

Typical artificial forest soil bacterial community response to the soil environment of Grain for Green area in the Loess Plateau, China

Xiaohua Liu

Beijing Forestry University

Tianxing Wei (✉ weitx@bjfu.edu.cn)

Beijing Forestry University

Research Article

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1 **Typical artificial forest soil bacterial community response to the soil** 2 **environment of Grain for Green area in the Loess Plateau, China**

3 Xiaohua Liu, Tianxing Wei*

4 School of Soil and Water Conservation, Beijing Forestry University, Beijing, 100083, PR China

5 Chinese National Forestry Ecosystem Observation and Research Station, CNERN, Ji Country, Shanxi, 6042200, PR China

6 **Abstract**

7 **Purpose:** This study aims to analyze the differences of soil bacterial community structure under
8 different vegetation restoration types, and to determine feasible strategies for revegetation of grain
9 for green area in the Loess Plateau.

10 **Methods:** High-throughput sequencing technology was used to analyze the alpha diversity of soil
11 bacteria, community structure characteristics and the correlation between soil environmental factors
12 and bacterial communities in different vegetation restoration types.

13 **Results:** The findings reveal that soil microbial C and N in different soil layers shows a decreasing
14 trend with an increase in soil layers. Across all samples, the dominant groups are Proteobacteria,
15 Acidobacteria, Chloroflexi, Gemmatimonadetes, Nitrospirae, Bacteroidetes, GAL15, and
16 Firmicutes. Soil nutrient supply, such as TOC, TN, AN, AP, and AK, had crucial roles in shaping
17 the composition and diversity of the bacterial communities.

18 **Conclusions:** Based on the results, our data provides a starting point for establishing soil bacterial
19 databases in the Loess Plateau, as well as for the plants associated with the vegetation restoration.

20 **Keywords:** Revegetation; High-throughput sequencing; Soil bacterial community; Soil
21 environment factor; The Loess Plateau

22 **1 Introduction**

23 Ecological restoration is a complex process that involves improvement of green vegetation
24 coverage, prevention of soil erosion, and reduction of natural disasters. Currently, revegetation is a
25 hot topic of research in ecology, encompassing soil physical and chemical properties and soil and
26 water conservation (Zeng et al. 2014), ecological stoichiometric characteristics of vegetation and
27 soil (Sun et al. 2020), soil nutrient element preservation (Ma et al. 2014), biodiversity of
28 undergrowth vegetation (Chen and Cao 2014), and soil microbial communities (Liu et al. 2018). As

* Corresponding author.

E-mail address: weitx@bjfu.edu.cn (Tianxing Wei)

29 the agents of decomposition in the ecosystem, soil microorganisms play a key role in forest
30 succession and nutrient cycling. They are the link between the aboveground and underground
31 components of the ecosystem and represent an important index for evaluation of ecosystem
32 restoration. Plant community in forest ecosystems by changing the species composition, biomass,
33 litter–earth system components, forest and soil environment factors, such as soil microbial
34 community structure and composition (Okubo et al. 2016a; Spohn and Widdig 2017a), and other
35 relevant studies have reported the species and soil physical and chemical properties of soil microbial
36 community composition and diversity (Ding et al. 2017; Nacke et al. 2016; Pei et al. 2016; Zhalnina
37 et al. 2015).

38 It is a long-term and arduous task to harness the ecological environment and restore the
39 vegetation in the Loess Plateau. Revegetation is not only the first step in ecosystem reconstruction
40 but also an important measure of ecological environment management and soil and water
41 conservation on the Loess Plateau, and it can effectively reduce soil erosion and improve soil quality
42 (You et al. 2020; Tan et al. 2019). Wuqi County in Shaanxi Province is located in an interlaced area
43 of wind erosion and water erosion on the Loess Plateau. In that area, the soil erosion is serious, and
44 the ecological environment is fragile. Since 1998, the government has adopted a development
45 strategy based on “closing hills and returning farmland, afforestation to grass, indoor-sheep
46 production, and planting trees to poverty eradication.” The change in vegetation development and
47 species composition in Wuqi County following the closure reflects the trend of increasing stability
48 of the degraded matrix and enhancement of the community environment (Shan et al. 2019).

49 In recent years, the revegetation project on the Loess Plateau has increased the area of
50 vegetation considerably, and an increasing number of studies have examined the revegetation of the
51 returned farmland in this area (Chen et al. 2015; Feng et al. 2016; Nie et al. 2015). In general, these
52 studies focus on the impacts of revegetation on soil properties, changes of species diversity, the
53 benefits of soil and water conservation, and the allocation mode of revegetation in the process of
54 land restoration. However, there has been no systematic analysis of the interaction between
55 revegetation, soils, and microorganisms in the process of land restoration.

56 With the development of soil microbial research methods, high-throughput sequencing
57 technology, also known as either "next-generation sequencing technology" or deep sequencing, has
58 gradually become a conventional method for analyzing the diversity of soil microorganisms.

59 Currently, this method is used widely to determine the effects of vegetation, crop cultivation and
60 land management, organic pollutants, heavy metals, seasonal changes, climatic conditions, and
61 other factors on soil microbial community structure and biomass (Ren et al. 2018; Zeng et al. 2016;
62 Zhang et al. 2018).

63 Therefore, to explore the role of microorganisms in the process of vegetation restoration on the
64 soil ecosystem of Grain for Green Project area in the Loess Plateau and to determine feasible
65 strategies for revegetation in this area, the present study investigated the dynamics of soil properties
66 and the structure of soil bacterial communities at sites on the Grain for Green are in the Loess Plateau
67 using 16S rRNA high-throughput sequencing.

68 **2 Materials and methods**

69 **2.1 Study site and sampling**

70 The vegetation types in the area are characterized by the transition from forest shrub steppe
71 vegetation to steppe forest shrub steppe vegetation. The vegetation composition is dominated by the
72 flora of North China. Since the project of returning farmland to forestry was implemented, there
73 vegetation coverage has increased significantly. Different artificial *H. rhamnoides* forests as the
74 research objects were selected. Typical sample plots were were selected in stands and 20 m × 20 m
75 arbor sample plots were set up to observe and record the position (coordinate information, altitude)
76 and site condition (such as slope and canopy closure) of the sample site in August 2017. Table 1
77 shows the specific situation. At each sampling site, soil was collected from three locations ca. 20m
78 apart from one another by pooling five soil cores from 1m×1m area at each location. The surface
79 soil samples were collected from each quadrat at random and mixed immediately to form a single
80 soil sample. Visible matter, such as loose gravel and plant debris, was removed and divided into two
81 parts. One part was passed through a 2 mm sieve into a 15 mL centrifuge tube, placed into a liquid
82 N canister, transferred to the laboratory, and stored in a freezer at −80 °C for microbial analyses.
83 The other part was subjected to air-drying treatment and used to determine soil nutrients.

84 **2.2 Soil sample determination**

85 **2.2.1 Sequencing of soil microbial genes**

86 Extraction of DNA: Following the manufacturer's instructions, the genomic DNA was isolated
87 from 0.5 g of each soil sample from each sample site using the FastDNA™ SPIN Kit for Soil (MP
88 Biomedicals, CA, USA). Extracts of three technical repeats were mixed into a single DNA sample.
89 Extracted genomic DNA was detected by 1% agarose gel electrophoresis. DNA concentration and
90 purity were measured with an ultramicro spectrophotometer (Nano Drop2000, US).

91 PCR amplification: Based on previous reports, the primers 338F-806R for the 16S rRNA of
92 bacteria were used. Amplified products were detected by 2% agarose gel electrophoresis, recovered
93 from the gel using the AxyPrep DNA gel extraction kit (Axygen Biosciences, Union City, CA,
94 USA), washed with Tris-HCl, and verified by 2% agarose gel electrophoresis.

95 PCR products were quantified by the QuantiFluor™-ST Fluorometer (Promega Biotech,
96 Beijing, China) and the samples adjusted as required for sequencing. High-throughput sequencing
97 was performed using Illumina Miseq PE300 platform (Shanghai Majorbio Bio-Pharm Technology
98 Co., Ltd, China).

99 **2.2.2 Soil microbial carbon and nitrogen determination**

100 Soil microbial biomass C and N were determined using fumigation extraction (Vance et al.,
101 1987). Briefly, 25 g of the oven-dry equivalent of field-moist soil was fumigated at 25.8 °C for 24 h
102 with CHCl₃. The soil was added to 100 ml of 0.5 M potassium sulfate by shaking at 200 rpm for 1
103 h and then filtered after fumigant removal. Another 25 g of non-fumigated soil was simultaneously
104 extracted. The TOC and TN contents of the extracts were determined using a TOC analyzer (multi
105 N/C 3100, TOC/TN analyzer, Analytik-Jena AG, Germany).

106 **2.2.3 Determination of soil physical and chemical properties**

107 The content of soil organic carbon (TOC) , total nitrogen (TN) , total phosphorus (TP), and
108 total kalium(TK) were determined using the vario EL cube CHNOS Elemental Analyzer(Elementar
109 Analysensysteme GmbH, Germany). Soil alkali-hydro N (AN) were measured using the alkali-
110 hydrolyzed diffusion method, and available phosphorus (AP) was determined by UV-Visible
111 Spectrophotometer (UV-2550, Shimadzu, Japan). And available kalium were determined by iCAP
112 6300 ICP-OES Spectrometer(Thermo Fisher, USA).

113 **2.3 Data Processing and Analysis**

114 2.3.1 Basic data processing

115 All data were analyzed using IBM SPSS statistics, R and the free online platform of Majorbio
116 Cloud Platform (www.majorbio.com).

117 2.3.2 Operational Taxonomic Units (OTU) Analysis

118 To obtain the species classification information of each OTU, RDP classifier was used to
119 analyze the representative sequences of OTU with 97% similarity level, and the species composition
120 of each sample was counted in each taxonomic level: domain, Kingdom, phylum, class, order, family,
121 genus, and species. The taxonomic identity of each phylotype was determined by the bacterial 16S
122 rRNA Silva reference database (<http://www.arb-silva.de>).

123 2.3.3 Alpha diversity analysis and Rarefaction curve

124 Indices of community richness Chao 1 estimator, abundance-based coverage estimator (ACE)
125 and indices of community diversity Shannon, Simpson estimator were calculated. And the coverage
126 index reflects whether the sequencing results represent the real situation of microorganisms in the
127 sample. A certain number of sequences were randomly selected from the samples, and the alpha
128 diversity index of the corresponding samples was counted. Taking the extracted data as the abscissa
129 and the alpha diversity index as the ordinate, the curve was drawn, and whether the sequencing data
130 volume was enough was judged according to whether the curve was smooth or not. In this study,
131 OTUs with 97% similarity were selected to calculate the alpha diversity index under different
132 random sampling by MOTHUR (<http://www.mothur.org/>), and the curve was made by using R
133 language tool.

134 2.3.4 Statistic analysis

135 Principal coordinates analysis (PCoA) was used to evaluate the overall differences in the
136 structures of the plant and microbial communities based on Bray-Curtis distances. Redundancy
137 analysis (RDA) using Monte Carlo permutation (999 repetitions) was used to test the relationships
138 among the soil properties. Correlations between the soil environment factor and composition of the
139 soil bacterial communities were investigated using Pearson's correlation analysis. Heatmap adopts
140 the Average level clustering method.

141 3 Results

142 3.1 Analysis of soil microbial carbon and nitrogen

143 **3.1.1 Characteristics of soil microbial carbon and nitrogen**

144 As shown in Table 2, what was revealed is that soil microbial C and N content in different soil
145 layers showed a trend of small change with increasing soil layers. That is, soil microbial activities
146 weaken gradually as the soil depth increased. There is a slight increase in soil microbial C content
147 in the 60–80 cm soil layer of *H. rhamnoides* pure forest (shady slope), the 80–100cm soil layer of
148 *H. rhamnoides* pure forest (sunny slope), the 80–100 cm soil layer of *H. rhamnoides* and Chinese
149 arborvitae mixed forest and the 60–80 cm soil layer of *H. rhamnoides* and *Pinus tabulaeformis*
150 mixed forest. There is no clear vertical variation of soil microbial N content, and there are no
151 significant differences between different soil layers. Over the four plots, the soil microbial C content
152 was highest in *H.rhamnoides* and *Pinus tabulaeformis* mixed forest, whereas there are few
153 differences in the soil microbial C contents of the other three plots. Soil microbial N content is
154 highest in *H. rhamnoides* pure forest (sunny slope), and the lowest value is in *H. rhamnoides* and *P.*
155 *tabulaeformis* mixed forest.

156 **3.1.2 Regression analysis of soil microbial carbon and nitrogen contents and soil** 157 **carbon and nitrogen contents**

158 Regression analysis, as a method for statistical analysis of the inter-dependent quantitative
159 relationship between two or more variables, can be used to analyze the relationship between one
160 variable and the change of another variable. The soil microbial C and N content and the soil C and
161 N content in different *H. rhamnoides* forests were analyzed via exponential and linear regression
162 analysis and established a regression model. Table 3 shows the results. Apart from the *H.*
163 *rhamnoides* and *P. tabulaeformis* mixed forest, the regression equations of the other three samples
164 are well fitted. Of the three well-fitting plots, the R^2 of the index model of soil microbial biomass N
165 and alkali-hydrolysable N of *H. rhamnoides* pure forest (sunny slope) is 0.73, whereas the remainder
166 are all above 0.8, with some even exceeding 0.9. Therefore, the model can well explain the
167 relationship between soil microbial C content and soil organic C content, soil microbial N content
168 and soil total N content, and soil microbial N content and alkali-hydrolyzed N content. Only the soil
169 microbial N content and soil alkali-hydrolyzed N content regression model shows a good fit in the
170 case of *H. rhamnoides* and *P. tabulaeformis* mixed forest.

171 **3.2 High throughput sequencing results**

172 By sequencing the V4-V5 region of bacterial genes, all samples were subsampled according to

173 the minimum number of sample sequences. And the OTU with sequence number ≥ 5 in at least one
174 sample is reserved. A total of 299,625 effective sequences were obtained from soil bacterial
175 community, with an average length of 438.21 bp. RDP classifier was used to cluster OTU
176 representative sequences with 97% similarity level. And the results showed that the bacterial
177 communities of the samples consisted of 24 phyla, 55 classes, 110 orders, 206 families, 348 genera,
178 680 species and 1989 OTUs. Combining OTUs of the same species and statistically analyzing the
179 changes of species composition among different samples enables the diversity of microflora among
180 different samples to be understood.

181 Table 4 shows the sequence number of test samples and alpha diversity index of microbial
182 community. The coverage of each sample is greater than 98%, which indicates that the probability
183 of gene sequence detection in soil samples is very high, and the sampling is basically reasonable,
184 which can truly reflect the microbial community characteristics of soil samples. According to the
185 rarefaction curves of each diversity index (Fig. 1), with the increase of random sequencing data, the
186 dilution curve basically tends to be flat, which indicates that the confidence degree of microbial
187 community in real environment is high, and the sequencing depth and data volume are enough.
188 What's more, the richness indices and diversity indices of bacterial community in arbor shrub mixed
189 forest are higher than that in shrub pure forest, and the indices of shrub forest on sunny slope are
190 higher than those on shady slope (Table 4).

191 **3.3 Soil bacterial community structure**

192 The dominant phylum in bacterial community is Actinobacteria (37.27% on average), followed
193 by Proteobacteria (23.91%), Acidobacteria (12.75%), Chloroflexi (12.27%), Gemmatimonadetes
194 (5.85%), Nitrospirae (1.66%), Bacteroidetes (1.26%), GAL15 (1.14%), and Firmicutes (1.1%) (Fig.
195 2). There are differences in relative abundance of dominant phylum between the pure and mixed
196 artificial forests, but the differences are not significant. Among the Proteobacteria, Alpha-, Beta-,
197 Gamma-, and Deltaproteobacteria are found in all samples (Fig.3). The members of
198 Alphaproteobacteria dominating this phylum, occupies 14.67% of all populations, highest at GL
199 compared to the others. Betaproteobacteria, Deltaproteobacteria, and Gammaproteobacteria
200 comprises 4.30%, 2.75%, and 2.19% of the total populations, respectively.

201 The plot of PCoA (Fig. 4) clearly identifies variations in bacterial community composition
202 among the sites, with the abundance of each group varying with different afforestation modes. At

203 the OTU level, the interpretation variance of PC1 and PC2 is 24.46% and 19.37%, respectively, and
204 the accumulated interpretation ability is 43.83%. The profiles at HrN and HrBo sites tend to group
205 together and are clearly separated from those at GL, HrS, and HrPt sites, which are clearly separated
206 from each other.

207 **3.4 Relationships among soil environment factor and bacterial communities**

208 The Pearson correlation coefficient were used to analyze the correlation between bacterial
209 communities and soil environmental factors. The results of the analysis are illustrated by a heatmap
210 in Fig. 5. TOC, AN, AP, TN, and AK contents are correlated positively with the abundance of
211 Alphaproteobacteria and Acidobacteria and negatively with the abundance of Gemmatimonadetes
212 and Bateproteobacteria. AP and AK are correlated negatively with the relative abundance of
213 Nitrospira.

214 The RDA confirmed the results of the correlation analysis and further identified the effect of
215 the soil on the bacterial communities at the class level(Fig. 6). The first two axes explain 87.77% of
216 the total variance, indicating that TOC, TN, AN, AP, and AK were the most influential factors
217 driving the changes in the composition and diversity of the bacterial communities.

218 **4 Discussion**

219 Restoration mainly affects the quantity and community of soil microorganisms by affecting the
220 litter, root morphology and secretion, and the material and energy transformation process in the
221 whole ecosystem(Okubo et al. 2016; Spohn and Widdig 2017). *Hippophae rhamnoides*, which is a
222 pioneer and associated species of vegetation structure in the Loess Plateau, has clear ecological and
223 economic benefits. Its root system can co-exist with nitrogen (N)-fixing actinomycetes to form N-
224 fixing nodules and fix free N. Furthermore, this root system can decompose insoluble organic matter
225 and minerals, thereby helping to improve the soil (Hu et al. 2014; Hu 2012).

226 Based on analysis of microbial biomass C and N in *H. rhamnoides* pure forest (shady slope),
227 *H. rhamnoides* pure forest (sunny slope), *H. rhamnoides* and Chinese arborvitae mixed forest, and
228 *H. rhamnoides* and *Pinus tabulaeformis* mixed forest, the driving capacity of the soil carbon (C)
229 cycle and the mineralization and fixation of soil microorganisms on nitrogen (N) were examined.
230 The soil microbial C and N analysis showed that soil microbial biomass N and N in different forest
231 layers exhibited a trend of a small change with increasing soil layers; that is, as the soil depth

232 increased, the soil microbial activity weakened gradually. Yao et al.(2016) studied the variation
233 rules of soil microbial biomass carbon in different vegetation restoration types in the Loess Plateau,
234 and the results showed that soil microbial biomass carbon content in mixed forest of trees was the
235 largest. Hu et al. (2010) studied the influence of vegetation patterns on soil microbial biomass and
236 functional diversity in a hilly area of the Loess Plateau. They also showed that, in terms of increasing
237 soil microbial biomass carbon and nitrogen content, the mixed mode of forest and grass was more
238 effective than the single vegetation and could effectively promote soil remediation. Therefore, it is
239 suggested that the mixed forest should be the main way to implement the Grain for Green Project
240 in the Loess Plateau.

241 Studies have shown that Actinobacteria is the main functional bacterium that degrades lignin
242 and cellulose (Pankratov et al. 2011). Alpha-, Beta-, Gamma-, and Deltaproteobacteria are the most
243 important microorganisms in saline alkali soil (Lin et al. 2019). In this study, we found that the
244 relative abundance of Actinobacteria and Proteobacteria was 37.23% and 23.91%, respectively,
245 which was the most dominant bacteria in the community, which was consistent with the research
246 results of Liu et al. (2016) on the characteristics of soil bacterial community in different arbor forests
247 in the Loess Plateau, and the vegetation type was an important factor affecting the diversity of soil
248 actinobacteria (Zhang et al. 2015). TOC, TN, AN, AP, and AK were the most influential factors
249 driving the changes in the composition and diversity of the bacterial communities. Xu et al. (2019)
250 reported the responses of soil *nosZ*-type denitrifying microbial communities to the various land-use
251 types of the Loess Plateau, and the results showed that AN was the key environmental factor
252 affecting microbial composition. An et al. (2017) reported that the loss of microbial diversity affects
253 the carbon mineralization rate and the pattern of carbon source utilization, leading to functional
254 changes in terrestrial ecosystems. Zhao et al.(2018) found that the change of soil carbon composition
255 after afforestation had a significant impact on the diversity and structure of soil microbial
256 community.

257 This study only addresses the soil microbial characteristics of different stands in growing
258 season during revegetation in Wuqi City, Shaanxi Province. To better analyze the mechanism of
259 influence of vegetation on microorganisms and the influence of site conditions and environmental
260 factors on soil microbial community diversity, the litter and root exudates under various artificial
261 forests can be studied further. To better analyze the microbial characteristics under the effect of

262 revegetation in the Loess Plateau, similar studies can be conducted in other areas of the Loess
263 Plateau. The results of these experiments can be analyzed to better explore the soil under *H.*
264 *rhamnoides* forest in the process of vegetation diversity of soil microbial community and is expected
265 to provide some data supporting the study of revegetation in the Loess Plateau.

266 **5. Conclusions**

267 In this study, the soil bacteria were analyzed as a function of different vegetation restoration
268 types on the Loess Plateau. Soil microbial biomass C and N in different forest layers exhibited a
269 trend of a small change with increasing soil layers; that is, as the soil depth increases, the soil
270 microbial activity weakens gradually. Across all samples, the dominant groups were Proteobacteria,
271 Acidobacteria, Chloroflexi, Gemmatimonadetes, Nitrospirae, Bacteroidetes, GAL15, and
272 Firmicutes. Soil nutrient supply, such as TOC, TN, AN, AP, and AK, have crucial roles in shaping
273 the composition and diversity of the bacterial communities. Furthermore, the data also provide a
274 starting point for establishing soil bacterial databases in the Loess Plateau, as well as for the plants
275 associated with the vegetation restoration.

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280 **Authors' contributions**

281 XL and TW conceived and designed the experiments. XL carried out the field work, analyzed the
282 data and drafted the manuscript. TW revised the manuscript and provided editorial advice. All
283 authors have read and agreed to the published version of the manuscript.

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287 **Availability of data and materials**

288 The datasets used and analyzed during the current study are available from the corresponding author
289 on reasonable request.

290 **Declarations**

291 **Ethics approval and consent to participate**

292 Not applicable

293 **Consent for publication**

294 All the authors have approved the manuscript that is enclosed.

295 **Competing interests**

296 The authors declare that they have no known competing financial interests or personal relationships
297 that could have appeared to influence the work reported in this paper.

298 **Authors' information**

299 School of Soil and Water Conservation, Beijing Forestry University, Beijing, 100083, PR China.

300 Chinese National Forestry Ecosystem Observation and Research Station, CNERN, Ji Country,

301 Shanxi, 6042200, PR China

302 **Figure Captions**

303 **Fig. 1.** Rarefaction curves at cutoff level of 3%.

304 **Fig. 2.** Relative abundance (a) and Circos (b) of bacterial community at the phylum level. Others:

305 Relative abundance <1% in all samples.

306 **Fig. 3.** Distribution of Proteobacteria populations at the class levels.

307 **Fig. 4.** Principal coordinates analysis (PCoA) of bacterial composition based on Bray-Curtis

308 distances.

309 **Fig. 5.** Heatmap of bacterial and soil environmental factors. R and P values are obtained by

310 calculation. The R value is shown in different colors in the graph. The color card on the right-hand

311 side of the heatmap is a color partition with different R values. $0.01 < p \leq 0.05$ * $0.001 < p \leq 0.01$

312 ** $p \leq 0.001$ ***. BC: microbial biomass C, BN: microbial biomass N, TOC: total organic C, TN:

313 total N, AN: alkali-hydro N, TP: total P, AP: available K, TK: total K, AK: available K, the same

314 as followings.

315 **Fig. 6.** Ordination plots of the results from the redundancy analysis to identify the relationships

316 among the bacterial populations (blue arrows) and soil environmental factors (red arrows).

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Figures

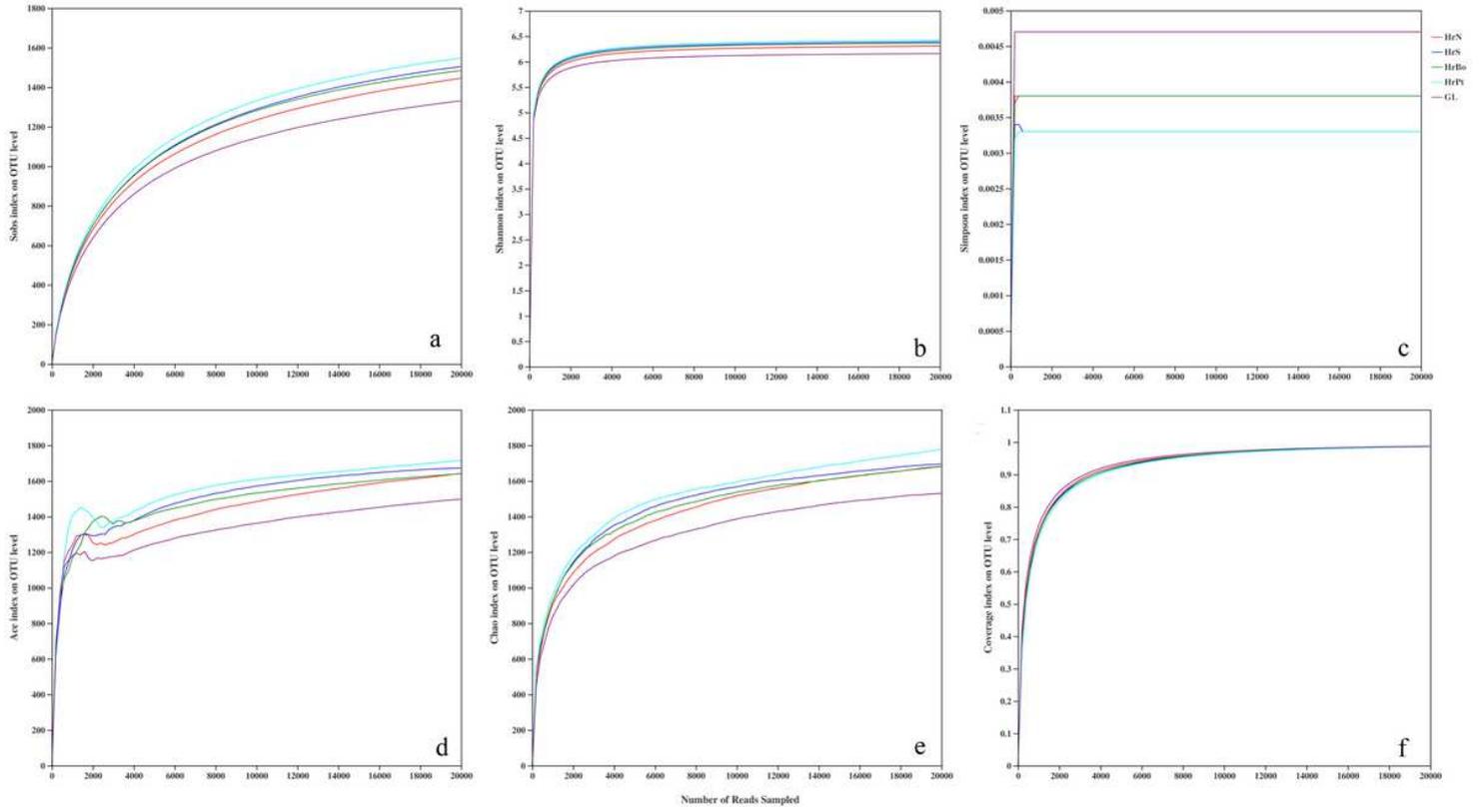


Figure 1

Rarefaction curves at cutoff level of 3%.

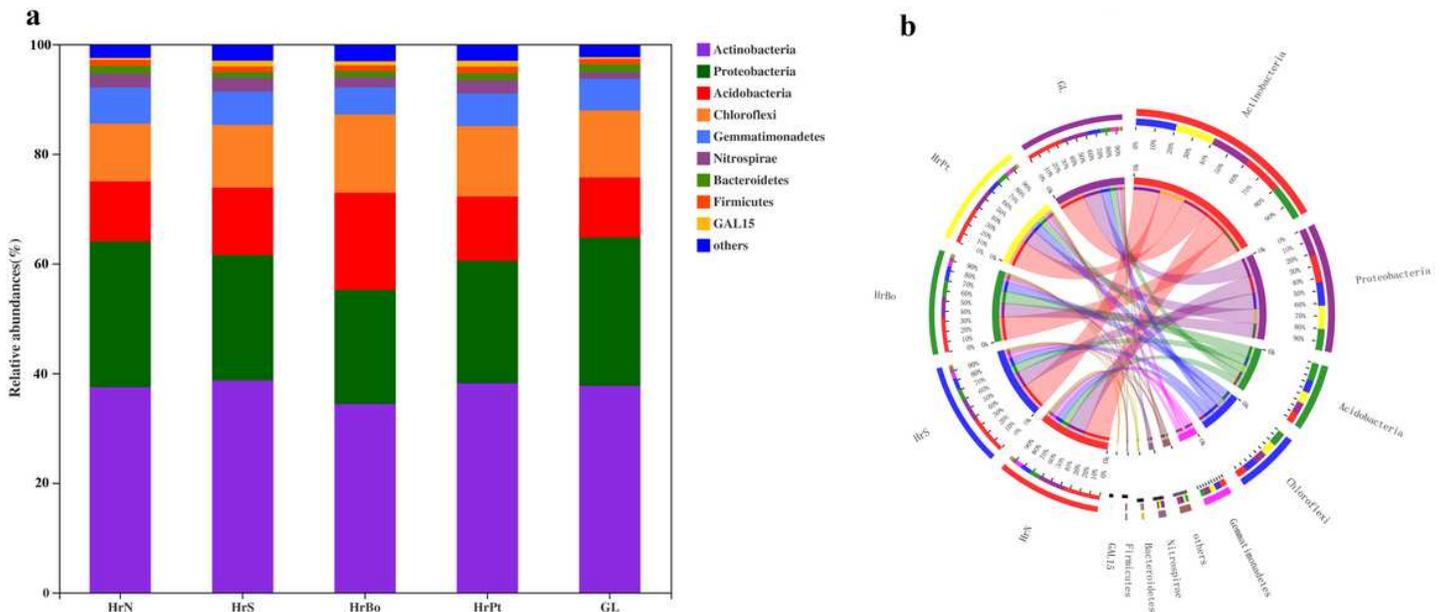


Figure 2

Relative abundance (a) and Circos (b) of bacterial community at the phylum level. Others: Relative abundance <1% in all samples.

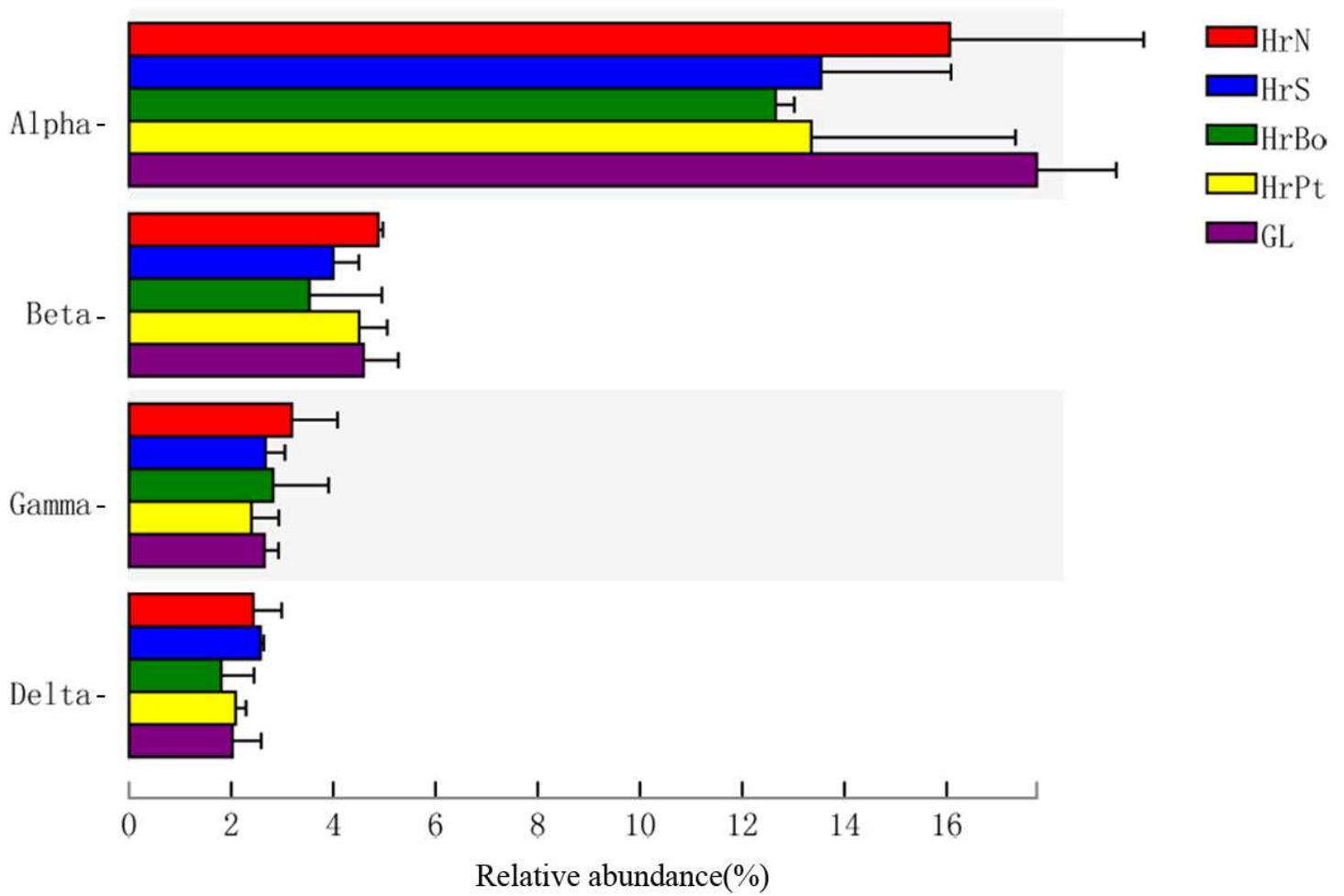


Figure 3

Distribution of Proteobacteria populations at the class levels.

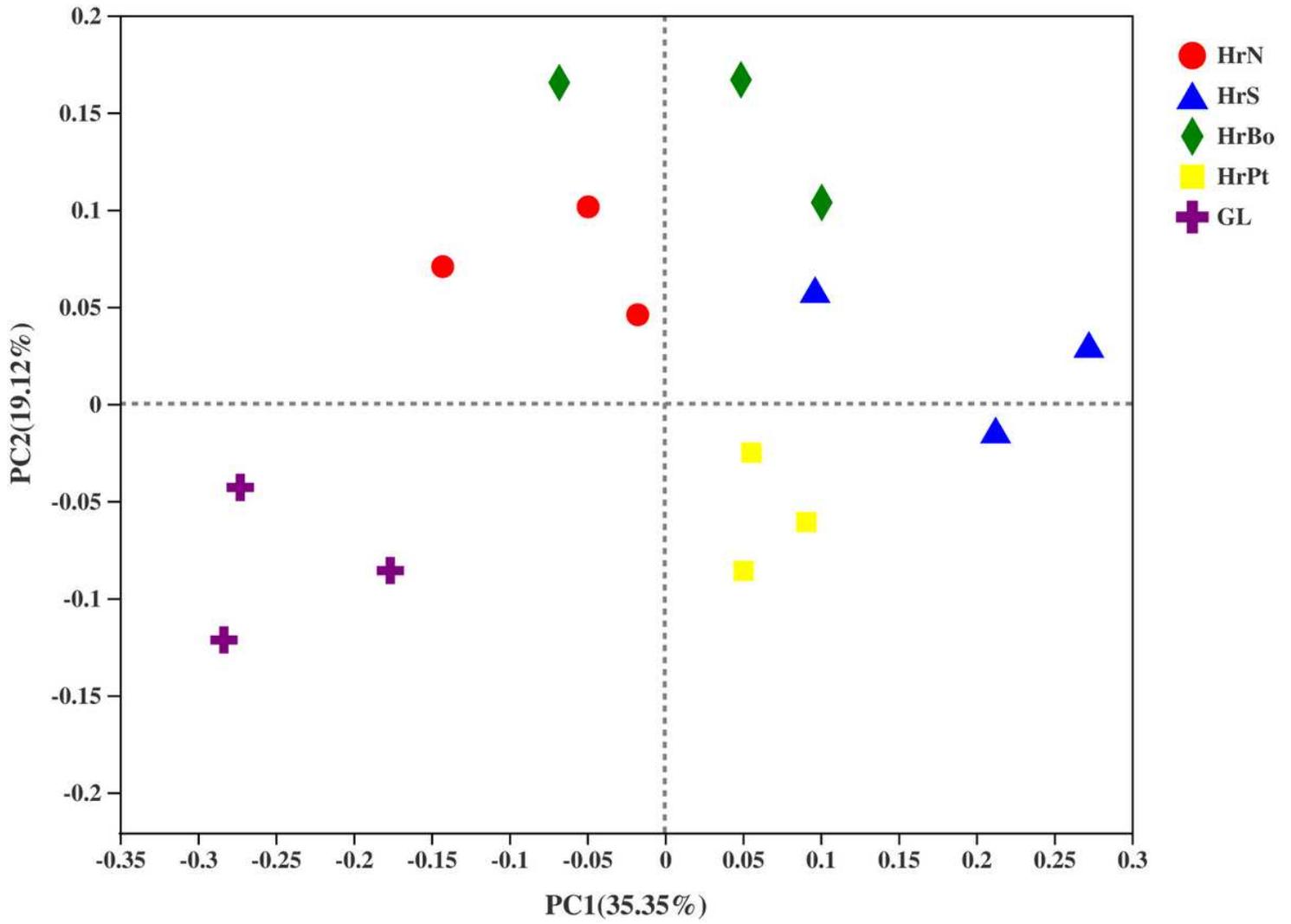


Figure 4

Principal coordinates analysis (PCoA) of bacterial composition based on Bray-Curtis distances.

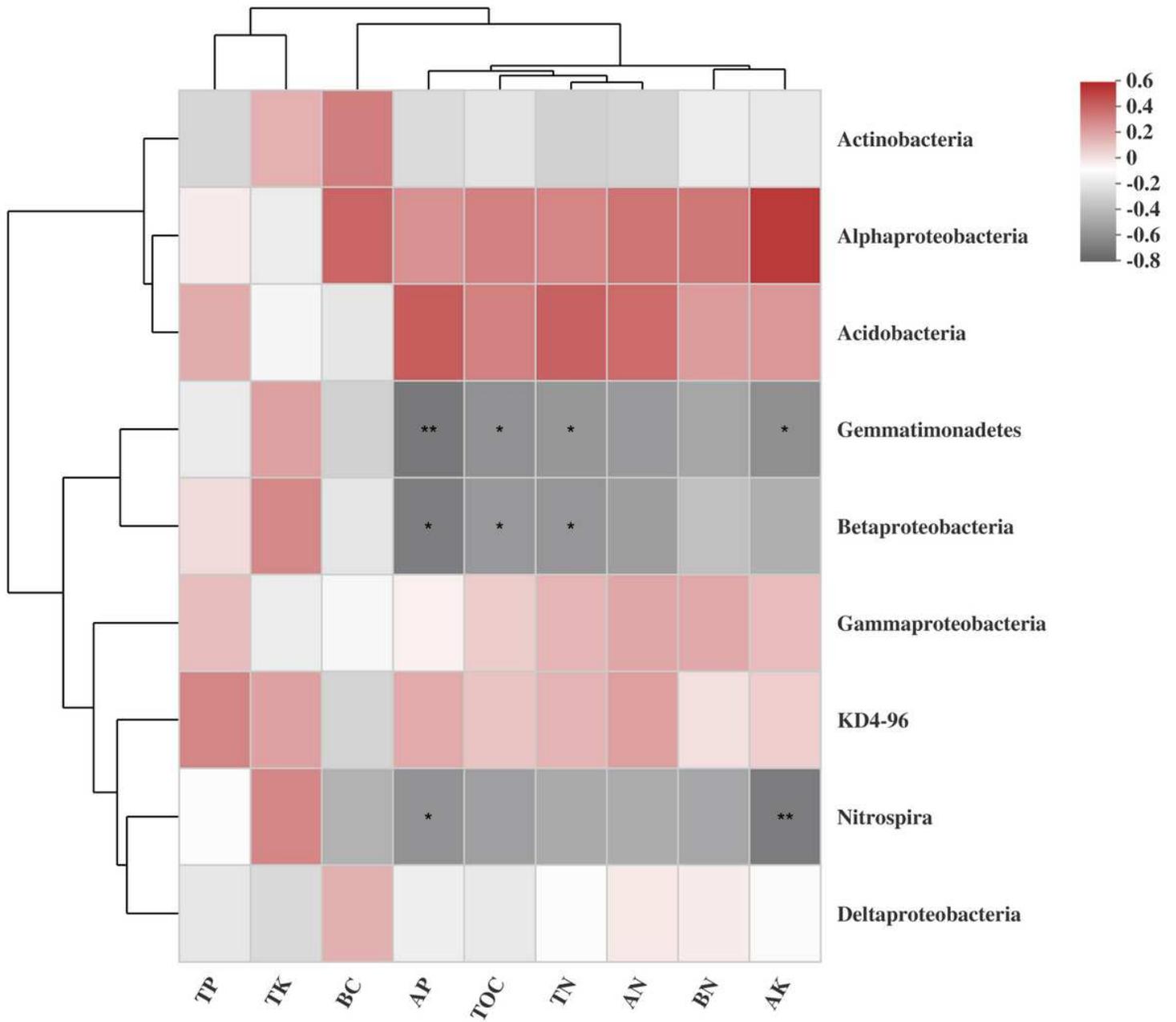


Figure 5

Heatmap of bacterial and soil environmental factors. R and P values are obtained by calculation. The R value is shown in different colors in the graph. The color card on the right-hand side of the heatmap is a color partition with different R values. $0.01 < p \leq 0.05$ * $0.001 < p \leq 0.01$ ** $p \leq 0.001$ ***. BC: microbial biomass C, BN: microbial biomass N, TOC: total organic C, TN: total N, AN: alkali-hydro N, TP: total P, AP: available K, TK: total K, AK: available K, the same as followings.

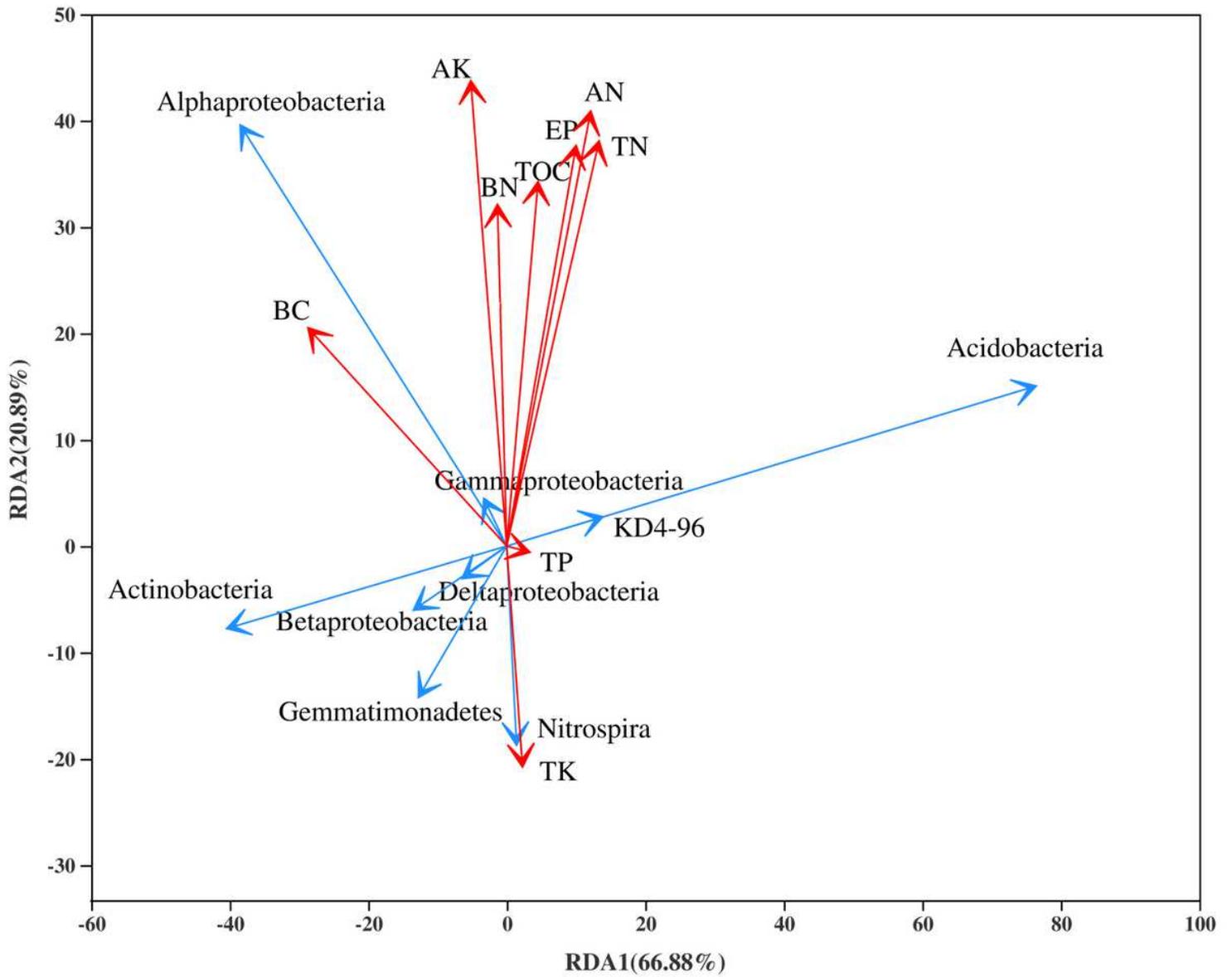


Figure 6

Ordination plots of the results from the redundancy analysis to identify the relationships among the bacterial populations (blue arrows) and soil environmental factors (red arrows).