

The effect of phosphate application in Pb-contaminated soil on the oxidative stress of leaves, Pb accumulation in biomass of maize and Pb speciation in rhizosphere soil

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Research Article

Keywords: antioxidant enzyme, heavy metal Pb, phosphate speciation, Pb speciation, accumulation of Pb in below- and above-ground biomass

Posted Date: March 21st, 2022

DOI: <https://doi.org/10.21203/rs.3.rs-1460757/v1>

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1 **The effect of phosphate application in Pb-contaminated soil on the oxidative**
2 **stress of leaves, Pb accumulation in biomass of maize and Pb speciation in**
3 **rhizosphere soil**

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Funding information: National Natural Science Foundation of China: 42167009, 31300349; Special Project of Basic Research in Yunnan Local Colleges and Universities: 2018FH001-004, 2018FH001-003; Scientific and Technological Innovation team Project of Agricultural Resources Utilization of Kunming University; Major Program for Basic Research Project of Yunnan Province: 202101BC070002.

25

26 **Abstract:** Reducing the bioavailability of Pb in soil is the key to alleviate the toxicity
27 of Pb to plant. Maize were exposed to Pb 100 mg/kg of soil with three fertilizer levels
28 of control (T1), nitrogen, phosphorus and potassium of 204 mg/kg (T2) and nitrogen,
29 potassium of 204 mg/kg (T3). The phosphate supplement lead to the reduction by
30 24.92%, 29.73% and 25.31% respectively in activity of total superoxide dismutase (T-
31 SOD), peroxidase (POD) and concentration of lipid peroxidation (MDA) in maize
32 leaves, and reduced Pb accumulation in above- and below-ground biomass of maize by
33 39.20% and 37.58%. In T2 treatment group, the water soluble Pb, ionic fraction and
34 carbonate fraction Pb in rhizosphere soil decreased by 37.57%, 36.36% and 43.24%,
35 and organic fraction Pb and residual fraction Pb was the highest with the value of 11.67
36 and 18.57 mg/kg; the soil aluminum bound (Al-P) and iron bound phosphate (Fe-P)
37 were the highest with 93.53 mg/kg and 230.32 mg/kg, indicating that the phosphate
38 supplement increases the soil ionic P and transforms the chemically mobilized P (such
39 as O-P, Ca-P) into the bioavailable P. Moreover, the soil organic fraction and residual
40 fraction immobilized Pb was positively correlated with the bioavailable Al-P and Fe-P,
41 indicating that the ionic fraction P (Al-P and Fe-P) react with Pb and produce residual
42 P-Pb compounds. Therefore, phosphate supplement to Pb contaminated soil could
43 transfer unstable fraction Pb into stable fraction Pb by P-induced Pb immobilization,
44 reduce the bioavailability of Pb and alleviate the toxicity of heavy metal to plant.

45 **Keywords:** antioxidant enzyme, heavy metal Pb, phosphate speciation, Pb speciation,
46 accumulation of Pb in below- and above-ground biomass

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48

49 **Introduction**

50 High rates of industrial and agricultural development have led to land
51 contamination by global pollutants such as heavy metal Lead (Pb), especially in Yunnan
52 province, southwest of China. Pb is a highly toxic environmental pollutant that has high
53 mobility in the soil-plant system and negatively impacts plants and animals. Its
54 concentration in some land types such as mining area and farmland around the mining

55 area has increased due to industry, mining and smelting (He et al. 2015; Kubier and
56 Pichler 2019; Deng et al. 2022; Chen et al. 2022; Zhang et al. 2021; Sun et al. 2020;
57 Yan et al. 2019). Pb can be retained in the soil through a series of reactions such as
58 adsorption, ionic exchange and precipitation, and formed various Pb-compounds which
59 have different effects on oxidative stress of plant and enrichment in plant organs (Selim
60 2015; Wan et al. 2015). In general, the Pb-compounds with high mobility are more
61 bioavailability and more environmental risk for plant. Heavy metal Pb has been divided
62 into four speciation: soluble/exchangeable fraction (F1), reducible fraction (F2),
63 oxidizable fraction (F3) and residual fraction (F4), and the mobility is decreased in the
64 order of: F1>F2>F3>F4 (Ure et al. 1993). The residual fraction of Pb has been
65 recognized as the most stable fraction which causes less effect on the oxidative stress
66 of plant than that of other fraction of Pb (Zeng et al. 2017). Based on this theory, some
67 researches and remediation strategies designed to promote the transformation of Pb
68 from unstable fraction (F1, F2 and F3) to mineral structure (F4) was focused, aiming to
69 protect against Pb-induce oxidative stress and enrichment for Pb in plant's organs
70 ((Awan et al. 2020; de Anicésio and Monteiro 2021; He et al. 2020; Hua et al. 2020;
71 da Silva et al. 2018; Moreira et al. 2018).

72 In recent years, phytoremediation for heavy metal contaminated some mining
73 areas, such as the hyperaccumulator plants of *Pteris vittata L.*, tanzania guinea grass
74 and *Arabis alpina L.var.parviflora Franch* has been studied under conditions of
75 phytotoxicity and the different fraction of heavy metals ((de Souza Cardoso and
76 Monteiro 2021; Wang et al. 2022; Guo et al. 2017; Zhang et al. 2020). These
77 hyperaccumulator plants has shown promise for phytoextraction of metals due to its
78 high amount of heavy metals absorption and antioxidant mechanism which have the
79 function of eliminating ROS, transforming them into non-toxic compounds (Nematian
80 and Kazemeini 2013; Xu et al. 2020; Kathal et al. 2016). However, the toxic effects of
81 Pb depend on speciation of Pb. The physiological changes of plant rhizosphere in Pb-
82 contaminated soil can affect the metal transformation through acidification effect and
83 mycorrhizae effect (Bolan et al. 2014; Zhang and Van Gestel 2019; Baruah et al. 2020;
84 Biliás et al. 2021). For farmland crops, such as maize, the supplement of fertilizer is

85 inevitable. Whether the application of fertilizer changes the speciation of Pb is also an
86 important aspect of its tolerance mechanism. The application of water-soluble P-
87 compounds of fertilizer could dissolve in soil with the formation of dicalcium phosphate
88 and phosphoric acid (H_3PO_4), and then H_3PO_4 dissociates into dihydrogen phosphate
89 and hydrogen ions, reducing the soil pH to a low level (Zeng et al. 2013; Basta and
90 McGowen 2004). Theoretically, the decrease of soil pH will promote the transformation
91 of residual fraction Pb into soluble/exchangeable fraction Pb and increase the
92 bioavailable Pb for plant (Cao et al. 2008). In addition, phosphorus, nitrogen and
93 potassium supplement can reduce the deleterious effects of Cd toxicity by increasing
94 photosynthesis in tomato (*Solanum lycopersicum L.*) plants (Naciri et al. 2021) and
95 antioxidant enzymes activities and proline and polyamines synthesis in tanzania guinea
96 grass (de Anicésio and Monteiro 2021). However, few studies have investigated the
97 physiological processes and the changes of speciation of Pb involved in the role of
98 phosphorous and nitrogen in alleviating Pb toxicity in plants, especially in farmland
99 crops, such as maize.

100 The phosphate supplement inhibits the absorption of Pb and improves the P/Pb
101 ratio in plants by the competitive absorption of phosphorus and heavy metal Pb in plants,
102 so as to alleviate the toxicity of Pb to plants, which shows that the plants has high
103 tolerance to heavy metal Pb (Lessl and Ma 2013; Han et al. 2017). However, phosphate
104 fertilizer supplement in Pb-contaminated soil will produce a large amount of
105 bioavailable phosphorus and Pb in the rhizosphere, and increases the chance of
106 chemical reaction of residual P-Pb compounds. Therefore, it is difficult to judge
107 whether the absorption of Pb by plants is blocked due to the formation of residual P-Pb
108 compounds in the rhizosphere, or the competitive absorption between phosphorus and
109 lead improves plant tolerance. Therefore, it needs to be further studied on the
110 relationship between phosphate application, and Pb speciation and physiological
111 changes of plants, especially in farmland crop. This study was carried out to evaluate
112 the effect of phosphate on rhizosphere soil Pb speciation, enrichment of Pb by plant and
113 changes in the antioxidant system of maize under Pb toxicity. The hypothesis is that

114 improving phosphorus nutrient can transform Pb speciation in rhizosphere soil and then
115 affects the antioxidant system, increasing the tolerance of maize to this heavy metal.

116 **Materials and methods**

117 **Experiment design**

118 Jinding lead zinc mining area in Lanping County of Yunnan province, southwest
119 of China has a long mining history, causing varying degrees of heavy metal pollution
120 to the surrounding land (Zhan et al. 2019). The mining area and wasteland, farmland
121 around the mining area, river beach in the mining area, farmland around the smelter in
122 the mining area and other land types have serious soil heavy metal pollution, especially
123 lead pollution (Fu et al. 2017). Among these land types, the highest content of heavy
124 metal Pb in soil is 2562.74 mg/kg, and the variation range of Pb content in farmland
125 around the mining area is 59.74-310.69 mg/kg (He et al. 2021; Zhang et al. 2002; Zhang
126 et al. 2021). According to the Pb content of the farmland soil around the mining area,
127 the concentration of heavy metal Pb (PbNO_3) was set as 100 mg/kg in pot experiments.
128 The soil used in pot experiments was red soil collected from farmland in Lanping
129 County. Red soil and humus are mixed by 1:1 weight, and then the heavy metal Pb
130 (PbNO_3) is added to make the Pb content in potted soil 100 mg/kg. After that, the soil
131 is kept moist and placed for 30 days, so that the heavy metal Pb can react with soil ions
132 to achieve even distribution of the pollutants.

133 Black plastic bags were used to cover the interior of the pots to prevent the loss of
134 soil solution including water soluble pollutants and nutrients. The maize variety
135 Qiandan 88 planted in large area was used as the research object. Firstly, 10 maize seeds
136 (Qiandan 88) are sterilized (75% $\text{C}_2\text{H}_5\text{OH}$ and 30% H_2O_2 solution is prepared in a
137 volume ratio of 4:1, soaked for 30s and then washed with distilled water). Then, these
138 seeds were transferred to the soil with PbNO_3 pollution 100 mg/kg soil pot and each Pb
139 treatment group was performed in quintuplicate. After 20 days of maize growth,
140 phosphate (P), nitrogenous (N) and potassium (K) fertilizer were supplied to the soil.
141 Therefore, the experimental treatments were as following: (i) 100 mg/kg Pb
142 contaminated soil (T1 treatment group); (ii) 100 mg/kg Pb contaminated soil supplied
143 with N, P and K of 204 mg/kg (T2 treatment group); (iii) 100 mg/kg Pb contaminated

144 soil supplied with N, K of 204 mg/kg (T3 treatment group). Soil moisture in the pots
145 was maintained at 50% of the full moisture capacity by daily watering, the need for
146 which was determined by weighing the pots. The plants were grown under greenhouse
147 conditions at 24–28/20–22 °C and under natural illumination for 60 days.

148 **Plant analyses**

149 At the growth of 2 months, the plants were removed from the pots, the above-
150 ground biomass (including stems and leaves) and below-ground biomass (roots) were
151 separated from the maize, and the roots were shaken and washed free of soil with tap
152 water. Plant biomass of maize was weighed and sampled for biochemical and chemical
153 analyses, and dried to constant weight in an oven at 75 °C. The dried plant biomass was
154 digested with HNO₃-H₂O₂, and the content of heavy metal Pb in above- and below-
155 ground biomass was determined by ICP-MS spectrometry. According to the methods
156 of Tara and Ewa (Tara and Urszula 2021; Ewa and Beata 2002), the enrichment
157 coefficient of heavy metal Pb and the content of Pb in above- and below-ground
158 biomass of maize were calculated.

159 Samples of leaves (1 g) were ground in a mortar with quartz sand and Na/K-
160 phosphate buffer (pH 7.0-7.4) with leaves samples/solution ratio of 1:4 and 1:9
161 respectively for total superoxide dismutase (T-SOD) and peroxidase (POD). 1 g leaves
162 of maize were ground in a mortar with quartz and 10 ml trichloroacetic acid (5%
163 C₂HCl₃O₂) for malondialdehyde (MDA). The homogenate was centrifuged for 10 min,
164 the supernatant liquid was decanted, the pellet was suspended in 2 mL of the same
165 buffer, and the suspension was centrifuged again. Then T-SOD and POD activities and
166 MDA concentration were determined by commercial kits produced by Nanjing
167 Jiancheng Bioengineering Instituted with mechanisms described in Li et al. (2013).

168 **Measure of chemical properties in maize rhizosphere soil**

169 The samples of rhizosphere soil were removed from maize rhizosphere for soil
170 enzyme activity, and air-dried and grounded in a ball mill. After thorough mixing, 200
171 mg of the sample, required for analysis, was collected in triplicate. The soil urease was
172 measured by sodium phenolate colorimetry, which was calculated from the mass of
173 NH₃-N (mg) produced by 1.0 g soil cultured at 37 °C for 24 h; the acid phosphatase was

174 measured by disodium diphenyl phosphate colorimetry, which was calculated by the
175 mass of phenol released by 1.0 g soil cultured for 24 h; the sucrase activity was
176 measured by 3,5-dinitrosalicylic acid colorimetry, which was calculated by the mass of
177 glucose (mg) produced by 1.0 g soil cultured at 37 °C for 24 h (Geng et al. 2017; Ma et
178 al. 2018). Each sample was performed in triplicate. The dried and grounded soil was
179 used to determine water-soluble phosphorus, aluminum bound phosphorus (Al-P), iron
180 bound phosphorus (Fe-P), occluded phosphorus (O-P) and calcium bound phosphorus
181 (Ca-P) by Sequential Extraction method (Zhang 1982; Lu 2000; Wang and Lu 2020).
182 Then the different soil extraction solution was determined by molybdenum-antimony
183 colorimetric method. Seven speciation Pb of soil, including water soluble Pb, carbonate
184 fraction Pb, ionic fraction Pb, humic acid fraction Pb, Fe-Mn oxide fraction Pb, organic
185 fraction Pb and residual fraction Pb, are extracted by Seven-step extraction method
186 according to the standard of China Geological Survey (Dong 2017). Then these soil
187 extraction solutions were determined by Inductively Coupled Plasma Mass
188 Spectrometry (ICP-MS). 10 g soil was sampled and dissolved by 25 mL of deionized
189 water at 20-25 °C, then soil pH was measured by pH meter.

190 **Statistics**

191 Data were processed by calculating the means of at least three replicates. Standard
192 deviations and confidence intervals were used at $p \leq 0.05$. SPSS 22.0 software was used
193 to analyze the effects of different fertilization treatments with the same Pb concentration
194 toxicity on rhizosphere soil enzyme activity, phosphorus and Pb speciation in maize
195 rhizosphere soil, antioxidant enzyme activity in maize leaves, and accumulation and
196 enrichment coefficient of heavy metal Pb by maize. And Duncan's multiple range tests
197 were applied in analysis of variance (ANOVA) ($p < 0.05$). Pearson parameter was
198 applied in analysis of the correlation among the changes of pH, different P speciation
199 and Pb speciation in rhizosphere soil (** $p < 0.01$; * $p < 0.05$). Originpro 2018 software
200 is used to perform linear fitting or nonlinear curve fitting (exponential) on the
201 relationship between phosphorus speciation and chemically stable Pb (organic fraction
202 Pb+ residual fraction Pb) of maize rhizosphere soil.

203 **Results**

204 **Changes of antioxidant enzyme system in maize leaves under Pb stress**

205 The total superoxide dismutase (T-SOD), peroxidase (POD) and lipid peroxidation
206 (MDA) in maize leaves under the stress of 100 mg/kg Pb concentration with the
207 different N, P and K supplement are shown in Fig.1. In T1, T2 and T3 treatments, the
208 activities of T-SOD, POD and MDA in maize leaves were T1 treatment group >T3
209 treatment group>T2 treatment group. The activities of T-SOD and POD in maize leaves
210 in T2 treatment group were significantly different from those in the other two groups (p
211 <0.05), and there was no significant difference between T1 and T3 treatment groups (p
212 < 0.05). The maximum value of T-SOD and POD appeared in T1 treatment with the
213 value of 60.48 U/mgprot, 72.11 U/mgprot, respectively. Compared to T1 treatment
214 group, T-SOD and POD in maize leaves of T2 and T3 treatment decreased by 24.92%
215 and 2.15%, 29.73% and 23.20%, respectively. There was significant difference in
216 concentration of MDA in maize leaves among T1, T2 and T3 treatments ($p < 0.05$),
217 and the highest value of it appeared in T1 treatment group with the value of 103.30
218 nmol/mgprot. The concentration of MDA of T2, T3 decreased by 25.31% and 13.14%,
219 respectively, comparing to T1 treatment group.

220

221 Fig.1 Maize superoxide dismutase (SOD), peroxidase (POD) and lipid peroxidation
222 (MDA) in leaves

223

224 **Accumulation characteristics of Pb in maize**

225 The accumulation of Pb in above- and below-ground biomass of maize is shown
226 in Fig.2. The accumulation of Pb in below-ground biomass of maize was higher than
227 that of the above-ground biomass among these treatment groups. In both aboveground
228 and below-ground biomass of maize, the accumulation of Pb in T1 treatment group was
229 significantly higher than that in the other two groups, and there was no significant
230 difference between T2 and T3 treatment groups ($p < 0.05$). In T2 and T3 treatment
231 groups, the accumulation of heavy metal Pb by above- and below-ground biomass of
232 maize decreased by 39.20% and 37.58%, 30.82% and 29.18%, respectively, compared
233 with T1 treatment group. The enrichment coefficient for Pb in maize was similar to the

234 accumulation of Pb in above- and below-ground biomass of maize. The highest
235 enrichment coefficient for Pb in maize appeared in T1 treatment group with the value
236 of 0.80, and this parameter of maize in T2 and T3 treatment groups has no significant
237 difference and were significantly lower than that of T1 treatment group ($p < 0.05$). The
238 enrichment coefficient for Pb in maize in T2 and T3 treatment groups decreased by
239 12.03% and 6.22%, respectively, compared with T1 treatment group.

240

241 Fig.2 Enrichment coefficient of heavy metal Pb in above- and below-ground biomass
242 of maize

243

244 **Pb and phosphorus speciation and enzyme activities in soil**

245 The speciation of Pb in maize rhizosphere soil is shown in Fig.3. The soil water
246 soluble Pb, ionic fraction Pb and carbonate fraction Pb content have significant
247 difference ($p < 0.05$) and showed T1 treatment group > T3 treatment group > T2 treatment
248 group. However, the contents of organic fraction and residual fraction Pb in T2 and T3
249 treatment groups were higher than those in T1 treatment group and both speciation of
250 Pb showed T2 treatment group > T3 treatment group > T1 treatment group. The highest
251 content of water soluble Pb in the rhizosphere soil of maize appeared in the T1 treatment
252 group with the value of 0.05 mg/kg. Compared with T1 treatment, the water soluble Pb
253 in maize rhizosphere soil in T2 and T3 treatments decreased by 37.57% and 28.65%
254 respectively, and ionic fraction and carbonate fraction Pb in both treatment groups
255 decreased by 36.36% and 28.16%, 43.24% and 36.37%, respectively, which were
256 significantly lower than those in T1 treatment group. In T2 and T3 treatments, there
257 was no significant difference in ionic Pb in maize rhizosphere soil, but there was
258 significant difference in carbonate fraction Pb in rhizosphere soil ($p < 0.05$). The content
259 of humic acid fraction Pb in maize rhizosphere soil in T2 treatment group was lower
260 than that of the other two groups. The content of Fe-Mn oxides Pb in maize rhizosphere
261 soil of T2 treatment group was significantly lower than that in T1 treatment group, and
262 the contents of organic fraction Pb and residual fraction Pb in rhizosphere showed the
263 maximum value with 11.67 and 18.57 mg/kg, respectively, which were significantly

264 higher than those in other groups.

265

266 Fig.3 heavy metal Pb speciation in maize rhizosphere soil

267

268 The speciation of phosphorus in maize rhizosphere soil are shown in Fig.4. The
269 contents of aluminum bound phosphate (Al-P) and iron bound phosphate (Fe-P) in
270 maize rhizosphere soil showed T2 treatment group > T3 treatment group > T1 treatment
271 group. The content of Al-P and Fe-P in maize rhizosphere soil showed T2 treatment
272 group > T3 treatment group > T1 treatment group. The highest value of maize rhizosphere
273 soil Al-P appeared in T2 treatment group with the value of 93.53 mg/kg, and increased
274 by 109.62% compared with T1 treatment group. In T2 treatment group, the content of
275 Fe-P in maize rhizosphere soil showed the maximum value with 230.32 mg/kg. The
276 contents of occluded phosphate (O-P) and calcium bound phosphate (Ca-P) in maize
277 rhizosphere soil were T1 treatment group > T3 treatment group > T2 treatment group, and
278 the rhizosphere soil O-P contents in T1, T2 and T3 treatment groups were 39.38, 29.89
279 and 38.17 mg/kg, respectively. And the content of Ca-P in maize rhizosphere soil in T1
280 treatment group was the highest with the value of 922.38 mg/kg, but in T2 and T3
281 treatment groups, the rhizosphere soil Ca-P content decreased by 57.05% and 20.51%
282 respectively compared with T1 treatment.

283

284 Fig.4 Soil phosphorus speciation

285

286 The maize rhizosphere soil pH of T1, T2 and T3 treatment groups is shown in
287 Fig.5A. The soil pH was T1 treatment group > T2 treatment group > T3 treatment group
288 of maize rhizosphere soil, indicating that the supplement of N, P, K to the soil reduces
289 the pH of the soil. Compared with T1 treatment group, the rhizosphere soil pH in T2
290 and T3 treatment groups decreased by 0.46 and 0.50, respectively. The activities of
291 urease, acid phosphatase and sucrase in maize rhizosphere soil are shown in Fig.5B, 5C
292 and 5D. The range of soil urease activity was 8.88-10.26 mg/g.24h⁻¹, and there was no
293 significant difference among these groups ($p < 0.05$). The activity of acid phosphatase

294 in maize rhizosphere soil was the highest in T3 treatment group, followed by T2
295 treatment group and the lowest in T1 treatment group. The soil acid phosphatase activity
296 in the T1 treatment group was the lowest with the value of 1443.18 mg/g.24h⁻¹, and the
297 activity of soil acid phosphatase in T2 and T3 treatment groups increased by 18.11%
298 and 117.43%, respectively, compared with T1 treatment group. The maize rhizosphere
299 soil sucrase activity was T3 treatment group>T1 treatment group>T2 treatment group,
300 and there was significant difference among these groups ($p < 0.05$), and the highest soil
301 invertase activity in T3 treatment group was 41.66 mg/g.24h⁻¹.

302

303 Fig.5 The maize rhizosphere soil pH and enzyme activity

304

305 **Correlation analysis of soil Pb and phosphorus speciation**

306 The correlation analysis between soil pH, different phosphorus speciation and
307 different Pb speciation is shown in Table 1. The soil Al-P, residual fraction Pb and
308 organic fraction Pb were significantly negatively correlated ($p < 0.01$) or significantly
309 negatively correlated ($p < 0.05$) with the maize rhizosphere soil pH, but water soluble
310 Pb, ionic fraction Pb and carbonate fraction Pb were positively correlated with soil pH
311 ($p < 0.01$). The soil Al-P was negatively correlated with Ca-P, water soluble Pb, ionic
312 fraction Pb, and carbonate fraction Pb ($p < 0.01$), positively correlated with organic
313 fraction Pb and residual fraction Pb ($p < 0.01$), and negatively correlated with Fe-Mn
314 oxide fraction Pb ($p < 0.05$). There was a significant positive correlation between O-P
315 and Ca-P ($p < 0.05$). The Ca-P was positively correlated with water soluble Pb and ionic
316 fraction Pb ($p < 0.01$), and negatively correlated with organic fraction Pb and residual
317 fraction Pb ($p < 0.01$). The water soluble Pb has a significantly positive correlation with
318 ionic fraction Pb and carbonate fraction Pb ($p < 0.01$), indicating that these three
319 speciation of Pb is extremely easy to transform each other. Water soluble Pb, ionic
320 fraction Pb, carbonate fraction Pb and Fe-Mn oxide fraction Pb all have a extremely
321 significantly negative correlation with organic fraction Pb and residual fraction Pb ($p <$
322 0.01), because the chemical properties of organic fraction Pb and residual fraction Pb
323 are stable and it is difficult to convert other speciation of Pb.

324

325

Table 1 Correlation analysis of Pb and P

326

327 P compound in soil reacts directly or indirectly with heavy metal Pb. In this study,
328 regression analysis is carried out on the data of P and Pb speciation to explore the
329 relationship between P speciation and Pb speciation (Fig.6). The contents of
330 immobilized Pb (organic fraction+residual fraction) in soil were positively correlated
331 with Al-P and Fe-P, and the correlation coefficient R^2 were 0.91 and 0.92 respectively
332 and negatively correlated with Ca-P. The change of P-Pb relationship model is
333 consistent with the correlation analysis (Table 1), which indicates that mobilized
334 phosphorus (such as Al-P, Fe-P) may react with easily absorbed Pb (such as water
335 soluble Pb, ionic fraction Pb and carbonate fraction Pb), and Pb will transform to a
336 stable state Pb (organic fraction Pb, residual fraction Pb, etc.) to reduce the mobility
337 and bioavailability of Pb.

338

339 Fig.6 Regression analysis of residual fraction and organic fraction Pb and P speciation

340

(Al-P、Fe-P、O-P、Ca-P)

341

342 **Correlation analysis of antioxidant enzyme system, accumulation characteristics**
343 **and Pb speciation in maize leaves**

344 The correlation analysis between enrichment coefficient for Pb in maize and Pb
345 speciation in soil is shown in Table 2. The enrichment coefficient for Pb in maize was
346 significantly positively correlated ($p < 0.05$ or $p < 0.01$) with water soluble Pb, ionic
347 fraction Pb, carbonate fraction Pb and Fe-Mn oxide fraction Pb, and extremely
348 significantly negatively correlated with organic fraction Pb and residual fraction Pb (p
349 < 0.01). These results indicated that the enrichment of heavy metal Pb by plants depends
350 on the content of bioavailable Pb in soil to a certain extent, which can promote the high
351 enrichment of Pb by plants.

352

353 Table 2 Correlation between enrichment coefficient and Pb speciation

354

355 The correlation analysis between enzyme activities of T-SOD, POD, concentration
356 of MDA in maize leaves and soil different Pb speciation is shown in Table 3. The
357 activity of T-SOD was significantly positively correlated ($p < 0.01$ or $p < 0.05$) with
358 MDA content in maize leaves, soil water soluble Pb, carbonate fraction Pb and Fe-Mn
359 oxide fraction Pb. The activity of POD was extremely significantly positively correlated
360 with MDA content in maize leaves, soil water soluble Pb ($p < 0.01$), ionic fraction Pb
361 and carbonate fraction Pb ($p < 0.05$). The activity of T-SOD, POD and the content of
362 MDA in maize leaves were extremely significantly negatively correlated with organic
363 fraction Pb and residual fraction Pb ($p < 0.01$). This result indicated that if the
364 bioavailable Pb for plant in soil (such as water soluble Pb, ionic fraction Pb and
365 carbonate fraction Pb) is transformed into stable Pb (such as organic fraction Pb and
366 residual fraction Pb), the toxicity of heavy metal Pb on plants will be alleviated.

367

368 Table 3 Correlation between SOD、POD、MDA and soil Pb speciation

369

370 Discussion

371 Most studies have shown that aluminum bound phosphate (Al-P) and iron bound
372 phosphate (Fe-P) in soil are bioavailable phosphorus for plant, and Fe-P is converted
373 into phosphate by chemical reduction of Fe^{3+} in the reduction reaction process (Qin et al.,
374 2021). In the reduction process in soil, the calcium bound phosphate (Ca-P) in the
375 soil is easy to transform into Al-P and Fe-P to increase the proportion of bioavailable
376 phosphorus sources (Al-P, Fe-P) in the soil and then improves the potential phosphorus
377 supply capacity of the soil for plant (Fan et al. 2021). The occluded phosphate (O-P) in
378 soil is amorphous aluminum iron phosphate wrapped by Fe_2O_3 to form stable
379 compounds. The main soil type in Yunnan province is red soil, which is rich in iron,
380 aluminum compounds, which provides more adsorption sites for P adsorption and
381 increases the Fe-P content. Moreover, PO_4^{3-} will replace -OH groups and adsorb on the
382 surface of aluminum oxide, and Al^{3+} will combine with organic matter to form "Al-

383 organic matter-P" complex, resulting in the increase of Al-P content (Baquero et al.
384 2012; Song et al. 2020; Asenso et al. 2018). With the decrease of soil pH, amorphous
385 and crystalline ferric phosphate were hydrolyzed to release phosphorus, and a large
386 amount of H⁺ reduced the variable negative charge on the soil surface and promoted the
387 desorption of adsorbed phosphorus (Temporetti et al. 2019; Rehman et al. 2000; Xu et
388 al. 2015; Ansari and Thakur 2015). In this study, maize rhizosphere soil pH decreased
389 in the T2 and T3 treatment groups (see Fig. 5) because the chemical fertilizer increase
390 H⁺ in the soil and the absorption of salt-based cations by plants releases the same
391 amount of H⁺ into the soil, resulting in soil acidification (Zhang et al. 2017; Zhang et
392 al. 2019; And and Sridhar 2004; Dong et al. 2019). The O-P and Ca-P content of the
393 maize rhizosphere soil in T2 and T3 treatment groups were lower than that of T1
394 treatment group. Moreover and the Al-P of maize rhizosphere soil was significantly
395 negatively correlated with soil pH ($p < 0.05$) while the Ca-P in soil has a significant
396 positive correlation with soil pH ($p < 0.01$, in Table 1). This result indicates that the
397 application of chemical fertilizers in soil not only increases the content of ionic P in
398 maize rhizosphere soil, but also transforms the chemically mobilized P (such as O-P,
399 Ca-P) into the ionic state due to the decrease of soil pH. In this process, although
400 mobilized P is conducive to plant absorption (Schillaci et al. 2021; Istenič and Božič
401 2021; Pedro et al. 2021; Suresh et al. 2021), and can also increase the risk of soil
402 phosphorus loss with surface runoff (Boero 1977; Li et al. 2013; Temporetti et al. 2019).
403 Some studies have shown that the organic acids exuded by plant roots reduce the pH of
404 the rhizosphere soil and promotes the release of ionic P. And with the decrease of H⁺
405 and organic acid in the soil, the insoluble phosphorus compounds (such as calcium
406 phosphate compounds, hydroxy aluminum phosphate (iron) compounds) in the soil are
407 hydrolyzed and dissociated into H₂PO₄⁻, HPO₄²⁻ and other ionic states ((Hayes et al.
408 2000; Boafo et al. 2018). It is generally believed that the increase of ionic P will promote
409 the absorption of by plants, and high phosphorus in rhizosphere soil will inhibit the
410 activity of acid phosphatase and reduce the transformation of soil organic phosphorus
411 into inorganic phosphorus (Su et al. 2019; Wang and Lu 2020; Peng et al. 2015). But
412 when the phosphorus is deficient in soil, plants can promote the hydrolysis of organic

413 phosphorus compounds by increasing the activity of acid phosphatase in rhizosphere
414 soil, generate inorganic phosphorus that can be absorbed by plants and improving the
415 ability of soil to supply phosphorus (Hayes et al. 2000; Wang et al. 2021; Cheng et al.
416 2021; Zhan et al. 201). However, in this study, the soil phosphatase activity in the T2
417 and T3 treatment groups was higher than that in T1 treatment group, indicating that the
418 ionic P in the soil was still deficient relative to the plants. This may be due to some
419 chemical reaction between the ionic fraction P and the heavy metal Pb in rhizosphere
420 the soil, which reduced the bioavailability of P.

421 The water soluble, ionic fraction and carbonate fraction of heavy metal Pb are easy
422 to be absorbed by plants, Fe-Mn oxide fraction and organic fraction will be released
423 under oxidation and reduction, and the residue Pb is insoluble, which can be
424 transformed into soluble state by chemical reaction (Eyenubo 2021; Wang et al. 2021;
425 Ida and Eva 2021; Wang et al. 2021). The soil pH will affect the change of heavy metal
426 Pb form to a great extent. Soil acidification will catalyze the complex biochemical
427 reaction in the soil and increase the ionic fraction of heavy metal Pb, so as to transform
428 them into water soluble, ionic fraction and carbonate fraction, and correspondingly
429 enhance the toxicity to plants. Based on this, many scholars applied alkaline compounds,
430 such as lime, to the soil to increase the soil pH (Cheng et al. 2021; Han et al. 2021), and
431 the Fe^{2+} and Mn^{2+} are hydrolyzed and combined into $\text{Fe}(\text{OH})_2$ and $\text{Mn}(\text{OH})_2$, so that
432 the Fe-Mn oxide fraction Pb content increases. In this process, Pb combines with
433 organic fraction to form insoluble complex (organic combined state) to reduce the
434 availability of heavy metals Pb (Wan et al. 2022; Ming et al. 2022; Meng et al. 2021;
435 Xiao et al. 2021; Zhang et al. 2021). In this study, the soil pH in T2 and T3 was lower
436 than that in T1 treatment, but the soil water soluble Pb, ionic fraction Pb and carbonate
437 fraction Pb in the two treatment groups were lower than that in T1 treatment group, and
438 the contents of organic fraction and residual fraction Pb increased, indicating that of the
439 supplement of ionic fraction phosphate fertilizer in maize rhizosphere soil (see above
440 discussion) produces more chemically stable P-Pb compounds (organic and residual
441 fraction Pb). In table1, soil pH was significantly negatively correlated with chemically
442 stable Pb (organic and residual fraction Pb) ($p < 0.05$). The pH decreased in soil, but

443 the residual fraction Pb increased. The reason may be that the supplement of chemical
444 fertilizer acidifies the soil, and then a large amount of ionic Pb is produced in the soil,
445 which increases the opportunity of chemical reaction with ionic fraction P and generates
446 residual P-Pb compound precipitation. In addition, a large amount of ionic fraction P
447 (Al-P and Fe-P) in the soil provides an opportunity for its reaction with Pb. Correlation
448 analysis of P and Pb form in soil showed that the mobilized Pb (water soluble, ionic
449 fraction, carbonate fraction) and chemically stable Pb (organic and residual fraction)
450 show significantly negative correlation and significantly positive correlation with the
451 soil Al-P respectively ($p < 0.01$, in table 1). And the soil chemically stable Pb (organic
452 and residual fraction) were extremely significant negative correlation with the Ca-P
453 content ($p < 0.01$, in table 1). Through regression analysis of the speciation of
454 chemically stable Pb (organic bound Pb+residual Pb) and P, it is found that the amount
455 of organic fraction and residue fraction Pb increases linearly with Al-P and Fe-P, and
456 decreases with the increase of O-P and Ca-P. The application of phosphate fertilizer in
457 Pb-contaminated farmland can alleviate the toxicity of heavy metals on maize plants by
458 changing the speciation of Pb. The result of this study is consistent with these findings
459 that the supplement of hydroxyapatite, dipotassium hydrogen phosphate (K_2HPO_4) and
460 potassium dihydrogen phosphate (KH_2PO_4) to Pb-contaminated soil can form a large
461 amount of residual fraction lead phosphate precipitation in the rhizosphere soil of plants,
462 reduce the mobility of Pb and their bioavailability (Wang and Song 2019; Narwade and
463 Khairnar 2018; Aliabadi et al. 2014).

464 Heavy metal Pb stress can make plants produce reactive oxygen species (ROS),
465 destroy membrane lipids and hinder the formation of proteins, pigments and nucleic
466 acids (Kasperczyk et al. 2015). Moreover, a large amount of malondialdehyde (MDA)
467 accumulated in the plant, and the increase of its content indicates that the plant is under
468 high-level antioxidant stress (Rehman and Sheikh 2000). In order to eliminate the harm
469 of oxidative stress on cells, plants catalyze generation of hydroxyl radical (OH^\cdot) and
470 the transformation of superoxide (O_2^\cdot) to H_2O_2 through Haber Weiss and Fenton
471 reaction or other mechanisms, so as to eliminate the harm of reactive oxygen. Therefore,
472 the change of antioxidant enzyme system of plant is the emergency response

473 mechanism of plants under heavy metal stress (Rahbari et al. 201). The activity of plant
474 antioxidant enzyme depends on the degree of plant tolerance, heavy metal ion
475 concentration, exposure time and toxicity, and reflects the redox state of cells to a
476 certain extent (Bahrami et al. 2016; Muradoglu et al. 2015). The activities of SOD
477 and POD of plants, such as *Dendrobium officinale* (Guan et al. 2016), *Medicago*
478 *sativa* L. (Zhu et al. 2018), *Ricinus communis* L. (Yi et al. 2016) increased linearly
479 with the increase of heavy metal content. When the concentration of heavy
480 metals exceeded the regulation range of plant antioxidant enzyme system, the
481 activities of SOD and POD decreased, indicating that the mechanism of
482 removing pollutants by upregulating antioxidant enzymes system is limited
483 under the stress environment of high concentration of heavy metals. In this study,
484 under the same heavy metal Pb stress (100 mg/kg), the activities of SOD, POD and
485 MDA in maize leaves of T2 and T3 treatment groups were lower than those of T1
486 treatment group, especially the three parameter values of T2 treatment group were the
487 lowest. And the accumulation amount and enrichment coefficient for Pb in the above-
488 and below-ground biomass of maize in T2 and T3 were lower than those in T1 treatment
489 group, especially the maize of T2 treatment group has the lowest of accumulation
490 amount and enrichment coefficient for Pb. This may be due to the supplement of N, P
491 and K in T2 and T3 treatment groups reduced the content of bioavailable Pb in soil,
492 thus reducing the toxicity on maize plant, which is consistent with the results that the
493 application of phosphate fertilizer reduced the absorption and accumulation of Pb in
494 roots, stems, leaves and grains of *Oryza sativa* L. plants (Gao et al. 2015). However,
495 some studies believe that the inorganic salt ions such as phosphorus compete with heavy
496 metal Pb for adsorption (Hua et al. 2020), which reduces the absorption of Pb by plants.
497 In this experiment, it was found that a large amount of residual fraction Pb was
498 produced in maize rhizosphere soil (see above discussion), which reduced its toxicity
499 on plants. The application of N, K elements can improve the antioxidant enzyme
500 defense system of plants under heavy metal Pb stress, enhance the ability of eliminating
501 reactive oxygen species (ROS) and reduce the accumulation of malondialdehyde (MDA)
502 in plant (Lin 2011). And the content of soil urease in T3 treatment group was higher

503 than that of T1 treatment group, which may be because the increase of N and K elements
504 promoted the increase of plant biomass, resulting in the increase of urease content in
505 soil, which was conducive to the circulation of soil nitrogen. Potassium application
506 increased the contents of glutamate acid, proline and arginine in *guinea* grass roots
507 under Cd pollution stress; thereby improve the activities of SOD, POD and other
508 enzymes (de Anicésio and Monteir 2021). However, in this study, SOD and POD activity
509 of maize leaves decreased in T3 (supplement of N, K) treatment group (Fig. 1), and its
510 mechanism needs to be further discussed.

511

512 **Conclusion**

513 The supply of phosphorus fertilizer can reduce the Pb-induced oxidative stress in
514 maize, which is evidenced by the decrease in MDA concentration in the leaves of maize.
515 In addition, the activities of antioxidant enzymes (SOD and POD) in maize leaves
516 decreased, and the accumulation amount and enrichment coefficient of Pb in above-
517 and below-ground biomass of maize decreased after the application of phosphorus
518 fertilizer. Mitigation of Pb toxicity and accumulation amount of Pb in biomass of maize
519 is because soil ionic phosphorus reacts with Pb to form residual P-Pb compounds, which
520 is evidenced by the increase in maize rhizosphere soil organic fraction Pb and residual
521 fraction Pb with supply of phosphorus fertilizer, which reduces the mobility and
522 bioavailability of heavy metal Pb. Therefore, the form of Pb in soil can alleviate the
523 stress of Pb on maize, change its physiological indexes, and finally reduce the
524 enrichment of Pb in plants.

525

526 **Acknowledgement**

527 This work is funded by the National Natural Science Foundation of China
528 (42167009, 31300349), special project of Basic Research in Yunnan Local Colleges and
529 Universities (2018FH001-004, 2018FH001-003), Scientific and Technological
530 Innovation team Project of Agricultural Resources Utilization of Kunming University
531 and Major Program for Basic Research Project of Yunnan Province (202101BC070002).

532

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Figures

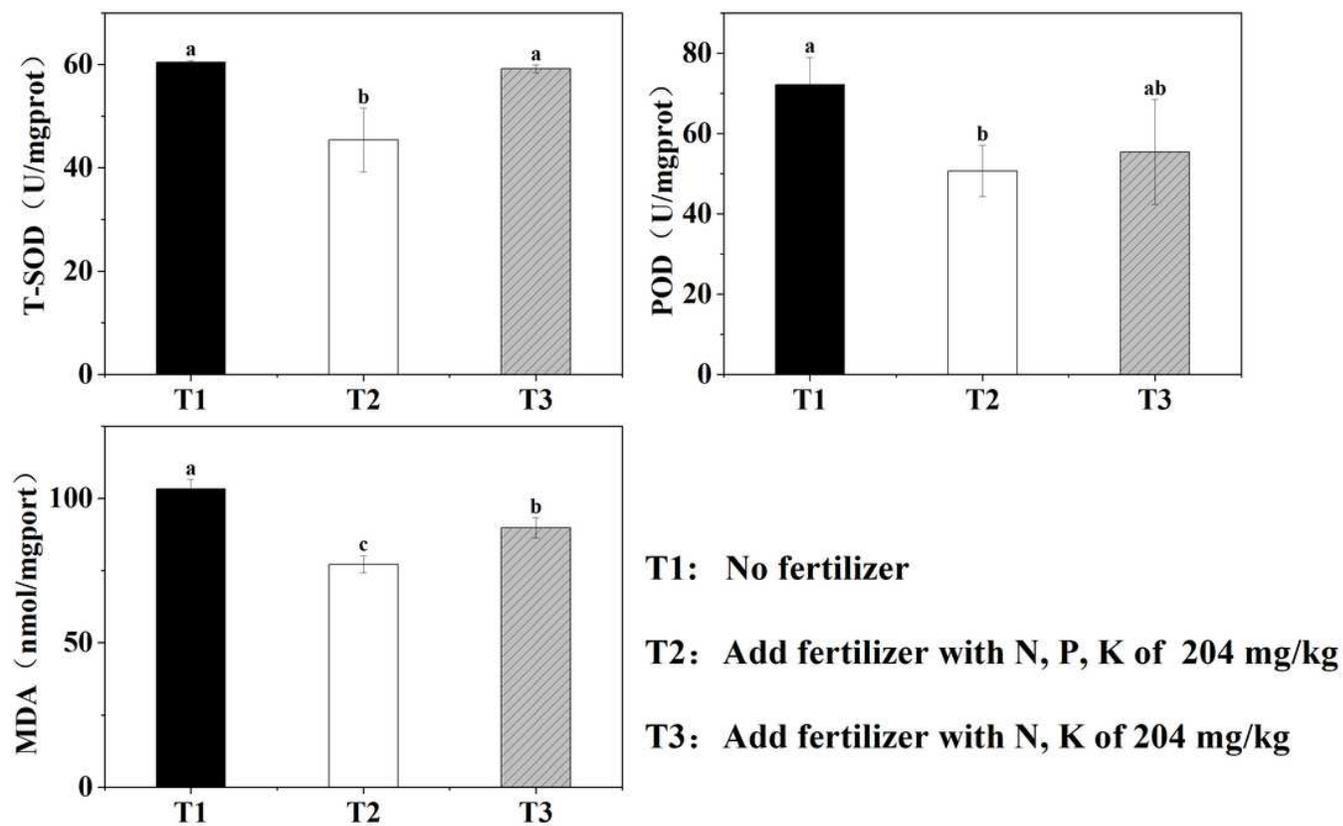


Figure 1

Maize superoxide dismutase (SOD), peroxidase (POD) and lipid peroxidation (MDA) in leaves

Figure 2

Enrichment coefficient of heavy metal Pb in above- and below-ground biomass of maize

Figure 3

heavy metal Pb speciation in maize rhizosphere soil

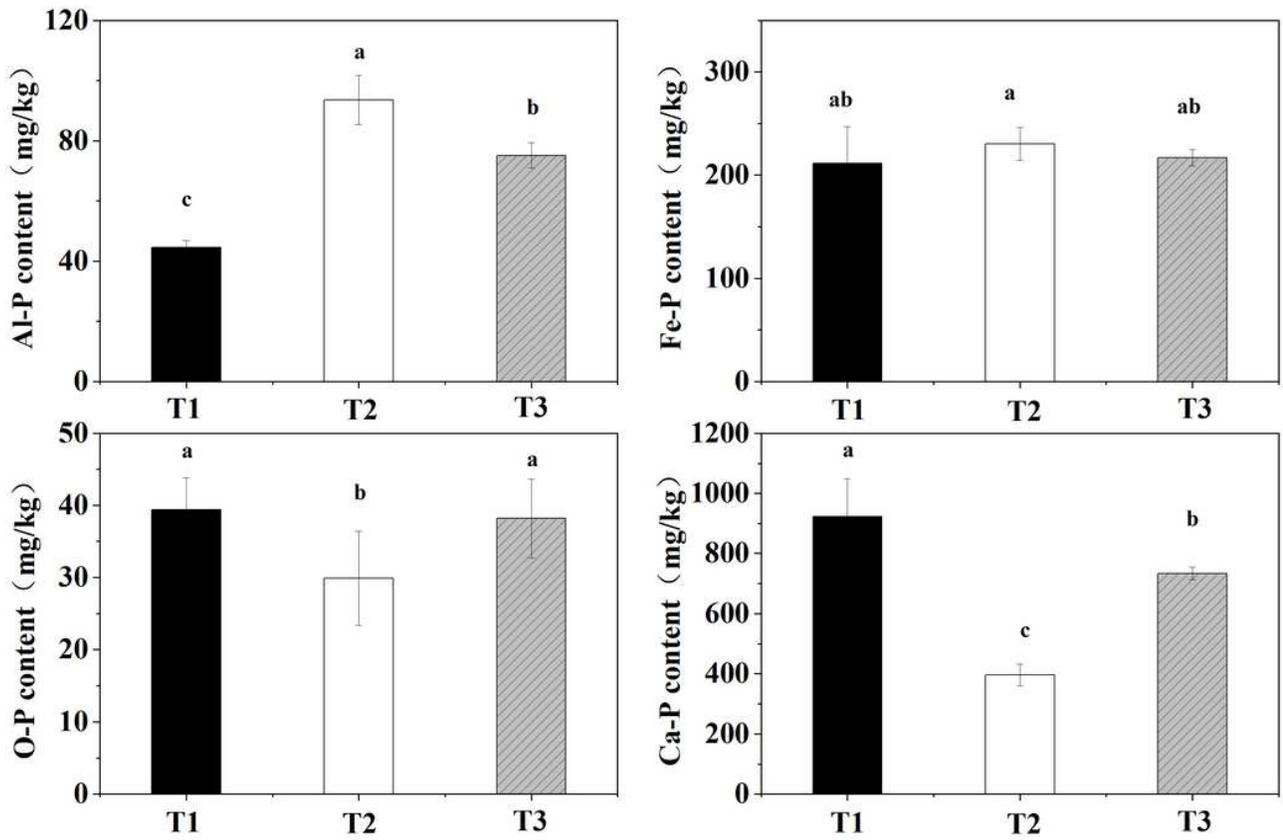


Figure 4

Soil phosphorus speciation

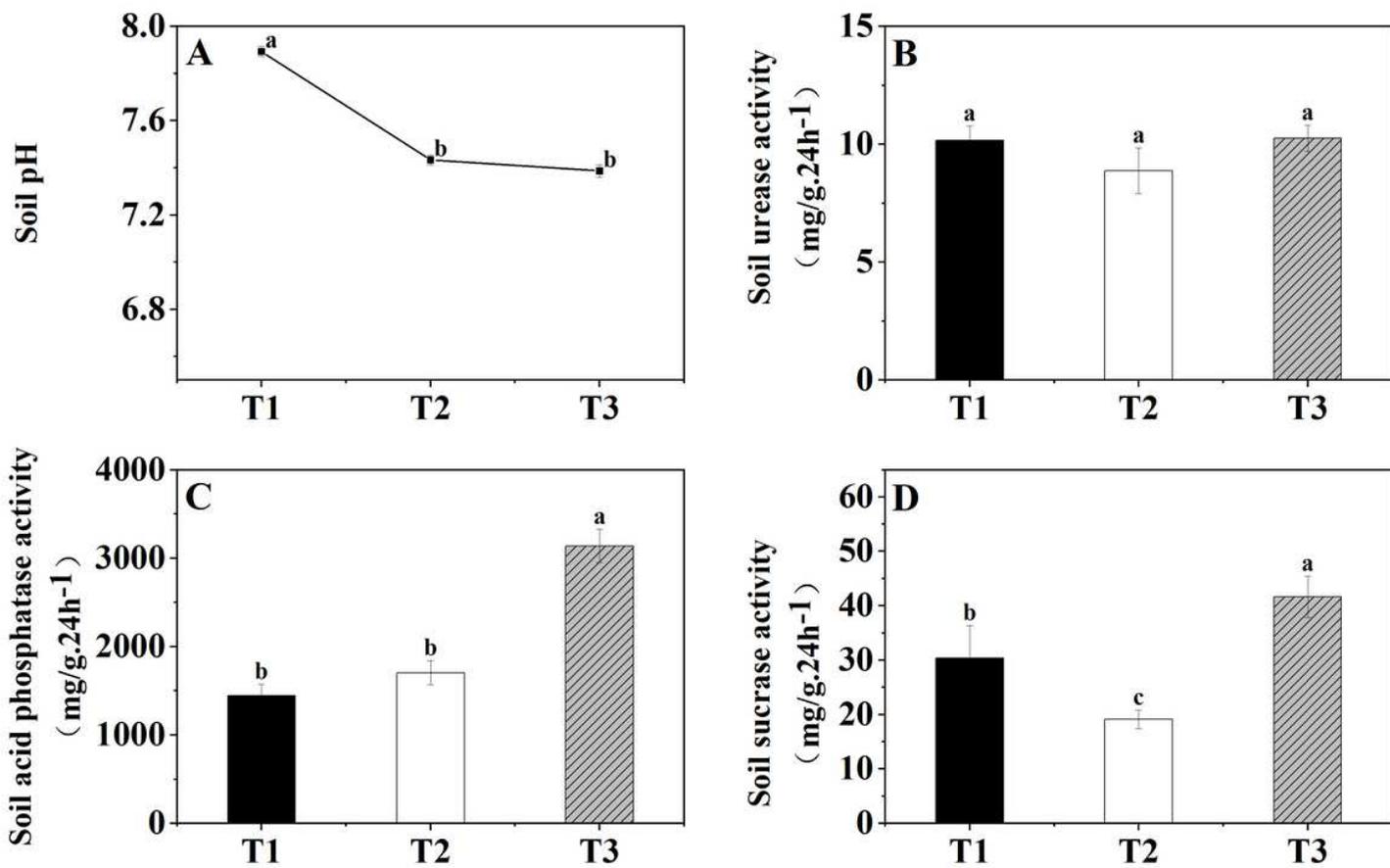


Figure 5

The maize rhizosphere soil pH and enzyme activity

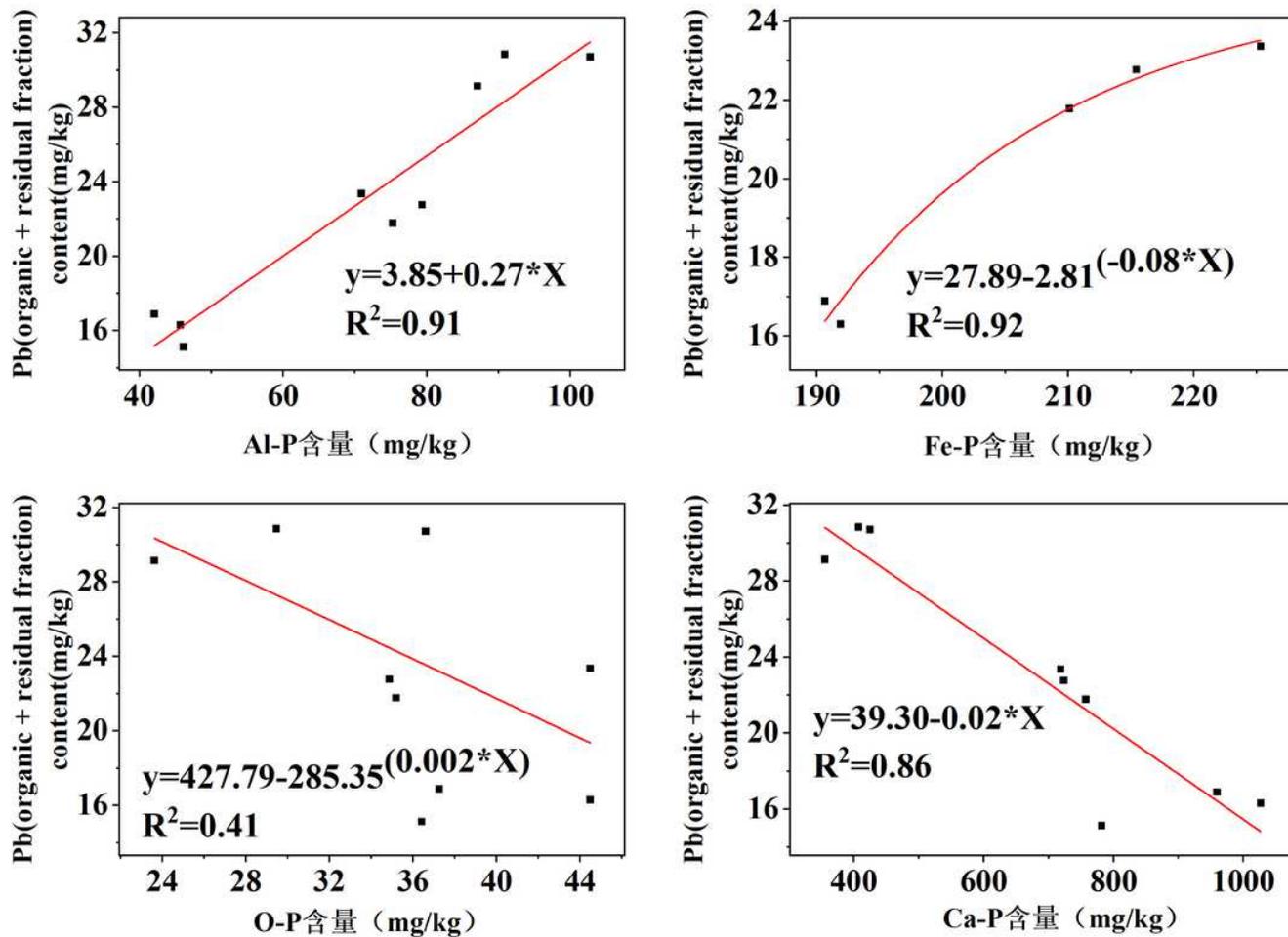


Figure 6

Regression analysis of residual fraction and organic fraction Pb and P speciation [Al-P][Fe-P][O-P][Ca-P]