

Ultra-efficient Copper Ions Adsorption of Chitosan-montmorillonite Composite Aerogel for Wastewater Treatment

Xin Ye

Nanjing Tech University

Sisi Shang

Nanjing Tech University

Yifan Zhao

Nanjing Tech University

Sheng Cui (✉ cui2002sheng@126.com)

Nanjing University of Technology

Ya Zhong

Nanjing Tech University

Longjin Huang

Nanjing Tech University

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1 **Ultra-efficient copper ions adsorption of Chitosan-**
2 **montmorillonite composite aerogel for wastewater treatment**

3 **Xin Ye^{1,2}, Sisi Shang^{1,2}, Yifan Zhao^{1,2}, Sheng Cui^{1,2*}, Ya Zhong^{1,2*}, Longjin Huang^{1,2}**

4 *1 College of Materials Science and Engineering, Nanjing Tech University, 211800 Nanjing, China*

5 *2 Jiangsu Collaborative Innovation Center for Advanced Inorganic Function Composites, Nanjing*

6 *Tech University, 211800 Nanjing, China*

7 ***Corresponding Author**

8 Sheng Cui

9 E-mail address: *cui2002sheng@126.com*

10 Ya Zhong

11 E-mail address: *yzhong@njtech.edu.cn*

12
13 **Abstract**

14 The modified montmorillonite(MMT) has a two-dimensional stable and ordered lamellar
15 structure. The addition of chitosan(CS) cross-links the two-dimensional sheets to build a three-
16 dimensional network structure with a high specific surface area. We have prepared the best MMT-
17 based water treatment materials that have been reported. This new type of aerogel can efficiently
18 adsorb heavy metal ions in wastewater. The structure and performance of the composite material
19 were characterized in this article. Besides, the adsorption kinetics, adsorption thermodynamics, pH
20 influence, and recycling performance are all focused on. The adsorption equilibrium time of CS-
21 MMT2 is 50 min. The removal rate of Cu²⁺ is as high as 98.21%. The maximum adsorption capacity
22 is 86.95 mg/g. The adsorption process of Cu²⁺ by CS-MMT composite aerogel conforms to the

23 quasi-second-order kinetic model and the Langrangian adsorption isotherm. After three cycles, the
24 removal rate of Cu^{2+} by CS-MMT2 remained above 80%. This article also involves the discussion
25 of the material's adsorption mechanism for Cu^{2+} . This is a kind of environmentally friendly material
26 that can be mass-produced, cheap, efficient, and excellent, which is of great significance to the
27 development of environmental protection.

28 **Key Words:** Chitosan, Montmorillonite, Cu^{2+} , Aerogel, Environmental protection

29

30 **0.Introduction**

31 With the exploitation of natural resources, the pollution of water resources has attracted much
32 attention(Kolpin et al. 2002). In several countries and regions, it has already exceeded the self-
33 purification capacity of natural water bodies, causing much severe environmental pollution(Fu and
34 Wang 2011). Heavy metal pollution is one of the most important parts of the water environment
35 pollution(Jarup 2003; Zou et al. 2016). In industry and agriculture, a large number of pollutants,
36 including heavy metals, have been discharged into rivers, which resulted in deteriorating water
37 quality, causing a series of serious consequences(Bailey et al. 1999).

38 Copper is an essential trace element in the human body. It can be absorbed from various sources,
39 such as food, dust, and water. Excess copper may lead to several symptoms such as nausea, vomiting,
40 abdominal muscle pain. Accumulating in the liver, it may cause various diseases such as
41 gastrointestinal discomfort, kidney, and liver damage (Thit et al. 2020). It can be fatal at extreme
42 doses (Shabbir et al. 2020). Therefore, it is essential to remove residual copper ions in water.
43 Numerous articles have reported various sewage treatment methods containing heavy metal ions,
44 including adsorption, ion exchange, chemical precipitation, membrane filtration, electrochemical

45 techniques, and reverse osmosis (Moja et al. 2020). Compared with others, adsorption is a simple,
46 easy-to-operate, and inexpensive way to treat heavy metal ions (El-Kousy et al. 2020). Recent
47 research has focused on relatively low-cost adsorbents such as zeolite(Bailey et al. 1999),
48 montmorillonite(Uddin 2017), kaolinite clay(Yuan et al. 2015), mineral silicate(Luo et al. 2010),
49 sepiolite(Sun et al. 2013), chitosan polymer(Wan et al. 2010), and biological packing(Wu et al.
50 2004), etc. For any adsorbent, the adsorption capacity is the most critical technical indicator.
51 Especially for low-concentration heavy metal ion wastewater, the development of nano-adsorption
52 materials has a long way to go(Liu et al. 2020b). Therefore, it is necessary to obtain a new adsorbent
53 with high adsorption capacity, high specific surface area, rich pore structure, abundant sources, low
54 cost, good stability, easy recovery, and regeneration.

55 In recent years, aerogel has been considered a promising insulation material. (Baetens et al.
56 2011) At the same time, high porosity and large specific surface area show superior adsorption
57 capacity (Cui et al. 2018). It can be widely used in sewage treatment, air purification, nuclear waste
58 treatment, and other environmental protection fields(Pierre and Pajonk 2002).

59 Modified clay, such as montmorillonite, is regarded as one of the best inorganic adsorbents
60 because of its high utilization rate, friendly environment, and high negative charge, which can
61 absorb positively charged metal ions(Mahouachi et al. 2020). Chitosan is hydrophilic,
62 biocompatible, biodegradable, non-toxic, has high mechanical strength and film-forming properties,
63 as well as antibacterial properties(Li et al. 2016). The amino (-NH₂) and hydroxyl (-OH) groups
64 inherent in the chemical structure of chitosan are the main functional groups that adsorb various
65 heavy metals in water. By modifying the structure of chitosan with -COOH group, its solubility at
66 pH=7 can be increased without affecting the above properties(Wang et al. 2017). Studies have

67 shown that ionic imprinted montmorillonite nanosheet/chitosan (IIMNC) gel beads provide active
68 adsorption sites through their highly efficient cation exchange performance for the selective
69 adsorption of Cu^{2+} (Qin et al. 2020). The montmorillonite-chitosan composite material can provide
70 hydrophobicity and functional groups to improve the performance of montmorillonite in wastewater
71 treatment (Hu et al. 2017b). The interaction sequence of adsorption-related groups in the chitosan-
72 Montmorillonite composite showed that the adsorption capacity of bonding with NH_3^+ may be
73 greater than that of bonding with hydroxyl (Hu et al. 2017a). The clay-chitosan nanocomposite
74 material is expected to achieve a major breakthrough in the treatment of heavy metal ion wastewater.

75 Our team combined chitosan aerogel and montmorillonite designed a new high-efficiency
76 adsorbent. Montmorillonite can be used as ideal templates for preparing nanomaterials and
77 supporting chitosan hydrogels due to their high surface area and stable structure. We use the
78 LiOH/urea dissolving system to process chitosan raw materials and the chemical precipitation
79 method to compound Montmorillonite. By adjusting the concentration of the initial reactants, the
80 size, pore size, and mechanical properties of the synthesized CS-MMT aerogel can be determined.
81 CS-MMT aerogel nanocomposite can be used as an environmentally friendly adsorbent for the
82 removal of Cu^{2+} from wastewater. With its outstanding performance, it will occupy a place in the
83 field of wastewater treatment. The adsorption kinetics and adsorption thermodynamics were used
84 to systematically study the maximum adsorption capacity. The separation and reusability of CS-
85 MMT aerogel were also studied in detail.

86 **1.Experimental Section**

87 **1.1 Fabrication of CS-MMT aerogel**

88 CS-MMT aerogel was prepared by sol-gel method (SI Fig. 1), which could be broken and

89 regenerated at low temperature(SI Section 1). According to the absorbance value of Cu^{2+} in the
90 supernatant, the concentration, adsorption capacity, and removal efficiency of Cu^{2+} in the
91 remaining solution were calculated using equations ((1)-(4)).

$$92 \quad A = 0.0191C + 0.0124 \quad (1)$$

$$93 \quad C_e = \frac{c}{V_1} \quad (2)$$

$$94 \quad Q_e = \left[\frac{(C_0 - C_e) \times V_2}{m_1} \right] \quad (3)$$

$$95 \quad S(\%) = \frac{C_0 - C_e}{C_0} \times 100\% \quad (4)$$

96 Where C is the content of Cu^{2+} in the residual solution after adsorption, which can be calculated
97 by different absorbance values(μg). C_0 before for adsorption of Cu^{2+} solution concentration
98 (mg/mL). C_e for the adsorption of Cu^{2+} solution concentration (mg/mL). m is obtained from the
99 standard curve of Cu^{2+} dosage (μg). m_1 as adsorbent dose (mg). V_1 for copper water volume (mL).
100 Q_e for equilibrium adsorption Cu (II) the amount of (mg/g). V_2 to join Cu^{2+} the volume of a solution
101 (mL), S for Cu^{2+} removal efficiency.

102 **2. Results and discussion**

103 **2.1 XPS, XRD and FT-IR analysis**

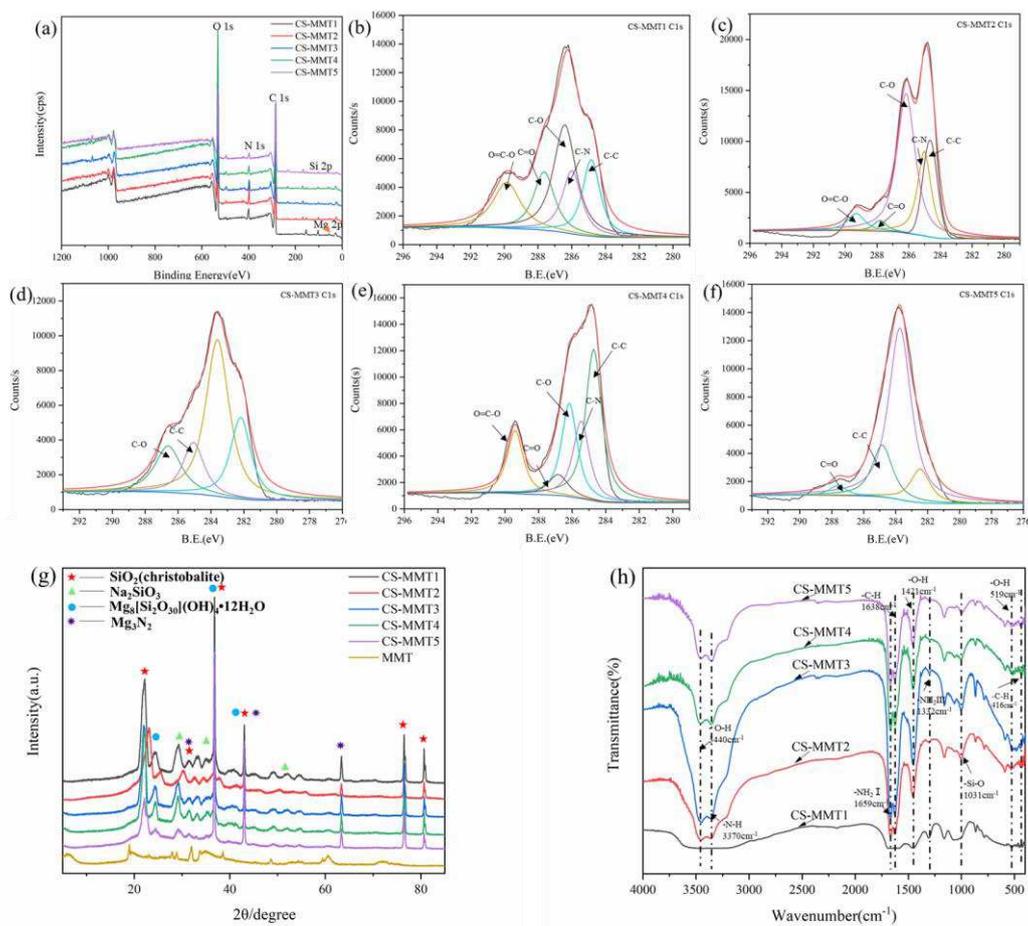


Fig. 1 (a) XPS survey spectrum, (b~f) C1s spin-orbit peaks of CS-MMTs, (g) XRD patterns of CS-MMTs and MMT, (h) FTIR spectra of CS-MMTs

To further analyze the interaction between MMT and CS matrix, X-ray photoelectron spectroscopy is used to analyze the surface chemical state of different components of CS-MMTs (Fig. 1(b~f)). The C spectrum shows that the CS-MMTs are rich in C-C, C=O, and C-O/C-N groups. The peak at 283.9 eV corresponds to the binding energy of the C-N bond, and the peaks of C=O and C-C mainly appear at about 286.8 eV and 285.0 eV. (Wu et al. 2020). Compared with other components, the C1 s of CS-MMT2 is significantly enhanced, and the intensity of the C-N peak is also significantly increased, proving that MMT was successfully modified by CS. The O1s (531.8 eV) peak is attributed to hydroxyl oxygen, indicating that siloxane was successfully converted to hydroxyl oxygen. (Guo et al. 2019). However, the N-element XPS spectrum showed a certain

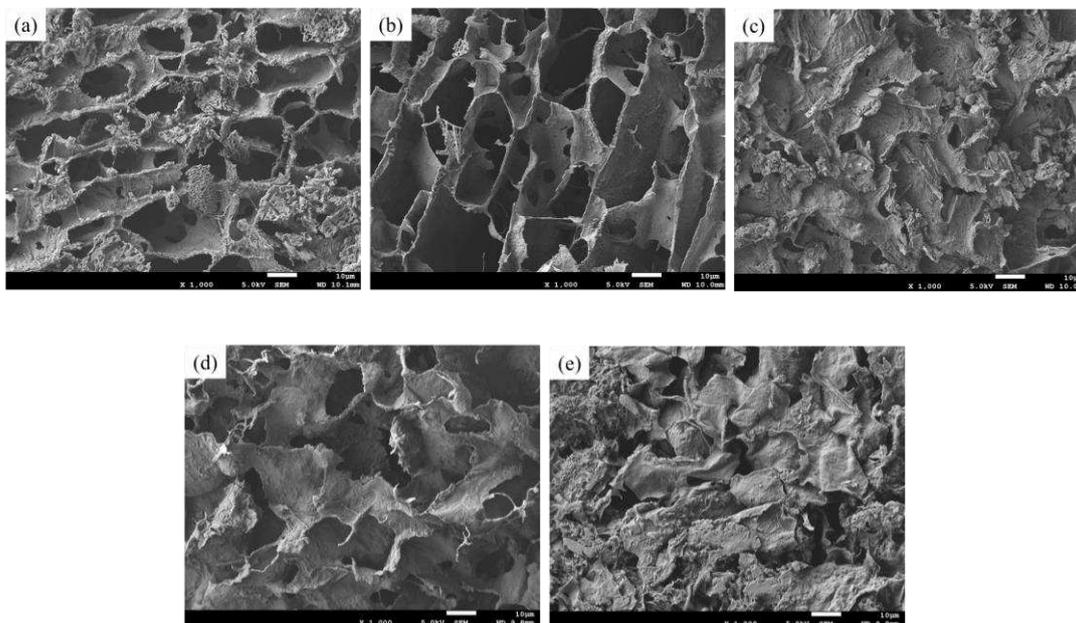
116 amount of C-N and N-O in CS-MMT. Besides, peaks of Si 2p(103eV) and Mg2p (55eV)
117 corresponding to MMT were found in the spectrum of CS-MMT2(Akar et al. 2020), indicating the
118 successful combination of CS and MMT.

119 The crystal phase and phase composition of CS-MMT composite aerogel were analyzed by
120 XRD (Fig. 1 (g)). According to extensive studies, MMT is a clay mineral in the form of octahedral
121 crystals. The characteristic spikes show high crystallinity of nano-MMT at 5.9°, 19°, and 26°.
122 According to the Scherer equation, we can easily calculate that the interlayer spacing of nano-
123 montmorillonite is 1.593nm. In contrast, the XRD patterns of the chitosan-modified
124 montmorillonite composite aerogels have obvious changes, indicating that the crystal structure of
125 the composite aerogel has changed significantly. By comparing and analyzing the standard spectra,
126 we can infer that the crystal structure of CS-MMT is roughly the same as that of square quartz.
127 Besides, we can notice that there is a broad peak at 29°. As the ratio of nano-montmorillonite
128 increases, the peak intensity gradually decreases. We attribute this to the cation exchange between
129 the protonated amino groups of chitosan and the cations in the montmorillonite layer, which results
130 in such formation of disordered exfoliation structures.

131 As shown in Fig.1(h), the hydroxyl groups of amide or hydroxide combined with amino groups
132 in montmorillonite and chitosan mainly appear at about 3400 cm⁻¹, which is consistent with previous
133 reports (Liu et al. 2020a). The amide stretching of the chitosan-modified montmorillonite composite
134 appeared at about 1640 cm⁻¹. All the CS-MMT composite aerogels showed characteristic silica sharp
135 bands around 1031 cm⁻¹. The bending vibrations of Si-O-Mg and Si-O-Si appeared at 1527 cm⁻¹ and
136 470 cm⁻¹, respectively(Zhu et al. 2017). There are strong peaks at 3440 cm⁻¹ and 1659 cm⁻¹, which
137 correspond to the overlapping stretching vibrations of hydroxyl and amino groups and the stretching

138 vibrations of carboxyl groups. The peak at 1332 cm^{-1} is the stretching vibration of amino. The FT-
139 IR is a great proof of the successful preparation of CS-MMT composite aerogels.

140 2.2 SEM analysis



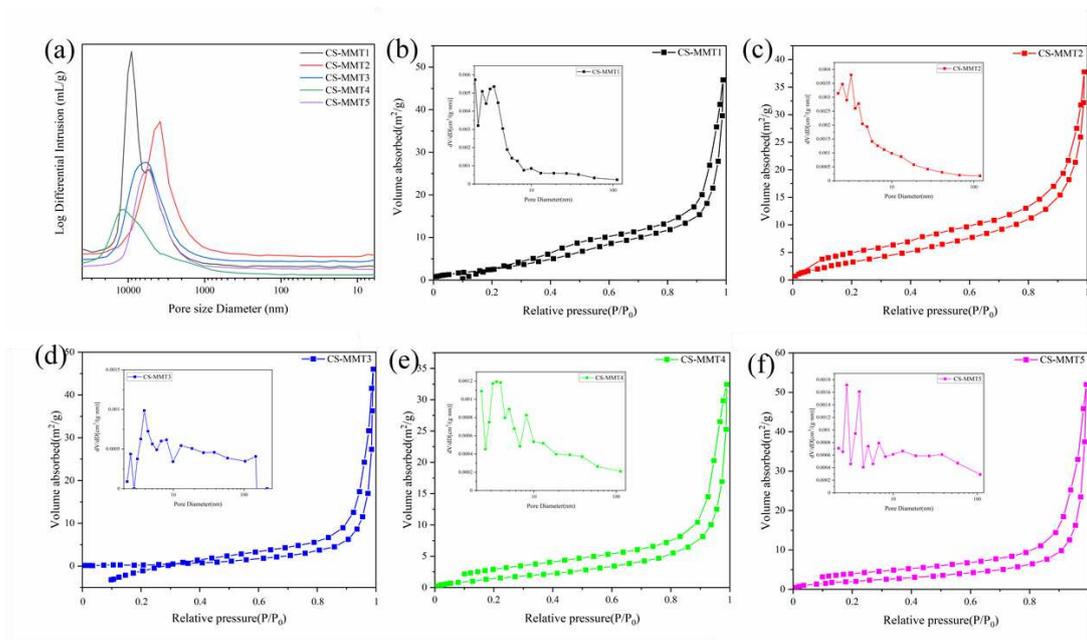
141

142

143 **Fig. 2** SEM image of (a)CS-MMT1, (b)CS-MMT2, (c)CS-MMT3, (d)CS-MMT4, (e)CS-MMT5

144 As is shown in Fig. 2 , the CS-MMT composite aerogels of different components all present a
145 clear layered structure. CS connects the montmorillonite laminates in a fiber-like cross-linking
146 manner. As shown in Fig. 2(b), CS-MMT2 composite aerogel has a large number of thin-layer
147 structures and a certain interlayer space. The doping of CS cross-links the layers of MMT to build
148 a micron-level pore structure. This three-dimensional porous structure can greatly increase the
149 specific surface area of the material, laying a solid foundation for the adsorption of heavy metal
150 ions. The C, N, O, and Si elements of CS-MMT2 aerogel are uniformly distributed. It contains a
151 large number of oxygen-containing functional groups, oxides, and nitrides. By TG analysis, CS-
152 MMT showed higher degradation temperature (SI Section2, Fig. S2, and S3).

153 2.2 The Specific surface area and pore size analysis



154

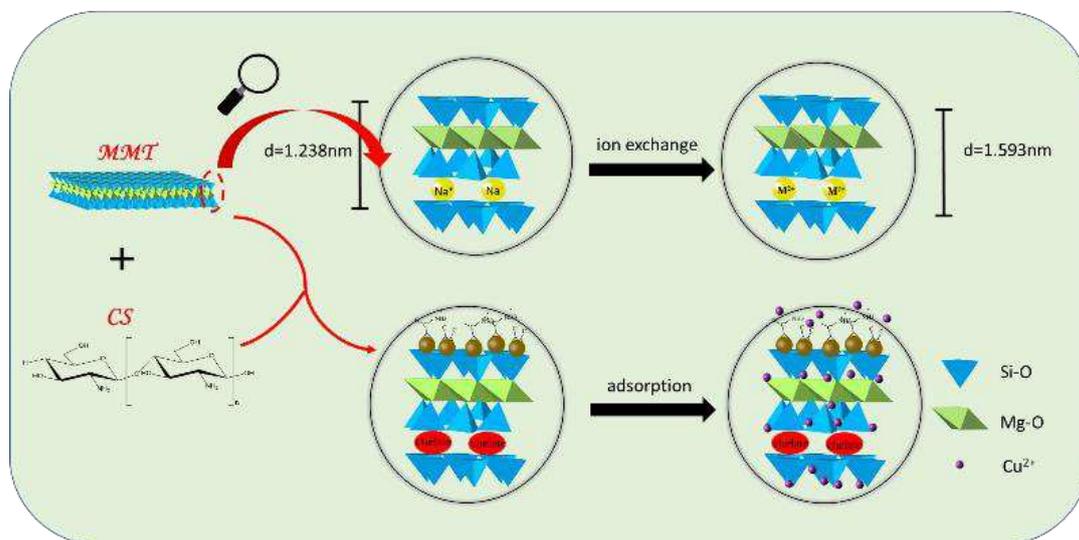
155 **Fig. 3** (a) Mercury intrusion test, (b~f) N₂ Adsorption–desorption isotherms and pore size

156 distributions of CS-MMT composite aerogels with different components

157 The shape of this type of hole should be a slit-like tubular capillary structure with both ends
 158 open. (Zhao et al. 2020). As is shown in Fig.3(a~f), when the relative pressure P/P_0 was 0-0.6, the
 159 N₂ adsorption increased slowly, which proved that there were a few micropores in the samples. In
 160 this stage, nitrogen molecules are gradually adsorbed from monolayer to multilayer into the porous
 161 structure. Subsequently, when the relative pressure was between 0.6-0.99, the nitrogen adsorption
 162 capacity at the high-pressure end increased sharply, indicating the presence of mesopore and
 163 macropore in the material. According to the physical property parameters in SI Table S1, the specific
 164 surface area of CS-MMT2 is relatively high, which is 14.133m²/g, and the pore diameter distribution
 165 was uniform. Combined with SEM images(Fig.1(b)), it was found that there was a stable three-
 166 dimensional network structure. (Chu et al. 2019). Because the nitrogen adsorption and desorption
 167 specific surface and pore size test can be widely used in the test of porous materials with mesoporous
 168 structure, it is only of reference value for macroporous materials. We conducted mercury intrusion

169 tests on all five CS-MMT samples. According to the analysis of mercury injection pore size, the
 170 pores are mainly mesopore and macropore. The average pore sizes of composite aerogels are
 171 1242.89 nm(CS-MMT1), 889.29 nm(CS-MMT2), 1012.60 nm(CS-MMT3), 3942.97 nm(CS-
 172 MMT4), and 1159.19 nm(CS-MMT5). The pore size ranges from 50-100 μm . Undoubtedly, large
 173 specific surface area and porous structure provided more abundant active sites for rapid absorption
 174 of Cu^{2+} .

175 2.3 Structure model and adsorption mechanism



176

177 **Fig. 4** Schematic diagram of structure model and adsorption mechanism

178 According to detailed characterization data, we can get the following material structure model
 179 (Fig. 4). The montmorillonite layer is modified by organic quaternary ammonium salt ions.
 180 Combined with the XRD diffraction pattern, it can be seen that the interlayer spacing of
 181 montmorillonite increases from 1.236 to 1.593 nm. XRD, FT-IR, and XPS are used for detailed
 182 elements and surface state analysis. We speculate that strong alkali treatment and ball milling break
 183 the Si-O and Mg-O bonds of the unit cell in montmorillonite. The amino groups of chitosan cross-
 184 link the active bonds and the -OH and amino groups carried by the surface sheet, forming such CS-

185 MMT composite aerogel. Cu^{2+} will not only chelate and react with the broken bonds on the surface
186 of CS-MMT, at the same time, the pores and lamella structure of the aerogel are also positively
187 related to the adsorption capacity.

188 **2.4 Adsorption property of CS-MMTs towards Cu^{2+}**

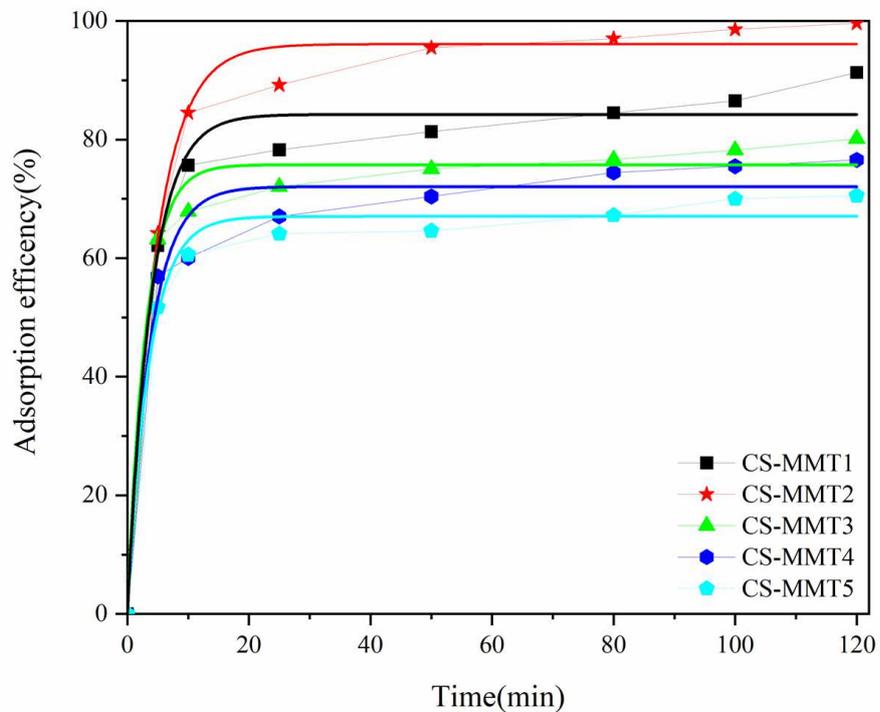
189 **Adsorption equilibrium time and adsorption kinetics**

190 CS-MMT composite aerogel adsorbed Cu^{2+} at different times (SI Section 2). The adsorption rates
191 of the five adsorbents at different times are sorted in Fig. 5. It can be seen from the adsorption curve and
192 adsorption efficiency that all adsorbents have a high adsorption efficiency of Cu^{2+} in the initial stage of
193 the adsorption reaction. As time goes on, the adsorption process slows down until an adsorption
194 equilibrium is finally reached. In the initial stage of adsorption in this experiment, the Cu^{2+} adsorption
195 material causes concentration diffusion in the solution, and the active adsorption sites on the
196 montmorillonite can combine with Cu^{2+} at different adsorption rates and numbers. As the concentration
197 gradient of Cu^{2+} decreases, the adsorption behavior on the adsorbent gradually decreases, which in turn
198 causes the adsorption rate to slow down. At the same time, as the chelation reaction progresses, the
199 adsorption sites are gradually occupied, and the adsorption reaction reaches equilibrium. From the
200 relationship between adsorption time and adsorption efficiency, it can be seen that the adsorption
201 efficiency of different components of CS-MMT composite aerogel on Cu^{2+} is not significantly different
202 under the same experimental conditions. The adsorption efficiency of CS-MMT2 can be around 99%,
203 and the adsorption equilibrium was reached at about 50min, indicating that CS-MMT2 composite aerogel
204 has a strong adsorption capacity for Cu^{2+} . To elucidate the adsorption kinetics of CS-MMT composite
205 aerogel on Cu^{2+} , the quasi-first-order model (Eq5) and quasi-second-order model (Eq6) were used to fit
206 the adsorption data of Cu^{2+} as follows:

207
$$\lg(Q_e - Q_t) = \lg Q_e - \frac{k_1 t}{2.303} \quad (5)$$

208
$$\frac{t}{Q_t} = \frac{t}{Q_e} + \frac{1}{k_2 Q_e^2} \quad (6)$$

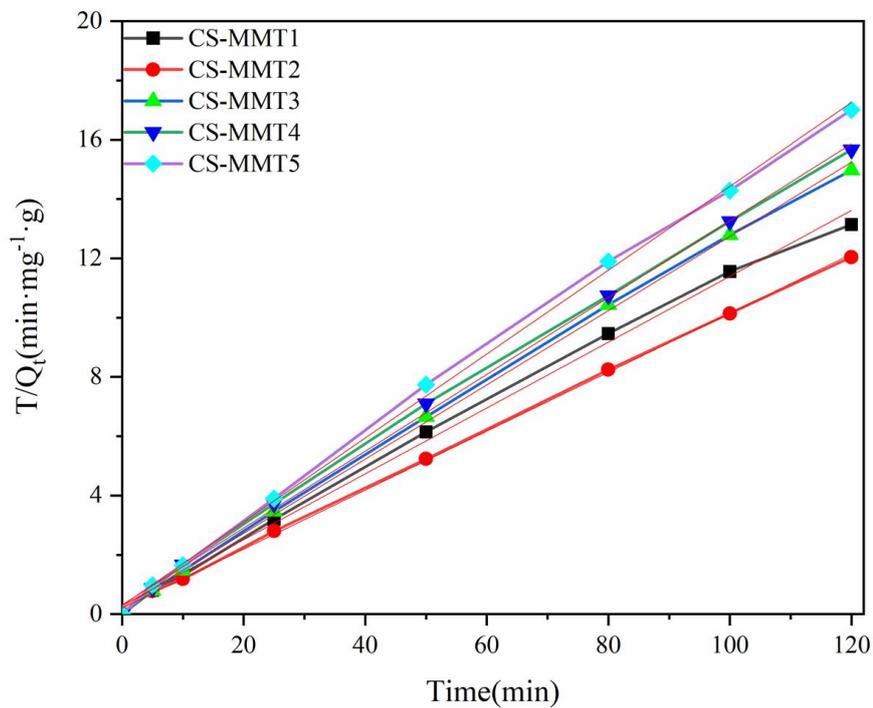
209 Where Q_e is the adsorption amount when the adsorption process reaches adsorption equilibrium
 210 (mg/g), Q_t is the adsorption amount corresponding to t at a certain moment in the adsorption process
 211 (mg/g), k_1 and k_2 are pseudo-first-order (min^{-1}) and pseudo-second-order rate constants ($\text{g}/(\text{mg} \cdot \text{min}^{-1})$),
 212 respectively.



213
 214 **Fig. 5** Adsorption efficiency of CS-MMT composite aerogels for Cu^{2+} at different time

215 The parameters of the two models are shown in SI Table S2. By comparing the correlation
 216 coefficient R^2 of the two models, it is obvious that the pseudo-second-order kinetic equation has a good
 217 correlation with the adsorption process. This indicates that the adsorption rate is more related to the
 218 number of unoccupied active sites. As the layered space of montmorillonite increases, the porosity
 219 increases. This will directly increase the number of active sites and increase the adsorption capacity

220 of Cu^{2+} . Based on the quasi-second-order kinetic equation model, the quasi-second-order rate equation
 221 of CS-MMT was shown in Fig. 6. According to the fitting curve of CS-MMTs, the corresponding
 222 correlation coefficient R^2 was calculated. The R^2 of CS-MMTs indicates that the adsorption of Cu^{2+} in
 223 this system follows a pseudo-second-order kinetic model. CS-MMT2 has the best fit, which can reach
 224 0.99937, while Q_e is 9.906 and K_2 is 0.04942.



225
 226 **Fig. 6** Pseudo-second-order model for adsorption of Cu^{2+} on CS-MMT composite aerogels

227 **Adsorption isotherms and adsorption thermodynamics**

228 As can be seen from Fig. 7, the equilibrium adsorption capacity of the three adsorbents to copper
 229 ions increased with the increase of the initial concentration of Cu^{2+} solution until the adsorption was
 230 completed. When Cu^{2+} solution is at a low concentration, the adsorption curve rises very fast. When the
 231 initial concentration increases to a certain concentration, the increasing rate of adsorbent adsorption
 232 gradually slows down, and the curve flattens out to near saturation adsorption.

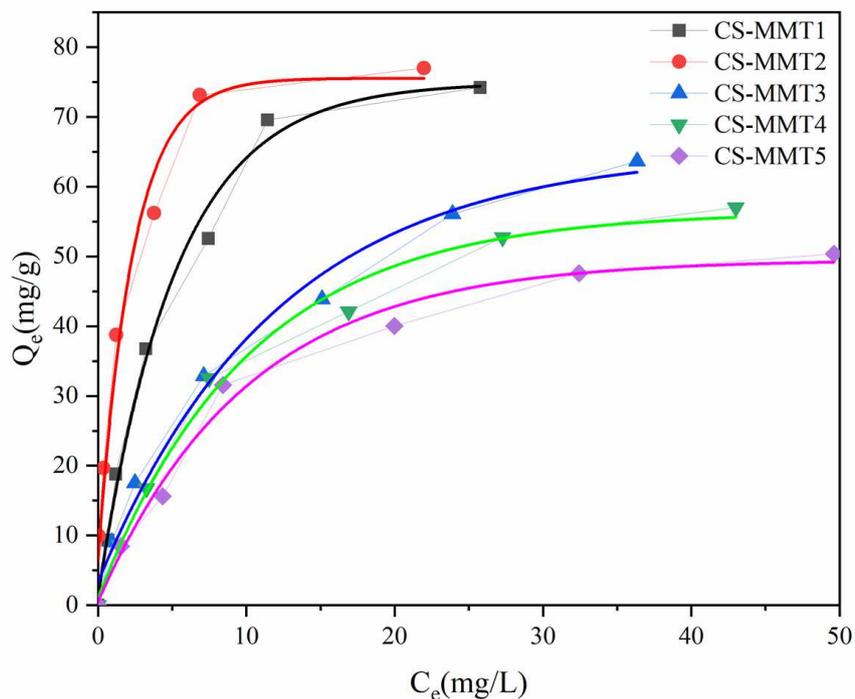
233 In order to obtain the thermodynamic mechanism of adsorption of Cu^{2+} by CS-MMT composite
234 aerogels at different initial concentrations, the Langmuir equation and Freundlich equation were used
235 to fit the experimental data. The Langmuir isotherm model (Eq. 7) and the Freundlich isotherm model
236 (Eq. 8) are as follows:

$$237 \quad \frac{C_e}{Q_e} = \frac{C_e}{Q_m} + \frac{1}{K_L Q_m} \quad (7)$$

$$238 \quad \ln Q_e = \ln K_F + \frac{1}{n} \ln C_e \quad (8)$$

239 Where Q_e is the adsorption amount of Cu^{2+} per unit mass of adsorbent at the adsorption equilibrium
240 (mg/g) Q_m represents the maximum adsorption capacity of Cu^{2+} (mg/g), and C_e represents the equilibrium
241 concentration of Cu^{2+} after adsorption (mg/L), where K_L and K_F are the adsorption equilibrium constants
242 of Langmuir isotherm equation and Freundlich isotherm equation, and $1/n$ is the heterogeneity factor.

243 As can be clearly seen in Fig.10, the equilibrium adsorption capacity of the five adsorbents on Cu^{2+}
244 showed a different increasing trend with the increase of the initial concentration of Cu^{2+} solution. When
245 Cu^{2+} is at a low concentration, the adsorption curve rises very fast. When the initial concentration
246 increases to a certain concentration, the increasing rate of adsorbent adsorption volume gradually slows
247 down, and the curve flattens out until it approaches saturation adsorption.

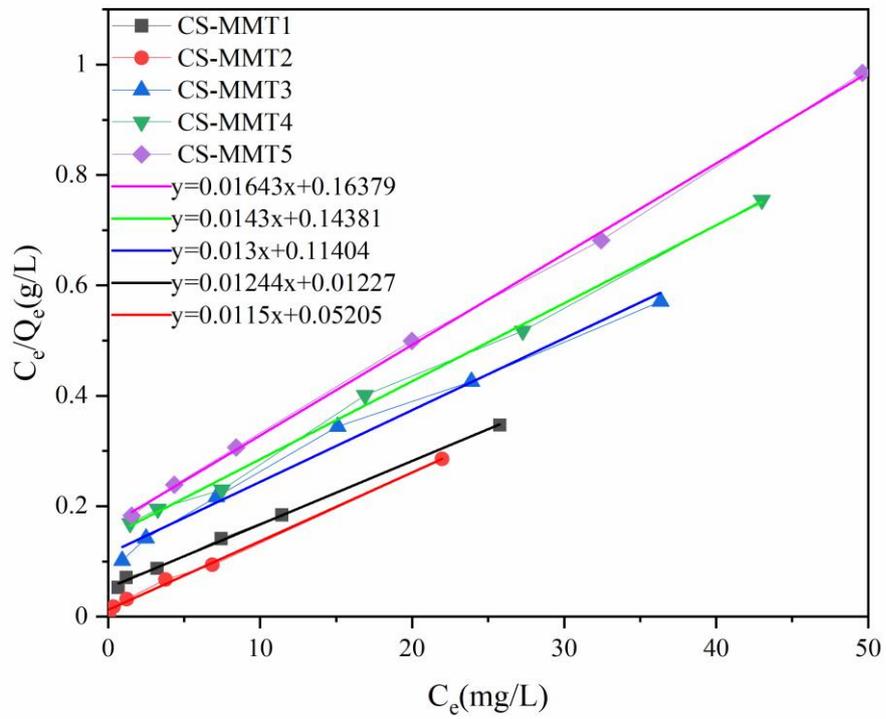


248

249 **Fig. 7** Adsorption capacity of CS-MMT composite aerogels for Cu^{2+} in different initial solution
 250 concentration

251 The adsorption data were correspondingly fitted to the Langmuir model. The fitting equations of
 252 CS-MMTs are shown in Fig.8. The Langmuir and Freundlich parameters for adsorption of Cu^{2+} are
 253 sort in SI Table S3, respectively. The maximum adsorption capacity of the CS-MMTs on Cu^{2+} can be
 254 estimated according to the Langmuir adsorption model. Under the optimal conditions, the maximum
 255 adsorption capacities of CS-MMT composite aerogels for Cu^{2+} were estimated to be 80.38 mg/g,
 256 86.95mg/g, 76.92mg/g, 69.93mg/g, and 60.86mg/g, respectively. Compared with the montmorillonite
 257 adsorbent reported in the literature, the CS-MMT2 composite aerogel prepared by this experiment
 258 showed quite outstanding adsorption capacity for Cu^{2+} (Table 1). The digital image of the adsorption
 259 effect is shown in SI Fig. 4. With the addition of the adsorbent for about 30s, the blue complex will
 260 initially form until it becomes stable. After adsorption and centrifugation, the solution was obviously

261 close to the colorless state.



262

263 **Fig. 8** Langmuir isotherms plots of Cu^{2+} adsorption on CS-MMTs

264

Table 1 Comparison with other published work

Researchers	Adsorbent	Cu^{2+} Adsorption performance	Ref.
Chen et al.	Fe_3O_4 -CS/EDTA	36.52 mg/g	(Chen et al. 2019)
Chu et al.	Arg-Mt	29.15 mg/g	(Chu et al. 2019)

Chu et al.	Histidine modifies montmorillonite	30.72 mg/g	(Chu et al. 2020)
Chu et al.	Montmorillonite was modified by diamino surfactant-containing tetra ammonium cation	29.03 mg/g	(Chu et al. 2020)
Zhang et al.	Magnetic bentonite/carboxymethyl chitosan/sodium alginate hydrogel sphere	56.79 mg/g	(Zhang et al. 2019)
Wu et al.	Magnetic polysaccharide/Go @Fe ₃ O ₄ gel beads	55.96 mg/g	(Wu et al. 2019)
This work	CS-MMT	86.95 mg/g	

265

266 **2.5 The influence of pH on the adsorption effect**

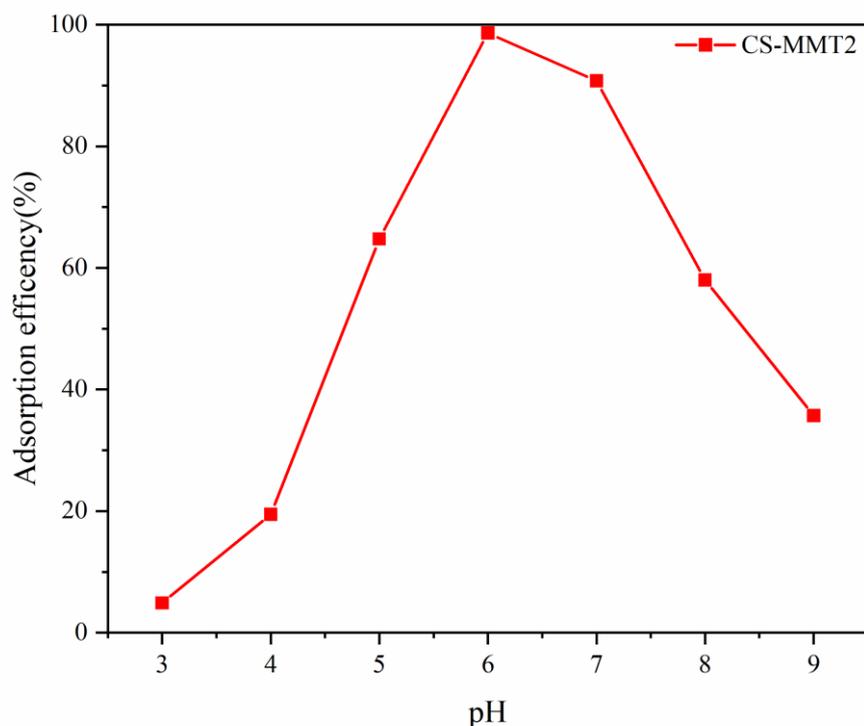


Fig. 9 The effect of solution pH value on the adsorption capacity of Cu^{2+} by CS-MMT2

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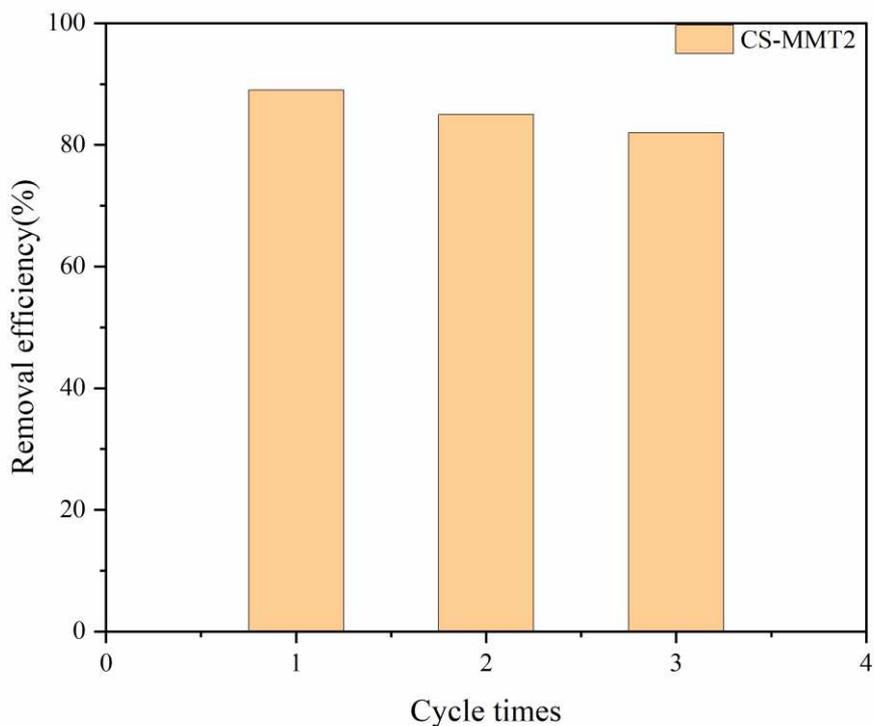
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279

It can be seen from Fig. 9 that the Cu^{2+} adsorption capacity of CS-MMT2 increases first and then decreases with the increase of the pH of the solution. When the solution is in a low pH range (3.0~4.0), the adsorption effect of CS-MMT2 is limited. This phenomenon can be interpreted as the presence of excess hydrogen ions under a low pH condition, which limits the reaction of chitosan and NH_3 groups. Besides, CS is easy to dissolve in an acidic environment, which is the main reason for the limited adsorption capacity. When the pH value is higher than 4, the adsorption capacity of Cu^{2+} increases significantly. As the pH of the solution increases, the number of hydrogen ions decreases, and protonation increases. There will be more adsorption sites on the surface of CS-MMT2, and the adsorption capacity will also show an increasing trend. At pH = 6, the adsorption capacity reached its peak and then began to decline. Cu^{2+} will form $\text{Cu}(\text{OH})_2$ precipitation at higher pH values. This shows from the side that the adsorption capacity of CS-MMT2 for Cu^{2+} decreases. Therefore, pH = 6 is the best service condition for

280 CS-MMT. Compared with the montmorillonite adsorbent reported in the literature, the CS-MMT
281 composite aerogel prepared in this experiment has a better adsorption capacity for Cu^{2+} .

282 2.6 Adsorption cycle performance



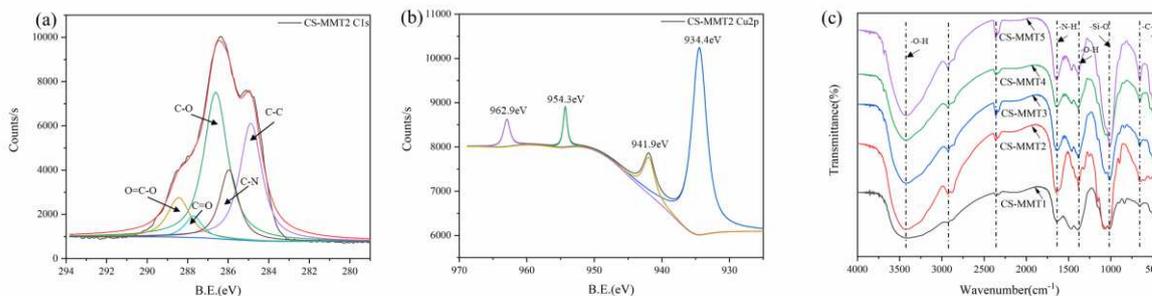
283

284 **Fig. 10** Circulation capability of CS-MMT2

285 The regeneration efficiency of adsorbent is an important index to measure the actual use effect of
286 an adsorbent. Generally speaking, effective adsorbent materials not only have high adsorption capacity
287 but also have a great reuse effect. In this cycle experiment, we used a 0.2 mol/L EDTA solution as the
288 desorption solution. The three-cycle effect of CS-MMT2 was measured. The experimental results are
289 shown in Fig. 10. After 3 times of adsorption-desorption, the adsorption efficiency of CS-MMT2 aerogel
290 changed slightly and remained at about 80%. EDTA mainly plays the role of chemically eluting Cu^{2+} in
291 the desorption process and has a certain group repair effect on CS-MMT2. CS-MMT2 aerogel can be
292 regenerated during desorption and incomplete vacuum desorption. But at the same time, certain structural

293 shrinkage and hole blockage are inevitable. This is due to the massive loss of CS-MMT surface groups
294 during the adsorption and desorption process.

295 2.7 Discussion on adsorption principle



296

297 **Fig. 11** (a) C1s, (b)Cu2p spin-orbit peaks of CS-MMT2 after adsorption, (c) FT-IR spectra of CS-
298 MMTs after adsorption

299 To explore the principle of adsorption, we characterized the adsorbed samples. As shown in
300 Fig. 11(a), the C 1s spectra of CS-MMT2 were fitted with three components centered at 284.8, 285.9,
301 286.6, 287.7, and 288.4 eV, which can be assigned to C-C, C-N, C-O, C=O, and O=C-O,
302 respectively. The emergence of a new characteristic peak at 934.4 eV, corresponding to the presence
303 of Cu2p orbital. At the same time, the metal oxidation peak area was changed. Cu²⁺ reacted with the
304 adsorbent to combine part of O. It is the oxygen-containing functional group that participates in the
305 adsorption process. The Cu2p spin-orbit peaks showed that Cu²⁺ replaced the original hydrogen ion
306 to form the corresponding complex. During the adsorption process, C-N, C-O, O=C-O content
307 increased significantly, indicating that Cu²⁺ was successfully adsorbed on CS-MMT2. Therefore, it
308 can be considered that the high adsorption capacity of CS-MMT2 to Cu²⁺ is due to chelating and
309 stacking, and electrostatic attraction. As shown in Fig. 11(c), the intensity of the characteristic peak
310 of the infrared spectrum changed to different degrees after copper ions were adsorbed, which
311 indicates that amino, carboxyl, siloxane, and hydroxyl groups in CS-MMTs were involved in the

312 adsorption of Cu^{2+} .

313 **3. Conclusions**

314 CS-MMT composite aerogels were prepared by freeze-drying adsorption of Cu^{2+} . Montmorillonite
315 has a layered structure. Modifier and physical methods are used to enlarge the distance between the
316 montmorillonite, and CS is grafted on the surface of montmorillonite. The surface contains a large
317 number of active sites, which promote the cross-linking of chitosan and montmorillonite, forming a
318 stable three-dimensional network structure. The experimental results show that CS-MMT2 has good
319 adsorption performance for Cu^{2+} . CS-MMT2 composite aerogel has a highly developed pore structure,
320 high adsorption capacity, fast adsorption and desorption, and can effectively remove Cu^{2+} with
321 remarkable effect. MMT first provides adsorption sites through cation exchange, and the amino group
322 on the surface forms hydrogen bonds with the hydroxyl group of the chitosan, which promotes the
323 cross-linking of the chitosan and montmorillonite in the soil. After 3 times of adsorption-desorption,
324 the removal effect of CS-MMT2 composite aerogel on Cu^{2+} remains about 80%. A feasible method was
325 provided for the selective removal and effective recovery of Cu^{2+} from heavy metal wastewater.

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342

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Figures

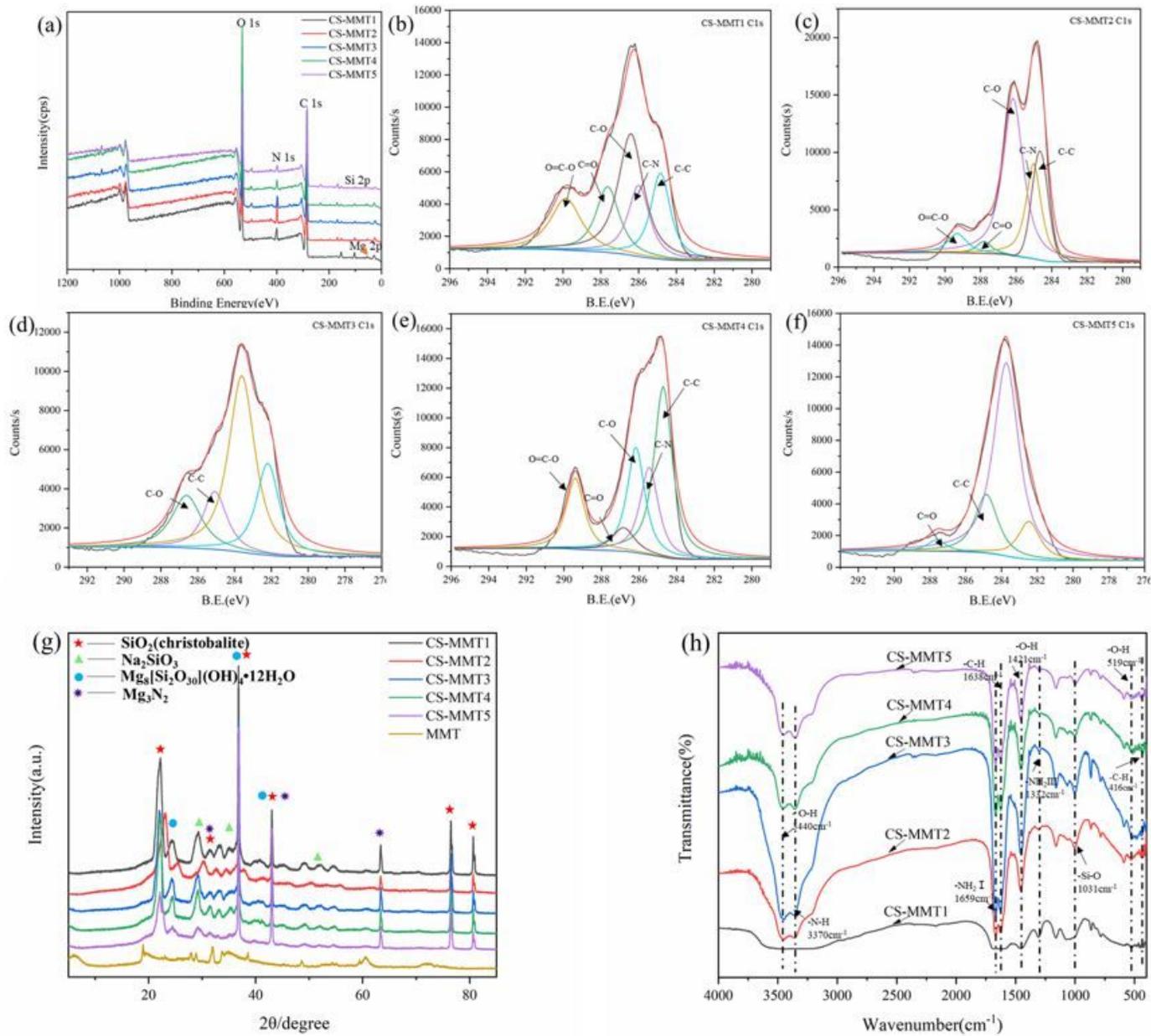


Figure 1

(a) XPS survey spectrum, (b~f) C1s spin-orbit peaks of CS-MMTs, (g) XRD patterns of CS-MMTs and MMT, (h) FTIR spectra of CS-MMTs

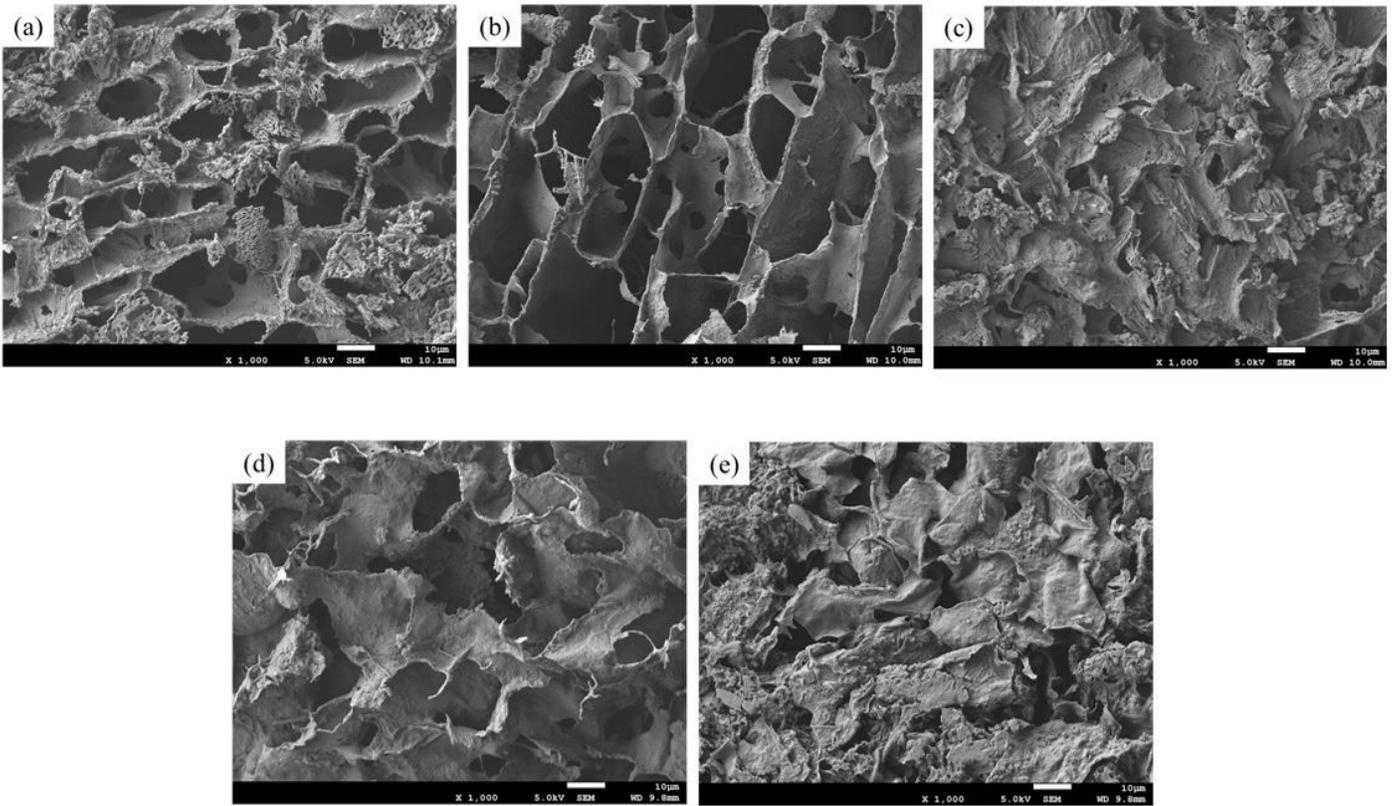


Figure 2

SEM image of (a)CS-MMT1, (b)CS-MMT2, (c)CS-MMT3, (d)CS-MMT4, (e)CS-MMT5

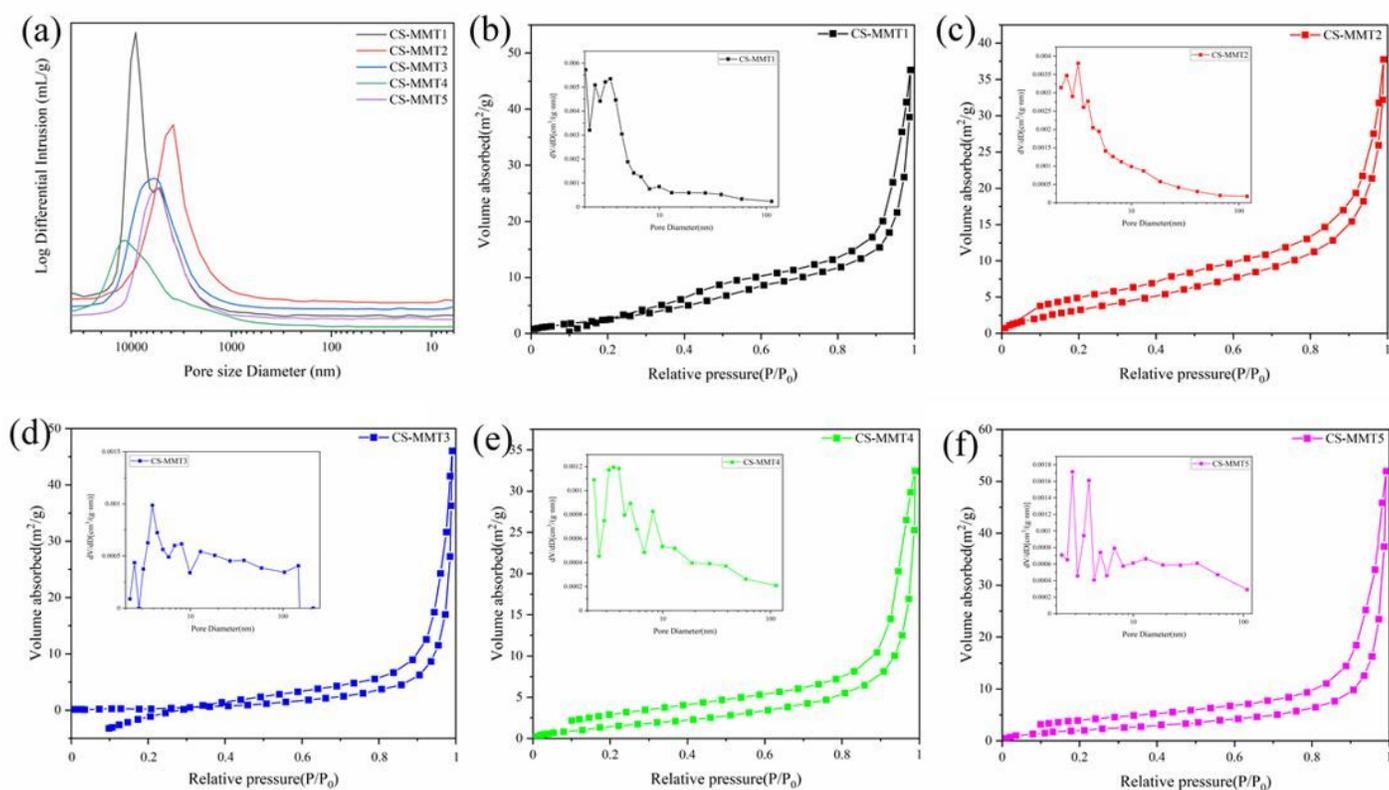


Figure 3

(a) Mercury intrusion test, (b~f) N₂ Adsorption–desorption isotherms and pore size distributions of CS-MMT composite aerogels with different components

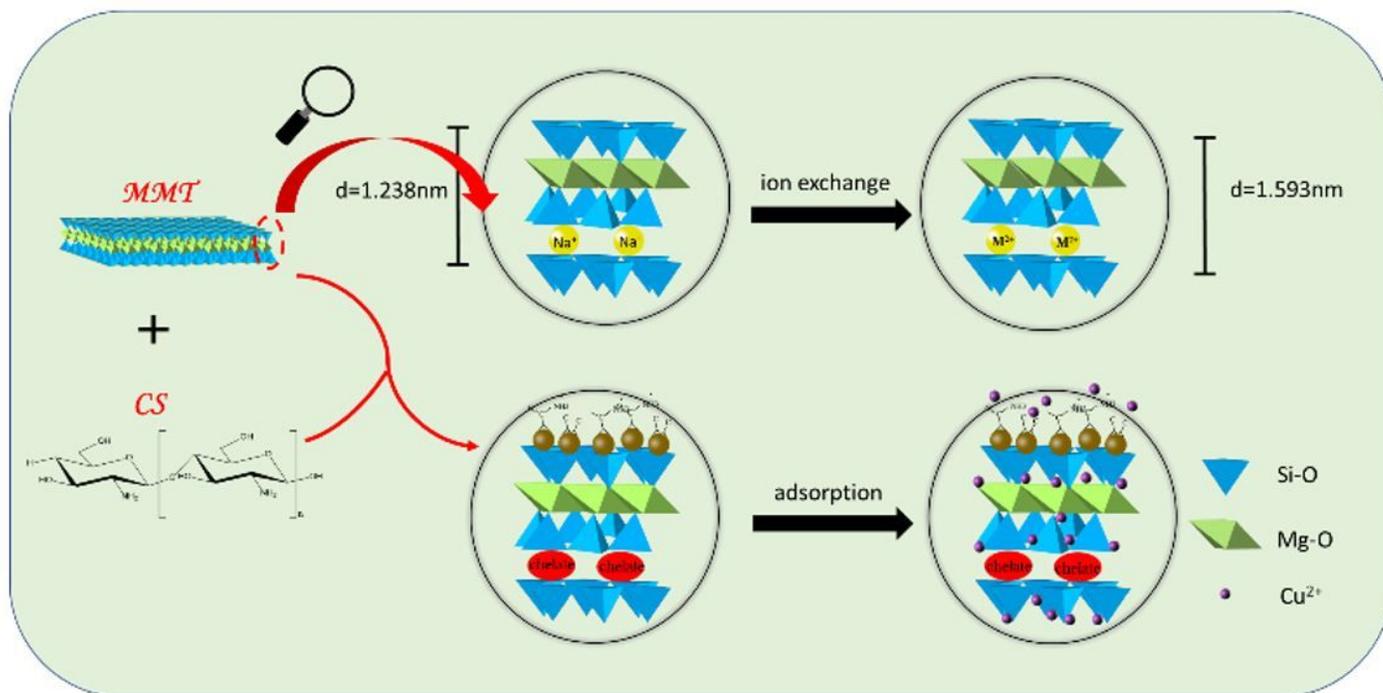


Figure 4

Schematic diagram of structure model and adsorption mechanism

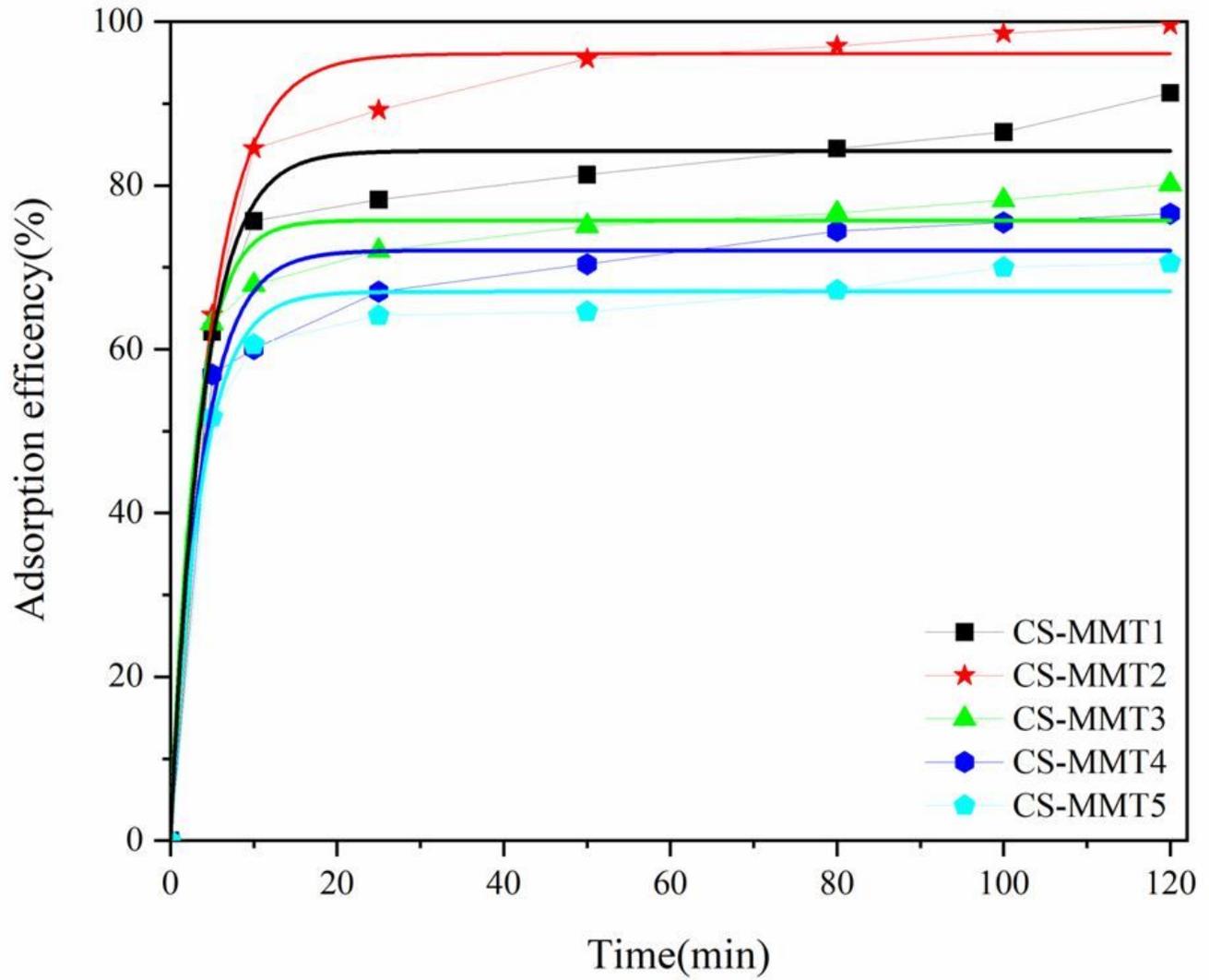


Figure 5

Adsorption efficiency of CS-MMT composite aerogels for Cu^{2+} at different time

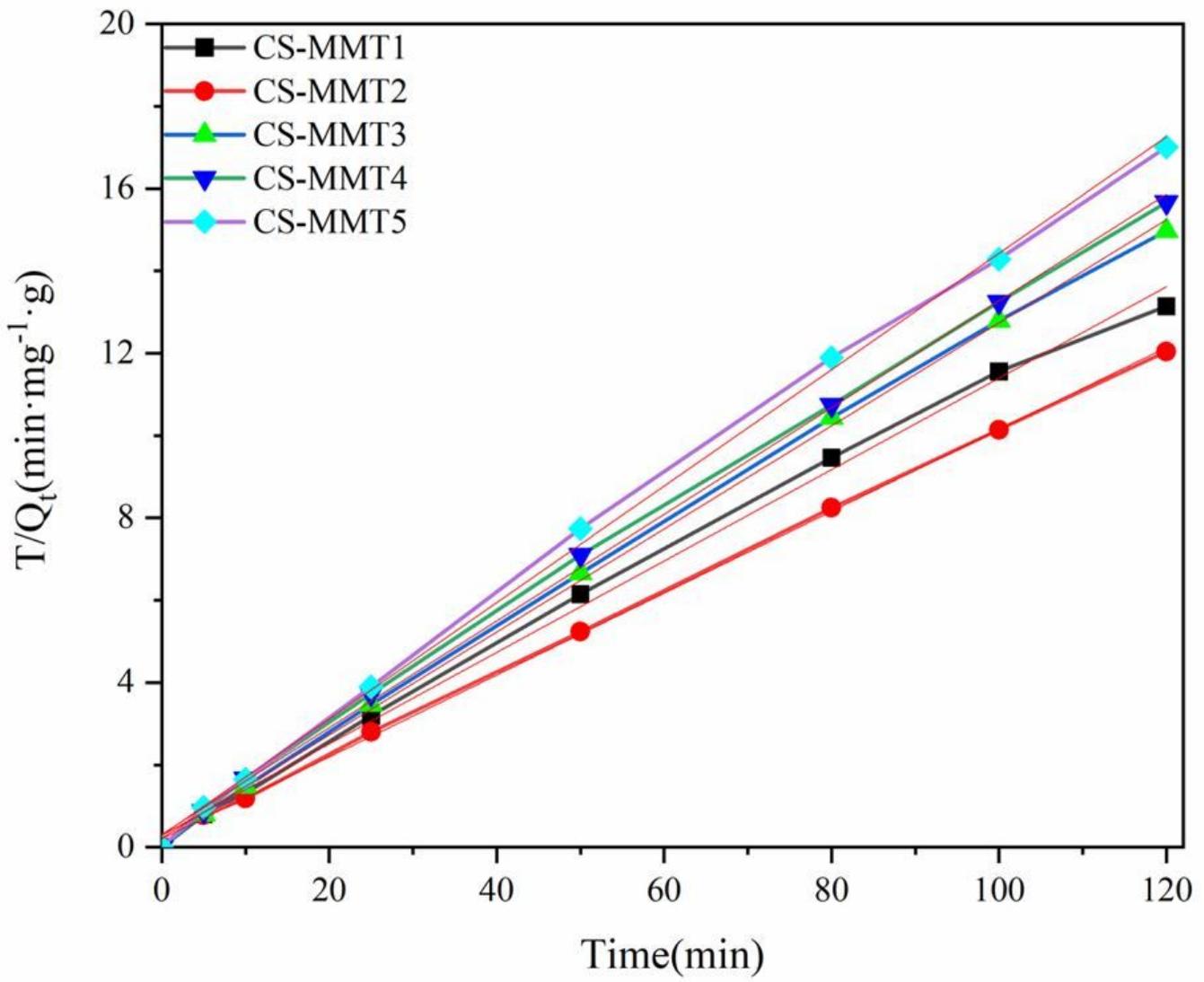


Figure 6

Pseudo-second-order model for adsorption of Cu²⁺ on CS-MMT composite aerogels

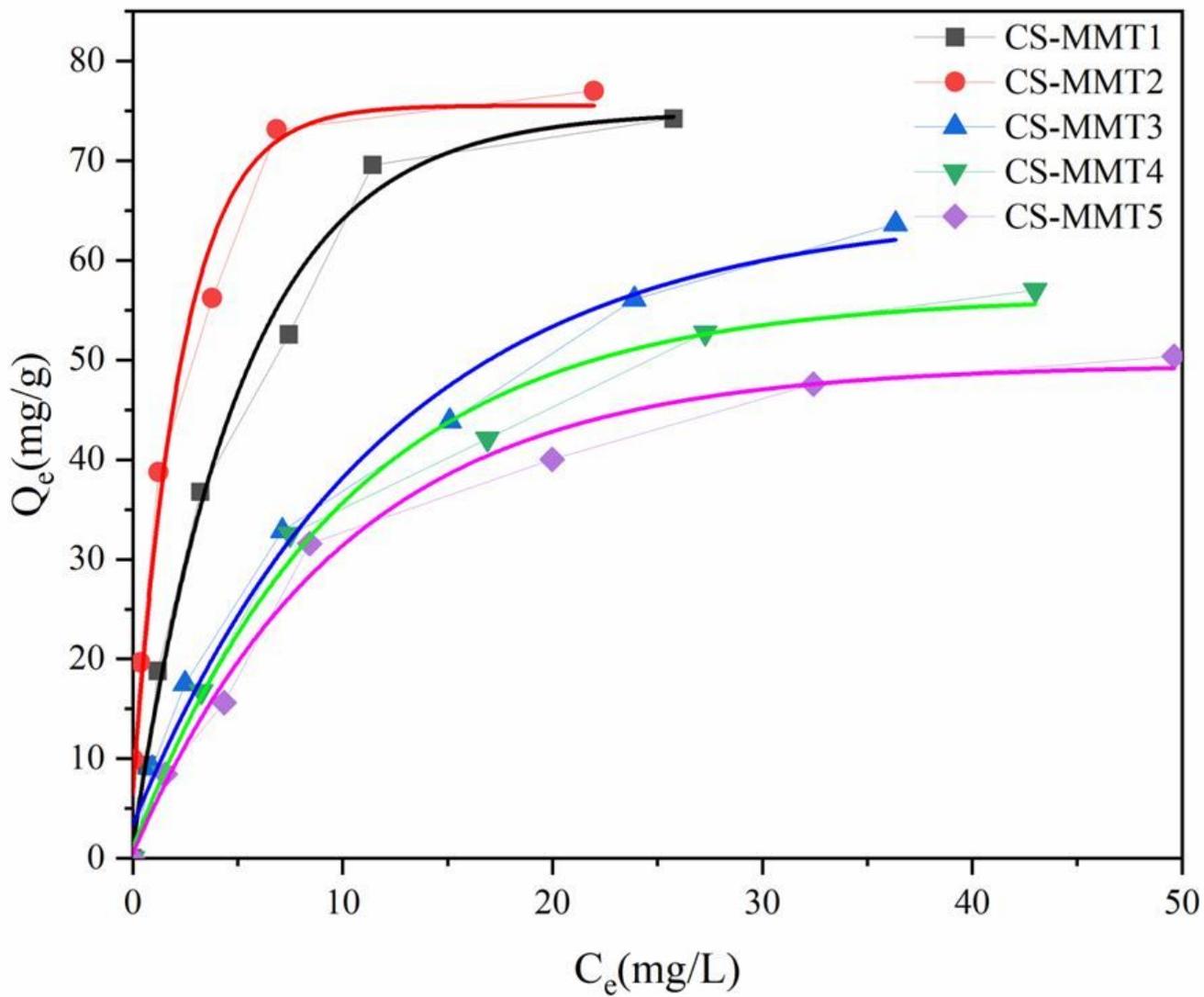


Figure 7

Adsorption capacity of CS-MMT composite aerogels for Cu^{2+} in different initial solution concentration

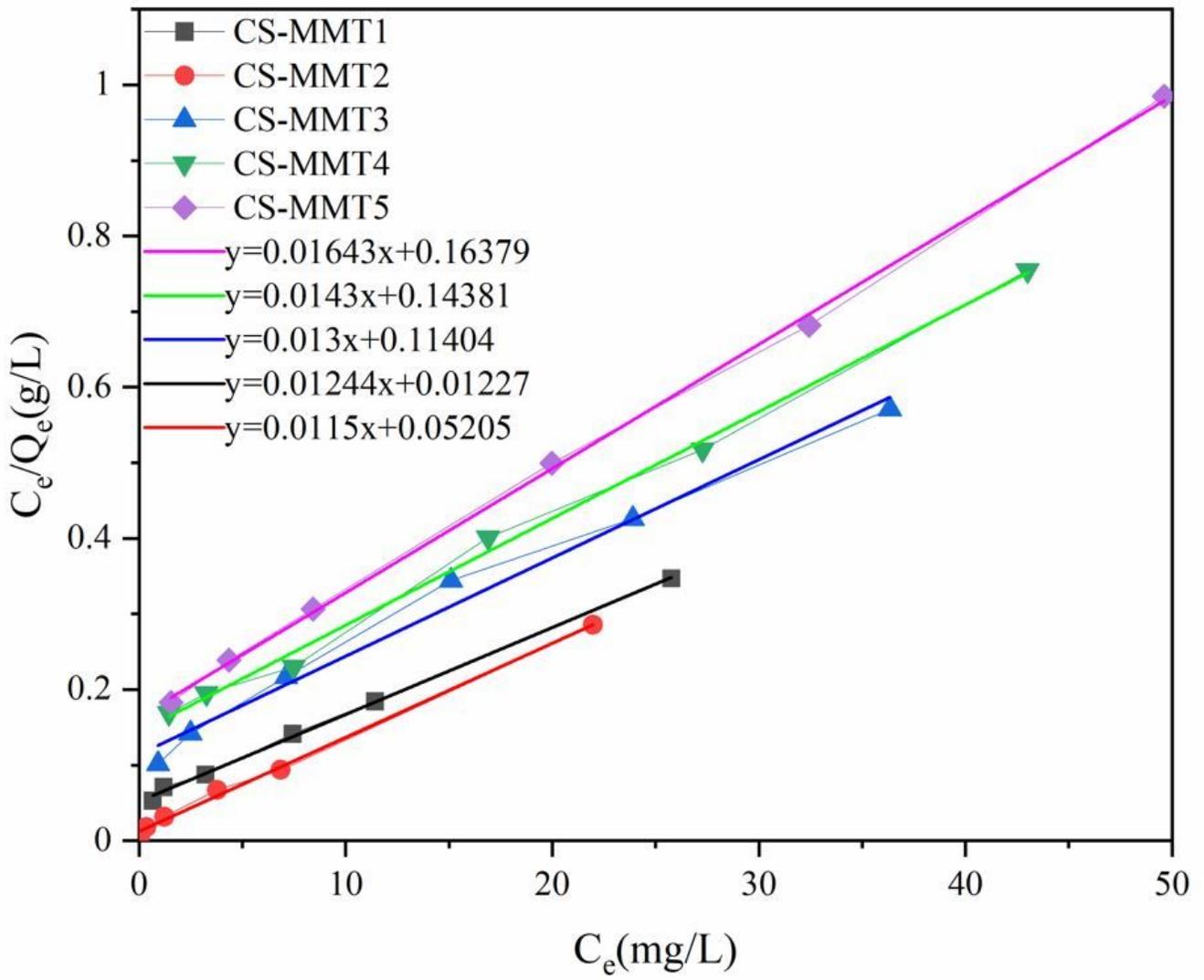


Figure 8

Langmuir isotherms plots of Cu²⁺ adsorption on CS-MMTs

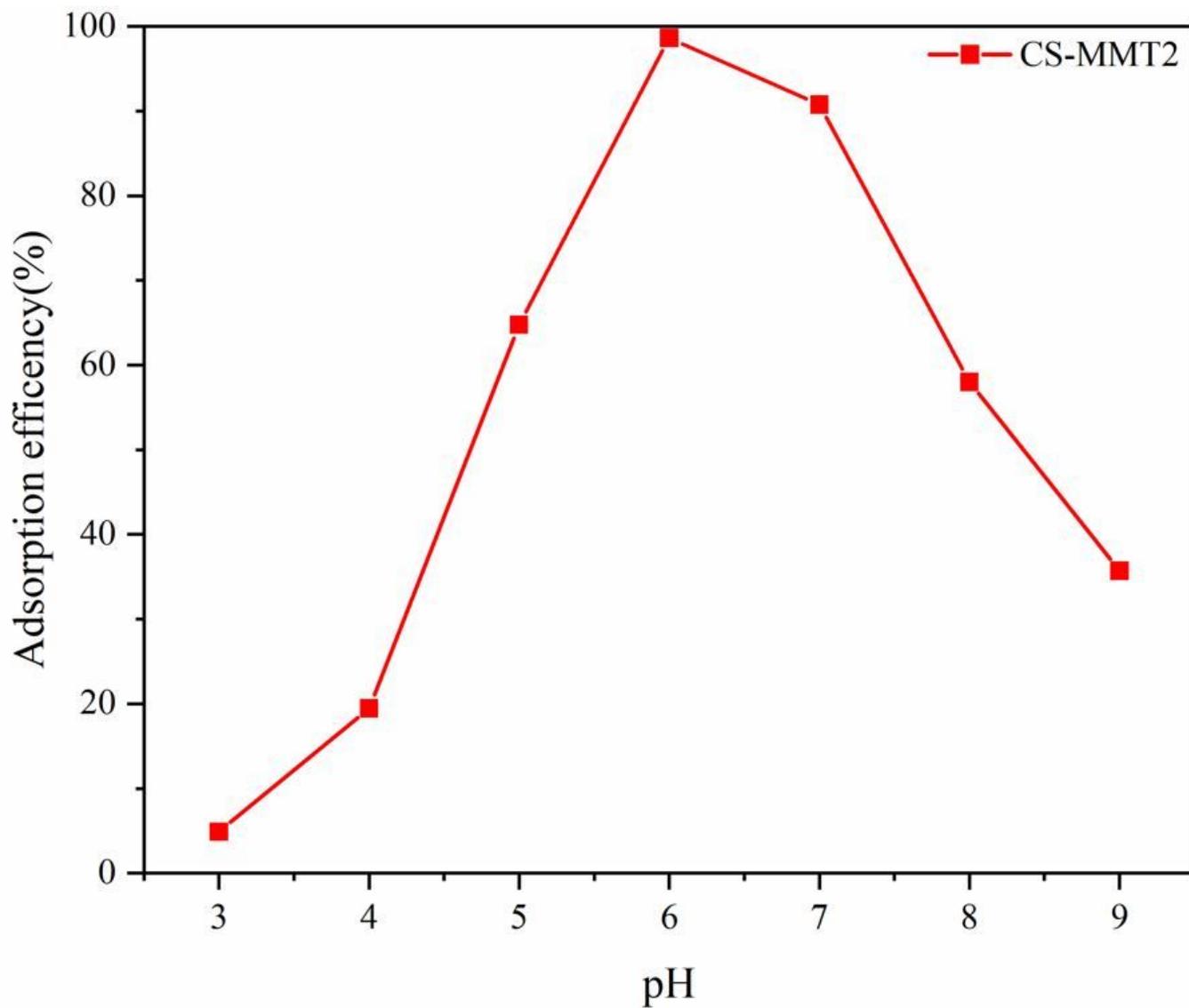


Figure 9

The effect of solution pH value on the adsorption capacity of Cu^{2+} by CS-MMT2

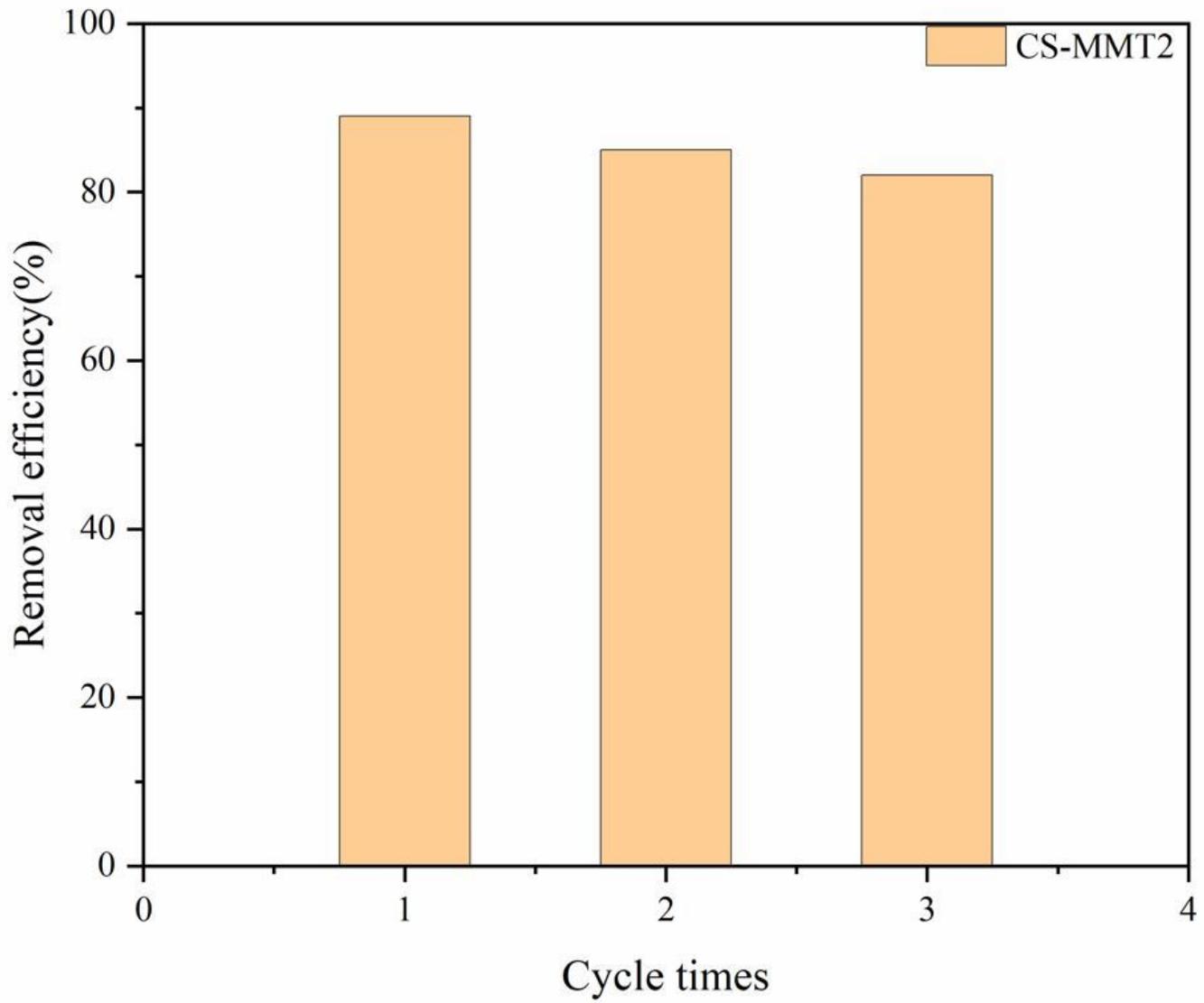


Figure 10

Circulation capability of CS-MMT2

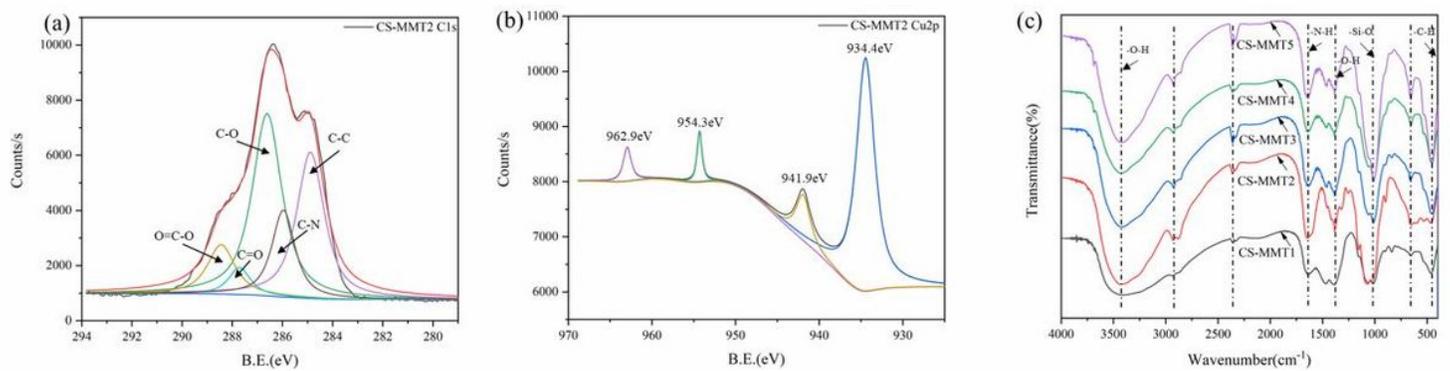


Figure 11

(a) C1s, (b)Cu2p spin-orbit peaks of CS-MMT2 after adsorption, (c) FT-IR spectra of CS-MMTs after adsorption

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