

Efficient photodegradation of Azucryl Red by copper doped TiO₂ nanoparticles – Experimental and modelling studies

Ahmed Kerrami

University of Bejaia Laboratory of Materials Technology and Process Engineering: Universite de Bejaia
Laboratoire de Technologie des Materiaux et Genie des Procedes

Lotfi Khezami (✉ lkhezami@gmail.com)

Al Imam Mohammad Ibn Saud University <https://orcid.org/0000-0001-5747-3114>

Mohamed Bououdina

University of Bahrain College of Health Sciences

Laila Mahtout

University of Bejaia Laboratory of Materials Technology and Process Engineering: Universite de Bejaia
Laboratoire de Technologie des Materiaux et Genie des Procedes

Abueliz Modwi

Qassim University College of Science and Arts in Alrass

Souhila Rabhi

University of Bejaia Laboratory of Materials Technology and Process Engineering: Universite de Bejaia
Laboratoire de Technologie des Materiaux et Genie des Procedes

Faycal Bensouici

Universite M'Hamed Bougara Boumerdes Faculte des Sciences

Hayet Belkacemi

University of Bejaia Laboratory of Materials Technology and Process Engineering: Universite de Bejaia
Laboratoire de Technologie des Materiaux et Genie des Procedes

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1 **Efficient photodegradation of Azucryl Red by copper doped TiO₂**
2 **nanoparticles – Experimental and modelling studies**

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4 Ahmed Kerrami ^a, Lotfi Khezami ^{b,*}, Mohamed Bououdina ^c, Laila Mahtout ^a, Abueliz Modwi ^d,
5 Souhila Rabhi ^a, Faycal Bensouici ^e, Hayet Belkacemi ^a

6 ^a Materials Technology and Process Engineering Laboratory, University of Bejaia, Algeria.

7 ^b Department of Chemistry, College of Sciences, Imam Mohammad Ibn Saud Islamic University (IMSIU),
8 P.O. Box 5701, Riyadh 11432, Saudi Arabia.

9 ^c Department of Physics, College of Science, University of Bahrain, PO Box 32038, Kingdom of Bahrain

10 ^d Department of Chemistry, College of Science & Arts, Qassim University, Ar Rass Saudi Arabia.

11 ^e Department of Physics, URMPE Unite, UMBB University, 35000 Boumerdes, Algeria.

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13
14 * Corresponding author: Lotfi Khezami (lkhezami@gmail.com)

22 **Abstract**

23 This work aims to investigate the effect of copper doping on the photocatalysis performance of
24 TiO₂ nanoparticles for disposal wastewater from organic pollutants. X-ray diffraction analysis
25 manifests the crystallization of a rutile phase for pure and copper-doped TiO₂ except for 2%
26 resulting in a rutile-to-anatase phase transformation. The crystallite size is found less affected by
27 Cu doping; i.e. ~30 nm. BET analysis indicates a decrease in the specific surface area as the doping
28 loading increases. Scanning electron microscopy observations reveal spherical particles at the
29 nanometer range for pure TiO₂ then larger agglomerates of ultrafine particles with Cu doping.
30 FTIR analysis notifies the existence of hydroxyl groups that will promote the photocatalysis
31 process. The photodegradation of Azucryl Red (AR) has been investigated under different
32 conditions; i.e., Cu-loading, initial concentration of AR, and pH. The kinetics of the
33 photodegradation process is further found to comply with the Lagergren kinetic law, regardless
34 the experimental conditions. Nevertheless, the photodegradation process is not only controlled by
35 the intra-particle diffusion mechanism, but also by mass transfer through a liquid film boundary.
36 The maximum degradation of AR, i.e. 86%, has been achieved at pH = 5.0 during 60 minutes of
37 contact time for the 2% Cu-doping, with effective regeneration. The Freundlich model exhibits a
38 better fitting for AR dye photodegradation equilibrium data, compared to Langmuir, Temkin and
39 Dubinin-Radushkevich.

40 **Keywords:** Sol-gel; Cu-doping; Rutile; Anatase; Phase transformation; Photodegradation;
41 Azucryl red.

42 **1. Introduction**

43 Industrial activities generate significant volumes of water contaminated by polluting
44 organic and inorganic matter. The organic contaminants include synthetic dyes in the industrial
45 sector that are widely used in the paper industry, cosmetics, food and especially in the textile
46 industry (El Haddad, Regti, Laamari, Mamouni, & Saffaj, 2014; Mahdavinia & Zhalebaghy, 2012;
47 Perez-Urquiza, Prat, & Beltran, 2000). These non-biodegradable dyes are likely to contaminate
48 the environment, besides being responsible for various environmental alterations, especially water
49 resources, their presence in the water, their pollution rate and their problem of water pollution.

50 Although several techniques have been adopted for organic dyes remediation, the
51 heterogeneous photocatalysis currently appears as a very promising and potential technological
52 alternative. Indeed, during the process of degradation, there will be the destruction of organic
53 pollutants, such as dyes, while leading to a complete mineralization (Fernández-García et al., 2000;
54 Joshi & Shrivastava, 2011). It has been reported that the photocatalytic process depends mainly on
55 the characteristics of the molecule to be degraded, the light source and the type of semiconductor
56 acting as a photocatalyst (gap and texture). In the literature, the widely investigated photocatalysts
57 are zinc oxide (ZnO) and titanium dioxide (TiO₂) (Hoffmann, Martin, Choi, & Bahnemann, 1995;
58 Kormann, Bahnemann, & Hoffmann, 1988; Liu & Yang, 2003; Pelizzetti, Minero, Piccinini, &
59 Vincenti, 1993; Villaseñor, Reyes, & Pecchi, 1998; William IV, Kostedt, Ismail, & Mazyck,
60 2008).

61 Titanium dioxide is currently considered to be the most efficient semiconductor owing to
62 its enhanced photocatalytic performance, non-toxicity, low cost, durability and chemical stability
63 to corrosion (Hoffmann et al., 1995), as well as for its large and tunable bandgap (3.2 eV for
64 anatase and 3.0 eV for rutile). Moreover, it uses UV rays containing only a small part (about 3 to

65 5%) of solar energy (Linsebigler, Lu, & Yates Jr, 1995). Many efforts have recently been devoted
66 to optimize the characteristics of TiO₂ in the visible (Wu, Dong, Zhao, & Guo, 2008) such as
67 through the doping TiO₂ with transition metal ions (Anpo & Takeuchi, 2003; Chen et al., 2002;
68 Einaga, Harada, Futamura, & Ibusuki, 2003; Thaminimulla, Takata, Hara, Kondo, & Domen,
69 2000; Xie & Yuan, 2003) and non-metallic anions (Asahi, Morikawa, Ohwaki, Aoki, & Taga,
70 2001; Burda et al., 2003; Gómez et al., 2003; Yu, Yu, Ho, Jiang, & Zhang, 2002). Moreover, other
71 approaches have been also proposed, such as the coupling with narrow-band semiconductors
72 (Bartl, Boettcher, Frindell, & Stucky, 2005; Bartl, Puls, Tang, Lichtenegger, & Stucky, 2004;
73 Kumar & Jain, 2001) and sensitization with dyes (Viseu, Hungerford, & Ferreira, 2002). Besides,
74 it is well-known that the doping process of TiO₂ by metals and transition metals is much easier
75 than using non-metallic elements (Iwazuk & Nolan, 2011).

76 This research work, aims to investigate the influence of Cu as an effective transition metal
77 dopant in order to improve TiO₂ photocatalytic properties (Srinivas et al., 2011). TiO₂ and Cu-
78 doped nanopowders are prepared using the sol-gel method then characterized by XRD, SEM/EDX,
79 BET, and FTIR. The photocatalytic performance of the as-prepared nanopowders is studied using
80 azucryl red (AR) molecule as a textile dye. The photodegradation data were modeled using
81 different kinetics and equilibrium models. A mechanism of the photodegradation process was
82 proposed and discussed.

83 **2. Experimental Part**

84 **2.1. Synthesis of Photocatalysts**

85 Pure and Cu-doped TiO₂ powders are prepared via a cost-effective and straightforward
86 procedure. The precursor tetraethyl-orthotitanate ((C₂H₅O)₄Ti, Merk 95%) was dissolved in
87 ethanol, water, and nitric acid (HNO₃ Biochem 69%) solution. The mixture was roughly stirred

88 using a magnetic agitator for one hour. After 24 hours of soaking, the mixture was oven-dried at
89 100°C and continuously stirred in air. The obtained powder was amorphous. The sol of Cu-doped
90 TiO₂ was prepared using the identical method as previously detailed, with the addition of copper
91 nitrate (Cu(NO₃)₂) to the TiO₂ solution. The as-produced powders are subsequently calcinated for
92 one hour at 600°C.

93 **2.2. Characterization of Photocatalysts**

94 The structure of the as-prepared powders was checked by X-ray powder diffraction (XRD)
95 using Rigaku Ultima IV diffractometer with Cu-K α X-ray source ($\lambda = 1.5418 \text{ \AA}$). The
96 morphological examinations have been accomplished by field emission scanning electron
97 microscopy (FESEM) employing LYRA3 TESCAN provided with a dispersive electron X-ray
98 (EDX) spectrometer for the specification of the chemical composition. The BET (The Brunauer,
99 Emmett and Teller) specific surface area (SSA), average pore diameter and volume were
100 determined from nitrogen adsorption-desorption isotherms performed at the boiling point (-
101 195.8°C) of N₂ through an ASAP 2020 (Micromeritics) apparatus. Each sample was outgassed
102 using He for 2h at 250°C to take off moistness and impurities prior to the adsorption experiment.
103 The BET model and t-plot scheme of Lippens and de Boer (Lippens & De Boer, 1965) were used
104 to calculate the porosity and surface area. The FTIR chemical bonding were analyzed using an
105 Nicolet 6700 spectrometer, FTIR spectra were achieved in the interval of 4000–400 cm⁻¹ with a
106 resolution of 4 cm⁻¹.

107 **2.3. Photocatalytic Degradation Measurements**

108 The photocatalysis test was carried out at atmospheric pressure and room temperature in a
109 catalytic photoreactor, as described elsewhere (Kerrami et al., 2019). 0.2 mg of catalyst was added

110 to 2 mL of azucryl red (AR) solution having an initial concentration $C_0 = 1,247 \times 10^{-5} \text{ mol.L}^{-1}$ in a
111 quartz container. The heterogeneous mixture was then irradiated with UV light with a cut filter
112 (365 nm) using a UV light lamp (Spot Light Source 300-450nm, L9588-01, Hamamatsu, Japan)
113 and kept under magnetic agitation. The absorbance was measured for different exposure periods
114 of time using a spectrophotometer scan (39919.FEB /10) at a maximum absorption wavelength of
115 about 532 nm.

116 The amount of adsorbed AR dye was calculated according to the formula:

$$117 \quad q_t = \frac{V(C_0 - C_t)}{m} \quad (1)$$

118 where q_t (mg.g^{-1}) is the adsorbed dye amount per gram of adsorbent at time t (min), V is the solution
119 volume (L), C_0 and C_t are the initial concentration and concentration of AR dye (mg.L^{-1}) in solution
120 at any time t , and m the mass of the adsorbent (g).

121 The quantity adsorbed (mg dye/g material) at equilibrium q_e was estimated through an analogue
122 expression:

$$123 \quad q_e = \frac{V(C_0 - C_e)}{m} \quad (2)$$

124 where C_e is the liquid-phase concentrations of AR dye at equilibrium time.

125 **3. Results and Discussion**

126 **3.1 X-ray diffraction analysis**

127 Figure 1 illustrated the evolution of X-ray diffraction patterns of pure and Cu 2-10% doped
128 TiO_2 . Qualitative phase analysis indicates the formation of rutile phase for pure TiO_2 , followed by
129 phase transformation into anatase phase for $\text{Cu}_{2\%}$, which in turn undergoes a second phase

130 transformation into original rutile phase for Cu_{4%} and Cu_{10%} doped TiO₂. Figure 1 also displays
131 that all diffraction peaks are comparatively broad and intense, implying the formation a nanoscale
132 phase with high crystallinity. For pure TiO₂, the observed peaks at $2\theta = 27, 36, 41, 44, 54, 56, 62,$
133 64 and 69° are indexed as (110), (101), (111), (211), (220), (002), (311), (301), and (112)
134 reflections of the tetragonal rutile phase, in conformity with JCPDS card No. JCPDS card No. 01-
135 075-1750. For Cu_{2%} doped TiO₂, the newly emerged diffraction peaks detected at $2\theta = 25.3, 37.8,$
136 $48.1, 53.9, 55.1, 62.7, 68.8, 70.4$ and 75.07° are indexed as (101), (004), (200), (105), (211), (204),
137 (116), (220), and (215) reflections of the tetragonal TiO₂ (anatase phase) structure in accordance
138 with JCPDS card No. 00-021-1272 (Pongwan, Wetchakun, Phanichphant, & Wetchakun, 2016).
139 Whereas for Cu_{4%} and Cu_{10%} doping contents, all the diffractions peaks belong to the tetragonal
140 rutile phase similarly to pure TiO₂. However, it is important to mention that for Cu doping in the
141 range 4–10%, new low intensity diffraction peak appears around of $2\theta = 44^\circ$ attributed to the (111)
142 reflection belonging to Cu metal in agreement with JCPDS card No. 02-1225 (Del Carro, 2018).

143 In a similar work, Park et al. reported that anatase-rutile phase transformation of TiO₂
144 occurred at 2.5 wt.% Cu doping using low temperature process at 60 °C for 4 and 14 h (Park, Kim,
145 Kim, & Lee, 2006). Besides, due to the differences between the ionic radii (Ti⁴⁺ = 0.64 Å and Cu²⁺
146 = 0.73 Å) and the cationic charges of Ti (+4) and Cu (+2), oxygen vacancies may be formed for
147 maintaining charge neutrality (Sahu & Biswas, 2011).

148 The average crystallite size (D) of Cu_{x%}-doped TiO₂ nanoparticles have been calculated by
149 using the most intense peak (101) anatase and (110) rutile by applying the Scherrer equation:

$$150 \quad D = \frac{K \lambda}{\beta \cos \theta} \quad (3)$$

151 where λ is the X-ray wavelength (1.5417 Å), θ is the Bragg angle, and β is the full width at half
152 maximum of the corresponding peak.

153 According to Williamson–Hall approach (Zak, Majid, Abrishami, & Yousefi, 2011), strain
154 and crystallite size contribute significantly to the diffraction lines broadening (equation 4) (Bindu
155 & Thomas, 2014; Pandiyarajan & Karthikeyan, 2012). In the uniform deformation model (*UDM*),
156 it is assumed that a crystal is isotropic and subsequently its properties are independent of the
157 crystallographic direction along which the measurement is considered.

$$158 \quad \beta_{hkl} \cos \theta_{hkl} = \frac{k\lambda}{D} + 4\varepsilon \sin \theta_{hkl} \quad (4)$$

159 A plot of $\beta_{hkl} \cos \theta_{hkl}$ versus $4 \sin \theta_{hkl}$ represents a linear graph (Ungár, 2007), whereas
160 microstrain (ε) and the crystallite size (D) are calculated from the slope and intercept of the linear
161 plot, respectively.

162 The d-spacing and the lattice parameters have been estimated by following expressions:

$$163 \quad d = \frac{\lambda}{2 \sin \theta} \quad (5)$$

$$164 \quad \frac{1}{d^2} = \frac{h^2 + k^2}{a^2} + \frac{l^2}{c^2} \quad (6)$$

165 All calculated structural (d-spacing and lattice parameters) and microstructural (crystallite
166 size and microstrain) parameters are given in Table 1. The mean crystallite size values estimated
167 using the Scherrer equation and Williamson–Hall model are virtually identical, signifying that the
168 inclusion of strain in the modified formulae of W–H method has inconsequential impact on the
169 average (D) value. Yet, the average crystallite size estimated using Scherrer’s and W–H methods
170 exhibit little disparity that can be ascribed to the variation in particle size distribution averaging
171 (Zak et al., 2011). Both values of crystallite size have the same trend, with 4% Cu doping
172 exhibiting the smallest crystallite size of 26 (29) nm and the lowest microstrain with $48 \times 10^{-5} \%$.

173 The lattice parameters of pure TiO_2 are found to be vary anisotropically compared to the
174 standard values of rutile phase; a and b increase from 4.5912 to 4.5937 while c decreases from

175 2.9599 to 2.9587 Å without any significant variation in the unit cell volume; i.e. 62.39 compared
176 to 62.43 Å³. Concerning the influence of Cu doping on the lattice parameter, it can be noticed that
177 for 2% Cu loading, the calculated values (a=b=3.7834 Å and c=9.5181 Å) are marginally smaller
178 in contrast to that of the standard pure anatase phase with a tetragonal structure; i.e. a=b=3.7852
179 Å and c=9.5139 Å (JCPDS card No. 00-021-1272). However, insignificant change in the lattice
180 parameters is observed for the rutile phase, the lattice volume for 4 and 10% Cu-doping is around
181 62.42 and 62.45 Å³, respectively compared to 62.43 Å³ (a=b=4.5937 Å, c=2.9587 Å; JCPDS card
182 No. 01-075-1750). This can be explained as follow: (i) for smaller loading, Cu²⁺ (0.73 Å) dissolve
183 into TiO₂ host lattice by occupying Ti⁴⁺ (0.64 Å) sites resulting in lattice contraction/expansion;
184 (ii) however, for charge balance of stoichiometric TiO₂, the compensation of the missing positive
185 charge (+2 for each Cu²⁺ replacing one Ti⁴⁺) will be compensated by the creation of oxygen
186 vacancy (-2); (iii) the creation of oxygen vacancy will be accompanied by a contraction of the unit
187 cell. Therefore, the simultaneous occurrence of lattice expansion / contraction results in
188 insignificant change in lattice parameters of TiO₂ upon Cu doping. Moreover, due to relatively
189 large difference in ionic radii (0.64 and 0.73 Å) and charge difference (+4 and +2), a limit of
190 solubility is reached, resulting in the precipitation of Cu metal for higher doping concentrations.

191 **3.2 Morphological observations and chemical analysis**

192 Figure 2 displays the surface particles' morphology of pure and 2% Cu-doped TiO₂
193 powders have with the corresponding elemental chemical analysis. SEM image of pure TiO₂
194 powder (Fig. 2a) consists of spherical-shaped particles at the nanoscale with relatively
195 homogeneous size distribution and tendency to agglomeration. Upon 2%Cu doping (Fig. 2b), the
196 powder consists of relatively large agglomerates of ultrafine particles (melted surfaces) with the
197 formation of pores, resulting in improved specific surfaces and consequently very beneficial for

198 the photocatalytic activity. The EDX spectrum of pure TiO₂ indicates the existence of O and Ti
199 elements only, in addition to Cu after doping, signifying the purity of the as-prepared powders.
200 The elemental chemical composition indicates that the Ti/O (24.2 / 50.12 = 0.482) ratio is very
201 close to 0.5, thereby in compliance the stoichiometry of the starting precursor solution. The amount
202 of Cu (1.56 at. %) is also close to the intended content (2.0 at. %), in accordance with the initial
203 precursor solution. The grain size was determined using the imageJ software. The results showed
204 that doping with Cu decreases the particle size; from 29.23 nm for pure TiO₂ up to 19.58 nm for
205 2% Cu doped TiO₂, which is very beneficial for the photocatalytic reaction.

206 **3.3 FTIR analysis**

207 Figure 3 displays the FTIR spectra of pure and Cu-doped TiO₂ powders. All samples
208 exhibit a broad band setting in the wavenumber range 3250 - 3700 cm⁻¹, which corresponds to the
209 stretching vibrations of the hydroxyl groups of water molecules. The small band at 2350 cm⁻¹ for
210 Cu-doped TiO₂ is ascribed to the CO₂ molecule in the atmosphere (Baneshi, Haghghi, Jodeiri,
211 Abdollahifar, & Ajamein, 2014), while the band at 1635 cm⁻¹ corresponds to the bending and
212 stretching OH groups adsorbed on the catalyst surface (Baneshi et al., 2014). Surprisingly, such
213 bands are not detected in pure TiO₂ having, signifying that the surface of pure TiO₂ contains fewer
214 hydroxyl groups than Cu-doped TiO₂ which can be attributed to the loss of water during
215 calcination. In the region lower than 1071 cm⁻¹, the absorption bands are assigned to the Ti-O
216 stretching band and O-Ti-O flexural vibration modes (Lopez et al., 2002), which confirm the
217 formation of TiO₂ phase.

218 **3.4 BET Measurements**

219 The isotherm shape provides information on pore size, volume, and distribution as well as
220 whether its macro-, meso-, or micro-porous structure. The nitrogen adsorption-desorption

221 isotherms are plotted for pure and 2% Cu-doped TiO₂ powders, as illustrated in Figure 4. Both
222 curves display a sorption isotherm of type IV, which is associated with capillary condensation,
223 respectively of nitrogen and argon adsorbed at 77 K and 87 K, occurring in mesoporous solids
224 having cylindrical pores, as well as H1 hysteresis loops according to IUPAC classification
225 (Sotomayor, Cychosz, & Thommes, 2018; Thommes et al., 2015). This kind of isotherm is
226 obtained with mesoporous adsorbents having a pore radius ranging between 25 and 500Å
227 (Thommes et al., 2015). Such a hysteresis loop depicts uniformly arranged spherical or
228 agglomerated nanoparticles (as shown in SEM images Fig. 2), with respective to TiO₂ higher
229 specific surface area and pore volume (Table 2).

230 Surprisingly, the surface area has been dramatically reduced upon the addition of Cu, even
231 at low content (2%), during the doping process. Systematically, a decrease of the pores' average
232 diameter and volume is observed, ranging from (13-20 nm) to (19-24 nm) and 108×10^{-3} up to 4.5
233 $\times 10^{-3} \text{ cm}^3 \cdot \text{g}^{-1}$, respectively. Maurice et al. (Sorolla II, Dalida, Khemthong, & Grisdanurak, 2012)
234 found a similar effect of Cu-doping when preparing 2 wt% Cu-TiO₂ via the sol-gel route. The
235 decline in the surface characteristics can be explained by the blocking of the pore entrance at the
236 surface of TiO₂ (Sorolla II et al., 2012). However, the appearance of hysteresis loop in the isotherm
237 of Cu-doped nanopowder indicates that the pore blockage does not take place over the entire
238 material.

239 The BJH method has been used to determine the pore size distribution, as illustrated by the
240 inset graphs in Figure 4. As it can be depicted in Figure 4a, the pore size distribution of pure TiO₂
241 nanopowder is narrow and ranging between 15 - 30 nm with an average pore diameter of about
242 23.4 nm. Meanwhile, the pore size distribution of Cu-doped nanopowder is dispersed in a broad
243 interval from 10 to 40 nm with an average pore diameter equal to 19.6 nm (Fig. 4b). As it can be

244 observed, the incorporation of Cu within TiO₂ host lattice even at such low doping level generated
245 lowering of pore diameter and volume as well as broadening and uniformity of pore size
246 distribution. This finding implies that Cu is well dispersed into TiO₂ network causing some
247 modifications of the mesostructure at its surface, which corroborate with XRD analysis.

248 **3.4. Photodegradation kinetics study of Azucryl red**

249 The study of the photodegradation kinetics and mechanism is very essential in processing
250 aqueous wastewater as it avail precious information on the degradation process. The degradation
251 kinetics has been performed by varying the amount of catalyst, the initial pH of AR solution as
252 well as the initial AR dye concentrations. Equation (6) represents a pseudo-first-order model of
253 Lagergren (Shahwan, 2015; Sorolla II et al., 2012) with a rate constant k_{app} (min⁻¹) and q_e the
254 maximum quantity of AR adsorbed at equilibrium. The plot of C_t/C_0 against t allows the
255 determination of k_{deg} :

$$256 \quad \frac{C_t}{C_0} = \text{Exp}(-k_{deg} \cdot t) \quad (7)$$

257 The kinetics results of the photodegradation process under UV irradiation are displayed in
258 Figure 5 (a, b, and c). The values of k_{deg} and the linear regression coefficient (r^2) are computed by
259 the simulation of the model using Origin Pro software and are summarized in Tables 3. According
260 to the regression coefficient, mainly greater than 0.96, the experimental results are well simulated
261 by the Lagergren pseudo-first-order model for the tested adsorbents. As can be noted, from Figure
262 5, photolysis does not affect the degradation of AR; however, the presence of the photocatalyst
263 reduces extremely the concentration of the dye. Based on the obtained rate constants, the half-life
264 time ($t_{1/2} = \ln 2/k_{deg}$) has been also calculated and compared (Shahwan, 2015). As shown in Table
265 3, the value of $t_{1/2}$ obtained based on the rate constant of Eq. (6) ranges from 18 up to 65 min,
266 which is much closer to the experimental values.

267 The adsorption of AR on the surface of the catalyst and its photodegradation can be
268 governed by the mass transfer through the boundary liquid film and/or by the intra-particle
269 diffusion/transport process. The linearized equation of mass transfer kinetic model proposed by
270 Ketcha Mbadcam *et al.* (Ho & McKay, 1998) is expressed as follows:

$$271 \quad \ln(C_0 - C_t) = \ln(D) + k_0 \cdot t \quad (8)$$

272 where D is the mass transfer constant and k_0 (min^{-1}) the adsorption constant. AR may be carried
273 from the liquid phase to the solid particles by the intra-particle diffusion process. The mass transfer
274 kinetic model constants are obtained from the slope and the intercept of the straight line $\ln(C_0 -$
275 $C_t)$ vs. time. The linear plots illustrated in Figure 6 (a, b, and c) and mass transfer model constants
276 are given in Table 3.

277 Sometimes, such a mechanism model is a rate-limiting stage in the adsorption or
278 photodegradation process. The probability of intra-particle diffusion is examined through the
279 Weber and Morris diffusion model (Hameed, Salman, & Ahmad, 2009; Mbadcam, Anagho, &
280 Nsami, 2011):

$$281 \quad q_t = k_{dif} \cdot t^{1/2} + C \quad (9)$$

282 where C , the intercept yields more details regarding the thickness of the boundary layer. The intra-
283 particle diffusion constant, k_{dif} value (in $\text{mg} \cdot \text{g}^{-1} \cdot \text{min}^{1/2}$), for the tested adsorbent is obtained from
284 the slope of the graph (Fig. 7 (a, b and c)) as illustrated in Table 3. The validity of these models is
285 thereafter discussed on the basis of the regression coefficient r^2 .

286 3.4.1 Effect of Cu-doping concentration

287 Figure 5a displays the photocatalytic decomposition of the azucrylic dye in aqueous
288 solution by pure and Cu-doped TiO_2 nanopowders under UV irradiation. As it can see, the azucryl
289 red decays constantly throughout the irradiation time for all nanopowders. The photocatalytic

290 elimination efficiency of Cu-doped TiO₂ for 2% and 4% reaches approximately 85.26 and 77.64%
291 of the initial concentration after 60 minutes of irradiation, respectively. Whereas beyond this value,
292 the degradation rate decreases, and pure TiO₂ shows only 73.84% of AR elimination efficiency.
293 This indicates that Cu-doping even in small loading, promotes the photocatalytic reaction.
294 Therefore, 2%Cu represents the optimal metal charge for the effectiveness of strengthening the
295 photocatalytic activity of TiO₂.

296 From Table 3, it can be observed that the highest constant efficiency (K_{deg}) is obtained for
297 2 and 4% Cu, with 3.83×10^{-2} and $2.7510^{-2} \text{ min}^{-1}$, respectively. Then it decreases drastically to
298 $1.16 \times 10^{-2} \text{ min}^{-1}$ for 10% Cu beneath the obtained value for pure TiO₂. This finding may indicate
299 that a larger content of copper will deteriorate the catalytic photodegradation capacity.

300 There may be the probable causes for the reduction of the photocatalytic efficiency at high
301 Cu-doping content: (i) the active sites present at the surface of catalyst are covered with an excess
302 of Cu; (ii) the combination of the charge carriers increases the rate photocatalytic degradation with
303 higher amount of Cu charge (Kim et al., 2013; Morikawa, Irokawa, & Ohwaki, 2006).

304 3.4.2 Effect of initial AR solution pH

305 The influence of the initial solution pH (i.e., pH = 3.0, 5.0, 6.0, and 8.0 ± 0.1) on the
306 degradation of AR by pure and Cu-doped TiO₂ nanopowders under UV-irradiation, are presented
307 in Figure 5b and the calculated Lagergren model constant values are given in Table 3.

308 The obtained results show that the azucryl red photodecoloration is higher in the case of acidic pH
309 medium, and the value of K_{deg} constant increases for the pH change from 3.0 to 5.0, subsequently
310 decreases for higher pH value. The most significant degradation kinetics is observed at a pH of 5.0
311 (Table 3). Besides, the effect of the solution's initial pH on the photodegradation process is

312 substantial, since it influences the electric charge of the particles. Indeed, in acidic medium, the
313 dye adsorption occurs at the surface of TiO₂ particles, may be due to the electrostatic attraction
314 between the positively charged TiO₂ and the negatively charged dye. However, the rate of
315 photodegradation decreases with increasing pH (pH > 6). In the present study, the azucryl red is
316 deprotonated and negatively charged for the basic pH. The same applies to TiO₂ (Katz, McDonagh,
317 Tijjing, & Shon, 2015), and the repulsion between the azucryl red anion and TiO₂ particles
318 decreases the initial velocity of the dye pollutant degradation.

319

320 3.4.3 Effect of AR initial concentration

321 Figure 5c depicts the influence of the initial Azucryl red concentration (5, 10, 15, and 20
322 mg. L⁻¹) on the photocatalytic degradation process in the presence of 0.2 mg of pure and Cu-doped
323 TiO₂ catalysts. The Lagergren constant and r² values are summarized in Table 3. The obtained
324 results prove that the effect of the initial concentration of Azucryl red is inversely proportional to
325 the rate of degradation. Indeed, the higher the concentration of Azucryl red, the longer the
326 degradation time is, until 15%Cu doping concentration. Beyond this value, the effectiveness of
327 AR degradation has increased, which can be explained by the enhancement in the pollutant-surface
328 attraction to the photocatalyst. Thus, for the same period of irradiation, i.e., 60 min, the degradation
329 yield of Azucryl red in solutions of 5 and 10 mg. L⁻¹ is found to be 85.2 and 57.55%, respectively.

330 In contrast, for higher concentration of 15 and 20 mg. L⁻¹, the degradation yield reaches a
331 steady state; i.e. only 45.96 and 46.89%. Several authors reported this phenomenon related to the
332 efficiency of photocatalytic degradation and the decrease in constant velocity with increasing
333 substrate concentration (da Silva & Faria, 2003). This result can be explained by the fact that when

334 the dye concentrated in a solution is high; the phenomenon of diffusion of light by the dye
335 molecules will be essential and will have a negative influence on the degradation activity.

336 The photodegradation of AR at the surface of the tested adsorbents may be controlled by
337 the intra-particle rate model, since, q_t and $t^{1/2}$ convene a linear correlation (Fig. 7 (a, b, and c)).
338 Besides, the regression coefficient values, regardless of the studied conditions, are mostly higher
339 than 0.96, signifying that the data are strictly appropriate to the model. The thickness of the
340 boundary layer is highly matched up with the C values. Nevertheless, Figure 7 (a, b, and c), linear
341 curves do not pass across the origin. This annotation can be ascribed to some level of boundary
342 layer control. The mentioned phenomenon is an allusion that the intra-particle diffusion is not the
343 only rate-limiting stage, as supplementary kinetic processes may affect the photodegradation
344 activity. In other words, the as mentioned physic-chemical parameters are intervening concurrently
345 and cannot be neglected (Royer et al., 2009; Yazdani, Tuutijärvi, Bhatnagar, & Vahala, 2016).
346 Meanwhile, this finding can be correlated by the results obtained using mass transfer model,
347 namely, the high values of regression coefficients (Tables 3). Besides, linear relation has been
348 observed (Fig. 6 (a, b, and c)), and the regression coefficients are higher than 0.91, except in the
349 case of Cu-doping concentration 10% and pH 6.2 (0.7966, 0.8324, respectively). This finding
350 points out that the photodegradation kinetic process may also be controlled by mass transfer
351 through a liquid-particle interface, namely, the mass transfer by convection (Royer et al., 2009).

352 **3.5. Photodegradation Equilibrium study of azucryl red**

353 Photodegradation isotherms can represent the equilibrium established between the azucryl
354 red on the adsorbent and the unadsorbed component in the solution. The most widely used isotherm
355 equations for equilibrium data modeling are Langmuir (Langmuir, 1918), Freundlich (Freundlich,

1907), Temkin (Allen, Mckay, & Porter, 2004), and Dubinin-Radushkevich (D-R) (da Silva & Faria, 2003) isotherms and they are expressed as follow:

$$\frac{C_e}{q_e} = \frac{1}{q_m} C_e + \frac{1}{q_m \cdot b} \quad \text{linear form of Langmuir equation} \quad (10)$$

$$\text{Ln}q_e = \frac{1}{n} \text{Ln}C_e + \text{Ln}K \quad \text{linear form of Freundlich equation} \quad (11)$$

$$q_e = \frac{RT}{B} \text{Ln}K_T + \frac{RT}{B} \text{Ln}C_e \quad \text{linear form of Temkin equation} \quad (12)$$

$$\text{Ln}q_e = \text{Ln}q_m - \beta \varepsilon^2 \quad \text{linear form of D-R equation} \quad (13)$$

$$\text{with } \varepsilon = RT \text{Ln}\left(1 + \frac{1}{C_e}\right)$$

where q_m is the concentration of sorbed adsorbate at nanoparticle-phase, corresponding to the complete monolayer coverage of adsorption sites (Langmuir, 1918), and b is the Langmuir constant concerning the adsorption free energy. The q_m and b values are evaluated from the Langmuir model linear plot (Eq. 9): the slope is $1/q_m$ and the intercept $1/(q_m \cdot b)$. The Freundlich model's constants, k and n , are determined from the slope and intercept of the plot $\text{Ln}(q_e)$ against $\text{Ln}(C_e)$ (Eq. 10), respectively. K and n constants may be correlated to the strength of the adsorptive bond and the bond distribution, respectively (Freundlich, 1907). The Temkin's isotherm presumes that the heat of adsorption of all molecules increases linearly with coverage (Allen et al., 2004). The linear form of this isotherm can be given by Eq. (11), where B and K_T are Temkin constants, respectively related to the heat of sorption ($\text{J} \cdot \text{mol}^{-1}$) and Temkin's isotherm constant ($\text{L} \cdot \text{mg}^{-1}$). The constants B and K_T are determined from the slope and the intercept of the linear plot of q_e against $\text{Ln}(C_e)$. Dubinin-Radushkevich isotherm expression is represented by the equation (12) (Ayawei, Ekubo, Wankasi, & Dikio, 2015), where ε is Polanyi potential, β is D-R's constant, R is gas constant ($8.31 \text{ J} \cdot \text{mol}^{-1} \cdot \text{K}^{-1}$), T is absolute temperature, and $E = (2\beta)^{-0.5}$ is mean adsorption energy.

377 It is well-known that the adsorption isotherms help in describing the degradation process
378 at equilibrium conditions. Figure 8 depicts the photodegradation equilibrium isotherms of AR dye
379 (5, 10, 15, 20, and 25 mg.L⁻¹) by Cu_{2%}-TiO₂ nanopowder for 60 minutes at room temperature and
380 5.0 ± 0.2 pH. Besides, the batch experiments data were modeled by the four previously described
381 isotherm models using the least-squares method. The linearized Langmuir, Freundlich, Temkin
382 and D-R isotherms of AR dye in solutions are plotted. Consequently, the respective model
383 parameters (q_m , b , K , n , B , KT , ε , and β) with a correlation coefficient (r^2) are determined by the
384 simulation of the different models by using Origin 8 software and summarized in Table 4.

385 The substantial characteristic of the Langmuir model may be manifested in the expression of
386 equilibrium parameter R_L , which is a constant known as the separation factor (Langmuir, 1918):

$$387 \quad R_L = \frac{1}{1 + b.C_0} \quad (14)$$

388 where R_L value indicates the adsorption nature as irreversible if $R_L = 0$, favorable if $0 < R_L < 1$,
389 linear if $R_L = 1$ and unfavorable if $R_L > 1$.

390 Taking into account the high values of the regression coefficient ($r^2=0.9935$) as shown in
391 Table 4, it should be pointed out that the Freundlich model exhibits a better fitting for AR dye
392 photodegradation equilibrium data (Fig. 8 and Table 4). Based on this finding, it can be highlighted
393 that the photodegradation data are better described by Freundlich isotherm compared to the
394 remaining models' isotherms, particularly Langmuir. However, from the obtained results, the
395 maximum monolayer coverage capacity (q_m) from the Langmuir isotherm model has been found
396 to be 120 mg.g⁻¹ and R_L is about 0.03, indicating that the equilibrium photodegradation is explicitly
397 more favorable ($r^2 = 0.9651$).

398

399 **3.6 Photocatalytic degradation mechanism**

400

401 A plausible mechanism for the photocatalytic degradation of AR dye is illustrated in Fig.

402 9. By means of illumination of the as-prepared $\text{Cu}_x\%-\text{TiO}_2$ nanocatalysts (NC) with UV-light; the

403 electrons in the valence band (VB) are excited to the conduction bands (CB), consequently an

404 equivalent number of holes are formed in the VB.

405 By means of radiation, the as-prepared $\text{Cu}_x\%-\text{TiO}_2$ nanocatalysts (NC) with UV-light

406 wavelength greater than or equal to its bandgap, the electrons move from the valence band (VB)

407 to the conduction band (CB) forming an equivalent number of an oxidizing site (hole h^+) and a

408 reduction site (electron e^-) (Ajmal, Majeed, Malik, Idriss, & Nadeem, 2014). Meanwhile, TiO_2 is

409 a semiconductor with a wide bandgap. Furthermore, the incorporation of Cu^{2+} into the TiO_2 matrix

410 will produce intermediate energy levels that occasionally retard in the charge carrier's

411 recombination and, consequently, reinforce the photocatalytic efficiency. The latter process is the

412 oxidation-reduction reaction of the adsorbed reagents on the NC particle surface (Ajmal et al.,

413 2014), the charges arriving at the photocatalyst surface react with the adsorbed substances that can

414 accept or donate electrons. The electrons react with oxygen to yield the superoxide radicals (Kaur

415 & Singhal, 2014), which may interact with the cations H^+ to form the radical hydro-peroxide HO_2 :

416 (i) and the formation of hydrogen peroxide that decomposes HO^{\bullet} hydroxide radical (ii) to the

417 catalyst surface after excitation by lengths appropriate wave. The h^+ holes react with the reagents

418 adsorbed on the NC surface liable to give electrons such as water, adsorbed OH^- anions or directly

419 organic pollutants (AR), producing highly oxidizing hydroxyl radicals which can oxidize the

420 organic molecules adsorbed on TiO_2 surface until complete mineralization into CO_2 and H_2O as

421 shown in Figure 9.

422

423 **3.7 Regeneration of the adsorbent**

424 The reusability of photocatalysts nanocomposite is required for their feasible utilization on
425 a large scale for industrial applications. The procedure used to remove dyes molecules from
426 wastewater systems is considered upon numerous criteria like cost, efficiency, and reusability. The
427 recycling of adsorbents and the renewal of adsorbate is a cost-effective process. In this context,
428 the regeneration of the 2% Cu-doped TiO₂ nanoparticles from the dye mixture has been
429 investigated for five cycles. After the photodegradation process, the nanocatalyst was only
430 thoroughly washed with deionized hot water and oven-dried at 100 °C for three hours. The
431 measurements are realized similarly to the photodegradation kinetics and equilibrium process, i.e.,
432 pH = 5.0 during 60 minutes of contact time. The adsorption–desorption process has been repeated
433 up to five cycles to test the reusability of the nanoparticles for AR dye molecules degradation. As
434 depicted in Fig. 10, 2% Cu-doped TiO₂ nano-photocatalyst indiscernibly displays a higher degree
435 of stability, and its efficiency is found to be relatively significant for AR dye after five sequential
436 cycles. It is also important to highlight that 2% Cu-TiO₂ nanoparticles fabricated via an amended
437 sol-gel method can be recovered effectively for the reuse in photodegradation for more than five
438 cycles.

439 **4. Conclusion**

440 TiO₂ and Cu-TiO₂ nanopowders were synthesized, characterized, and tested as catalysts in
441 the azucryl red dye's photodegradation. XRD data indicated a rutile-anatase phase transformation
442 for 2% Cu doping, while the crystallite size was slightly affected by Cu loading (around 30 nm).
443 The morphology of the nanoparticles has transformed from spherical-shaped into agglomerated
444 large nanoparticles upon Cu doping. The BET analysis revealed a sudden drop in surface area from
445 16.5 m².g⁻¹ for TiO₂ to about 1.0 m².g⁻¹ after 2% Cu-doping content addition. Photocatalytic

446 efficiency up to 86 % was achieved at 2% Cu-doping content, at 25°C and a pH of 5.0. The kinetics
447 study of the dye photodegradation process was proved to obey the Lagergren first-order law,
448 regardless of experimental conditions (i.e., %Cu doping concentration, pH, and initial dye
449 concentration). The kinetic investigation indicated that the rate is not only controlled by the intra-
450 particle diffusion process, particularly at its initial phase. Instead, it was proved that this process
451 might also be controlled by mass transfer through the liquid phase's boundary layer. Besides, the
452 Freundlich isotherm yielded a better fitting of the equilibrium data of the photodegradation
453 process. Apart from being highly reusable, the photocatalyst showed excellent performance at a
454 low Cu concentration Cu also. The as-obtained findings on Cu-doped TiO₂ nanopowder
455 demonstrated it as a potential nanocatalyst contender for organic dyes photodegradation.

456 **Authors' contributions:**

457 Ahmed Kerrami: validation, investigation, visualization, resources, writing –original draft, writing
458 - review & editing.

459 Lotfi Khezami: methodology, validation, writing –original draft, writing - review & editing.
460 validation, investigation.

461 Mohamed Bououdina: conceptualization, validation, writing –original draft, writing - review &
462 editing, visualization, supervision.

463 Laila Mahtout: conceptualization, methodology, resources, supervision.

464 Abueliz Modwi: methodology, validation, investigation, review & editing.

465 Souhila Rabhi: validation, investigation, review & editing.

466 Faycal Bensouici, Hayet Belkacemi: conceptualization, methodology, resources, supervision.

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474

475 **References**

- 476 Ajmal, A., Majeed, I., Malik, R. N., Idriss, H., & Nadeem, M. A. (2014). Principles and mechanisms of
477 photocatalytic dye degradation on TiO₂ based photocatalysts: a comparative overview. *RSC*
478 *Advances*, 4(70), 37003-37026.
- 479 Allen, S., Mckay, G., & Porter, J. F. (2004). Adsorption isotherm models for basic dye adsorption by peat
480 in single and binary component systems. *Journal of colloid and interface science*, 280(2), 322-
481 333.
- 482 Anpo, M., & Takeuchi, M. (2003). The design and development of highly reactive titanium oxide
483 photocatalysts operating under visible light irradiation. *Journal of Catalysis*, 216(1-2), 505-516.
- 484 Asahi, R., Morikawa, T., Ohwaki, T., Aoki, K., & Taga, Y. (2001). Visible-light photocatalysis in nitrogen-
485 doped titanium oxides. *Science*, 293(5528), 269-271.
- 486 Ayawei, N., Ekubo, A. T., Wankasi, D., & Dikio, E. D. (2015). Adsorption of congo red by Ni/Al-CO₃:
487 equilibrium, thermodynamic and kinetic studies. *Oriental Journal of Chemistry*, 31(3), 1307.
- 488 Baneshi, J., Haghghi, M., Jodeiri, N., Abdollahifar, M., & Ajamein, H. (2014). Homogeneous precipitation
489 synthesis of CuO–ZrO₂–CeO₂–Al₂O₃ nanocatalyst used in hydrogen production via methanol
490 steam reforming for fuel cell applications. *Energy Conversion and management*, 87, 928-937.
- 491 Bartl, M. H., Boettcher, S. W., Frindell, K. L., & Stucky, G. D. (2005). 3-D molecular assembly of function in
492 titania-based composite material systems. *Accounts of chemical research*, 38(4), 263-271.
- 493 Bartl, M. H., Puls, S. P., Tang, J., Lichtenegger, H. C., & Stucky, G. D. (2004). Cubic mesoporous
494 frameworks with a mixed semiconductor nanocrystalline wall structure and enhanced sensitivity
495 to visible light. *Angewandte Chemie International Edition*, 43(23), 3037-3040.
- 496 Bindu, P., & Thomas, S. (2014). Estimation of lattice strain in ZnO nanoparticles: X-ray peak profile
497 analysis. *Journal of Theoretical and Applied Physics*, 8(4), 123-134.
- 498 Burda, C., Lou, Y., Chen, X., Samia, A. C., Stout, J., & Gole, J. L. (2003). Enhanced nitrogen doping in TiO₂
499 nanoparticles. *Nano letters*, 3(8), 1049-1051.
- 500 Chen, C., Li, X., Ma, W., Zhao, J., Hidaka, H., & Serpone, N. (2002). Effect of transition metal ions on the
501 TiO₂-assisted photodegradation of dyes under visible irradiation: a probe for the interfacial
502 electron transfer process and reaction mechanism. *The Journal of Physical Chemistry B*, 106(2),
503 318-324.
- 504 da Silva, C. G., & Faria, J. L. (2003). Photochemical and photocatalytic degradation of an azo dye in
505 aqueous solution by UV irradiation. *Journal of Photochemistry and Photobiology A: Chemistry*,
506 155(1-3), 133-143.

507 Del Carro, L. (2018). *Sintering of copper nanoparticle pastes for microelectronic packaging*. ETH Zurich,

508 Einaga, H., Harada, M., Futamura, S., & Ibusuki, T. (2003). Generation of active sites for CO

509 photooxidation on TiO₂ by platinum deposition. *The Journal of Physical Chemistry B*, 107(35),

510 9290-9297.

511 El Haddad, M., Regti, A., Laamari, M. R., Mamouni, R., & Saffaj, N. (2014). Use of Fenton reagent as

512 advanced oxidative process for removing textile dyes from aqueous solutions. *J Mater Environ*

513 *Sci*, 5(3), 667-674.

514 Fernández-García, M., Martínez-Arias, A., Iglesias-Juez, A., Belver, C., Hungria, A., Conesa, J., & Soria, J.

515 (2000). Structural characteristics and redox behavior of CeO₂-ZrO₂/Al₂O₃ supports. *Journal of*

516 *Catalysis*, 194(2), 385-392.

517 Freundlich, H. (1907). Über die adsorption in lösungen. *Zeitschrift für physikalische Chemie*, 57(1), 385-

518 470.

519 Gómez, R., López, T., Ortiz-Islas, E., Navarrete, J., Sánchez, E., Tzompanztzi, F., & Bokhimi, X. (2003).

520 Effect of sulfation on the photoactivity of TiO₂ sol-gel derived catalysts. *Journal of molecular*

521 *catalysis A: Chemical*, 193(1-2), 217-226.

522 Hameed, B., Salman, J., & Ahmad, A. (2009). Adsorption isotherm and kinetic modeling of 2, 4-D

523 pesticide on activated carbon derived from date stones. *Journal of hazardous materials*, 163(1),

524 121-126.

525 Ho, Y.-S., & McKay, G. (1998). Sorption of dye from aqueous solution by peat. *Chemical Engineering*

526 *Journal*, 70(2), 115-124.

527 Hoffmann, M. R., Martin, S. T., Choi, W., & Bahnemann, D. W. (1995). Environmental applications of

528 semiconductor photocatalysis. *Chemical reviews*, 95(1), 69-96.

529 Iwaszuk, A., & Nolan, M. (2011). Charge compensation in trivalent cation doped bulk rutile TiO₂. *Journal*

530 *of Physics: Condensed Matter*, 23(33), 334207.

531 Joshi, K., & Shrivastava, V. (2011). Degradation of alizarine red-S (A textiles dye) by photocatalysis using

532 ZnO and TiO₂ as photocatalyst. *Int. J. Environ. Sci*, 2(1), 8-21.

533 Katz, A., McDonagh, A., Tijing, L., & Shon, H. K. (2015). Fouling and inactivation of titanium dioxide-based

534 photocatalytic systems. *Critical Reviews in Environmental Science and Technology*, 45(17), 1880-

535 1915.

536 Kaur, J., & Singhal, S. (2014). Facile synthesis of ZnO and transition metal doped ZnO nanoparticles for

537 the photocatalytic degradation of Methyl Orange. *Ceramics International*, 40(5), 7417-7424.

538 Kerrami, A., Mahtout, L., Bensouici, F., Bououdina, M., Rabhi, S., Sakher, E., & Belkacemi, H. (2019).

539 Synergistic effect of Rutile-Anatase Fe-doped TiO₂ as efficient nanocatalyst for the degradation

540 of Azucryl Red. *Materials Research Express*, 6(8), 0850f0855.

541 Kim, C.-S., Shin, J.-W., Cho, Y.-H., Jang, H.-D., Byun, H.-S., & Kim, T.-O. (2013). Synthesis and

542 characterization of Cu/N-doped mesoporous TiO₂ visible light photocatalysts. *Applied Catalysis*

543 *A: General*, 455, 211-218.

544 Kormann, C., Bahnemann, D. W., & Hoffmann, M. R. (1988). Photocatalytic production of hydrogen

545 peroxides and organic peroxides in aqueous suspensions of titanium dioxide, zinc oxide, and

546 desert sand. *Environmental science & technology*, 22(7), 798-806.

547 Kumar, A., & Jain, A. K. (2001). Photophysics and photochemistry of colloidal CdS-TiO₂ coupled

548 semiconductors—photocatalytic oxidation of indole. *Journal of molecular catalysis A: Chemical*,

549 165(1-2), 265-273.

550 Langmuir, I. (1918). The adsorption of gases on plane surfaces of glass, mica and platinum. *Journal of the*

551 *American Chemical Society*, 40(9), 1361-1403.

552 Linsebigler, A. L., Lu, G., & Yates Jr, J. T. (1995). Photocatalysis on TiO₂ surfaces: principles, mechanisms,

553 and selected results. *Chemical reviews*, 95(3), 735-758.

554 Lippens, B. C., & De Boer, J. (1965). Studies on pore systems in catalysts: V. The t method. *Journal of*
555 *Catalysis*, 4(3), 319-323.

556 Liu, H.-L., & Yang, T. C.-K. (2003). Photocatalytic inactivation of Escherichia coli and Lactobacillus
557 helveticus by ZnO and TiO₂ activated with ultraviolet light. *Process biochemistry*, 39(4), 475-481.

558 Lopez, T., Moreno, J., Gomez, R., Bokhimi, X., Wang, J., Yee-Madeira, H., . . . Reyes, P. (2002).
559 Characterization of iron-doped titania sol-gel materials. *Journal of Materials Chemistry*, 12(3),
560 714-718.

561 Mahdavinia, G. R., & Zhalebaghy, R. (2012). Removal kinetic of cationic dye using poly (sodium acrylate)-
562 carrageenan/Na-montmorillonite nanocomposite superabsorbents. *Journal of Materials and*
563 *Environmental Science*, 3(5), 895-906.

564 Mbadcam, J. K., Anagho, S. G., & Nsami, J. N. (2011). Kinetic and equilibrium studies of the adsorption of
565 lead (II) ions from aqueous solution onto two Cameroon clays: Kaolinite and smectite. *Journal of*
566 *Environmental Chemistry and Ecotoxicology*, 3(11), 290-297.

567 Morikawa, T., Irokawa, Y., & Ohwaki, T. (2006). Enhanced photocatalytic activity of TiO₂- xNx loaded
568 with copper ions under visible light irradiation. *Applied Catalysis A: General*, 314(1), 123-127.

569 Pandiyarajan, T., & Karthikeyan, B. (2012). Cr doping induced structural, phonon and excitonic
570 properties of ZnO nanoparticles. *Journal of Nanoparticle Research*, 14(1), 647.

571 Park, H. S., Kim, D. H., Kim, S. J., & Lee, K. S. (2006). The photocatalytic activity of 2.5 wt% Cu-doped TiO₂
572 nano powders synthesized by mechanical alloying. *Journal of Alloys and Compounds*, 415(1-2),
573 51-55.

574 Pelizzetti, E., Minero, C., Piccinini, P., & Vincenti, M. (1993). Phototransformations of nitrogen containing
575 organic compounds over irradiated semiconductor metal oxides: Nitrobenzene and atrazine
576 over TiO₂ and ZnO. *Coordination Chemistry Reviews*, 125(1-2), 183-193.

577 Perez-Urquiza, M., Prat, M., & Beltran, J. (2000). Determination of sulphonated dyes in water by ion-
578 interaction high-performance liquid chromatography. *Journal of Chromatography A*, 871(1-2),
579 227-234.

580 Pongwan, P., Wetchakun, K., Phanichphant, S., & Wetchakun, N. (2016). Enhancement of visible-light
581 photocatalytic activity of Cu-doped TiO₂ nanoparticles. *Research on Chemical Intermediates*,
582 42(4), 2815-2830.

583 Royer, B., Cardoso, N. F., Lima, E. C., Vaggetti, J. C., Simon, N. M., Calvete, T., & Veses, R. C. (2009).
584 Applications of Brazilian pine-fruit shell in natural and carbonized forms as adsorbents to
585 removal of methylene blue from aqueous solutions—Kinetic and equilibrium study. *Journal of*
586 *hazardous materials*, 164(2-3), 1213-1222.

587 Sahu, M., & Biswas, P. (2011). Single-step processing of copper-doped titania nanomaterials in a flame
588 aerosol reactor. *Nanoscale Research Letters*, 6(1), 1-14.

589 Shahwan, T. (2015). Lagergren equation: Can maximum loading of sorption replace equilibrium loading?
590 *Chemical Engineering Research and Design*, 96, 172-176.

591 Sorolla II, M. G., Dalida, M. L., Khemthong, P., & Grisdanurak, N. (2012). Photocatalytic degradation of
592 paraquat using nano-sized Cu-TiO₂/SBA-15 under UV and visible light. *Journal of Environmental*
593 *Sciences*, 24(6), 1125-1132.

594 Sotomayor, F. J., Cychosz, K. A., & Thommes, M. (2018). Characterization of micro/mesoporous
595 materials by physisorption: concepts and case studies. *Acc. Mater. Surf. Res*, 3(2), 34-50.

596 Srinivas, B., Shubhamangala, B., Lalitha, K., Anil Kumar Reddy, P., Durga Kumari, V., Subrahmanyam, M.,
597 & De, B. R. (2011). Photocatalytic reduction of CO₂ over Cu-TiO₂/Molecular sieve 5A composite.
598 *Photochemistry and photobiology*, 87(5), 995-1001.

599 Thaminimulla, C., Takata, T., Hara, M., Kondo, J., & Domen, K. (2000). Effect of chromium addition for
600 photocatalytic overall water splitting on Ni-K₂La₂Ti₃O₁₀. *Journal of Catalysis*, 196(2), 362-365.

- 601 Thommes, M., Kaneko, K., Neimark, A. V., Olivier, J. P., Rodriguez-Reinoso, F., Rouquerol, J., & Sing, K. S.
602 (2015). Physisorption of gases, with special reference to the evaluation of surface area and pore
603 size distribution (IUPAC Technical Report). *Pure and Applied Chemistry*, 87(9-10), 1051-1069.
- 604 Ungár, T. (2007). Characterization of nanocrystalline materials by X-ray line profile analysis. *Journal of*
605 *Materials science*, 42(5), 1584-1593.
- 606 Villaseñor, J., Reyes, P., & Pecchi, G. (1998). Photodegradation of pentachlorophenol on ZnO. *Journal of*
607 *Chemical Technology & Biotechnology: International Research in Process, Environmental AND*
608 *Clean Technology*, 72(2), 105-110.
- 609 Viseu, T. M., Hungerford, G., & Ferreira, M. I. C. (2002). Optical and photophysical studies on porphyrin
610 doped TiO₂ matrixes. *The Journal of Physical Chemistry B*, 106(8), 1853-1861.
- 611 William IV, L., Kostedt, I., Ismail, A. A., & Mazyck, D. W. (2008). Impact of heat treatment and
612 composition of ZnO– TiO₂ nanoparticles for photocatalytic oxidation of an azo dye. *Industrial &*
613 *Engineering Chemistry Research*, 47(5), 1483-1487.
- 614 Wu, Z., Dong, F., Zhao, W., & Guo, S. (2008). Visible light induced electron transfer process over nitrogen
615 doped TiO₂ nanocrystals prepared by oxidation of titanium nitride. *Journal of hazardous*
616 *materials*, 157(1), 57-63.
- 617 Xie, Y., & Yuan, C. (2003). Visible-light responsive cerium ion modified titania sol and nanocrystallites for
618 X-3B dye photodegradation. *Applied Catalysis B: Environmental*, 46(2), 251-259.
- 619 Yazdani, M. R., Tuutijärvi, T., Bhatnagar, A., & Vahala, R. (2016). Adsorptive removal of arsenic (V) from
620 aqueous phase by feldspars: Kinetics, mechanism, and thermodynamic aspects of adsorption.
621 *Journal of Molecular Liquids*, 214, 149-156.
- 622 Yu, J. C., Yu, J., Ho, W., Jiang, Z., & Zhang, L. (2002). Effects of F-doping on the photocatalytic activity and
623 microstructures of nanocrystalline TiO₂ powders. *Chemistry of Materials*, 14(9), 3808-3816.
- 624 Zak, A. K., Majid, W. A., Abrishami, M. E., & Yousefi, R. (2011). X-ray analysis of ZnO nanoparticles by
625 Williamson–Hall and size–strain plot methods. *Solid State Sciences*, 13(1), 251-256.

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635 **Figure Captions**

636 **Fig. 1** X-ray diffraction patterns of copper-doped TiO₂ nanopowders.

637 **Fig. 2** SEM images with corresponding EDS spectra of (a) pure TiO₂, and (b) 2% Cu-doped TiO₂
638 nanopowders.

639 **Fig. 3** Infrared spectra of undoped and Cu-doped TiO₂ nanopowders.

640 **Fig. 4** N₂ adsorption-desorption Isotherm of (a) pure TiO₂, (b) 2%Cu-TiO₂ nanopowders and
641 inset figures are BJH pore size distribution of pure TiO₂, 2%Cu-TiO₂ catalysts.

642 **Fig. 5** Kinetics UV photocatalytic degradation at different conditions of (a) Cu-doping loading;
643 (b) concentration initial pH, and (c) initial concentration of AR.

644 **Fig. 6 (a)** Mass transfer model plots for the photodegradation of 0.5 mg.L⁻¹ AR by Cu_x%-TiO₂
645 nanopowder at pH 5.0; **(b)** Mass transfer model plots for the photodegradation of AR by Cu₂%-
646 TiO₂ nanopowder at different solution initial pH; **(c)** Mass transfer model plots for the
647 photodegradation of different initial doses of AR by Cu₂%-TiO₂ nanopowder at pH 5.0.

648 **Fig. 7 (a)** Intra-particle diffusion plots for photodegradation of 0.5 mg.L⁻¹ AR by Cu_x%-TiO₂
649 nanopowder at pH 5.0; **(b)** Intra-particle diffusion plots for photodegradation of AR by Cu₂%-
650 TiO₂ nanopowder at different solution initial pH; **(c)** Intra-particle diffusion plots for
651 photodegradation of different initial doses of AR by Cu₂%-TiO₂ nanopowder at pH 5.0.

652 **Fig. 8** photodegradation isotherm models for AR dye by Cu-doped TiO₂ nanopowders using
653 batch mode procedure at 25°C, adsorbent mass of 0.2 mg; pH fixed at 5.0; and using a contact
654 time of 60 min. Comparison of isotherm models non-linear plot.

655 **Fig. 9** Schematic AR photodegradation over Cu-doped TiO₂ nanocatalysts.

656 **Fig. 10** Reusability of 2% Cu-doped TiO₂ photocatalyst for five successive cycles for the
657 degradation of AR.

658 **Table captions**

659 **Table 1.** Structural and microstructural parameters of pure and 2% Cu-doped TiO₂ nanopowders
660 as obtained from XRD analysis.

661 **Table 2.** BET characteristics of pure and 2% Cu-doped TiO₂ nanopowders.

662 **Table 3.** Intraparticle diffusion, mass transfer, and Lagergren pseudo-first-order models'
663 constants for the photodegradation of AR by pure and Cu-doped TiO₂ nanopowders at different
664 conditions.

665 **Table 4.** Different equilibrium model isotherm constants for AR photodegradation by 2% Cu-
666 doped TiO₂ nanopowder [m = 0.2mg, pH = 5.0, T = 25°C].

Figures

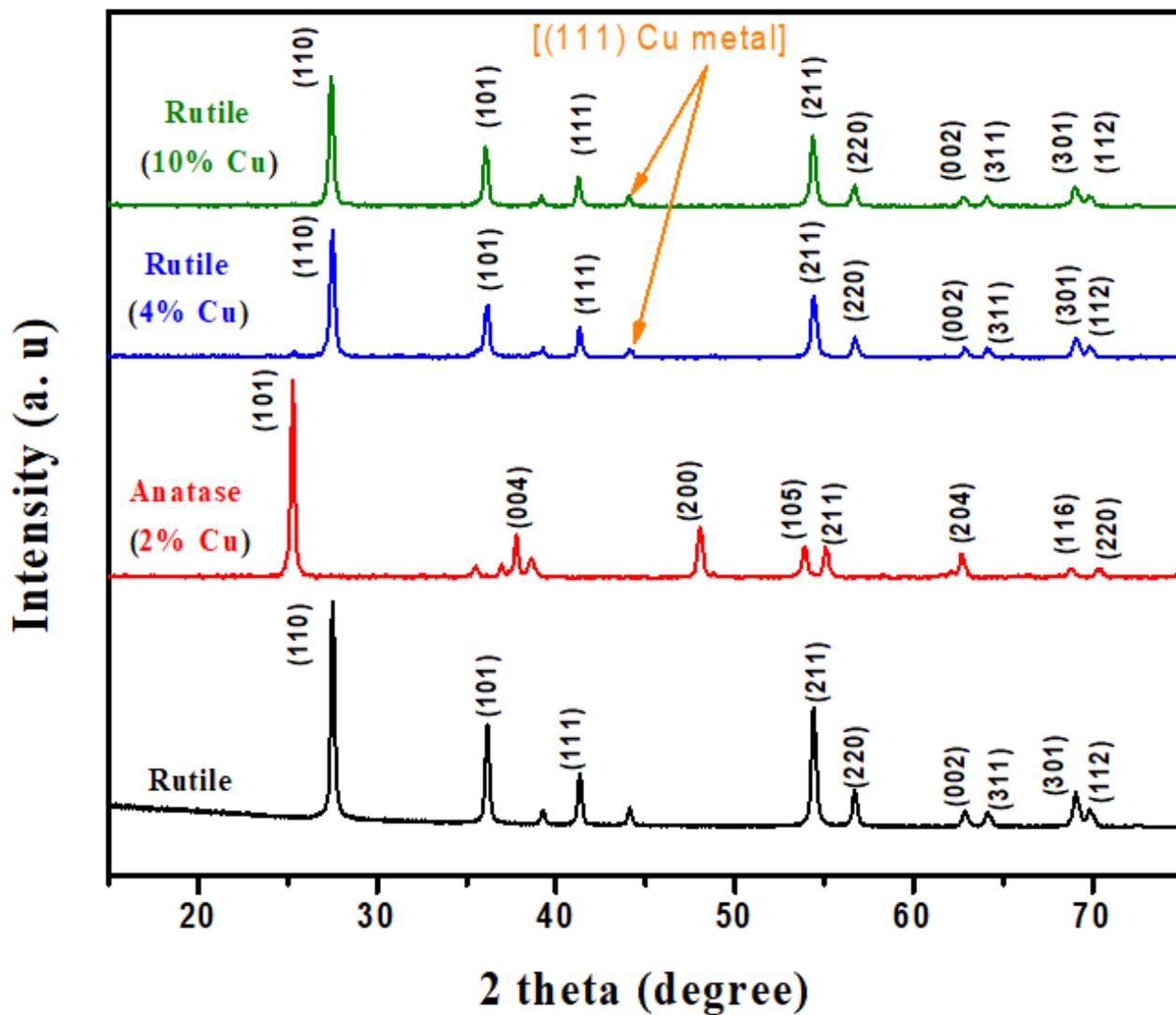


Figure 1

X-ray diffraction patterns of copper-doped TiO₂ nanopowders.

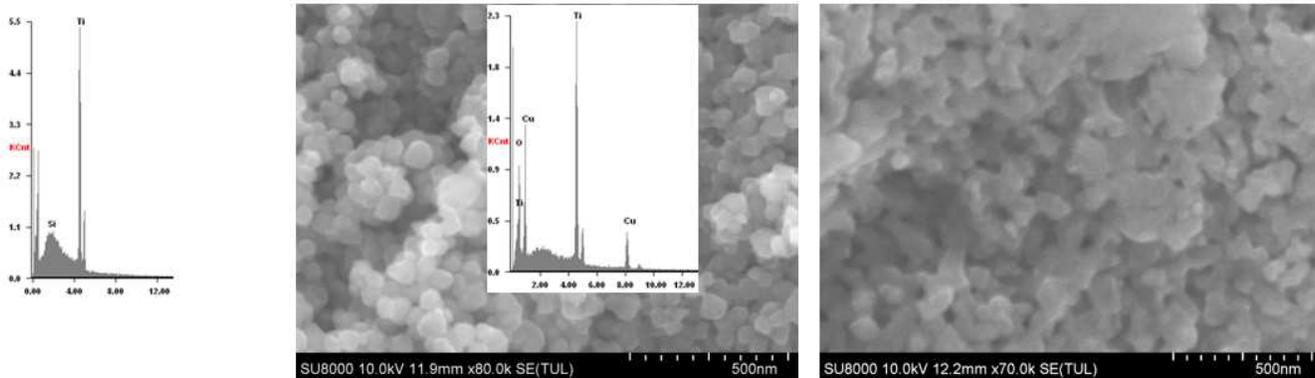


Figure 2

SEM images with corresponding EDS spectra of (a) pure TiO₂, and (b) 2% Cu-doped TiO₂ nanopowders.

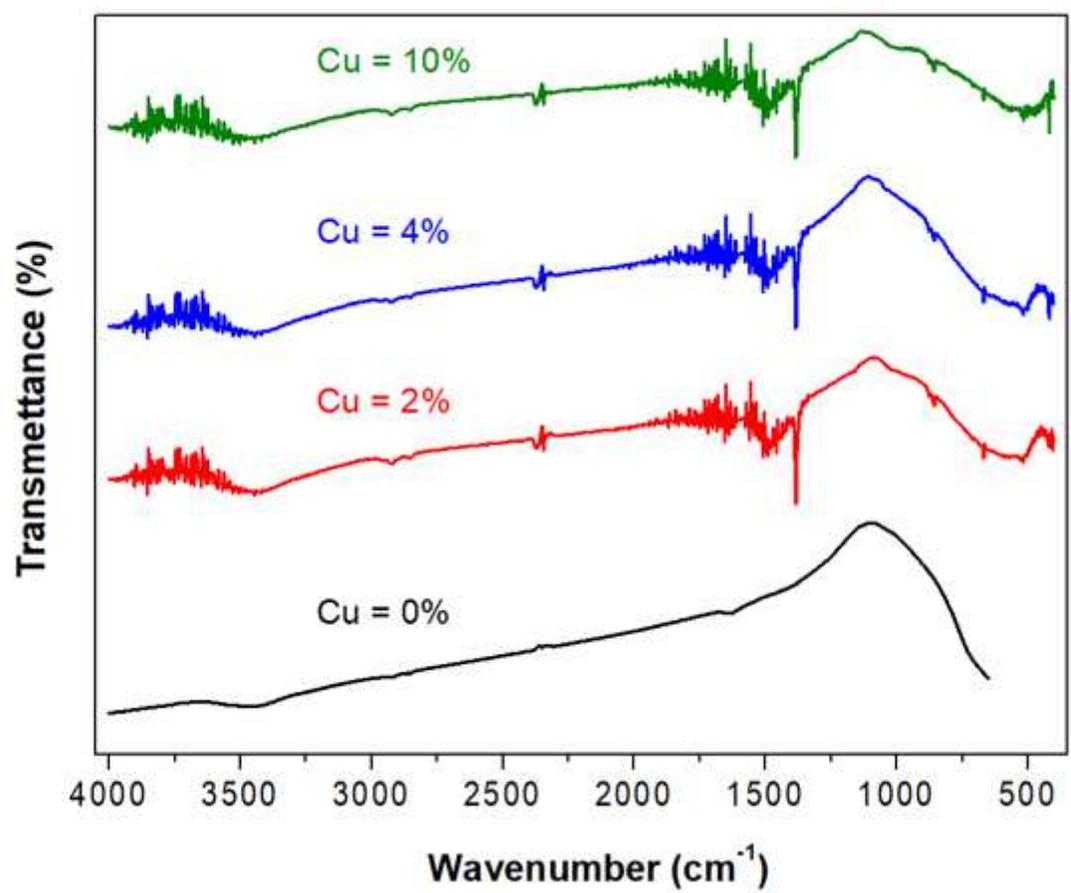


Figure 3

Infrared spectra of undoped and Cu-doped TiO₂ nanopowders.

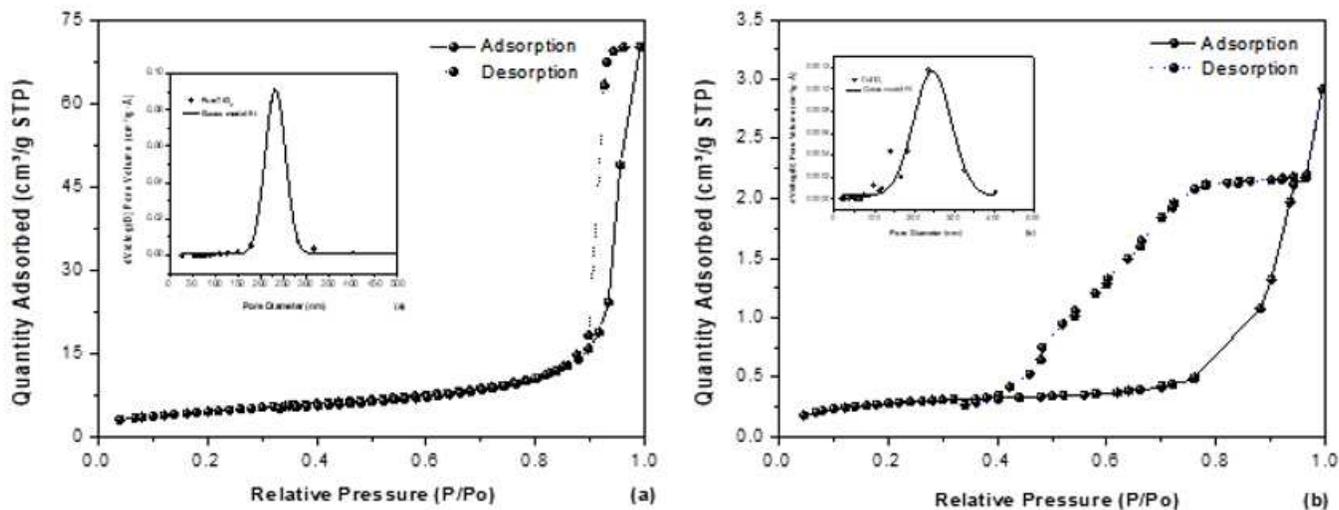


Figure 4

N₂ adsorption-desorption Isotherm of (a) pure TiO₂, (b) 2% Cu-TiO₂ nanoparticles and inset figures are BJH pore size distribution of pure TiO₂, 2% Cu-TiO₂ catalysts.

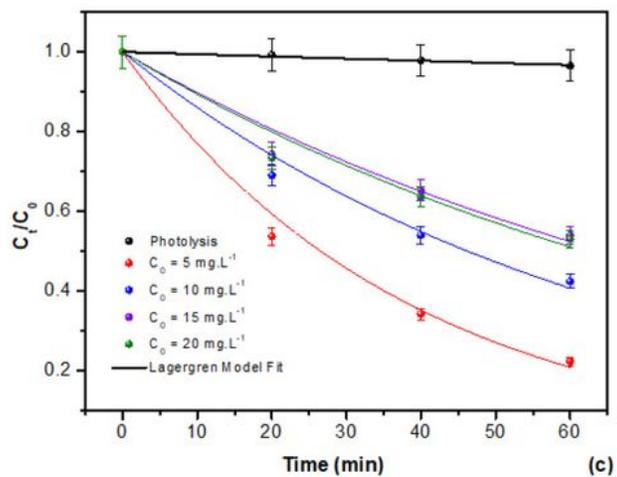
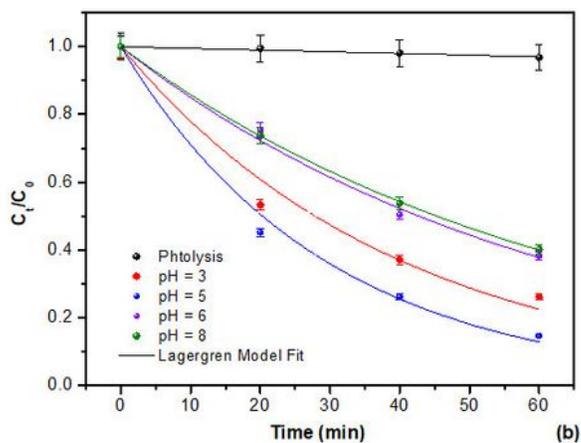
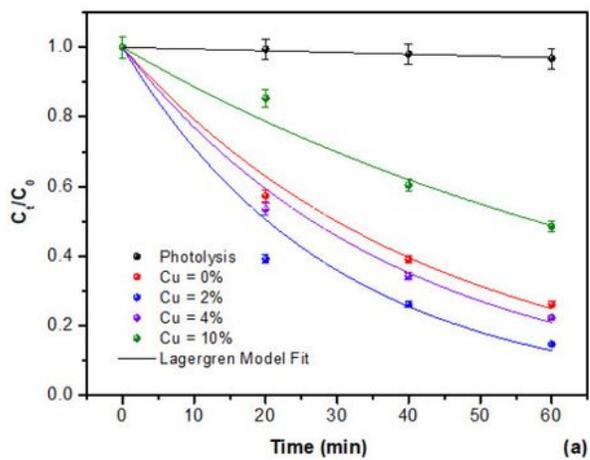


Figure 5

Kinetics UV photocatalytic degradation at different conditions, of (a) Cu-doping loading, (b) concentration initial pH, and (c) initial concentration of Ar

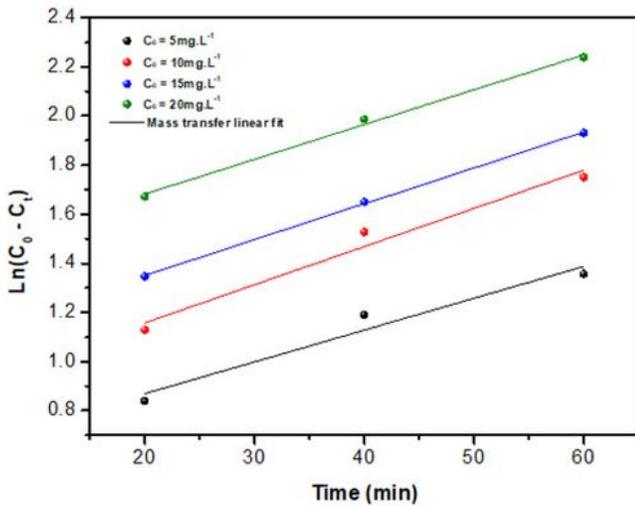
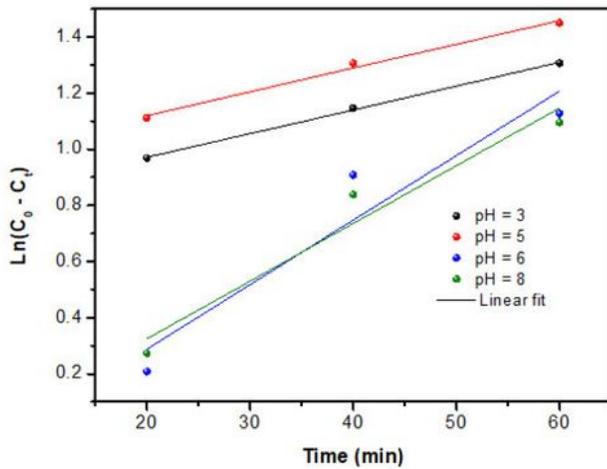
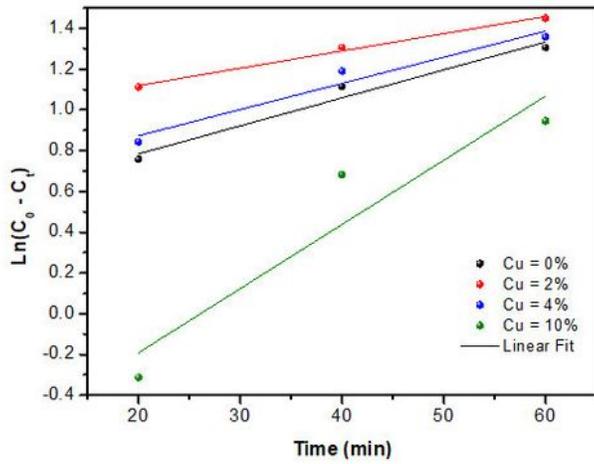


Figure 6

Top panel-a Mass transfer model plots for the photodegradation of 0.5 mg.L⁻¹ AR by Cux%-TiO₂ nanopowder at pH 5.0. Middle panel-b Mass transfer model plots for the photodegradation of AR by Cu₂%-TiO₂ nanopowder at different solution initial pH. Bottom panel-c Mass transfer model plots for the photodegradation of different initial doses of AR by Cu₂%-TiO₂ nanopowder at pH 5.2

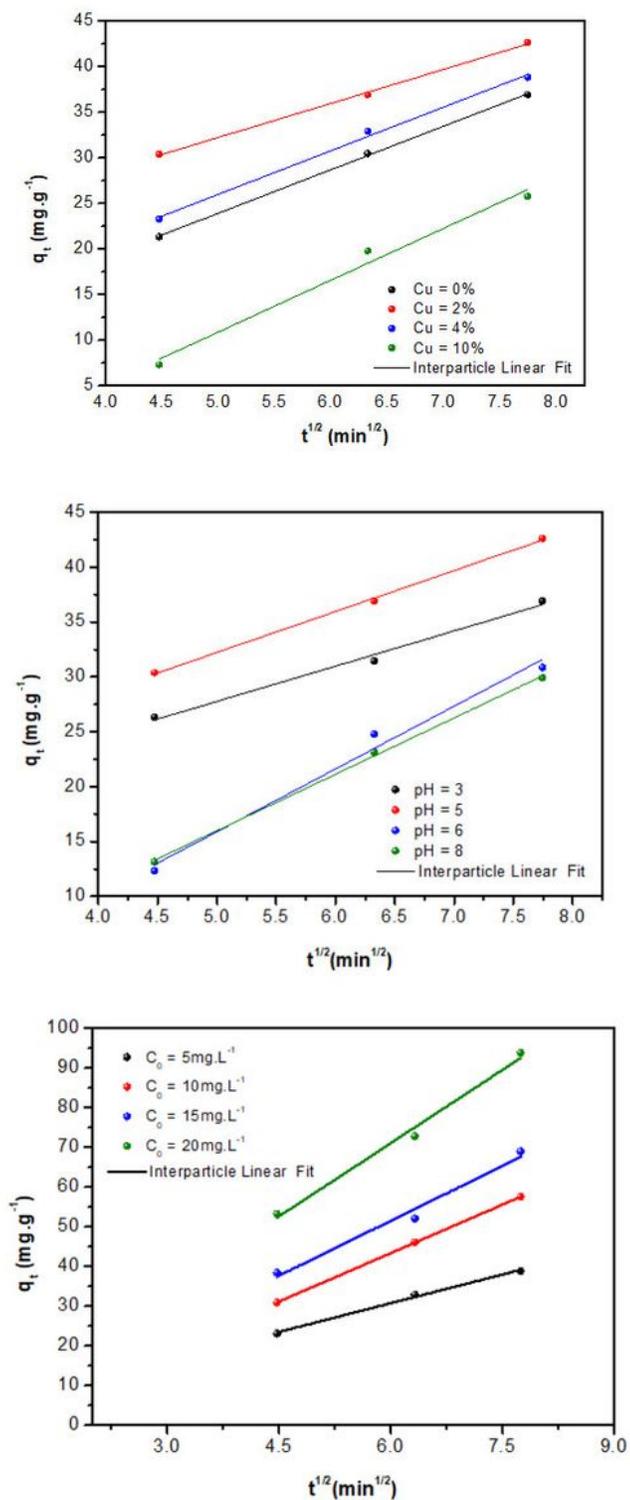


Figure 7

Top panel-a Intra-particle diffusion plots for photodegradation of 0.5 mg.L⁻¹ AR by Cu_x%-TiO₂ nanopowder at pH 5.0. Middle panel-b Intra-particle diffusion plots for photodegradation of AR by Cu₂%-TiO₂ nanopowder at different solution initial pH. Bottom panel-c Intra-particle diffusion plots for photodegradation of different initial doses of AR by Cu₂%-TiO₂ nanopowder at pH 5.0.

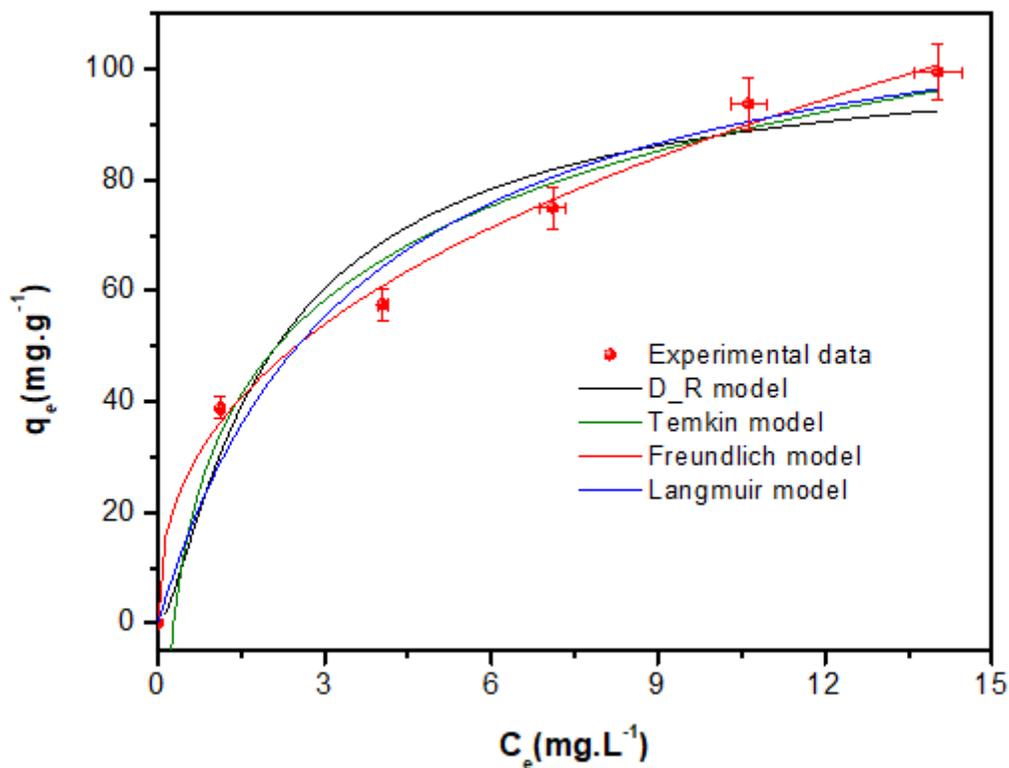


Figure 8

Photodegradation isotherm models for AR dye by Cu-doped TiO₂ nanopowders using batch mode procedure at 25°C, adsorbent mass of 0.2 mg; pH fixed at 5.0; and using a contact time of 60 min. Comparison of isotherm models non-linear plot.

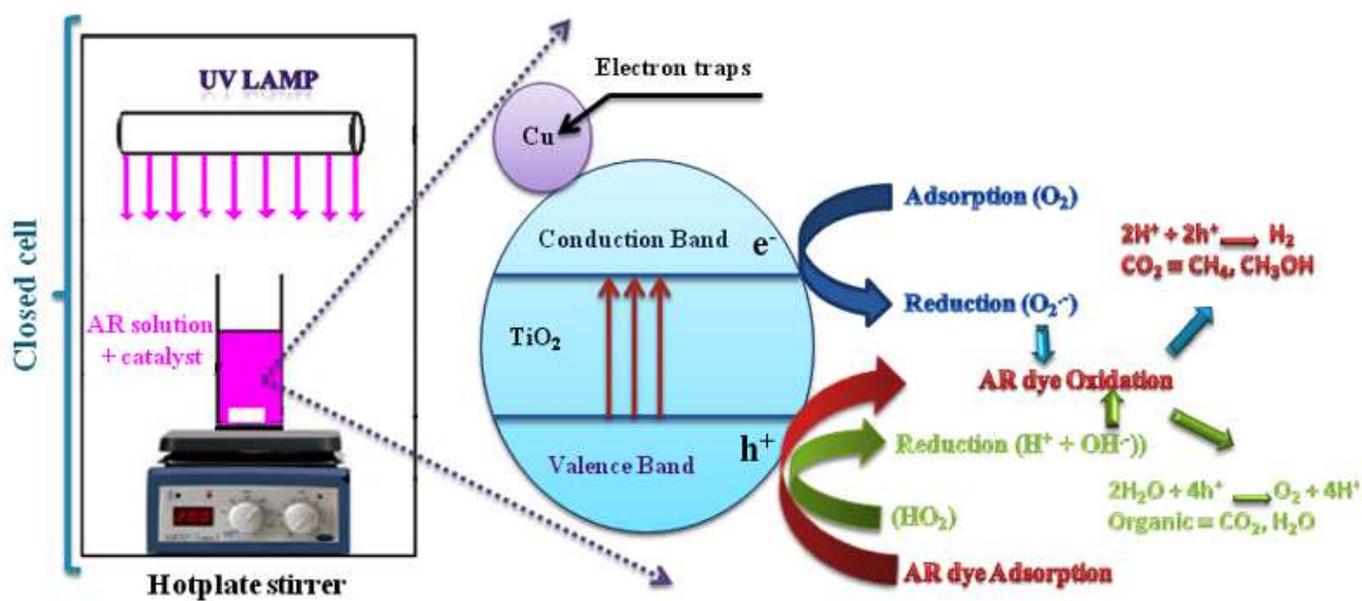


Figure 9

Schematic AR photodegradation over Cu-doped TiO₂ nanocatalysts.

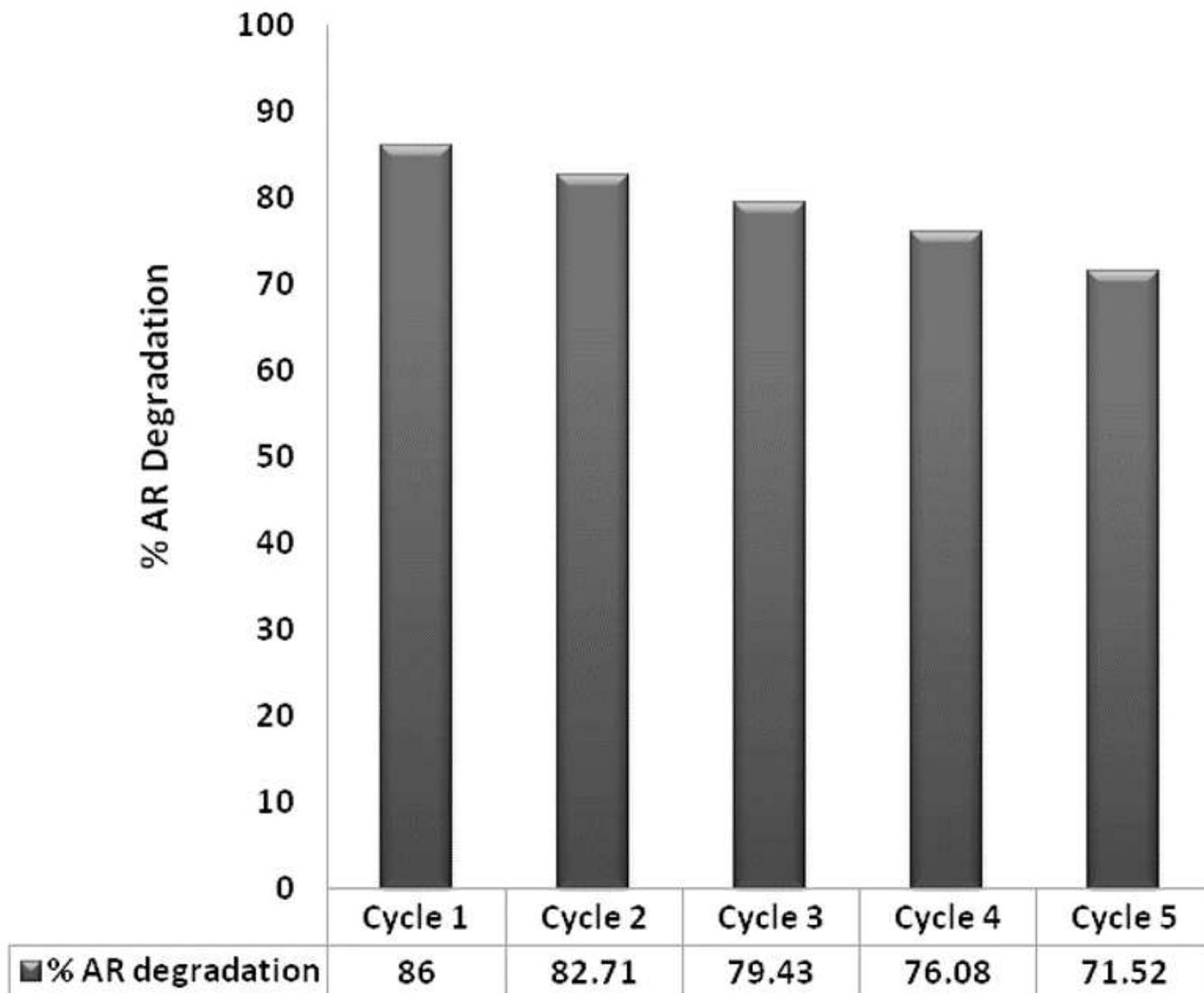


Figure 10

Reusability of 2% Cu-doped TiO₂ photocatalyst for five successive cycles for the degradation of AR.