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1 Regional disparity in clinker emission factors and their potential 2 reduction in China

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11 **ABSTRACT:**

12 Detailed analysis the disparity and reduction potential of clinker emission factors at the provincial level is
13 important for regional reduction policies. Using the surveyed data from 185 new suspension and pre-heater
14 (NSP) process lines and 69 Shaft kiln lines, this study firstly analyzed the disparity in emission factors
15 based on production process, production scale, and regional distribution in 2015. We found that the
16 emission factor of the Shaft kiln process (898.24 kg/t) is higher than that of the NSP process (858.59 kg/t),
17 and that small-scale production lines have higher emission factors than large-scale lines both for the two
18 process. China's clinker emission factors increase from the eastern to the western regions. Then we
19 estimated the reduction potential of structural adjustment, raw material substitution, and energy saving and
20 fuel substitution in regional emission factors by 2030. The result shows that emission factors of the

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21 surveyed provinces will decrease by 101.41-174.60 kg/t compared to the values in 2015, which is mainly
22 contributes by energy saving and fuel substitution (65.98%), and raw materials substitution (25.72%). And
23 structural adjustment contributes only a small part reduction for most investigated provinces. The national
24 average emission factor is estimated to be 715.33 kg/t in 2030, which indicates a reduction of 16.65%.
25 These results can provide valuable feedback to government officials on the effectiveness of existing
26 measures and also serve as a reference for future decisions on emission reduction polices.

27 **Keywords:** Cement clinker; Emission factor; Reduction potential; Disparity; China

28

29 1. Introduction

30 The cement industry is characterized by its high usage of mineral resources and fossil
31 energy and correspondingly high emissions of greenhouse gases. As a major carbon-emitting
32 industry, it accounts for approximately 7% of the total global carbon emissions (Deja et al.,
33 2010). China is both the largest producer of cement and the highest emitter of CO₂ in the
34 world. It has ranked first globally in cement output since 1985. In 2016, China's cement
35 output was 2410 million tons (Mt) (NBS, 2017), accounting for 59.16% of the total global
36 cement output (U.S.G.S, 2018). China's cement sector emitted approximately 1270 Mt of
37 CO₂ in 2011 (Gao et al., 2017a; Shen et al., 2015). The rapid expansion of the cement sector
38 is associated with increasing resource and energy consumption and pollution, thereby posing
39 a severe challenge to the sustainable economic and social development of China.

40 Several studies have investigated the carbon emission generated by China's cement
41 industry. These studies have estimated the carbon emissions based on different calculation
42 methods, production processes, and material or fuel consumptions; thus, the results differ

43 greatly from each other. Wei et al. (2012) and Liu et al. (2014) analyzed and compared the
44 methods for the calculation of carbon emission in China's cement industry proposed by
45 scholars domestically and internationally. Hu et al. (2015) analyzed the emissions generated
46 by two production processes, namely, a new suspension and pre-heater (NSP) kiln and Shaft
47 kiln. Based on production data from different enterprises, Shen et al. (2014) studied the
48 emission factors of clinkers and cement in the major production processes. Zhao and Wei
49 (2013) compared the emission factors between two production scales of 2500 tons/day (t/d)
50 and 5000 t/d. However, few studies have examined the disparity in regional carbon emissions.
51 Research into the regional disparity in the emission factors of cement clinkers and underlying
52 reasons for such variations will aid in the drafting of appropriate regional policies for
53 emission reduction.

54 To reduce the emission factor of cement clinker, scholars have conducted in-depth
55 research on emission reduction with respect to fuel consumption and production processes.
56 Several researchers have examined opportunities for reducing carbon emissions in the cement
57 sector by means of technological modernization and improved energy efficiency. Ali et al.
58 (2011) studied the energy intensity of different cement kiln processes. Compared with the
59 wet-process kiln, a dry-process kiln can reduce power consumption by 13% and fuel
60 consumption by 28% (Avami and Sattari, 2007). Energy-efficiency-asia (2010) compared the
61 heat consumed by different types of cement kilns and auxiliary devices and found that a
62 multi-stage cyclone preheating system could reduce energy intensity. Therefore, replacement
63 of outdated kilns with the NSP process kiln is effective for reducing carbon emissions.
64 Previous studies have analyzed the energy intensity of different types of kilns in terms of

65 production processes, but few have investigated the differences in energy intensity between
66 production lines at different production scales and the emission reduction potential of
67 large-scale production.

68 Wastes and by-products can be utilized as constituents of the final product and
69 components of the kiln feed for cement production (Trezza and Scian, 2000). Some industrial
70 wastes with high CaO, SiO₂, and Fe₂O₃ content can be used as calcium, silicate and iron
71 sources in the production of cement clinker. Numerous studies have reported the utilization of
72 carbide slag (Li and Li, 2010; Liu et al., 2014), steel slag (Carvalho et al., 2017; Iacobescu et
73 al., 2011; Iacobescu et al., 2013; Saade et al., 2015; Zhang et al., 2011), lead slag (Onisei et
74 al., 2012), phosphorous slag (Allahverdi et al., 2016; Gao et al., 2008; Li et al., 2000), copper
75 slag (Kalinkin et al., 2012), magnesium nickel slag (Maes and Belie, 2017; Tan et al., 2016;
76 Zhang et al., 2011; Zhang et al., 2017), coal fly ash (Darsanasiri et al., 2018; Shwekat and Wu,
77 2018; Xu and Shi, 2018), and waste sludge (Lin et al., 2017; Rodríguez et al., 2013;
78 Valderrama et al., 2013) as alternative materials in cement clinker production. The main mix
79 materials used in China include combustion ashes, metallurgical slag, chemical slag, and
80 other mineral components. The annual output of industrial byproducts that can be used in
81 cement production in China exceeded 2 Bt (Zhang et al., 2013). However, it is only used in a
82 small number of investigated enterprises, and the substitution rate is quite low; The current
83 rates of replacement by alternative raw materials and fuels in China's cement industry are
84 1.30% and 1.80%, respectively (Gao, 2018; Gao et al., 2017a; Ke et al., 2012), only a 2.37%
85 reduction has been achieved in calcination emissions (Gao et al., 2017a). There is a big
86 difference between the scale at which raw materials and fuels are substituted in China and

87 that in developed countries.

88 Detailed analysis the regional disparity and reduction potential of clinker emission
89 factors, which is important for both scientific research and design of emission control policies
90 has been lacking. Objective quantification of the CO₂ emission factors of regional clinkers,
91 identification of the disparity in their values, and evaluation of ways to reduce them are
92 warranted. Based on a sampling survey that considers the production processes and scales,
93 the variety and quantity of substitute materials and fuels, and the energy saving technologies
94 application plan in different regions, the current study aims to fill this gap by analyzing the
95 scale changes and regional differences of clinker emission factors, evaluating the potentials
96 and methods for emission reduction in the surveyed provinces in 2030, thus aiding in
97 decision-making regarding regional emission reduction.

98 To achieve these objectives, this study was divided into four sections. The survey data
99 and clinker emission calculation method are described in Section 2; in Section 3 comparison
100 of clinker emission in provinces and scales; Section 4 evaluates the emission reduction
101 methods and their reduction potential; and conclusions are discussed in Section 5.

102 2. Data sources and calculation methods

103 2.1 Data sources

104 During the years 2011–2015, a total of 254 production lines (185 NPS kilns and 69 Shaft
105 kilns), which accounted for 14.96% of China's clinker production, were surveyed in the 22
106 provinces. The locations of the plants are shown in Figure 1. The surveyed data consist of
107 four sections. The first section contained basic information questions related to enterprise
108 name, geographical location, number and scale of production lines, production process. The

109 second section mainly included production data: annual output of clinker, annual
110 consumption of fuels, power consumption by the three production stages, and power
111 generated by waste heat. The third section consisted of questions related to chemical
112 composition of raw materials, raw meals, clinkers, and fuels, fuel industry analysis, and
113 proportioning of raw materials. The fourth section requested information on whether the
114 facilities had adopted any of the 32 recommended energy-saving and 3 material and fuel
115 substitution measures and, if not, the reason for this and any future plans to adopt new
116 technologies.

117

118 Figure 1 Distribution of the surveyed cement production plants

119

120 In order to reflect the real carbon emissions and reduce errors, all data items used are
121 annual average values surveyed under normal production status. And the abnormal values
122 were identified and revised by considering the local realities. The clinker production, average
123 fuel intensity, raw meal intensity, chemical compositions of raw meal and clinker, raw
124 materials substitution and waste heat power generation (WHPG) in NSP kilns and Shaft kilns
125 are listed in Table S1. The NSP kilns accounted for 94.98% clinker production of the
126 surveyed samples, and the remaining 5.02% of production was from Shaft kilns. These
127 proportions were similar to the overall structure of China's cement industry. The survey data
128 for average fuel intensity, power intensity and WHPG of the NSP kilns and Shaft kilns were
129 close to the national published value. Additionally, the composition of raw meals and clinkers
130 data from both of the two production process are also fall within their ranges (Zhou and Peng,

131 2005). Thus, the surveyed samples represent the present performance of China's cement
132 industry.

133 2.2 Methods for the calculation of emission from production lines/scales

134 The carbon emissions from cement clinkers comprise three components: process-related,
135 fuel-related, and power-related emissions (Mikulčić et al., 2011). Process-related emissions
136 refer to CO₂ emissions arising from the decomposition of carbonate minerals. Calcium
137 carbonate and magnesium carbonate decompose into Cao, MgO, and CO₂ during clinker
138 calcination. The calculation approach is indicated in Formula (1).

$$139 \quad EF_{pr} = Ra_{co2} \times r_a \quad (1)$$

140 Here, EF_{pr} represents the process-related emission factor of clinker (kg/t), r_a is the raw
141 meal intensity (kg/t), and Ra_{co2} is the carbon content of raw meals (%), which is based on the
142 proportioning of raw materials and the chemical composition of their carbonate materials
143 (Gao et al., 2017a).

144 Fuel-related emissions refer to the CO₂ emissions arising from the burning of fuels
145 during the calcination of clinkers. This calculation can be made using Formula (2).

$$146 \quad EF_{fu} = FI_{coal} \times EF_{coal} \quad (2)$$

147 Here, EF_{fu} represents the fuel-related emission factor of clinker (kg/t), FI_{coal} is the fuel
148 intensity (kgce/t), and EF_{coal} is the fuel emission factor (kg/kgce) (IPCC, 2006).

149 Power-related emissions (EF_{po}) refer to the CO₂ emissions arising from the consumption
150 of power during the grinding of raw materials and calcination of clinker and reduce the power
151 generated by low-temperature waste heat. Usually, different regions use the same calculation
152 method (Formula 3).

153
$$EF_{po} = (PI_{ele} - PG_{was}) \times EF_{ele} \quad (3)$$

154 Here, EF_{po} represents the power-related emission factor of clinker (kg/t), PI_{ele} is the
 155 power intensity (kWh/t), PG_{was} is the power generated by low-temperature waste heat
 156 (kWh/t), and EF_{ele} is the regional power emission factor (kg/kWh) (NDRC, 2012).

157 The emission factors (EF_j) of production lines (j) can be made using Formula (4).
 158 Sequentially calculated the emission factors of production scales (EF^q) by considering the
 159 weight of clinker output (C_j^i) of each production line (j) with the same scale ($i=q$) (Formula
 160 5).

161
$$EF_j = EF_{jpr} + EF_{jfu} + EF_{jpo} \quad (4)$$

162
$$EF^q = \sum (EF_j^i \times \frac{C_j^i}{\sum_{i=q} C_j^i}) \quad (5)$$

162 2.3 Method for the calculation of emissions from regional clinkers

163 We proposed the three tiers of integration framework to estimate regional emission
 164 factors from the surveyed clinker production lines (Figure 2). After obtaining the emission
 165 factors relevant to all the surveyed lines, we sequentially calculated the emission factors of
 166 regional different processes by considering the weight of each of their clinker outputs.
 167 According to the proportion of clinker output of Shaft kilns and NSP kilns in the region, the
 168 emission factor (EF_r) for province was then integrated (Formula 5).

169
$$EF_r = \frac{C_{nr}}{C_r} \sum (EF_{jnr} \times \frac{C_{jnr}}{\sum C_{jnr}}) + \frac{C_{sr}}{C_r} \sum (EF_{jsr} \times \frac{C_{jsr}}{\sum C_{jsr}}) \quad (5)$$

170 Here, C_{nr} and C_{sr} represent the regional NSP (n) and Shaft kiln(s) clinker output (kt), C_r
 171 is the regional clinker production (kt), C_{jnr} and C_{jsr} stands for the clinker output of surveyed
 172 NSP and Shaft production line j in region r (kt), EF_{jnr} and EF_{jsr} are their production line
 173 emission factors (kg/t).

174

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Figure. 2 The framework of three tiers regional emission integrated system

176

177 3. Analysis of disparity between carbon emissions from cement clinkers

178 3.1 Analysis of disparity between emissions with respect to production processes and 179 scales

180 According to the survey data and the above calculation method, the main emission
181 indexes and emission factors of different processes and scales are listed in Table S2. Among
182 the major indexes of clinkers (Table S2), the CaO content (65.26%) for the NSP process is
183 higher than that (64.44%) for the Shaft kiln process, and the MgO content differs slightly
184 between the two production processes. With respect to substitute materials, the Shaft kiln
185 process reduces carbon emissions (by 10.59 kg/t) more effectively than the NSP process (by
186 7.58 kg/t). Therefore, the process-related emission factor (521.00 kg/t) of the Shaft kiln
187 process is lower than that (529.33 kg/t) of the NSP process. The power intensity of the NSP
188 kiln process are slightly higher than those of the Shaft process, whereas WHPG are greater.
189 Correspondingly, power-related emissions of the NSP kiln process are lesser than those of the
190 Shaft process, and the fuel-related emissions are also lesser. The fuel-related emission of the
191 Shaft kiln process is 319.67 kg/t and that of the NSP process is 291.74 kg/t. Thus, with
192 respect to production processes, the emission factor of clinkers for the Shaft kiln process
193 (898.24 kg/t) is higher than that for the NSP kiln (858.59 kg/t). This shows that the
194 replacement of Shaft kiln by NSP kiln is one of the methods to reduce clinker emission
195 factor.

196 In addition to the choice of process, the scale of production will have an impact on fuel
197 and power-related emissions. In order to study the effects of changes of scale upon the
198 emission factors, this paper classifies the production lines of the NSP kiln into five types, and
199 classifies the production lines of the Shaft kiln into four types (Table S2). As shown in Table
200 S2, in both production processes, fuel and power intensity tend to decrease at larger scales of
201 production (Gao et al., 2017b; Zhao and Wei, 2013). For NSP kilns, the fuel and power
202 intensity of >5000 t/d production lines were 104.79 kgce/t and 58.20 kWh/t, which were
203 15.17 kgce/t and 16.10 kWh/t lower, respectively, than those of <2000 t/d production lines.
204 The WHPG was also higher for larger production lines as they are basically equipped with a
205 low-temperature waste heat power generation system. Thus, the fuel-related and
206 power-related emissions were decreased to 276.65 and 31.23 kg/t for lines with production
207 of >5000 t/d. Thus, it's mass production that has reduced the fuel-related and power-related
208 emissions for both of the two processes.

209 As shown in Table S2 , the CaO or MgO content of the clinkers does not vary with the
210 changes in production scale; in the NSP process, average CaO content is >65% at all
211 production scales, whereas those of the Shaft kiln process are all <65%. The difference in
212 process-related emissions arising from the different CaO content of clinkers in the two
213 production processes is 6.43 kg/t. For the NSP process, the MgO content of the clinkers
214 varies by 2% across different production scales and is consistently lower than that for the
215 Shaft kiln process. As mentioned before, the chemical composition of clinker is different
216 between NSP kiln and Shaft kiln, but there is no obvious change under different scales of the
217 same process. The difference of process-related emissions is mainly reflected in the different

218 proportion of alternative raw materials. In terms of substitution rates for raw materials, both
219 production processes show higher substitution rates at smaller production scales. It means
220 that raw material substitution was relatively more common in small scale production. For the
221 two production scales (≤ 300 t/d and in the range of 500~900 t/d) in the Shaft kiln process, the
222 substitution rates of raw materials are 2.77% and 4.41%, respectively, with corresponding
223 emission reductions of 15.16 kg/t and 15.91 kg/t. For the two production scales (< 2000 t/d
224 and 2000~2500 t/d) in the NSP process, the rates of substitution with raw materials are 2.95%
225 and 3.81%, respectively, with corresponding emission reductions of 8.16 kg/t and 12.59 kg/t.
226 Due to the different in chemical composition of clinker and substitution rate of raw material,
227 the process-related emissions of various scale NSP process are little higher than that of Shaft
228 kilns.

229 3.2 Analysis of disparity in regional carbon emission factors

230 As shown in Figure 3(a), the emission factors of cement clinkers tend to increase from
231 the eastern coastal to the western inland regions. The western regions (including Tibet,
232 Xinjiang, Inner Mongolia, and Yunnan) have the highest emission factors, and the eastern
233 regions (including Zhejiang, Jiangsu, Guangxi, and Hebei) have lower emission factors than
234 the central regions (including Shanxi, Henan, Hubei, and Hunan). Hebei and Guangxi have
235 the lowest emission factors, which can be attributed to various factors. Specifically, in
236 Guangxi, the MgO content in clinkers is remarkably lower than the national average level,
237 and in Hebei the substitution rate of raw materials is very high, thus reducing process-related
238 emissions.

239

Figure 3 Emission factors and composition of regional clinker in 2015

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CaO and MgO in clinkers are decomposed from calcium carbonate and magnesium carbonate respectively, which are the main factors affecting the process-related emission. With respect to process-related emissions, the MgO content of clinker samples for the NSP process tends to decrease with increase in CaO content (as shown in Figure 4). In general, the MgO content of the clinkers across China is higher than the standard emission factor specified by the CSI (Gao et al., 2017b). Clinker samples from the NSP process and Shaft kiln have an average MgO content of 2.20% and 2.30%, respectively. In China, the high-MgO regions are mainly distributed in Shandong, Henan, Hebei, Shanxi, Fujian, and Jiangxi. As a result of the high MgO content, Fujian, Jiangxi, and Henan are areas of higher process-related emissions as shown in Figure 3(b). While Hubei, Hunan, Guizhou, and Yunnan are provinces with low MgO content where the process-related emissions are lower than the standard emission factor (525 kg/t) specified by the CSI (2011), as shown in Figure 3(b).

Figure 4 CaO and MgO contents in clinker for the surveyed NSP samples

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Note: The first two digits of the sample number are the province number, and the others are the provincial sample number.

260 Energy intensity is determined by diverse factors, including production processes,
261 production scales, production experience and management level. In China, clinkers from the

262 northern coast exhibit the highest energy efficiency, followed by those from the southern
263 coast. The northwest and southwest China are characterized by high energy intensity and high
264 fuel-related emissions, as shown in Figure 3(c). The above analysis shows that in China, the
265 fuel intensity (110.04 kgce/t) of the NSP process is lower than that (121.09 kgce/t) of the
266 Shaft kiln process, as described in Table S2. Judging by the relationship between production
267 processes and energy intensity, the energy intensity of clinkers is lower than the national
268 average in those regions (including Hebei, Henan, Liaoning, and Jiangsu) where the
269 proportion of NSP clinker production is higher. It can also be concluded that the application
270 of the NSP process, especially large scale NSP kiln production lines, is beneficial for
271 reducing the energy intensity of clinkers. Regional energy intensity and production scale (>
272 4000 t/d) showed a high correlation (Figure 5). Among the 22 provinces, energy intensity is
273 relatively low in the regions where production capacity of 4000 t/d accounts for a high
274 overall proportion, whereas energy intensity—and consequently, fuel-related emission
275 factors—are relatively high in the regions (including Guizhou, Tibet, Xinjiang, and Yunnan)
276 where production capacity of 4000 t/d accounts for a low overall proportion, as shown in
277 Figure 3(c). Thus large scale production can reduce energy emission intensity of clinker.

278

279 Figure 5 Relationship between regional energy intensity and production scales >4000 t/d

280

281 Substitution of steel slag, fly ash, phosphorous slag, or sulfate slag for natural calcium
282 and siliceous materials is the main way to reduce process-related emissions of clinkers. The
283 survey of the 22 provinces shows that the substitution rates of raw material are relatively high

284 in the regions where steel, phosphorite, and coal resources are widely distributed, including
285 Hebei (a large steel-producing province); phosphorite rich provinces (including Guizhou and
286 Yunnan); and large coal-producing provinces (including Shanxi, Shaanxi, and Shandong), as
287 shown in Figure 3(d). For the Shanxi and Hebei provinces, steel slag and fly ash are used to
288 substitute natural materials such as limestone and sandstone, resulting in emission reductions
289 of 16.89 kg/t and 15.87 kg/t, respectively. Southwest China is also the major region where
290 emission reductions are achieved by raw material substitution. For Guizhou, Yunnan, and
291 Sichuan, the main substitute materials are phosphorous slag, steel slag, and fly ash, which
292 reduce carbon emissions by 24.74 kg/t, 11.83 kg/t, and 10.12 kg/t, respectively. In some large
293 steel-producing and coal-producing provinces (including Jiangsu, Inner Mongolia, and
294 Xinjiang), the rates of substitution of raw materials are presently very low, thereby indicating
295 great potential for the reduction of carbon emissions by raw material substitution in these
296 regions.

297 4. Selection of methods for regional emission reductions and their potentials

298 It can be seen from Figure 2 and above analysis that we can use structural adjustment,
299 raw material substitution, and energy saving and fuel substitution to eliminate or reduce the
300 clinker emission factor; these results are slightly different from those established by previous
301 research. The three methods for reduction will be introduced below and basic information
302 relevant to our assessment will be presented.

303 4.1 Emission reduction potential of structural adjustment

304 In 2016, China's clinker production capacity was 2.02 billion tons, but the backward
305 production capacity, which is mainly attributed from Shaft kilns and small-scale NSP kilns

306 (<2000d/t), was approximately 19.46% (393 million tons). Phase out the obsolete production
307 is conducive to capacity reduction, energy conservation, and emission reduction, as the
308 fuel-related emissions and fuel intensity of Shaft kiln are higher than that of NSP kiln, and
309 those of small-scale NSP kilns are higher those that of large-scale NSP kilns (>4000d/t)
310 (Table S2). Thus, this policy was announced in the *Cement industry capacity reduction*
311 *action plan (2017-2020)*. Qiao Longde, the president of the China Cement Association
312 (CCA), said, As the energy conservation and emission reduction are not up to the reduction
313 standard, in the next ten years, a larger cement production line (≤ 2500 t/d) will enter a new
314 round of phase out backward production capacities. Therefore, we expect that the production
315 lines less than 2500 t/d in China will be eliminated by 2030.

316 The emission reduction method of structural adjustment reveals the different change
317 trends of process-related, fuel-related and power-related emissions in regional clinker
318 production. It results in an increase in process-related emissions and weakened the emission
319 reduction obtained from raw material substitution in most provinces (Figure 9). In most
320 provinces, the proportion of raw material substitution in small-scale kilns, especially in kilns
321 with 2000~2500 t/d, is higher than that of large-scale kilns. Just for this main reason result in
322 higher process-related emissions, compared to that of small-scale kilns. Thus, Guizhou
323 presented the highest increased process-related emissions (11.94 kg/t). Except for Sichuan,
324 Chongqing and Guangdong, structural adjustment increases the process-related emissions in
325 most of the surveyed provinces, compared with the value in 2015.

326 While, this adjustment decreases fuel- and power-related emissions due to improvement
327 in energy efficiency of large-scale kilns, the fuel and power intensity of the kilns > 5000 t/d is

328 decreased by 8.90% and 14.57%, respectively, as compared to that kilns with 2000~2500 t/d
329 (Table S2). The reduction in fuel-related emissions from structural adjustment method
330 exceeded 6 kg/t in half of the surveyed provinces, where in the proportion of large-scale kilns
331 was relatively lower or there were a large number of Shaft kilns. Power intensity have
332 reduced in a vast majority of provinces; in particular, these consumptions have reduced by
333 8.55, 7.44, and 11.89 kWh/t in Shanxi, Inner Mongolia, and Tibet, respectively. Moreover,
334 the large NSP line in China is equipped with a low-temperature waste heat power generation
335 system; therefore, shutting down backward and small production capacities will increase the
336 emission reduction potential of WHPG. In southwest China, such as in Sichuan, Chongqing,
337 and Guizhou, structural adjustments can increase the emission reduction of WHPG by more
338 than 12.53 kg/t; and other provinces were reduced in different degree (Figure 6).
339 Power-related emissions caused by structural adjustment have been decreased in all of the
340 surveyed provinces.

341 In general, structural adjustments can reduce the clinker emission factor of each
342 provinces. In Tibet, Chongqing, and Sichuan, emission factor reduced more than 30 kg/t,
343 where large scale NSP kilns account for a lower proportion. On the contrary, Zhejiang (1.75
344 kg/t), Hunan (2.90 kg/t), and Jiangxi (3.47 kg/t) emission reduced less because of the minor
345 difference in the energy efficiencies of different scale production lines in these regions
346 (Figure 6).

347

348

Figure 6 Emission reduction of structural adjustment

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351 4.2 Emission reduction potential of energy saving and fuel substitution

352 Owing to the high energy intensity of the calcination process and significant CO₂ release
353 from the fuel and power consumption, energy saving and low carbon emissions are major
354 concerns in clinker production. A total of 260 energy saving technologies and 27 low carbon
355 technologies have been promoted by the National Development and Reform Commission
356 (NDRC) in the *National Key Energy Conservation and Low Carbon Technologies Promotion*
357 *Catalog (2017)*, as significant opportunities to improve energy efficiency exist in China's
358 cement industry and other industrial sectors. Thirty-two potential energy efficiency
359 technologies, including 12 grinding technologies (R1-R12) and 20 calcination technologies
360 (C1-C20), and fuels substitution (A3) were evaluated for provinces of the future cement
361 industry. The description, fuel and power savings, investment, income, emission reductions,
362 and payback period of each of these measures can be found in Talaei et al. (2019), Gao et al.
363 (2017a), Huang et al. (2016), Madloul et al. (2011), Hasanbeigi et al. (2010b), Hasanbeigi et
364 al. (2010a), and *National key energy saving and low carbon technology promotion catalogue*
365 (NDRC, 2018). According to historical data and the present application of various
366 technologies obtained from our survey, studies of the above mentioned authors, and the
367 NDRC, the future applied share of the energy improvement technologies was estimated
368 according to their unit costs and penetration parameters in the surveyed provinces (Table S3).

369 The emission reduction of this method is reflected in three aspects: power saving, fuel
370 substitution and fuel efficiency improvement. The fuel intensity, fuel substitution, and power
371 intensity of surveyed provinces were subsequently been estimated. Compared with 2015, the

372 fuel intensities of all the provinces decreased by more than 10 kgce/t, and Inner Mongolia,
373 Shanxi and Tibet decreased the most, which was 21.63, 20.61, and 20.25 kgce/t respectively.
374 In 2030, the fuel substitution in eastern coastal provinces are relatively higher (>20%), while
375 that in Xinjiang and Tibet are the lowest, 9.18% and 4.69% respectively. This method
376 increases the WHPG and decreases the power consumption in clinker production in all of the
377 investigated provinces. After deducting the WHPG, the power intensity are less than 30
378 kWh/t in the investigated provinces, expect for Tibet (65.95 kWh/t), Xinjiang (53.78 kWh/t),
379 Shanxi (35.38 kWh/t), and Yunnan (34.05 kWh/t) .

380 For all the investigated provinces, except Tibet, the emission reduction potential of
381 energy saving and fuel substitution exceeded 70 kg /t (Figure 7). In Tibet, the slow diffusion
382 of energy saving technology and the insufficient availability of alternative fuels result in an
383 emission reduction potential of only 46.34 kg/t (Figure 7). In 2030, the fuel substitution rate
384 of the cement industry in Tibet is predicted to be less than 5%, whereas that of Guangdong,
385 Jiangsu, and Zhejiang is predicted to be close to 30%. In these provinces, the emission
386 reduction potential of fuel substitution is 72.06, 71.53, and 76.58 kg/t, respectively (Figure 7).
387 Due to the diffusion of fuel saving technologies, such as energy management and process
388 control systems, combustion system improvement, and low nitrogen decomposing furnace
389 system, the emission reduction potential of these provinces, such as Shanxi, Jiangxi,
390 Guangdong, and Fujian, is predicted to exceed 40 kg/t in 2030. In provinces with a high
391 proportion of ball milling and the third generation grate cooler, such as Shaanxi, Xinjiang,
392 and Inner Mongolia, the emission reduction potential of power saving is evident, and the
393 emission reduction potential of these regions are 25.29, 12.40, and 17.19 kg/t, respectively

394 (Figure 7).

395

396

Figure 7 Emission reduction of energy saving and fuel substitution

397

398 4.3 Emission reduction potential of raw material substitution

399 Utilizing industrial waste as an alternative material has been proven to be economically
400 and environmentally viable for the cement industry (Carvalho et al., 2017; Darsanasiri et al.,
401 2018; Iacobescu et al., 2011; Iacobescu et al., 2013; Maes and Belie, 2017; Taher, 2007; Xu
402 et al., 2012). With the implementation of China's comprehensive utilization of resources, the
403 substitution of raw materials will become the focus of process-related emission reduction in
404 the future. This study only considers substituting steel slag and phosphorus slag with
405 limestone (A1), fly ash, and copper slag for the sandstone or iron materials (A2) in the
406 surveyed regions because significant amounts of these materials are produced in China. The
407 survey data show significant regional differences in the substitution of raw materials, and that
408 the rates of substituting raw materials are affected by the availability and price of substitute
409 materials and their maximum substitution rate (Dai, 2017; Iacobescu et al., 2011; Uson et al.,
410 2013). The studies by Iacobescu et al. (2013), Feng and Li (2010), and Rao (2011) indicate
411 that the amount of steel slag mixed in cement raw materials can reach 10%. The addition of
412 phosphorus slag is mainly affected by P_2O_5 in the clinker. When the content of P_2O_5 in
413 clinker exceeds 0.5%, the compressive strength decreases rapidly (Li and Zhai, 2011). Thus,
414 the optimal concentration of phosphorus slag to be added to the raw meal is 6–10% (Yang et
415 al., 1995). Due to this restriction on the maximum addition amount, the substitution rate of

416 raw materials is obtained via a questionnaire rather than the popularity rate of low-carbon
417 technology. The possible alternative rate of the surveyed province was calculated based on
418 the availability of the technical, economic, and alternative raw material, according to the
419 questionnaire of production technicians, managers of cement plant, and staff of CCA.

420 In the areas with a high proportion of steel slag output to clinker output, such as Hebei,
421 Jiangsu, Liaoning, and Shanxi, the calcium and silica substitution ratios obtained from the
422 questionnaire was relatively high, ranging from 6.21–11.13% and 6.41–10.93% (Table S1),
423 respectively in 2030. The emission reduction of raw material substitution in these provinces
424 is 86.12, 56.23, 59.66, and 59.73 kg/t, respectively. Furthermore, in Hubei, Guizhou, Sichuan,
425 and Yunnan, which have higher phosphorus slag production, the replacement ratios of
426 calcium and silicon raw materials was in the range of 5.46–7.71% and 4.69–7.71% (Table
427 S1), respectively, and their emission reductions are 61.66, 45.16, 52.68, and 46.66 kg/t
428 (Figure 8), respectively. Additionally, other central and eastern provinces yield a lower
429 emission reduction potential due to the constraints of alternative raw materials. This
430 combined with the long transportation distance and high production cost in Xinjiang, Tibet,
431 and other northwest regions, results in a lower raw material substitution rate and reduction
432 potential in these regions (Figure 8).

433

434 Figure 8 Emission reduction of raw material substitution

435

436 4.4 Regional Emission Reductions Potential

437 Regional clinker emission factors in 2030 were calculated based on the assumption and

438 surveyed data of the three emission reduction methods. Figure 9 shows the provincial
439 changes in emission factors for the clinker production. From 2015 to 2030, emissions
440 reduction for all investigated provinces exceed 100 kg/t (Tibet has the lowest reduction
441 potential of 101.41 kg/t). Especially in Jiangsu, Hebei, Sichuan, and Guangdong with the
442 emission reduction potential of 174.60, 169.25, 165.82, and 162.05 kg/t, corresponding to a
443 reduction of 20.58%, 20.16%, 18.91%, and 18.99% respectively. The emission factor in north
444 coast, east coast, and southwest experience the greatest reduction, which is driven by a
445 combination of the use of raw materials and fuel substitution. The national average emission
446 factor estimated in 2030 is 715.33 kg/t, a decline of 16.65% compared to that in 2015. The
447 emissions reduction estimated by Wei et al. (2019) over the period 2018-2030 (158.50 kg/t)
448 were higher than those in the present study (142.85 kg/t), as a higher diffusion rate of raw
449 material (~21%) and fuel substitution (~42%) in 2030 was used. In 2030, the clinker emission
450 factors of the 13 surveyed provinces were higher than the national average level. These
451 provinces are mainly located in northwest, and middle reaches of the Yellow River and the
452 Yangtze River, and Tibet has the highest emission factor of 858.14 kg/t (Figure 9(d)). The
453 remaining 9 provinces have lower clinker emission factors, and they are mainly distributed in
454 east and north coast. In particular, Hebei has the lowest emission factor of only 670.34 kg/t;
455 raw material and fuel substitution each are responsible for over 40% of its total reduction
456 (Figure 9).

457 Across all the investigated provinces, the energy saving and fuel substitution account for
458 approximately 65.98% of the overall reduction in carbon intensity; the remaining two
459 methods contribute about 25.72% and 8.30% each. Structural adjustment has the minimum

460 proportion because only three provinces (Tibet, Chongqing, and Sichuan) have an emission
461 reduction potential of more than 30 kg/t (Figure 9(a)). Emission reductions via improvement
462 in energy saving and fuel substitution are mainly concentrated in Zhejiang, Fujian, Jiangxi,
463 Anhui, Henan, Hunan, Guangdong, Guangxi, and Shaanxi (Figure 9(b)), which account for
464 over 70% of the total reduction in these provinces, and over 45% in the other provinces.
465 Contribution of the emission reduction can differ from place to place. The emission
466 reductions of this method in Inner Mongolia, Shaanxi, Xinjiang, and Tibet are mainly
467 realized via energy saving technologies, while fuel substitution contributes to most of the
468 method reductions in Guangdong, Guangxi, Jiangsu, and Zhejiang. China's major iron and
469 steel producing provinces as well as large phosphorus chemical provinces lead to the high
470 amounts of raw material substitution reduction. Hebei has the largest raw material
471 substitution reduction of 73.20 kg/t (Figure 9(c)), which contributes approximately 43.25% of
472 the total reduction in the region. Five provinces mainly located in northwest China have raw
473 material substitution reduction of less than 30 kg/t, of which Tibet has the smallest reduction
474 of 15.83 kg/t (Figure 9(c)).

475
476

477 Figure 9 Clinker emission factor and emission reduction among provinces by 2030

478

479 In terms of emission composition, fuel-related emissions are the major factor
480 contributing to emission factor reduction, followed by process-related emissions and
481 power-related emissions, which reduce by 93.38, 35.14, and 14.34 kg/t, accounting for

482 65.37%, 24.60% and 10.04% respectively. Jiangsu, Zhejiang, Shandong, Guangdong, and
483 Guangxi have great potential (more than 100 kg/t) for fuel-related emissions. Process-related
484 emission reduction can be mainly attributed to the substitution of raw materials, while shut
485 down backward production facilities contribute a small increase in process-related emission,
486 especially in Guizhou, Hunan, Shandong, and Henan, where the raw material substitution rate
487 of Shaft kiln and small-scale NSP kilns is higher than that of large-scale NSP kilns. And, the
488 provinces, such as Zhejiang, Anhui, Fujian, and Guangxi, have the lowest power-related
489 emissions reduction (7.32-8.57 kg/t), due to the high efficiency of power at current stage.

490 The combined standard uncertainty and expanded uncertainty of our estimate for the
491 emission factor in the provinces are calculated by adopting the methodologies (JCGM, 2010).
492 The combined standard uncertainty of the national clinker emission factor is 30.04, and its
493 relative expanded uncertainties were within the range of 7.64-11.45%. The uncertainties of
494 process-related, fuel-related, and power-related emissions are also estimated at 0.96-1.45%,
495 5.67-8.51%, and 7.63-11.44%, respectively. The uncertainty of emissions factor from Tibet is
496 the smallest (4.54-6.81%), while the uncertainty for Jiangsu, having the lowest emission
497 factor, is the largest (9.27-13.90%), due to alternative fuel use, which is estimated to
498 contribute more to the variance of emissions.

499

500 5. Conclusions

501 As the second largest CO₂ emitter, China's cement industry has attracted much attention
502 in its emissions. However, the status and reduction potential of clinker emission factors at the
503 provincial level were not fully studied. Based on the survey production data, application of

504 energy saving technologies and emissions reduction policy of China's cement industry
505 obtained from the research project, a remarkable disparity in carbon intensity across different
506 production process and scales and regions was noted; and future emission factors from the
507 province's cement industry were forecasted while taking structural adjustment, raw material
508 substitution, and energy saving and fuel substitution into account. The main conclusions that
509 can be drawn from the present study are summarized as follows:

510 (a) The emission factor of the Shaft kiln process is higher than that of the NSP process;
511 for the two processes, the emission factor of small-scale production lines is higher than that
512 of large-scale production lines. This discrepancy is mainly determined by the fuel and power
513 intensity.

514 (b) Regional distribution shows that the emission factors of clinker tend to increase from
515 the eastern coast to the western inland regions. Fujian, Jiangxi, and Henan are areas of higher
516 process-related emissions; fuel-related emission are relatively high in Guizhou, Tibet,
517 Xinjiang, and Yunnan.

518 (c) Structural adjustment leads to the different change trends of process-related , fuel
519 -related and power-related emissions in regional clinker production. It increases the
520 process-related emissions in most of the surveyed provinces, while decreases the fuel-related
521 and power-related emissions of each provinces.

522 (d) Energy saving and fuel substitution is the major method contributing to emission
523 factor reduction. Fuel-related emission reduction in surveyed provinces mainly comes from
524 fuel substitution or energy saving technologies. Power saving technologies makes the
525 power-related emissions reduced in different degree in the investigated provinces.

526 (e) We estimated the national average emission factor in 2030 to be 715.33 kg/t,
527 indicating a decline of 16.65% compared with that in 2015, in which the fuel-related,
528 process-related, and power-related emissions reduced by 93.38, 35.14, and 14.34 kg/t,
529 respectively.

530 The conclusions from these studies, not only gave a comprehensive understanding of
531 regional distribution of clinker emission factors in China, but also show us a clearer picture of
532 the reduction potential of the three methods in the surveyed provinces. These results can be
533 effectively applied to evaluate the relative effectiveness of each emission reduction measure
534 and provide reference for making policies in reaching energy-saving and emission-reducing
535 objectives.

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547 Authors Contributions

548 **Tianming Gao:** Conceptualization, Writing draft preparation, Funding acquisition

549 **Lei Shen:** Software, Visualization

550 **Jianan Zhao:** Methodology, Supervision

551 **Limao Wang:** Investigation, Formal analysis

552 **Litao Liu:** Investigation, Data curation

553 Tao Dai: Reviewing and Editing

554 All authors read and approved the final manuscript.

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562

563 Competing Interests

564 The authors declare that they have no competing interests.

565

566 Availability of data and materials

567 The datasets used during the current study are available from the corresponding author
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Figures

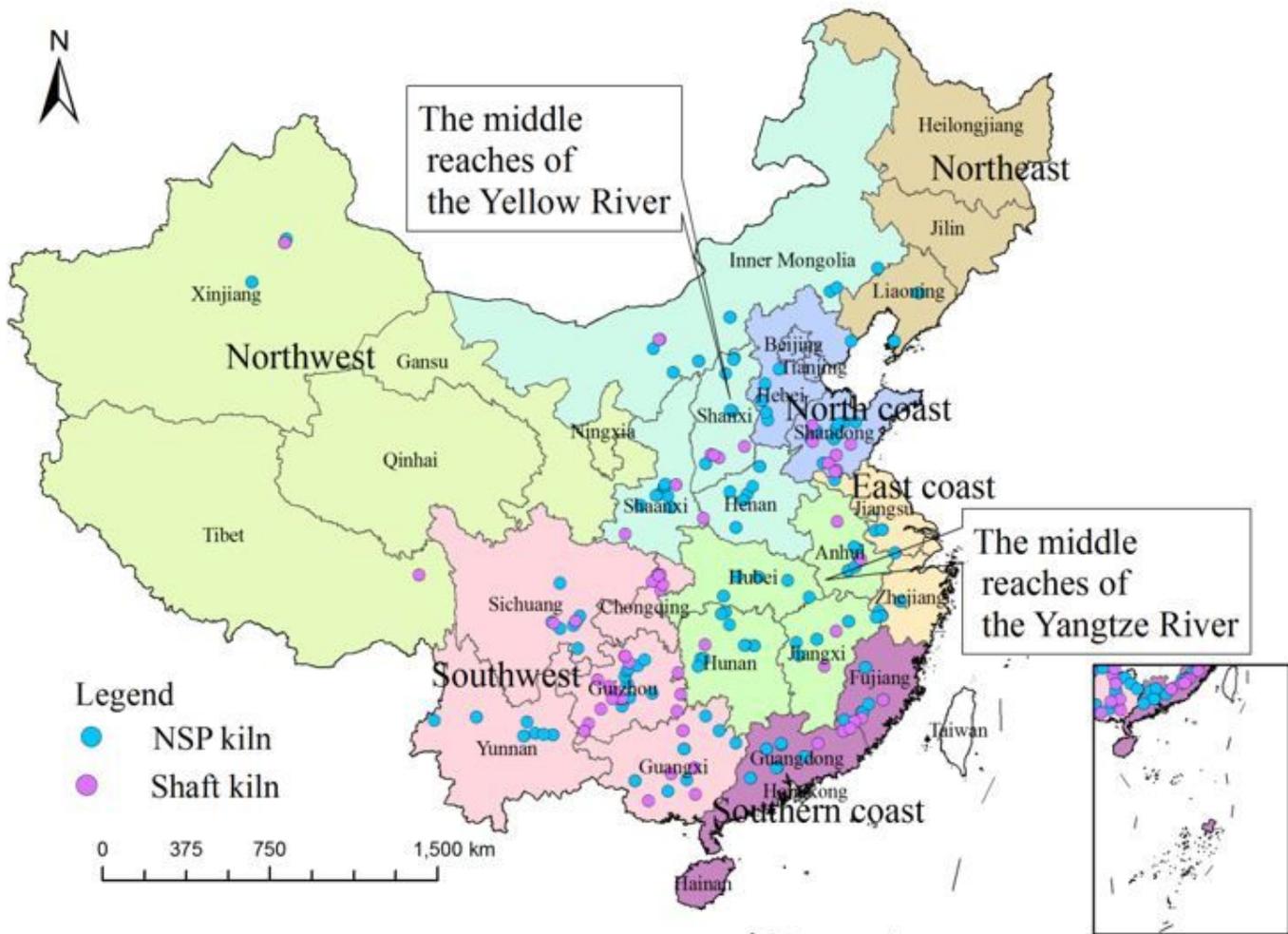


Figure 1

Distribution of the surveyed cement production lines. Note: The designations employed and the presentation of the material on this map do not imply the expression of any opinion whatsoever on the part of Research Square concerning the legal status of any country, territory, city or area or of its authorities, or concerning the delimitation of its frontiers or boundaries. This map has been provided by the authors.

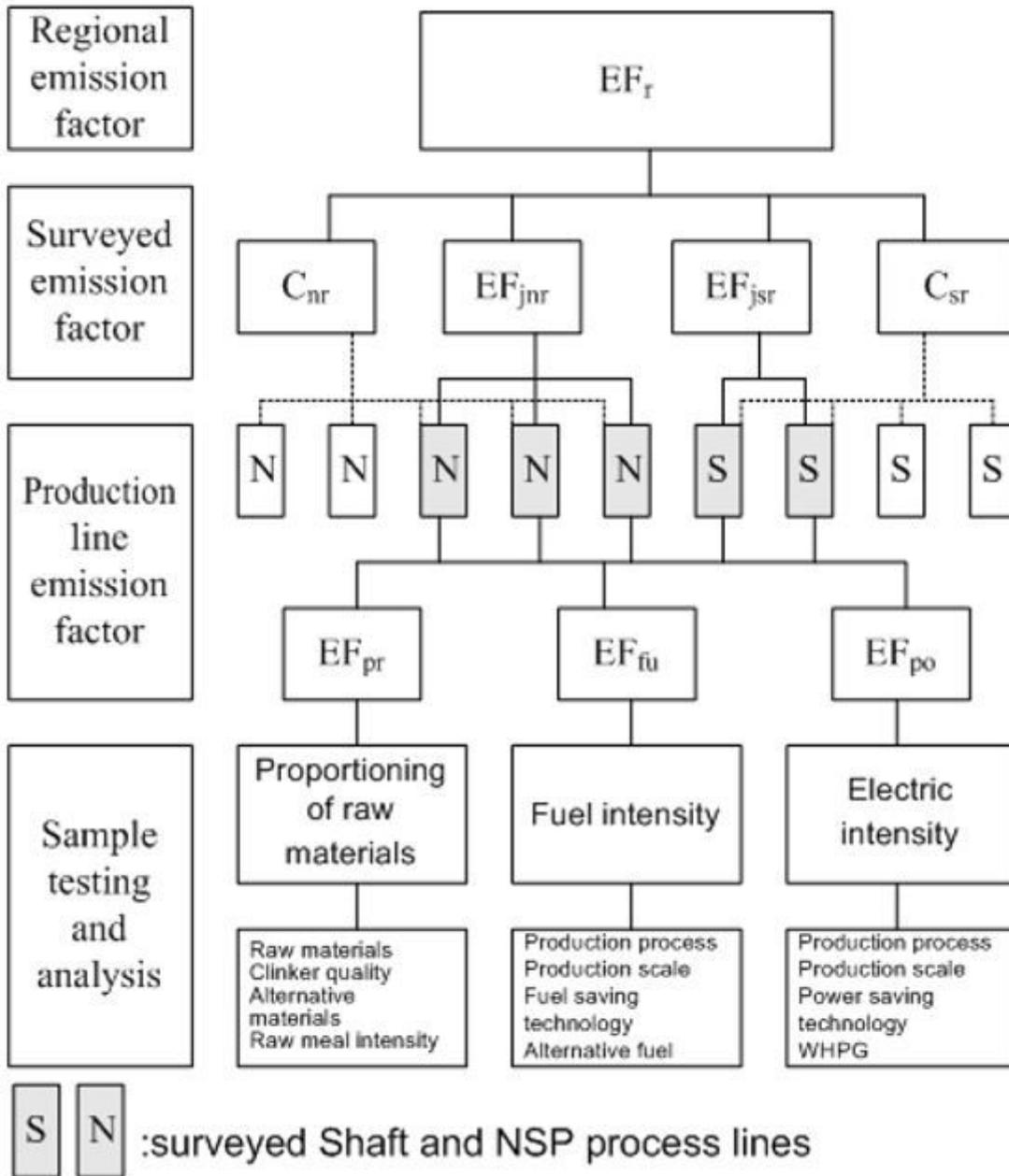


Figure 2

The framework of three tiers regional emission integrated system

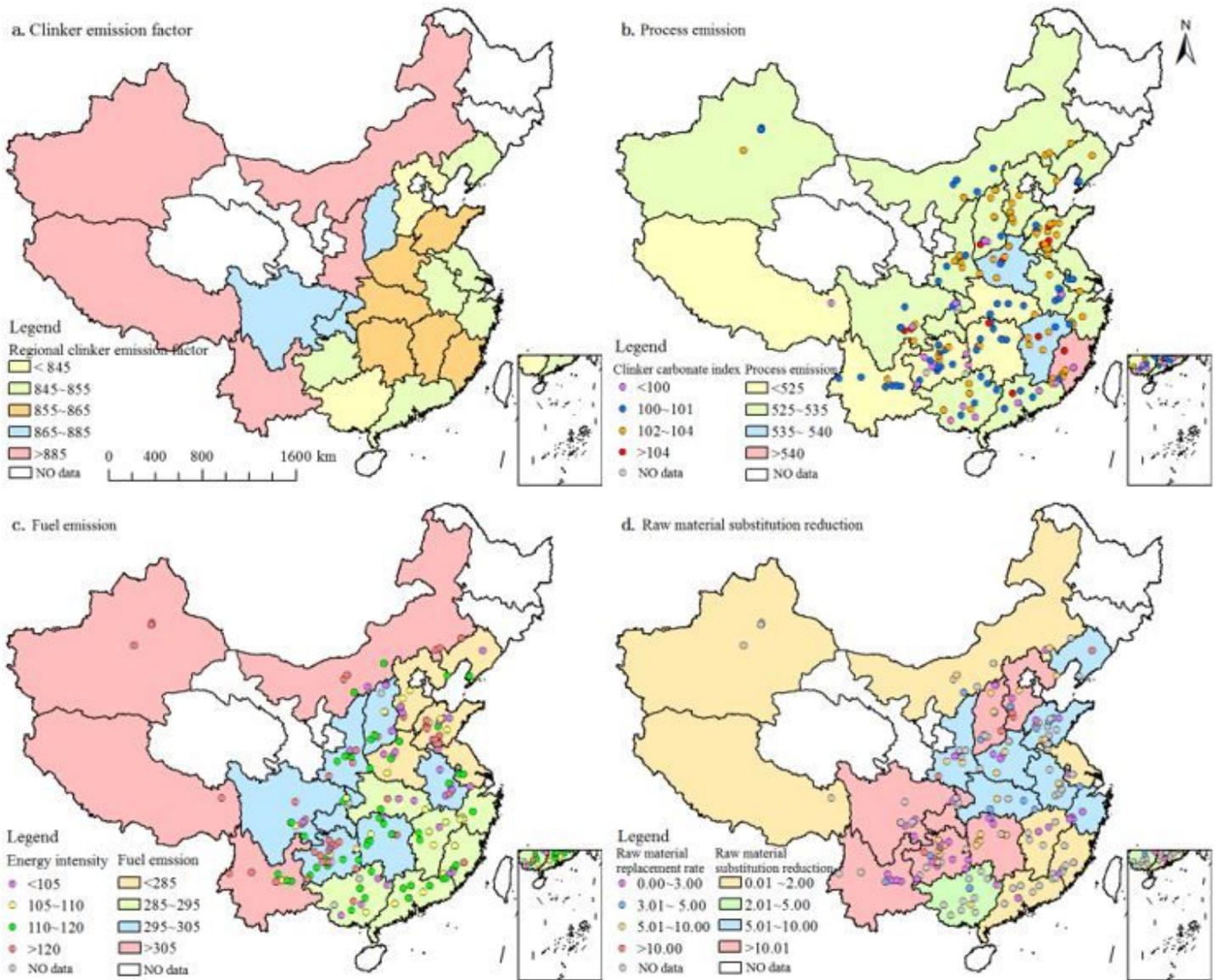


Figure 3

Emission factors and composition of regional clinker in 2015. Note: The designations employed and the presentation of the material on this map do not imply the expression of any opinion whatsoever on the part of Research Square concerning the legal status of any country, territory, city or area or of its authorities, or concerning the delimitation of its frontiers or boundaries. This map has been provided by the authors.

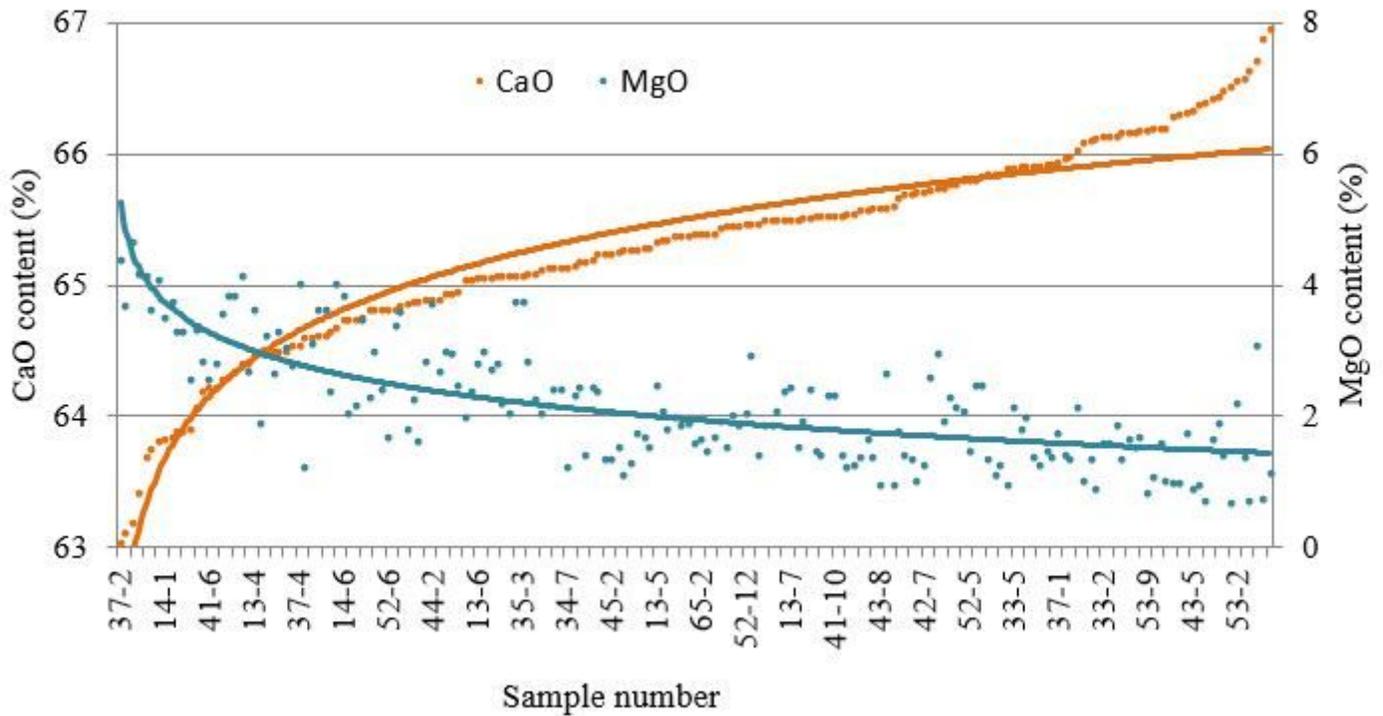


Figure 4

CaO and MgO contents in clinker for the surveyed NSP samples Note: The first two digits of the sample number are the province number, and the others are the provincial sample number.

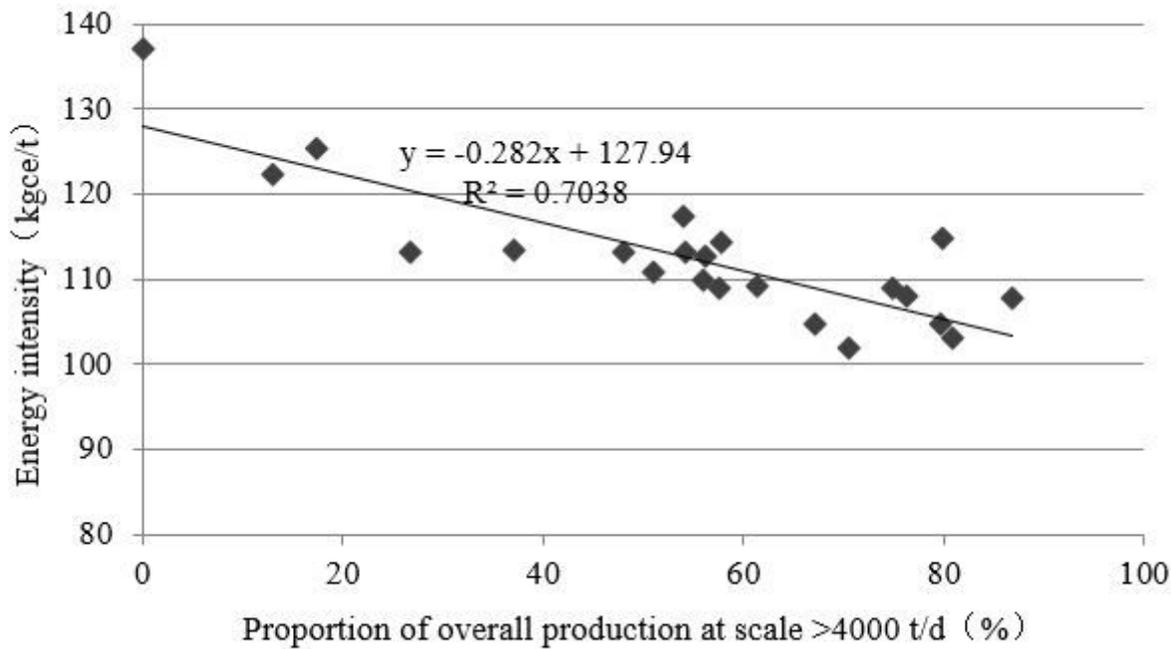


Figure 5

Relationship between regional energy intensity and production scales >4000 t/d

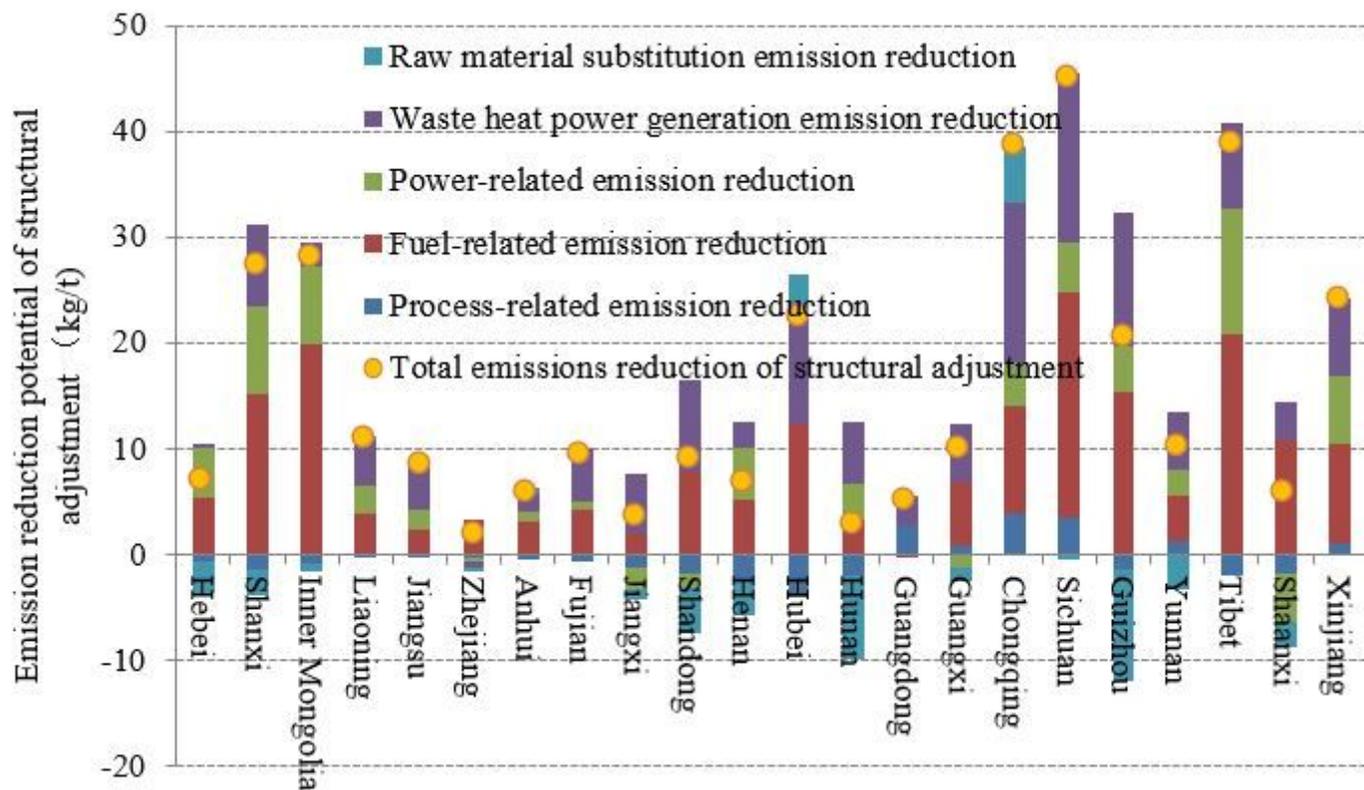


Figure 6

Emission reduction of structural adjustment

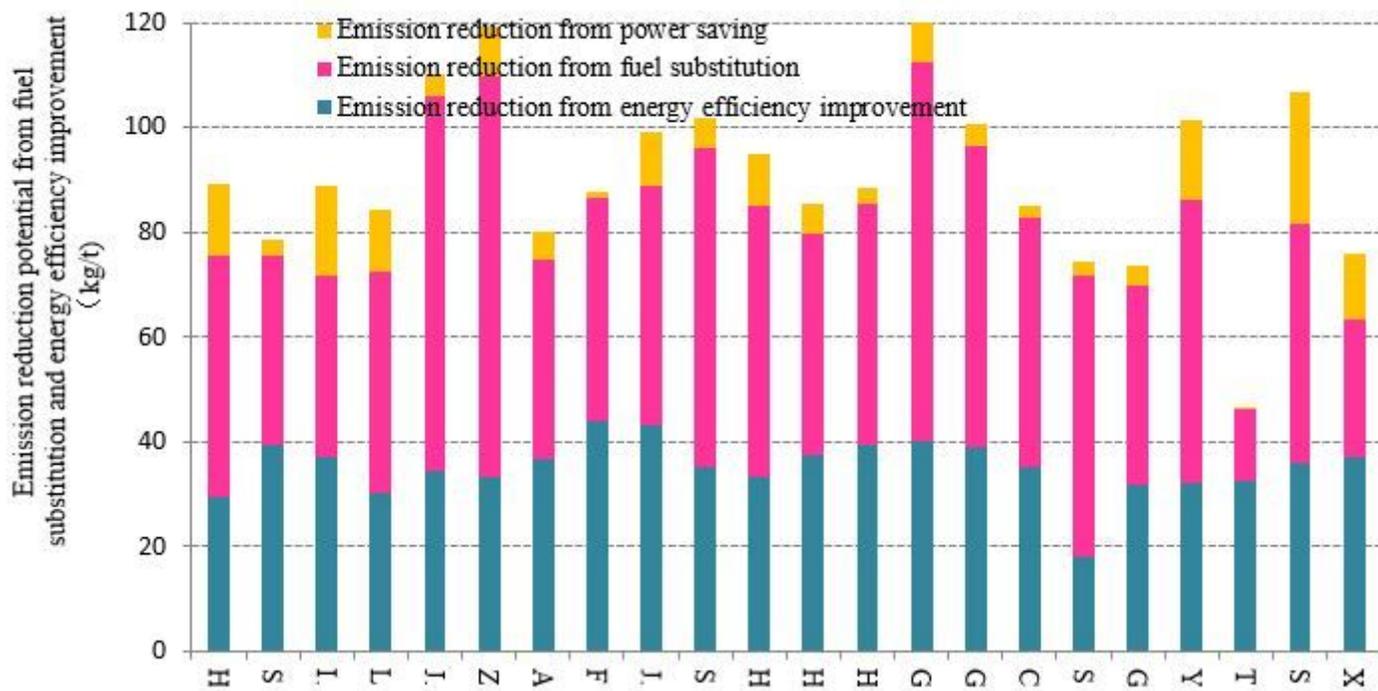


Figure 7

Emission reduction of energy saving and fuel substitution

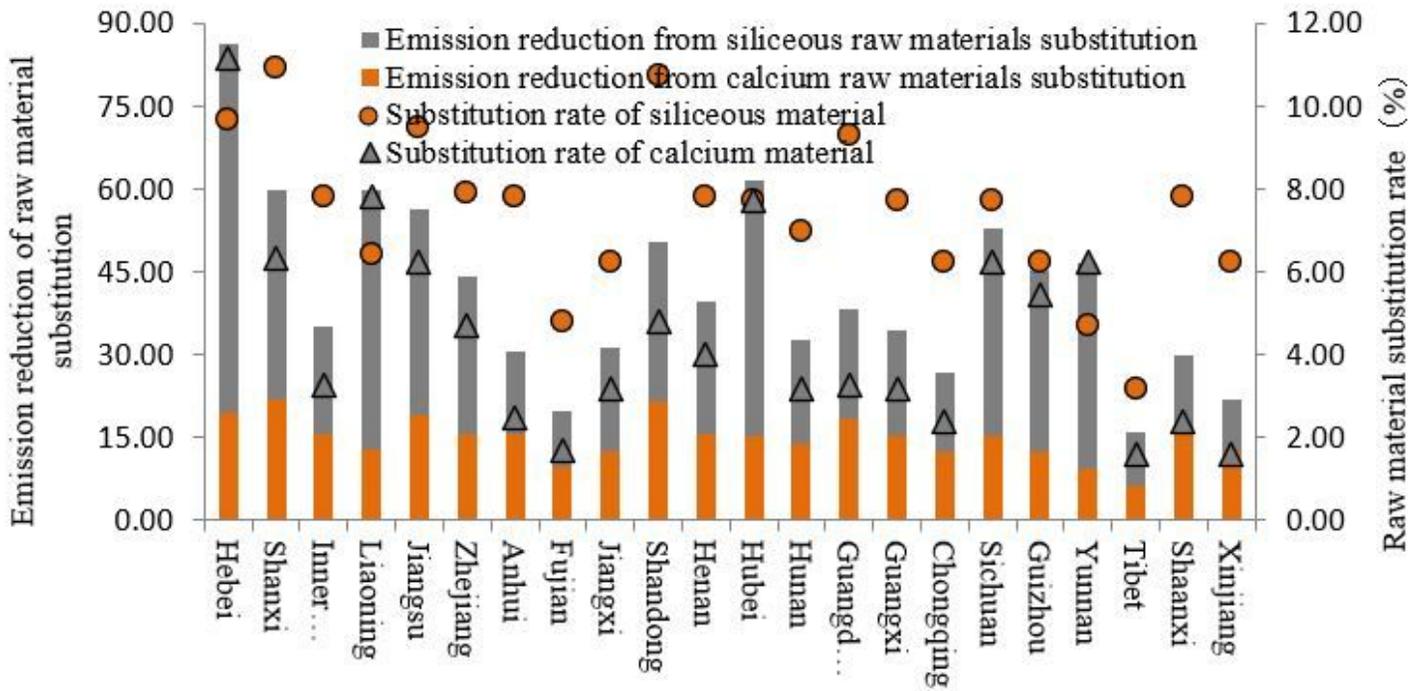


Figure 8

Emission reduction of raw material substitution

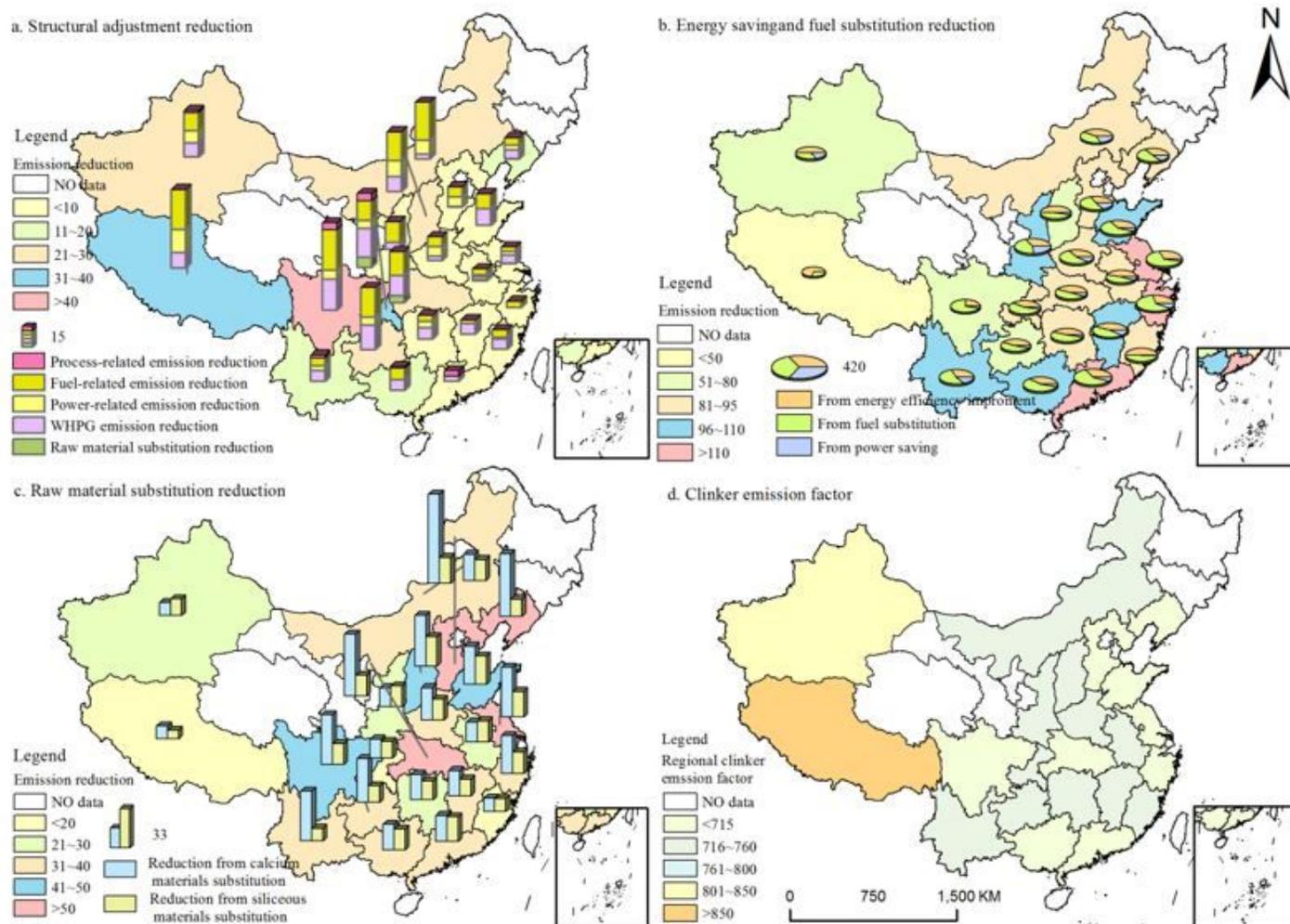


Figure 9

Clinker emission factor and emission reduction potential among provinces by 2030. Note: The designations employed and the presentation of the material on this map do not imply the expression of any opinion whatsoever on the part of Research Square concerning the legal status of any country, territory, city or area or of its authorities, or concerning the delimitation of its frontiers or boundaries. This map has been provided by the authors.

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