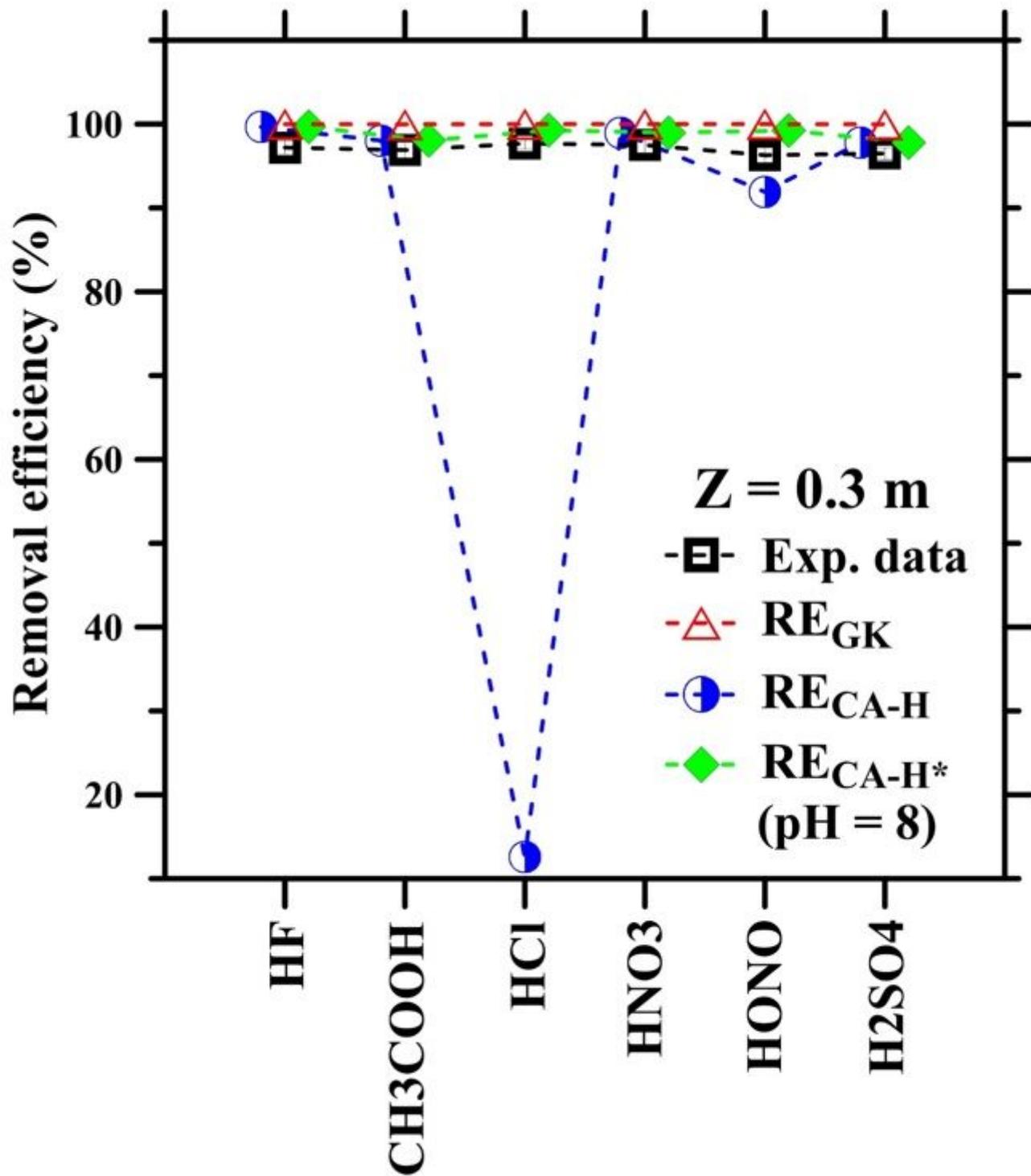


Figure 5

Comparison of the experimental and theoretical removal efficiencies for different acidic gases (Note: the data are slightly shifted to avoid overlapping).

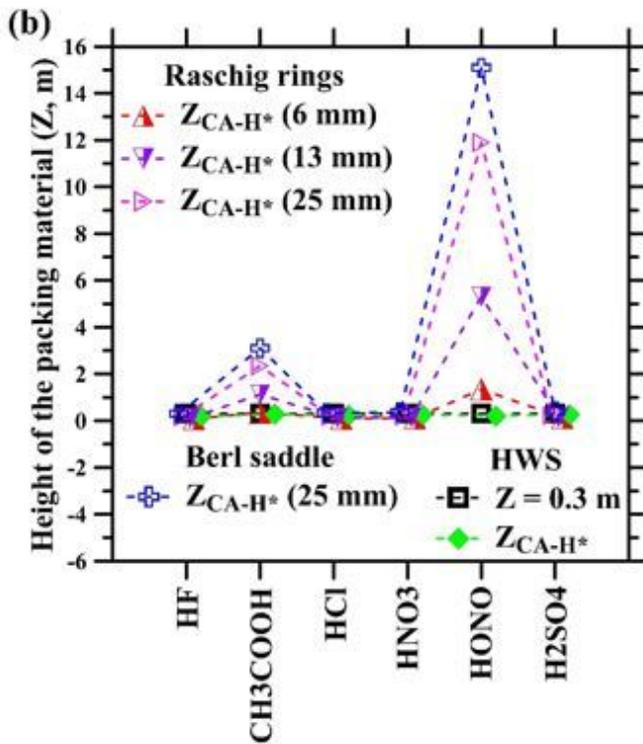
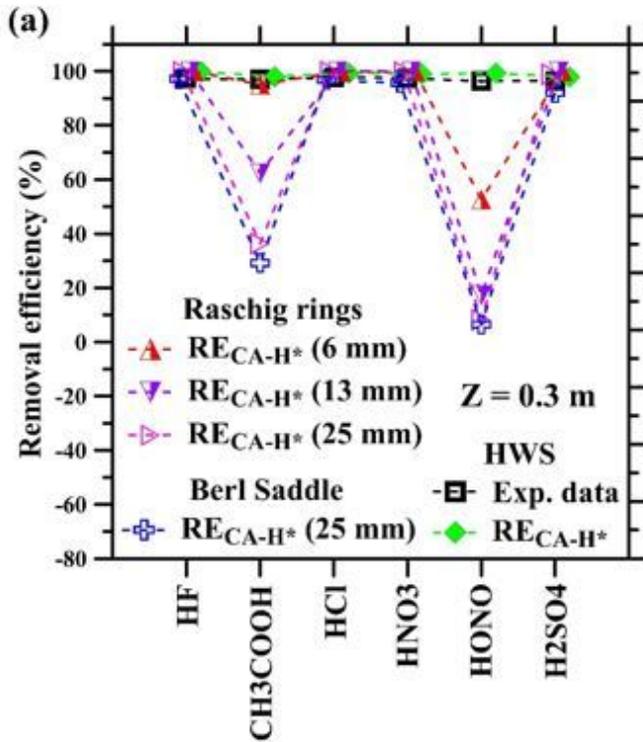


Figure 6

Comparison of a the theoretical removal efficiency, and b the theoretical height of the traditional WS and the HWS (Note: the data are slightly shifted to avoid overlapping).