

Waterbody loss due to urban expansion of large Chinese cities in last three decades

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1 Waterbody loss due to urban expansion of large Chinese cities in last three decades

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4
5 **Abstract:** Urban waterbodies are one of the most pertinent issues involved in multiple aspects of
6 Sustainable Development Goals (SDGs). However, waterbodies in large Chinese cities are highly
7 vulnerable to urban-land expansion, which is mostly due to economic development, population
8 growth, and rural-urban migration. In this work, we select 159 Chinese cities of over one million in
9 population to investigate the encroachment on waterbodies due to rapid urbanization from 1990 to
10 2018. Overall, 20.6% of natural waterbody area was lost during this period to urban expansion, and
11 this fraction varies from city to city and is related to waterbody abundance. With the acceleration of
12 urbanization, waterbody occupation is becoming more serious ($p < 0.01$). However, in all cities, this
13 encroachment has eased since 2010, which justifies the effective implementation of national-scale
14 policies to conserve urban waterbodies. In the future, ecological resources, including waterbody,
15 should be considered in urban planning to provide reasonable protection to waterbodies in the quest
16 for urban sustainability.

17 **Keywords:** Urban waterbody; urbanization; water encroachment; urban planning; China

18 19 Introduction

20 Waterbodies are vital for cities¹ by providing drinking water, transportation, and cultural
21 activities such as leisure tourism.^{2,3} The sixth objective of the Sustainable Development Goals
22 (SDGs) is sustainable water conservation and management. In particular, urban waterbodies
23 maintain the stability of local ecosystems and provide a wide range of ecosystem services, including
24 conservation of biodiversity, supplementation of groundwater, and climate adjustment.⁴⁻⁷
25 Furthermore, urban waterbodies are connected to networks that help regulate rainfall and contribute
26 significantly to the city's capacity to accommodate and recover rapidly from extreme weather and
27 other natural disasters.

28 Over the past decades, economic development and population growth in China fuelled by rural-
29 urban migration have driven the demand for urban land.⁸ From 2001 to 2018, the growth rate of
30 China's built-up land has been the highest in the world, contributing to nearly half of the total rise
31 globally.⁹ Complex human activities based on urban expansion are constantly changing the planet,
32 leading to deteriorated resources, frequent climate extremes, and severe perturbations of
33 biochemical cycles.¹⁰⁻¹³ China's rapid urbanization is characterized by extensive expansion based
34 on a high rate of consumption of natural resources.¹⁴ Impervious surfaces have replaced vast tracts
35 of ecological areas, leading to drastic ecological degradation.¹⁵ Studies have shown that changes in
36 land use and associated pressures have reduced the biodiversity of global land surfaces by up to
37 58%.¹⁶ Massive green space, water, and other natural habitats have deteriorated and/or been
38 converted to different uses, which now poses a significant challenge to urban ecological security.¹⁷
39 As a result, urban growth in China is now subject to the dual constraints of ecological space and
40 environment.

41 Widespread urban expansion is encroaching on waterbodies all over the world,^{18,19} particularly
42 in the large cities of Asia due to the recent rapid urbanization of developing countries.²⁰⁻²² On one
43 hand, the agglomerated population and the limited land of large cities make the phenomenon of

44 waterbody occupation more serious in large cities. On the other hand, urban planning in large cities
45 is generally more systematic and comprehensive. According to the literature, waterbodies in large
46 Chinese cities have been severely reduced and/or transformed into urban construction land.^{23–25}
47 Most previous studies explored spatial-temporal changes in waterbodies and made specific
48 suggestions for future urban planning on a regional basis.^{22,26–28} However, research is lacking on
49 how waterbodies are affected by urbanization on a national scale. Of particular importance is the
50 question of how much area have waterbodies lost on the national or regional scale?

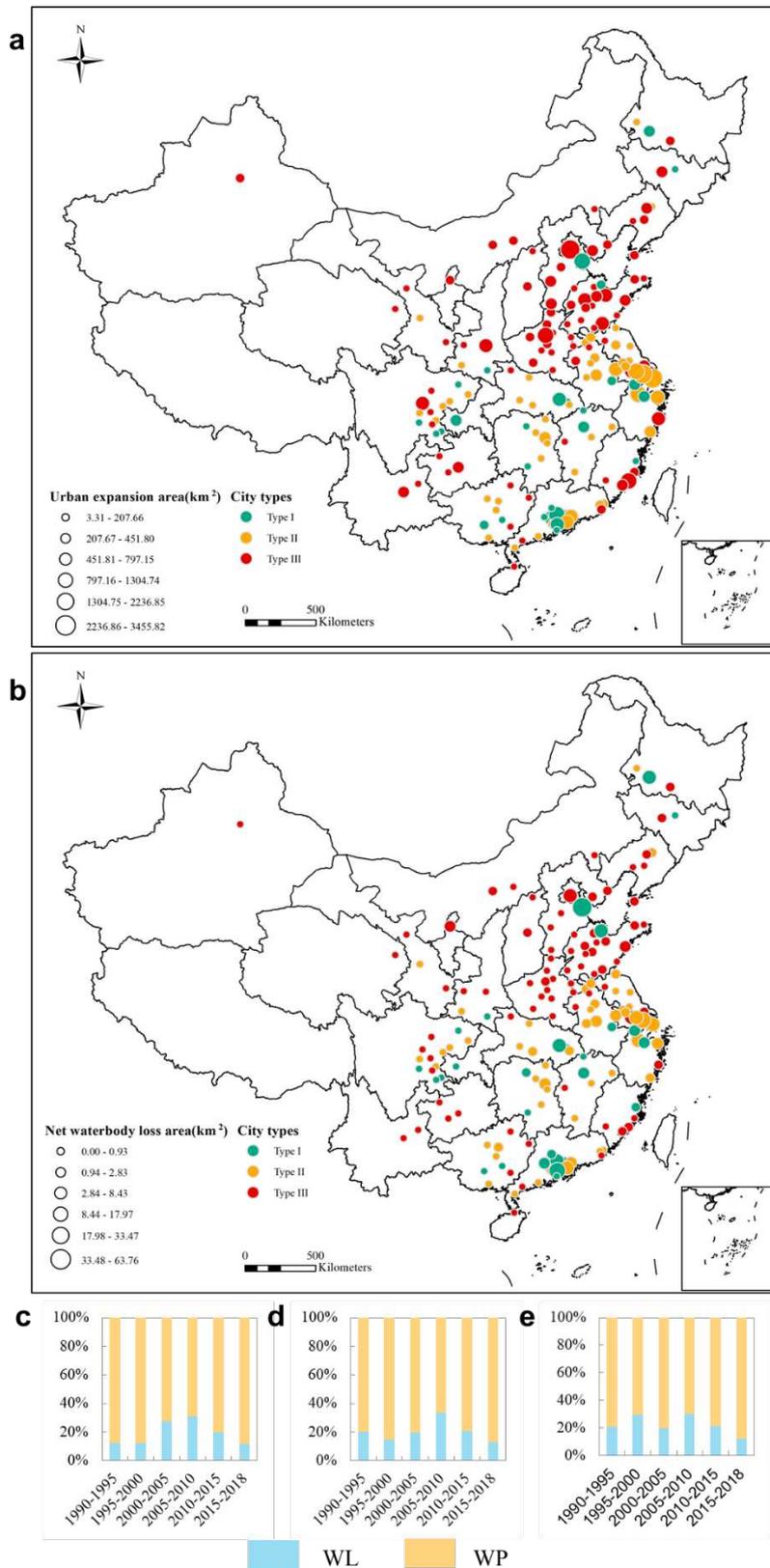
51 To deal with its rapid development, China is exploring a new path of sustainable urban planning.
52 In 2007, waterbody conservation was reinforced in Chinese urban planning, and the Administrative
53 Measures of the Urban Blue Line were issued to restrict development activities on urban
54 waterbodies. Reports have also discussed artificial water and water protection as cities strive for
55 liveable space. Has a turning point been reached in people’s concept of waterbody protection? This
56 paper aims to quantitatively assess how urban expansion affects waterbodies in Chinese cities of
57 over one million in population. Towards this end, we analyse the dynamics of urban boundaries and
58 waterbodies over different periods from 1990 to 2018 and explore whether and how waterbodies
59 are affected by urban expansion.

60 **Results**

61 **Unbalanced distribution of waterbodies and waterbody loss**

62 Only about 2% of urban area is devoted to waterbodies in cities on average out of a total of
63 2090.89 km². Zhaoqing city, Guangdong Province, has the highest fraction of waterbodies at 10.3%,
64 and 16 cities devote less than 0.1% of their area to waterbodies (Fig. S4a). This reveals the
65 significant regional differences in waterbody development in highly urbanized cities. From 1990 to
66 2018, these cities lost 464.3 km² of waterbodies, which is 20.6% of the original area devoted to
67 waterbodies (Table S1). According to the waterbody abundance in city, we classified the cities into
68 three categories from abundant to deficient, including Type-I, -II and -III, and more detail in
69 Material. This waterbody loss is mainly concentrated in Type-I and -II cities and accounts for 245.1
70 and 143.9 km² of waterbody loss. Type-III cities devote the least amount of surface area to
71 waterbodies (only 75.3 km² in total). In general, a city with abundant waterbodies will encroach
72 more on them as a result of its expansion. Furthermore, the occupation of waterbodies has increased
73 continuously from 1990 to 2010, following which it decreased drastically in urban areas. The period
74 2005–2010 saw the greatest loss of 183.7 km², of which over half occurred in Type-I cities.

75 Although the three types of cities differ significantly in terms of waterbody loss, these
76 differences are less significant in proportion to original waterbody. Significant areas of waterbodies
77 were preserved despite urban expansion, but the rate of waterbody loss rate increased before 2010,
78 particularly for Type-I and -II cities (Figs. 1b and 1c). The peak 5-year waterbody loss in all three
79 types of cities was from 2005 to 2010 (31.2%, 33.5%, and 30.3% for Type-I, -II, and -III cities,
80 respectively). In contrast, the 5-year waterbody loss of Type-I cities was only 12.1% from 1990 to
81 1995. A similar waterbody-loss trend occurred in Type-II cities. The rate of waterbody loss in Type-
82 III cities was much greater from 1990 to 2000 and fluctuated. After 2010, however, the rate of
83 waterbody loss for all three types of cities declined significantly.



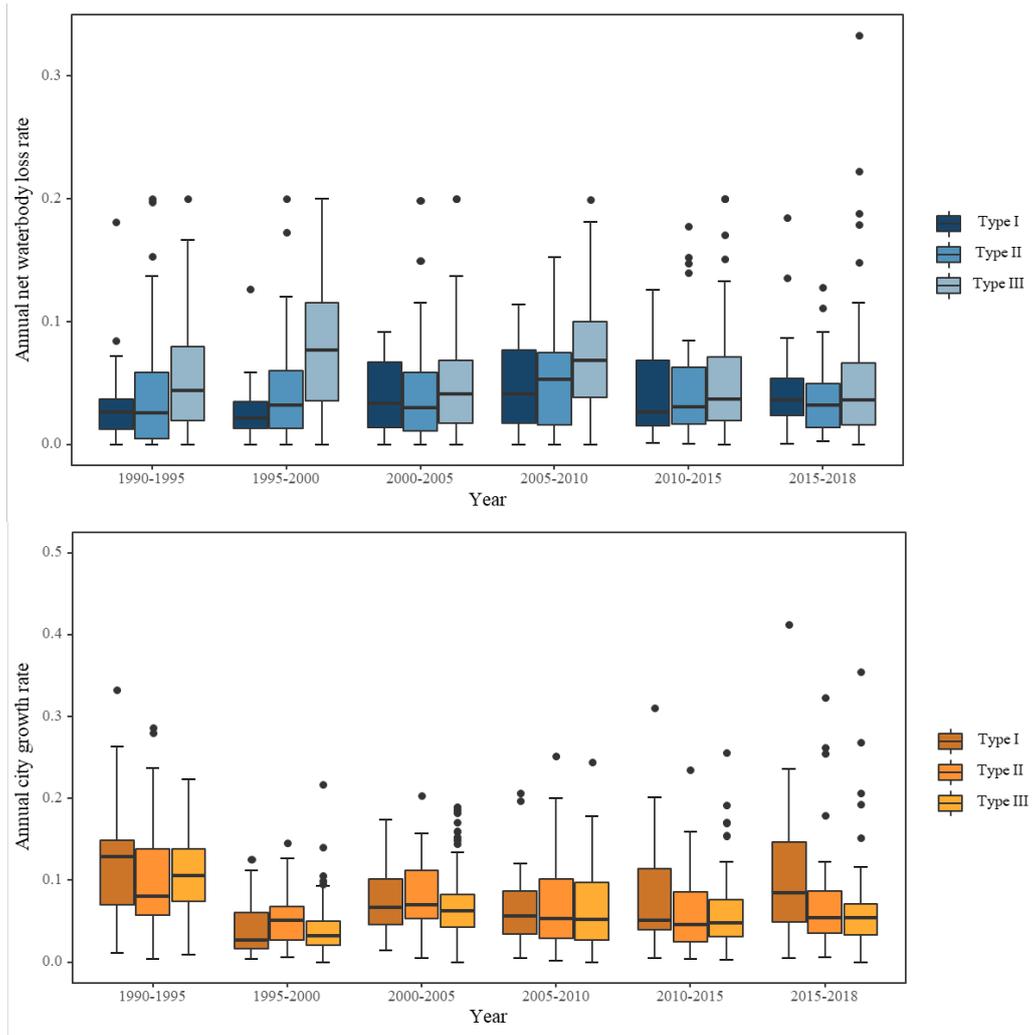
84

85 Fig. 1. Waterbody loss in urban expansion: (a) Spatial distribution of urban expansion in different

86 types of cities. (b) Spatial distribution of waterbody loss in different types of cities. (c)–(e)

87 Waterbody loss (WL) and waterbody preservation (WP) rate as a function of time in areas of

88 urban expansion in cities of Type I–III, respectively.



89

90 Fig. 2. Annual net waterbody loss rate and city growth rate as a function of time for cities of Type
 91 I–III.

92 **Waterbody encroachment due to urban expansion**

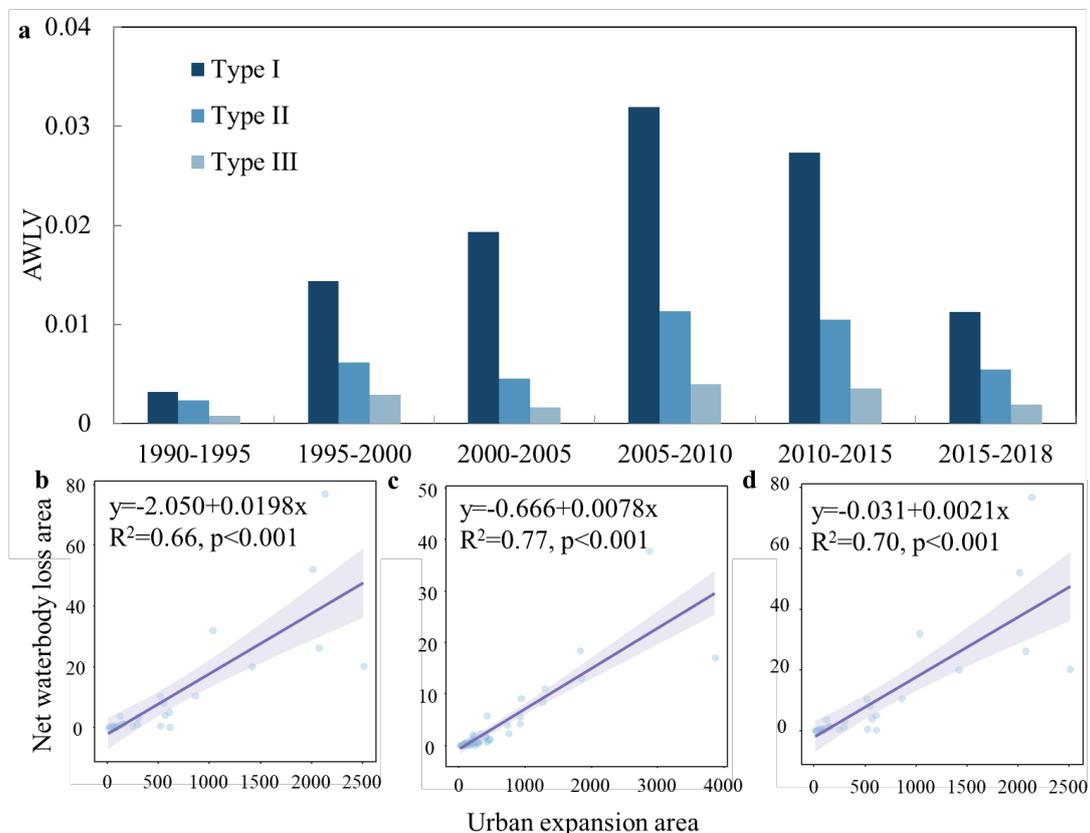
93 In Chinese cities of over one million in population, urban area increased from 13 194.8 km² in
 94 1990 to 81 622.1 km² in 2018, for an average expansion of 23.9 km² per year overall (Fig. S2). The
 95 overall rate of urban growth slowed steadily during this time, decreasing from 11.59% in 1990–
 96 1995 to 7.43% in 2015–2018 (Fig. 2). In addition, the growth rate was essentially the same for cities
 97 with different waterbody abundance. Cities that underwent significant urban expansion are
 98 concentrated mainly in the four urban agglomerations with high rates of economic development.
 99 These cities were large in 1990, and so underwent large-scale expansion; the expansion area reached
 100 1000 km² over the past 30 years. However, many cities in China achieved a high rate of urban
 101 expansion on a smaller scale (i.e. urban areas of fewer than 50 km² in 1990). For example, the built-
 102 up area of Putian City in Fujian Province increased nearly 67-fold from 5.92 km² in 1990 to almost
 103 400 km² in 2018 (Table S2).

104 The annual average net waterbody loss (ANWL) rate in every city varies significantly over
 105 time (Fig. 2b). The ANWL for each type of city is 3.9%, 4.2%, and 6.1%, respectively, and peaked
 106 at 6.0% in 2005–2010. Type-III cities underwent the largest variation in ANWL, mainly because of
 107 a lack of original waterbody. Mild waterbody loss can lead to dramatic changes. The ANWL rate

108 increased from 1990 to 2010, following which it decreased and then stabilized.

109 Variations in waterbody coverage may be affected by the rate of urban expansion. We illustrate
110 the actual waterbody coverage by using the waterbody loss per unit urban expansion area, which is
111 independent of city size and expansion rate. A Spearman correlation analysis shows that waterbody
112 loss correlates strongly with the rate of urban expansion (statistical significance p -value < 0.01) (see
113 Figs. 3b–3d). For Type-I cities, the average net waterbody loss velocity of urban expansion (AWLV)
114 was only 0.003 in 1990–1995, which means 0.003 km² of waterbodies were occupied when the city
115 expanded by 1 km² (Fig. 3a). However, the AWLV increased 9-fold from 2005 to 2010, which
116 represents a huge waterbody loss. Since 2015, the AWLV has decreased to 0.01, which is the 2000–
117 2005 level. The AWLV for the three types of cities increased first and then decreased and tended to
118 be less for cities with rich waterbody resources at the outset. Wuhan, known in China as the city of
119 a thousand lakes, has abundant waterbody resources, covering 139.3 km² and accounting for 8.4%
120 of the urban area. However, the waterbody area in Wuhan decreased from 1990 to 2018, with 28
121 km² of waterbody area being occupied due to urban expansion. The fraction of waterbody loss for
122 the period 2005–2015 even exceeds the fraction of waterbody retained (Table S1).

123 In contrast with Type-I cities, the AWLV for Type-II and -III cities decreased from 2000 to
124 2005. In general, the indices for Type-II and -III cities are much lower than that for Type-I cities,
125 which peaked at 0.011 and 0.003, respectively. The phenomenon of water encroachment rarely
126 occurred during urban expansion in water-deficient cities due to the small size of waterbodies (Fig.
127 6c). The velocity of waterbody loss for the three types of cities all increased first and then decreased.
128 Meanwhile, the rate of waterbody loss decreased significantly for cities with an initial lack of
129 waterbodies. Water-rich cities thus may have neglected to protect their waterbodies from rapid urban
130 expansion.

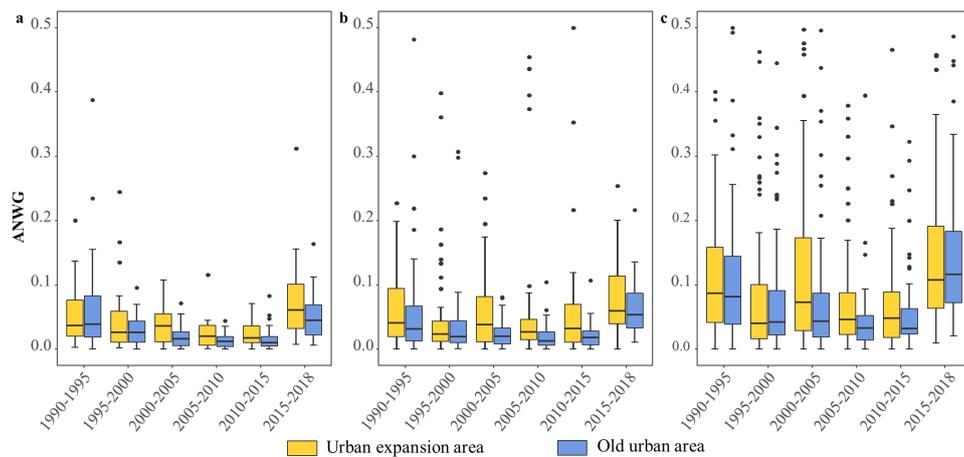


132 Fig. 3. Rate of waterbody loss as a function of urban expansion area. (a) Overall AWLV. (b)–(d)
 133 Spearman correlation analysis of waterbody loss as a function of city expansion area for cities of
 134 Type I–III, respectively.

135 **Overall change in urban waterbody area**

136 Up to 38.4% of waterbody area was encroached upon from 1995 to 2000, which represents a
 137 significant loss of waterbody area in water-deficient cities (Fig. 1c). Although waterbody area was
 138 significantly encroached upon due to urban expansion, it was also augmented in other ways. From
 139 1990 to 2018, 407.3 km² of waterbody area was added in urban expansion areas, of which 154.4,
 140 129.1, and 123.8 km² were added in cities of Type I–III, respectively. Type-III cities have a greater
 141 rate of annual average net waterbody gain (ANWG), with an average of 36.6%, because these cities
 142 lacked waterbodies at the outset (i.e. in 1990). Type-I and -II cities experienced a slightly greater
 143 rate of urban expansion (16.6% and 25.9%, respectively).

144 In addition to the urban expansion area, we also found that waterbody area was added in old
 145 urban areas (Fig. 4). In old urban areas, the ANWG index decreased gently from 1990 to 2010 and
 146 then increased from 2010 to 2018. In Type-I and -II cities, the ANWG of old urban areas is less than
 147 that of urban expansion areas, although this is not the case in Type-III cities, where waterbody
 148 replenishment in old urban areas outstripped even the urban expansion area. This demonstrates not
 149 only a large gain in waterbody area during urban expansion but also the importance of protecting
 150 and developing waterbodies in old urban areas.

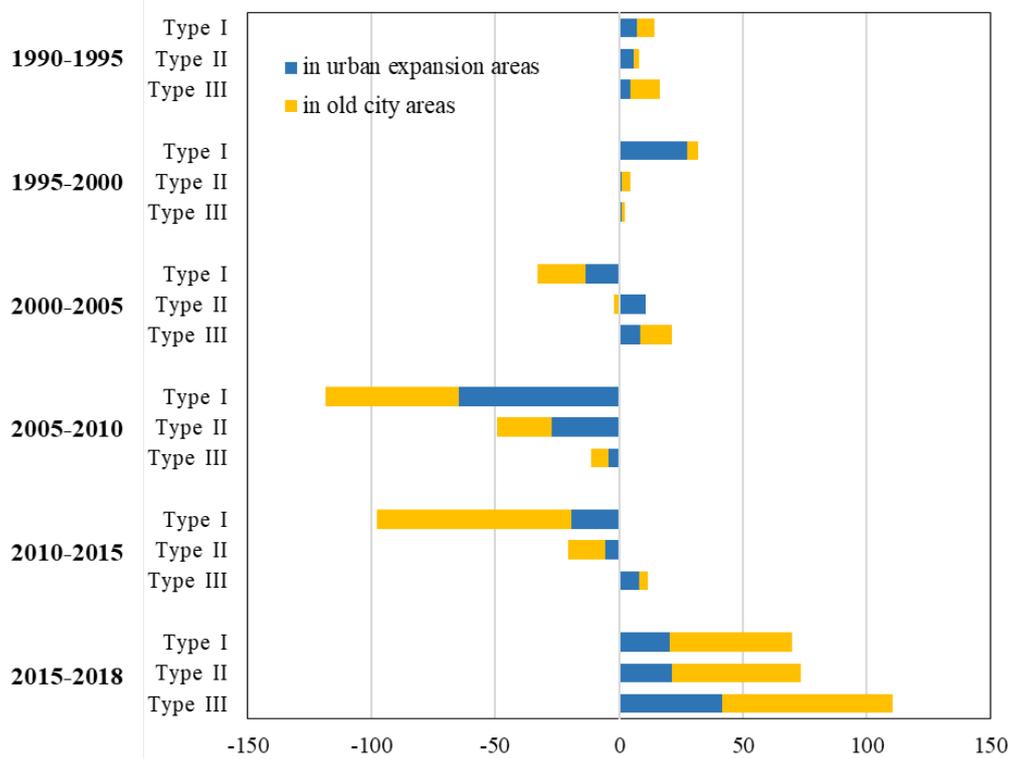


151
 152 Fig. 4. Box diagram showing ANWG index in urban expansion area and old urban areas for (a)–
 153 (c) cities of Type I–III, respectively. For visual, some ANWG values above 0.5 in Type-III cities
 154 are not shown in the figure.

155 The total waterbody changes in urban areas can be obtained by summing the urban expansion
 156 area and the old urban area at each stage. The result is that waterbodies in all cities mainly decreased
 157 from 2005 to 2015 and then increased from 2015 to 2018 (Fig. 5). The waterbody area decreased
 158 from 2000 to 2015 in Type-I cities, with gross waterbody loss reaching as high as 118.5 km² within
 159 city boundaries from 2005 to 2010. From 1990 to 2018, gross waterbody loss reached 133.5 km².
 160 In Type-II cities, the overall waterbody area remained constant from 1990 to 2005, with an
 161 occasional increase, until suffering a significant loss in 2005–2015. As a typical Type-II city,
 162 Changsha has relatively little waterbody with a waterbody area that accounts for 2% of its urban
 163 area. However, it suffers from negligible waterbody encroachment. The preserved waterbody area
 164 is 22.7 km², and the loss was only 3.8 km² due to urban expansion from 1990 to 2018, with the most

165 serious loss occurring in 2005–2010. From 1990 to 2018, the waterbody gain increased for Type-II
 166 cities, exceeding the loss by about 25.3 km² on average.

167 While in Type III cities, waterbody gain thus exceeds waterbody loss. Waterbody gain occurred
 168 in various periods, except for 2005–2010. In other words, although a net waterbody loss occurred
 169 overall, waterbody gain occurred in other areas of the cities. Zhengzhou is a typical water-deficient
 170 city, where the urban waterbody occupies only nine-thousandths of the city’s area. Therefore,
 171 waterbodies are rarely involved in urban expansion in Zhengzhou, and only 1.3 km² of waterbodies
 172 were encroached upon from 1990 to 2018. In contrast, up to 12.3 km² of new waterbodies were
 173 developed in the urban expansion. Thus, significant waterbody gain occurred in Zhengzhou from
 174 2010 to 2018, far exceeding the waterbody loss.



175
 176 Fig. 5. Urban waterbody variation for each period for the three city types.

177 **Discussion**

178 Massive waterbodies are occupied and converted to other land-use types in the process of urban
 179 expansion in China. In general, waterbody loss correlates positively with waterbody abundance. In
 180 addition, the waterbody distribution is not balanced between the city types, with Type-I and -II cities
 181 and their more abundant waterbodies losing significantly more waterbody due to urban expansion.
 182 Existing research focuses on the protection and management of mostly water-deficient regions.²⁹
 183 However, serious problems currently exist regarding waterbodies, requiring more attention to
 184 waterbodies all over the world. More than half of the world’s rivers or streams currently are non-
 185 perennial,³⁰ and the connectivity of rivers is also weakening, producing a risk of flow interruption.³¹
 186 Lakes are also shrinking worldwide.³² Although these changes are strongly affected by natural
 187 factors, urban expansion cannot be ignored.^{33,34} Human activities such as river diversion and
 188 occupation may affect the future spatial distribution of waterbodies.

189 Waterbody loss is related to the scale of urban expansion. The larger the scale of urban
 190 expansion, the more likely is waterbody encroachment. For instance, severe waterbody

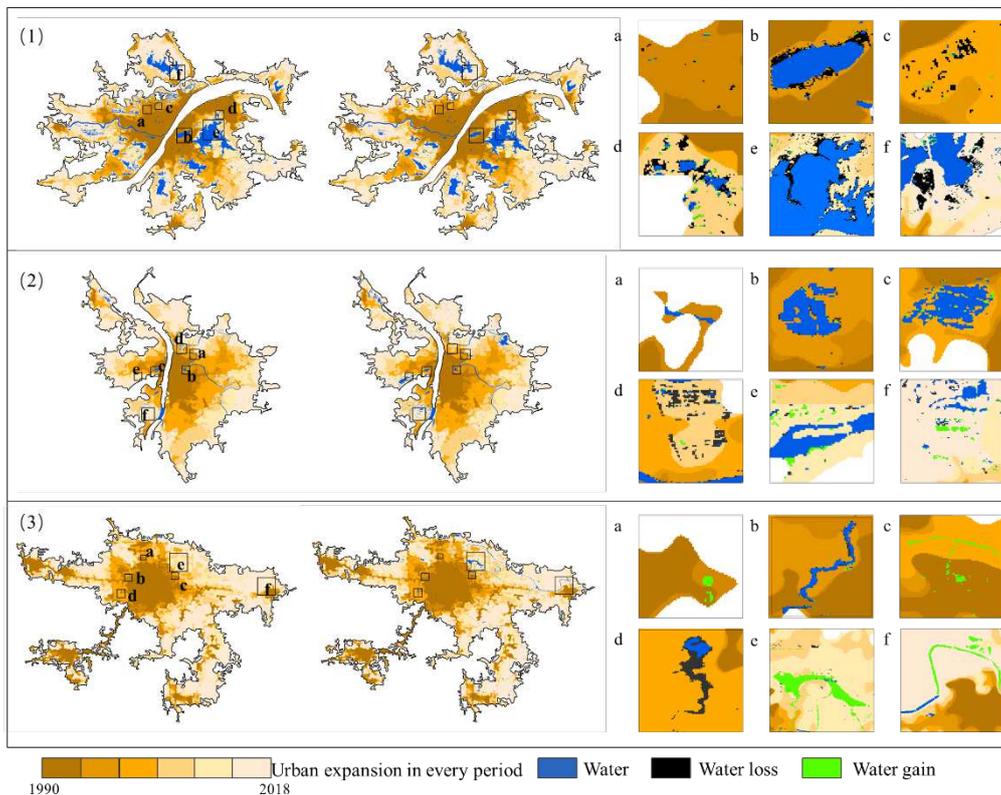
191 encroachment occurred in several urban agglomerations in China with rapid social and economic
192 development, mainly in the Beijing-Tianjin-Hebei Region, the Yangtze River Delta, the Wuhan
193 metropolitan area, and the Pearl River Delta, although these areas are not with the same extent of
194 urban waterbodies in China. Other investigators report similar results for nationwide wetland loss,
195 which includes lakes, rivers, and reservoirs.³⁵ The primary reason for these tragedies is the policies
196 and regulations introduced recently to promote urban development. Urban infrastructure and other
197 urban development projects support local urbanization but require land for construction, which leads
198 to encroachment on waterbodies.³⁶ In addition, the urban population has increased rapidly in China,
199 with the fraction of urban population increasing from 17.9% to 52.6% from 1978 to 2012,³⁷
200 increasing the demand for construction land. Given the visible benefits of land finance, the boom in
201 economic development zones and real estate has led to lake filling and reckless housing construction.
202 Finally, the increase in impervious surface area caused by urban expansion also negatively affects
203 waterbodies.³⁸

204 The overall changes in waterbody area over time for a large city clearly show that the concept
205 of water protection must be strengthened, despite the waterbody loss per unit urban expansion
206 suffered by all three city types decreasing since 2010 (Fig. 3). Moreover, the observed turning point
207 in waterbody loss (ca. 2010) demonstrates that the protection of waterbodies and the environment
208 in general has increased since the reckless era of economic development in China. Starting in the
209 1990s, the government has issued policies and regulations that strive to protect rivers, lakes, and
210 wetlands. For example, in 1988, the Regulations on River Management of China prohibited the
211 reclamation of land from lakes and rivers. In 2007, the “Blue Line” governance further emphasized
212 waterbody protection within city limits, restricting landfills in waterbodies during urban
213 development. Various provinces and cities also imposed special regulations for their regions. For
214 example, Jiangsu province implemented the Lake Protection Regulations in 2004, which
215 emphasized the protection of waterbodies. In 2006, Zhejiang Province promulgated measures for
216 the management of waterbodies involved in construction projects, which initiated water protection
217 in urban construction. Many regions then promulgated regulations to protect waterbodies from
218 urban expansion. In addition, to construct an eco-friendly city in the Xiong’an New Area, the plan
219 proposed 70% fraction of blue and green spaces and proposed strictly protecting waterbodies.³⁹ In
220 2017, the Ministry of Housing and Urban-Rural Development issued guidelines to strengthen
221 ecological restoration of urban areas, stressing the concept of “sponge-city” construction and the
222 ecological restoration of rivers, lakes, wetlands, and other waterbodies, and two ecological-
223 restoration projects and urban restoration were undertaken to protect the natural urban water
224 system.⁴⁰ To a certain extent, these measures reflect the change in attitudes towards waterbodies,
225 gradually shifting from economically oriented rapid expansion to the joint development of urban
226 economy and ecology, thereby enhancing the sustainability of the city.

227 Many scholars predict that cities will continue to expand in the future.^{41,42} Cities will also face
228 increasing uncertainties and risks, including climate change and natural disasters. Recently, serious
229 urban waterlogging disasters related to the dramatic urban expansion in China have occurred with
230 increasing frequency in many large- and medium-sized cities, causing incalculable losses to society.
231 Although waterbodies generally occupy only a small part of a city, they sustain the urban ecological
232 environment and regulate runoff within the urban area.⁴³ It is thus urgent to focus more attention on
233 coordinating urban development of construction land with ecological spaces. It is clear now that
234 natural resources must be protected to ensure the sustainable development of cities and to construct

235 ecological cities.⁴⁴

236 Waterbodies in Type-III cities have generally increased in scope during the study period (1990–
237 2018; see Figs. 5 and 6). For example, Changsha constructed artificial lakes and created water
238 landscapes to sustain the ecology of the city and to provide new avenues of local tourism. People
239 who propose ecological cities believe that the inefficient use of resources degrades urban
240 development.⁴⁵ The quantitative research presented herein reveals the shift in China’s development
241 strategy, which is crucial to the future ecological development of cities. This strategy holds that
242 urban planning must now consider the ecological function of waterbodies, which includes flood
243 control, drainage, water storage, and biodiversity sustenance. In addition, the importance of urban
244 blue and green infrastructure should be paramount, as dictated by science for the construction of
245 sustainable cities.⁴⁶



246

247 Fig. 6. Variations in urban waterbodies of every period in typical cities: (1) Wuhan, (2) Changsha,
248 and (2) Zhengzhou.

249 This analysis provides a valuable argument for the protection of surface waterbodies in
250 urbanization; however, a challenge remains to evaluate encroached waterbodies because of the
251 difficulty of eliminating the influence of natural conditions such as climate on waterbodies. To
252 negate this effect, we used herein the 3-year waterbody frequency index to describe surface
253 waterbodies. Limited by the data on waterbodies with 30 m resolution, waterbodies obscured by
254 trees are difficult to identify, making it difficult to detect all waterbodies.

255

256 **Methods**

257 This study quantitatively assesses waterbody loss due to urban expansion of large Chinese
258 cities. We first extracted multi-temporal urban boundaries to determine the expansion of cities of
259 over one million in population from 1990 to 2018. The monthly surface-water dataset was then used

260 to identify surface waterbodies in the study period. Depending on the ratio of surface waterbody
261 area to urban area, cities were further divided into three categories (i.e. water-abundant, water-
262 medium, water-deficient). Finally, we quantified the rate of waterbody loss and evaluated the spatial
263 and temporal variation of waterbody loss as a function of urban expansion and according to city
264 type.

265 **GUB dataset**

266 The Global Urban Boundary (GUB) dataset (<http://data.ess.tsinghua.edu.cn>) was used to
267 determine urban expansion. GUB provides data on built-up areas over 30 years, with a spatial
268 resolution of 30 m. In the GUB dataset, nonurban areas (such as green space and water space)
269 surrounded by artificial impervious areas are filled within the urban boundary and removed by the
270 algorithm, which is consistent with global mapping methods. The continuous urban boundary was
271 demarcated by morphological image processing methods, which have an overall accuracy of over
272 90%. In this dataset, extensive water and forests are excluded, and the impervious surface within
273 the urban boundaries accounts for about 60% of the total surface area.⁴⁷ Compared with urban
274 boundaries obtained from night-time light, GUB better separates urban areas from surrounding
275 nonurban areas.

276 **Monthly waterbody dataset**

277 We selected the JRC Monthly Water History V1.3 dataset([https://global-surface-
278 water.appspot.com/](https://global-surface-water.appspot.com/)), which is available from the Google Earth Engine, as the basis for representing
279 surface waterbodies.⁴⁸ This data collection, which was produced by using images from the Landsat
280 series, contains 442 images of global monthly waterbody area from March 1984 to December 2020.
281 In this dataset, the validation confirmed that fewer than 1% of waterbodies were incorrectly detected,
282 and fewer than 5% of waterbodies were missed altogether. We chose this dataset due to the long-
283 term spatial distribution of waterbodies and due to mountain shadows and urban-constructions
284 masking, which reflects the real changes in waterbodies.

285 **Extracting the extent of large Chinese cities from GUB dataset**

286 Cities have high concentrations of population and resources and expand spatially during their
287 development. GUB data are selected as the original data for urban boundary selection to characterize
288 urban expansion. In our study, cities were defined as municipal districts excluding the extensive
289 countryside within the administrative boundaries of prefecture-level cities. We identified urban
290 areas based on the physical boundaries from the perspective of remote sensing, which can precisely
291 track urban expansion.⁴⁹

292 In this work, we selected 159 cities with a population of over one million in 2018 based on the
293 average annual population of urban districts from the *2019 China City Statistical Yearbook* (Fig.
294 S1). Taiwan, Hong Kong, and Macau are not included. According to statistics, China had 160 cities
295 with populations exceeding one million in 2018. However, due to the lack of data for the built-up
296 area of Guang'an in 1990, Guang'an was not included in the study. We thus obtained 159 cities from
297 the GUB dataset. Due to numerous fragmented patches within the administrative boundary, the main
298 urban areas were identified by the population, and max patch areas were comprehensively based on
299 the urban boundaries. Through manual detection and adjustment of the map, we determined that the
300 location of the extracted urban area was consistent with that of the municipal government, and the
301 boundary was extracted for each time period. We took the growth area as the expansion area, with

302 the original area being the city at the onset of each time period (Fig. S3).

303 We used the average annual urban growth (AUG) rate to characterize the rate of urban
304 expansion, as is widely done to evaluate urban expansion.^{50,51} It is calculated as

$$305 \quad \text{AUG} = \left[\frac{\text{Land}_{t_1}^{\frac{1}{t_1-t_0}}}{\text{Land}_{t_0}} - 1 \right] \times 100\%,$$

306 where Land_{t_0} and Land_{t_1} represent the urban land area at time t_0 and t_1 , where t_0 and t_1 are the start
307 and end of the given study period.

308 **Identification of urban waterbodies**

309 Urban waterbodies contain all the components of urban flow networks above the ground and
310 include natural waterbodies such as lakes, rivers, streams, and wetlands and artificial waterbodies
311 such as parks and ponds.⁴⁸ Considering the dryness or wetness of each year, we used the data for 3
312 years (36 months) around each time period (1990, 1995, 2000, 2005, 2010, 2015, and 2018) to
313 describe the waterbody. Not all waterbodies could be detected for each month of the year; for
314 example, freezing may prevent waterbodies from being detected. To cover seasonal and permanent
315 waterbodies, we used the waterbody frequency index (WFI), which is calculated as the fraction of
316 waterbody months within the 3 years to identify stable waterbodies pixel by pixel. The spatial
317 distribution of each waterbody was then mapped comprehensively for each time period. By
318 comparing the extracted waterbody with the long-time-series high-resolution remote-sensing
319 images from Google Earth, we found that the extracted waterbodies fit the actual waterbody
320 distribution quite well (Fig. S2):

$$321 \quad \text{WFI}(i) = \frac{\text{WM}(i)}{\text{DM}(i)}$$

322 where $\text{WFI}(i)$ is the water occurrence for pixel i in the images before and after the given year, and
323 i is the pixel number for the study area. $\text{WM}(i)$ is the number of months during which the waterbody
324 is detected in i pixel over the 3 years. $\text{DM}(i)$ is the number of months during which the data are
325 available in pixel i . If the waterbody frequency index of a pixel is greater than 25%, this pixel is
326 considered as a waterbody; otherwise, it is not.

327 **City classification based on waterbody surface**

328 Cities with over one million in population may not be short of waterbodies, but significant
329 differences remain in surface waterbody abundance. Due to large differences in city size, it is
330 inappropriate to use waterbody area as a criterion. Considering the influence of urban expansion,
331 we rank 159 cities according to the fraction of the baseline surface water within the urban boundary
332 in 2018, which uses the average surface water from 1985 to 1991 as baseline. We used the natural
333 break method to divide cities into abundant, moderate, and deficient levels (referred to as Type I,
334 Type II, and Type III, respectively) and evaluate the abundance of waterbodies in cities.

335 **Temporal characteristic of waterbody loss and gain**

336 To understand the spatial-temporal features of surface waterbodies, we use five normalized
337 indicators to compare waterbody variations between cities during urban expansion from the overall
338 perspective and from the city perspective.

339 The variation in original natural waterbodies reflects the intensity of the natural resource
340 development in urban expansion. We summarize the reduction and preservation of original

341 waterbodies in urban expansion areas with a population of over one million to represent the
342 encroachment of urban expansion on waterbodies:

$$343 \quad WL = \frac{\sum NWL_{t_0,t_1}(i)}{\sum W_{t_0}(i)} \times 100\%$$

$$344 \quad WP = \frac{\sum (W_{t_0}(i) - NWL_{t_0,t_1}(i))}{\sum W_{t_0}(i)} \times 100\%$$

345 where i labels the city within the 159 cities, WL and WP are the fractions of waterbody loss and
346 preservation in urban expansion areas of all cities, NWL_{t_0,t_1} is the net waterbody loss during period
347 t_0-t_1 , and W_{t_0} is the natural waterbody in the urban expansion area at time t_0 .

348 To estimate the net waterbody loss caused by urban expansion at various stages, we used the
349 standardized indicator, annual average net waterbody loss rate (ANWL), to compare waterbody loss
350 speeds over time. This indicator is independent of the difference in waterbody abundance and can
351 be compared over time. Waterbody loss is one part of the impact of urbanization; the other is
352 waterbody gain. We used the same method to evaluate the annual average net waterbody gain rate
353 (ANWG). The formulas are

$$354 \quad ANWL = \frac{NWL_{t_0,t_1}}{W_{t_0}(t_1 - t_0)} \times 100\%$$

$$355 \quad ANWG = \frac{NWG_{t_0-t_1}}{W_{t_0}(t_1 - t_0)} \times 100\%$$

356 where NWL and NWG are the net waterbody loss and gain, respectively, and the other abbreviations
357 are the same as above.

358 Considering the direct impact of urban expansion, we used a normalized indicator, the average
359 net waterbody loss velocity of urban expansion (AWLV), which refers to the amount of waterbody
360 encroachment per unit urban expansion area. It quantifies the time-heterogeneity of waterbody loss
361 due to urban expansion and is calculated as follows:

$$362 \quad AWL = \frac{NWL_{t_0,t_1}}{Land_{t_1} - Land_{t_0}}$$

363 We calculated these indicators for the six expansion periods (1990–1995, 1995–2000, 2000–
364 2005, 2005–2010, 2010–2015, and 2015–2018) (Fig. 3). In the study, if the waterbody count is zero
365 at the onset of the period, the indicator for the period is abnormal and thus excluded.

366

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