

# In-Situ Monitoring of Xenobiotics using Genetically Engineered Whole-Cell-Based Microbial Biosensors: Recent Advances and Outlook

Syed Azmal Ali (✉ [alisyedazmal@gmail.com](mailto:alisyedazmal@gmail.com))

National Dairy Research Institute <https://orcid.org/0000-0003-3024-9379>

Deepti Mittal

National Dairy Research Institute

Gurjeet Kaur

University of New South Wales

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## Research Article

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## Abstract

Industrialisation, directly or indirectly, exposes humans to various xenobiotics. The increased magnitude of chemical pesticides and toxic heavy metals in the environment, as well as their intrusion into the food chain, seriously threatens human health. Therefore, the surveillance of xenobiotics is crucial for social safety and security. Online investigation by traditional methods is not sufficient for the detection and identification of such compounds because of the high costs and their complexity. Advancement in the field of genetic engineering provides a potential opportunity to use genetically modified microorganisms. In this regard, whole-cell-based microbial biosensors (WCBMB) represent an essential tool that couples genetically engineered organisms with an operator/promoter derived from a heavy metal-resistant operon combined with a regulatory protein in the gene circuit. The plasmid controls the expression of the reporter gene, such as *gfp*, *luc*, *lux* and *lacZ*, to an inducible gene promoter and has been widely applied to assay toxicity and bioavailability. This review summarises the recent trends in the development and application of microbial biosensors and the use of mobile genes for biomedical and environmental safety concerns.

## Introduction

The xenobiotic is a chemical substance found within an organism that is not naturally produced or expected to be present within the organism. It includes substances that are present in much higher concentrations than are usual. With this definition, the excessive industrialisation and urbanisation in the past few decades have resulted in a drastic increase in the rate of xenobiotic compounds being released into the environment, with long-term impacts on nutrient cycling. The effluent released from industries contain a wide range of environmental compounds such as heavy metals, with various health and environmental hazards (Gursahani et al. 2011). Heavy metals, including cadmium (Cd), lead (Pb), mercury (Hg), arsenic (As), among others are essential constituents of xenobiotics and are being extensively used at industrial levels (Gueu et al. 2007; Karbassi et al. 2008). The likelihood of these heavy metals bioaccumulating in different ecological levels poses a severe threat to biological systems (i.e. organismal organ systems) (Singh et al. 2017). According to the Agency for Toxic Substances and Disease Registry Reports (ATSDR), most of the aforementioned heavy metals are highly hazardous (<https://www.atsdr.cdc.gov/spl/previous/07list.html>).

Many xenobiotics either degrade slowly or are highly resistant to degradation and therefore remain in the environment for an extended period of time. Several different reasons are responsible for the slow degradation of xenobiotics in the natural environment. The foremost is unfavourable physicochemical conditions such as temperature, pH, redox potential, oxygen concentration, and others, including nutrient limitation and auxiliary organic growth substrates (Bachmann et al., 1988) and various substrates such as solubility or predation (Swindoll et al., 1988). On the other hand, the incapacity of microorganisms present in the leading contaminant site to actively metabolise pollutants leads to their persistence in the environment for a more extended time (Katarína et al., 2018). This affects human health by damaging various organ systems (Singh et al. 2017). In this regard, environmental toxicology allows to study the adverse effects of anthropogenic and naturally occurring stressors, both qualitatively and quantitatively (Pranjali et al. 2012). There are several tools available that involve the quantification of xenobiotics, such as gas chromatography-mass spectroscopy (GC-MS) or high-performance liquid chromatography (HPLC). However, these techniques are likely to overestimate the risks as only a part of the total concentration of the pollutant will affect living organisms; specifically, the bioavailable part cannot be determined (Xuemei Liu et al. 2010). Other analytical techniques, such as inductively-coupled plasma mass spectrometry (ICP-MS), inductively-coupled plasma atomic electron spectrometry (ICP/AES), use strong acids during sample digestion, which themselves are pollutants, thereby making it difficult to distinguish between bioavailable and non-bioavailable fractions (Branco et al. 2013). In this sense, it is imperative to develop a practical, efficient and inexpensive method for actively monitoring and determining the concentrations of bioavailable constituents of xenobiotics in the environment. This review therefore focuses on the current knowledge for the detection of heavy metals and xenobiotics using whole-cell-based microbial biosensors (WCBMB).

The selective and sensitive assays were developed to evaluate the real exposure risk scenario of specific pollutants. One of the promising molecular-based tools is the use of whole-cell-based microbial biosensors (WCBMB) that possess considerable value in monitoring environmental contaminants (Ron E 2007). Microbes are considered excellent biosensing components for the construction of a biosensor to detect the bioavailable portion of a pollutant. Moreover, they are cost-effective, have a long lifespan and can survive in an array of suitable pH and temperature values, which significantly facilitates the improvement of the product (Mulchandani et al. 1998). Microbial biosensors might represent the most reliable and sensitive detection technology and can be used without expensive instruments and sophisticated skills. Despite the early prediction of microbial biosensors, there is still a large gap in the understanding and modernisation

of the technology. However, several genotoxicity tests have been developed, such as the Umu-Chromotest, the Ames test and the SOS Chromo test, but only the Umu-Chromotest was proved to be reliable, sensitive and reproducible in terms of assessing environmental genotoxicity (Oda et al. 1985) and was certified according to ISO standardisation (ISO 13829). Microbial biosensor research has focused on three significant applications. First, bacterial monitoring for competition ability and survival (Abby et al. 2003), second, tracking of xenobiotic-sensing or -degrading bacteria for complex environmental samples (Boldt et al. 2004), third, monitoring the degrees of specific environmental pollutants (Boldt et al. 2004; Lejon et al. 2008). Interestingly, the WCBMB technology is currently becoming the most exciting area of research for the detection of environmental pollutants, and it also provides the status of pollutant bioavailability, bioremediation and toxicity.

The biosensor is a device containing immobilised biological material in intimate contact with a compatible transducer, converting the produced biochemical signal into a quantifiable signal (Gronow 1984; Chien and Shih 2007) (Fig. 1). In the biosensor, the bioavailable fraction of the environmental pollutant is identified using a promoter and reporter gene construct. The readout depends on the gene expression, which is usually achieved by creating transcriptional fusions between a promoter and the reporter gene. The downstream expression of marker protein provides a measure of the availability of particular xenobiotics in the contaminated environment (Liu et al. 2004; Nouri et al. 2011). Biosensors supply an inexpensive and straightforward way of ascertaining contaminants; moreover, living organisms in the biosensor also provide information related to the toxicology of different substances and depict the mutagenic consequences of samples with high sensitivity (Fig. 1).

A great deal of concern has been devoted to introducing WCBMB in comparison to traditional methods (Hansen et al. 2001). In this review, we focus on the application of WCBMB as they are more sensitive, selective and specific for the detection of pharmacological pollutants; they provide information about the physiology and toxicology of a sample (Fig. 1). Here, we describe the recent advances in the field of whole-cell biosensors specific to heavy metals and xenobiotic compounds. More specifically, we explore several whole-cell biosensor examples established in biomedical and environmental sensing for the identification of bioavailable fractions of compounds in food specimens, drug screening and pharmacology-related samples (Gui et al. 2017; Mittal et al. 2020). Finally, we underline the future application of microbial biosensors in terms of biomedical and environmental safety concerns.

## **Main Text**

### ***Whole-Cell-Based Microbial Biosensors (WCBMB)***

A great advantage in the field is the use of whole microorganisms, such as bacteria or fungi, for biosensors, along with the bioreceptor component (Vo-Dinh et al. 2000). These microbial biosensors are genetically designed by coupling with a reporter gene to produce a signal with a contaminant-sensing part that replies to chemical or physical changes, such as the influence of a particular analyte (Biran et al. 2003). When these biosensors are brought out to such a change, the detecting component induces the reporter gene through a biochemical pathway inside the cell. Subsequently, the reporter gene generates a measurable activity, such as releasing visible light, which reveals the magnitude of chemical or physical change (Biran et al. 2003; Tauriainen et al. 2000, 2003). Usually, the ability of the cell to sense is altered to integrate the transducer activity. Several parameters can be assayed, such as bioavailability, toxicity and genotoxicity. In this way, it becomes straightforward to estimate the amounts of xenobiotics present at the polluted site. When compared with the use of isolated and purified enzymes, the use of whole cells as biocatalysts presents various benefits, the most significant being the increased stability and security from intervening contents. Therefore, it becomes desirable to use whole-cell biosensors for the assessment of pollution levels.

#### ***(i) Reporter Genes and Proteins for Whole-Cell-Based Microbial Biosensors (WCBMB)***

The expression of any gene for the development of whole-cell biosensors requires a vector or plasmid that acts as a vehicle for transferring a particular gene and further helping in its expression (Kaur et al., 2017). In this scenario, the desired gene must be connected to the reporter gene with a strong or weak promoter. A reporter gene specifically needs to be highly sensitive, have a non-invasive detection capability, and must not be present in the native organism itself (Daunert et al. 2000). During the process, the altered organism with the potential to produce sensing enzymes is induced in the presence of heavy metals or xenobiotics. The whole event activates the reporter gene that provides the amplified signal for the presence of a pollutant in the environment. Table 1 shows some of the biosensors with their

specific host organisms and the relevant heavy metals and xenobiotics. The selectivity of a reporter assay largely depends on the type of target to be detected.

Green Fluorescent Protein (GFP) is a reporter gene derived from jellyfish (Chalfie et al. 1994). It has a specific auto-fluorescence and therefore does not require a substrate, cofactors or ATP and can remain stable for a long duration (Soboleski et al. 2005). Furthermore, the firefly luciferase (*luc*) reporter has been widely incorporated into WCBMB because of its broad linear range (7-8 orders of magnitude) and higher sensitivity (Welsh et al. 1997). Likewise,  $\beta$ -galactosidase (*lacZ*) is considered a highly efficient reporter with an excellent transfection efficiency monitoring capability (Jain et al. 1991). It offers other unique advantages such as rapid and straightforward sample usage via colorimetric and fluorescent methods. In addition to this, the availability of colorimetric or fluorescent substrates for *lacZ* provides the clear advantage of a decreased detection range. A regulatory protein must be able to interact with the target analyte or the contaminant to produce the quantifiable signal, making the entire cell-based biosensor a successful detection platform. Some of the transcription factors have also proved to be responsible for sensing certain pollutants. For example, Kim et al. (2016) developed heavy metal-detecting microbial biosensor by using a *cadC* regulator and a *cadO* operator found in *Bacillus oceanisediminis* 2691, as bacterial populations of ocean sediments contain heavy metal resistance operons. Another biosensor for Cd was fabricated using the GFP reporter gene under the control of *cad* promoter and *cadC* gene of *Staphylococcus aureus* plasmid pI258, with a better response time and a sophisticated detection limit (Kumar et al. 2017). In another report, a metal-responsive operon (*cop*-operon) was used to generate a cell-based biosensor against copper (Cu). Also, a genetically engineered *E. coli* was developed, with enhanced sensitivity and accurate bioavailable Cu quantification (Kang et al. 2018). Furthermore, a whole-cell biosensor harbouring fusion genes responsive to metal toxicants has been prepared based on pZntA and eGFP. The metal efflux system was disrupted to enhance the metal-binding property of ZntR, and the selectivity and sensitivity of the biosensor were increased by changing the amino acid sequence of ZntR (Kang et al. 2018).

### ***(ii) Advancement of Synthetic Biology-Centred Genetic Circuits for Whole-Cell-Based Microbial Biosensors***

A plethora of research demonstrates viable alternatives to conventional and costly analytical methods; however, their realisation under field conditions is challenging. The experimental scenarios in the laboratory are confined to limited and controlled conditions, which are quite different from the field conditions. Accidental release or gene transfer from genetically modified organisms are possible. With the advent of advanced technologies, synthetic biology can be employed to overcome limitations associated with WCBMB made with recombinant DNA technology.

Synthetic biology uses engineering concepts to construct natural genetic circuits to process information that considers cellular components such as biomolecules including proteins and nucleic acids. This genetic element acts as molecular Lego units that can be assembled into functional synthetic circuits with broad applicability (Ali et al., 2017; Saltepe et al. 2017; Wagner et al. 2019). The smart cellular system can navigate complex molecular events during a response signal generated by any external stimuli. It is comprised of diverse transcription factors, promoters, operators, regulators and genes of interest, and the complex networking among various genetic elements can be used to design synthetic genetic circuits to perform novel functions ranging from environmental monitoring to the detection or killing of pathogens (Brophy and Voigt 2014). In this context, the analogy to logic gate signals is essential in setting up a fine-tuning between regulatory systems of the target organism.

One of the promising aspects of synthetic biology is the programming of cellular entities in such a manner that makes them more effective in the real-time monitoring and reporting of analytes. The prime importance of designing a synthetic circuit with the advent of biocomputing increases the number of signals produced in a controlled manner. The emerging field of biocomputing employs computing information to assemble molecular biology parts into the whole cell to purposefully link the genetic circuit to the physiological state of the cell (Goni and Nikel 2019).

### ***Overcoming limitations with genetic engineering***

We have previously discussed the various aspects of WCBMB, and despite their advantages, there are concerns regarding their biosafety, low sensitivity, low limit of detection and confined dynamic ranges (Dana et al. 2012; Shemer et al. 2017). Therefore, substantial research has been diverted in creating suitable interconnected networks with predicted and tuneable behaviour. The significant components of gene expression regulation include ligand-inducible promoter engineering (Chen et al. 2018), translational efficiency tuning (Wang et al. 2014) and control of protein degradation (Camerone and Collins 2014). These components have been widely described in improving the performance of genetically encoded biosensors (Chen et al. 2018; Merulla et al. 2015; Wang et al. 2015). The synthetic biologically driven WCMB hold a promising position in sensing several environmental contaminants with enhanced sensitivity and robustness. Another possible point is the attachment of the nanoparticles which will further improve the efficacy of biosensors in a biological system using the nano-bio interaction (Holzinger et al. 2014; Mittal et al. 2020; Yadav et al. 2021).

Recently, Jia et al. (2019) designed the arsenic (Ar) whole-cell-based biosensor, which significantly provides 10–20 times more robust responses in the presence of an analyte. The authors used a well-studied Ar resistance operon and prepared a whole-cell-based biosensor engineered by synthetic biological approaches. The positive feedback improves the sensitivity of the WCBMB and amplifies the signals to a significant level. The *E. coli* consists of an ars operon with a regulator (arsR), permease (arsB) and reductase (arsC). In the absence of Ar, arsR remains bound and inhibits the transcription from the promoter. They have been constructed purposefully through the mechanism of genetic circuit engineering with an enhanced positive feedback amplifier, using the LuxR auto-regulatory element to the Ar. The two plasmids were created to improve the signal and intensity of the output. One of the plasmids contains Pars and arsR with an mCherry reporter gene, providing the signal intensity with an increased concentration of Ar.

On the other hand, in the second plasmid, mCherry, was replaced with a variant of LuxR placed with a transcriptional activator that is controlled by an arsR- Pars circuit, whereas mCherry, together with LuxR, was placed under the control of PluxI. These plasmids are coupled in the presence of Ar; LuxR in the first plasmid gets activated, which in turn activates the mCherry and LuxR from the second plasmid. Additionally, they both can activate their expression. As a result, a positive feedback amplifier was maintained that potentially enhances the output signal (Jia et al. 2019).

### ***Measure for using the WCBMB biosensors***

Physical biocontainment is a biosafety alternative for utilising cell confinement to bioreactors or encapsulation in alginate or silica beads (Chang and Prakash, 2001; Papi et al., 2005). However, physical containment alone is not sufficiently effective. Therefore, to avoid the uncertainties related to biosafety, risk governance comes into a picture that develops specific strategies to minimise the risks.

Several researchers use plasmids that carry only the gene of interest in genetically engineered biosensors instead of inserting the complete genome. Additionally, the designed vector does not share the homology with mobile elements and the host's genome, which creates another safety possibility (Bensasson et al., 2004). Although these strategies minimise the risks, plasmid transfer is also a chance through other organism's natural transformation mechanisms (Thomas et al., 2005).

In order to provide further safety assurance, we collected seven significant points from the literature for considering measures to be taken while working with genetically engineered biosensors. The detailed points are as follows,

1. A complete understanding of auxotrophy for WCBMB provides the necessary biosafety assurance. Therefore, auxotrophy could be used as an excellent method to reduce the escape risk since engineered bacteria can only be grown in complementing substance. It was found that the amino acid auxotrophy is predominantly utilised, but other auxotrophies, including genes for carbohydrate and lipid metabolism, could also be considered into account for the biosafety assurance of WCBMB (Baba et al., 2006). Nonetheless, auxotrophy is also associated with certain drawbacks such as toxic effects due to overexpression of complement (Vidal et al., 2008) or metabolic cross-feeding where plasmid-free cell relies on their neighbour prototroph for metabolic compounds. The problem was

overcome using the modular multilayered biosafety platforms, including safeGuard or GeneGuard. Furthermore, a more complex four-layered, robust system integrated with riboregulators, engineered modules, auxotrophies and supplemental repressors were prepared that demonstrated minimised escape frequency (Gallagher et al., 2015). In theory, auxotrophy can be extended with engineered hosts that rely on synthetic cofactor for synthetic metabolism. This could create a potential barrier to information transfer. An example is that utilising non-canonical amino acids as a synthetic cofactor for engineered *E. coli* renders six essential genes for ncAA L-4,4'-biphenylalanine (BFA) (Lajoie et al., 2013).

2. Inducible systems could be another strategy of biocontainment achieved by the inducer of gene expression uncommon to the environment. Thus, engineered traits will be displayed under the conditioned and controlled environment.
3. Kill switches or induced lethality is an active containment strategy for biosafety purpose. The engineered microorganism becomes dependent upon the addition of an inducer signal in order to survive. Once the Kill switch gets activated, it leads to the cell membrane disruptive protein; Hok activation further activated cell death (Gerdes et al., 1990). The artificial cell division advancement brought us towards limiting engineered cells life span by automatic cell death. A synthetic gene network includes a riboregulated transcriptional cascade and a recombinase-based cascade of memory units that counts three molecular events (Friedland et al., 2009). These counting circuits can be linked to the intracellular cell cycle cues that limit their life span.
4. Additional regulatory elements could be employed to constrain the gene expression or microbial replication to prevent undesired consequences (Gallagher et al., 2015). The utilisation of the dependency devices provides a clear advantage, for instance, toxin-antitoxin pair in which the activity of small toxin is nullified by cis-encoded antitoxin (Hayes et al., 2011). It is a combination of stable toxin and unstable antitoxin that can be placed at different parts, i.e. toxin on plasmid and antitoxin on the cell's genome (Hayes, 2003).
5. Another approach is based on the DNA barcoding system that makes the task simpler by tracking the engineered microorganism employed in the field. With advancements in DNA sequencing technologies, new synthetic DNA can be sequenced directly in environmental samples. The genetic barcodes can be employed in the synthetic operon or DNA 'watermarks' embedded in multiple genomic locations that could help identify modified microbes in the environment (Gibson et al., 2010). Likewise, we can also incorporate watermarks directly into the synthetic genes via codon usage optimisation (Liss et al., 2012).
6. Refactoring the existing genetic code involves designing novel genomes utilising non-canonical genetic codes (Lajoie et al., 2016, Mukai et al., 2017). It is also known as codon reassignment that takes three routes- sense to stop, stop-to-sense and sense-to-sense or triplet-to-quadruplet with modified aminoacyl tRNA synthetases (aaRS)/tRNA pairs creating alternative code (Bezerra et al., 2015). Such engineered organisms are sufficiently disruptive so that the translated protein is non-functional under the natural organism as it provides natural code to the system, thus preventing information transfer. Sense-to-stop or stop codon suppression can be exploited for codon reassignment through engineered aaRS and tRNA (or orthogonal pair) that can easily recognise non-canonical amino acid (ncAA) without interfering with cellular aaRS/tRNA pair (Chin, 2014, Des Soye 2015). Stop codon assignment within an essential gene provides additional benefit as the absence of charged suppressor tRNA prevents engineered information transfer. Sense-to-sense reassignment is not direct but employed via sense-stop-sense and removing multiple codons (Ostrov et al., 2016), but this strategy is not widely explored. The semantic containment or orthogonality in the genetic code offers double containment benefit, engineered information via altered genetic code and ncAA dependent auxotrophy. Furthermore, engineered microbes can no longer exchange their genetic parts with natural counterparts due to their unique biochemical language.
7. Xenobiology aims to design Xeno nucleic acid (XNA) that do not exist in nature, and the engineered microorganism that harbours it will not survive outside their destined environment. Also, the sugar moiety of the DNA backbone is changed by extensively modifying their functional groups, size and isomerisation states to form anhydrohexitol nucleic acid (HNA) (Hendrix et al., 1997), threose nucleic

acid (TNA) (Schöning et al., 2000), locked nucleic acid (LNA) (Anosova et al., 2015), glycerol nucleic acid (GNA) (Zhang et al., 2005), cyclohexene nucleic acid (CeNA) (Wang et al., 2000). A xenobiotic host would require engineered RNA polymerases and other additional replication and transcription components compatible with XNA. A modification in the nucleobase is also resulting from the diversification of the nucleotide pairs.

### ***Critical parameters used in synthetic biology: biocomputation aspect***

Mechanical model predictions based on the previous assumptions are possible by using various algorithms that can further be applied to make a reliable genetic circuit. Boolean logic gates are helpful in designing predictable computational gene circuits, where genes and proteins can respond according to the changing environmental signals (Nielsen et al. 2016). In this regard, a conceptual mechanical model of *arsR* regulation was developed based on the *Ar* resistance element of *E. coli*. Here, the author tested two types of circuits, one with the feedback circuit hold *arsR* controlling its own as well as the EGFP reporter expression and another one with the uncoupled circuit in which *ArsR* is controlling the expression of EGFP from the promoter of *ars*. The efficiency of the predicted model design was confirmed experimentally, and the model may further be useful for preparing detailed modelling of not only regarding the *Ar* sensor, but also other types of bioreporters (Berset et al. 2017).

Jia et al. (2018) introduced six gene circuits for the whole-cell-based biosensor for Pb by rearranging various regulatory elements in the Pb resistance operon and additionally incorporating positive feedback loops. The operon resistant for Pb encodes six genes and consists of regulatory *pbrRT* on one side and *pbrABCD* on the other side. The native operon was reconfigured by rearranging the position of *pbrRT* and GFP on the same hand under the control of different promoters with positive a feedback loop design, which resulted in an enhanced output signal; this demonstrates the advantage of introducing synthetic biology for purposely reconfiguring the genetic elements for increased sensitivity and specificity of the biosensor (Jia et al. 2018).

Synthetic biological circuits are a genetic engineering framework where biological components are configured to execute logical functions within a cell that resemble electronic circuits. One of the critical categories is the Logic AND gate circuits. The logical AND gate is the signalling channel, where if Signal A AND Signal B are connected, only the anticipated gene product will produce.

The environmental signals depend on the presence or absence of small molecules (*lac*, *trp* operons) (Setty et al., 2003). A complex array of multiple signals integrated by the bacteria enhances the sensing sensitivity and specificity. The bacterial system integrates environmental signals utilising various mechanisms, including a two-component system, transcription factors and small RNA (Hoch and Silhavy, 1995). The introduction of synthetic biology allows the designing of synthetic cellular circuits often intended to implement required functions. For instance, genetic logic gates can be utilised to control the response against stimulus input signals. Various logic gates can be built using transcription factor genes, regulatory elements or protein-protein interactions. The development of these circuits depends upon the particular set of inputs and outputs that can be made modular if connected to different inputs to drive different outputs. Various environmental pollutant-like arsenic (Joshi et al., 2009), xylene (Paitan et al., 2004) and *Pseudomonas aeruginosa*-responsive (Saeidi et al., 2011) sensors can be wired to a multiple-input AND gate, allowing cells to report and neutralise a toxic combinatorial circumstance. It is often possible for transcriptional logic gates to combine promoter or operator sequences to control transcription of the output (Gander et al., 2016, Stanton et al., 2013). In this genetic AND gates, each AND gate integrates two promoter inputs and controls one promoter output. This allows the gates to be layered by having the upstream circuit's output promoter serve as the input promoter for a downstream circuit. The output of the circuit is ON only when all the inputs are ON and the output will be OFF if any of the input is OFF. In one of the AND gate circuit, the orthogonal system from *P. syringae* was utilised to construct *E. coli* with co-activating genes *hrpR/hrpS* (regulated by separate promoters) regulating  $\sigma^{54}$ -dependent transcription. The final output response was generated only when two individual environmental responsive transcriptional inputs expressed (Wang et al., 2011). The AND gate can also be prepared using similar inputs and outputs as promoters. Additionally, a set of promoters having the capability to sense different environmental signals can be concatenated to and resulting AND gates activate only in the presence of all the conditions (Anderson et al., 2007).

Genetic circuits, such as AND gates that include multiple sensory inputs, can also be used to develop the sensitivity and selectivity of the sensor. In this approach, regulated promoters are integrated with NOT gate output to produce a combined NAND gate, thus allowing

accurate measurement of biological control (Wang et al. 2011). In this context, Park and Taffet (2019) developed a uranium (U)-responsive whole-cell-based bioreporter in *Caulobacter crescentus* by integrating a two-component signalling system along with UzsRS and UrpRS and the signal amplifier protein UzcY within the AND gate circuit (Park and Taffet 2019). In another study, the multiple heavy metal ion-sensing heterologous transcriptional modules of *E. coli* were made in AND gate fashion that was able to produce fluorescent signals. The authors prepared an AND logic gate that was based on the previously known hrp (hypersensitive response and pathogenicity) operon from *P. syringae*. First, the two-input AND gate was prepared, in which HrpR and HrpS, under the control of the Ar-responsive promoter, produced from the As<sup>2+</sup> promoter, is noninverted. The HrpV, under the supervision of the Hg-responsive promoter, is produced from inverted Hg<sup>2+</sup> input. The output will be turned ON only in the presence of Ar. In the next strategy, the two-input AND gate was extended to a three-input gate in which HrpS was controlled by the Ar-responsive promoter and HrpR by the QS-responsive promoter (Wang et al. 2014).

Another critical part of the genetic circuits is a toggle switch that can also be introduced into the biosensor to reduce the background noise created by the fluorescent signals. Wu and co-workers (2009) constructed a toggle switch for a whole-cell Cd sensor. The toggle circuit was designed with the Cd (III)-sensitive regulatory promoter PcadR of *Pseudomonas putida* 06909 fused with gfp and lacIq genes. The Ptac, which is an IPTG-inducible promoter, was placed upstream of the cadR gene encoding CadR protein that, in the presence of IPTG, represses the PcadR promoter. The presence of Cd ions leads to the expression of LacI and GFP proteins. The LacI represses the Ptac promoter, which consequently leads to the decreased expression of the cadR gene and the increased expression of the GFP gene. The detection limit of the toggle switch was 20 times less than that of the non-toggle circuit with lower background fluorescence (Wu et al. 2009).

### ***Enhanced sensitivity: multilayering approach***

The growing field of synthetic biology offers new avenues for the precise construction of novel engineered biosensors, but the problem of sensitivity persists, depending on several factors. Most of the reports only focused on a single aspect of the engineered biosensor, leading to a trade-off between different performance parameters; for example, decreased LOD may increase the background expression, leading to high background noise. Therefore, it is crucial to keep all parameters in mind to enhance sensor performance in a real scenario. An ultrasensitive *E. coli*-based sensor was devised by employing the signal amplification methodology (Wang et al. 2015).

A multilayer sensing module was created by Wan et al. (2019), integrating three significant parts: the sensing module, the computing module and the output module. The signal output was enhanced by tuning different parts of the module, such as maintained densities of receptor and ligand in the sensor module, the activator-based amplifier in the computing module and the multiple amplifier cascade in tandem. The authors used a well-known Ar sensor (J101-arsR-ParsR-gfp) that was modified purposefully. The constitutive promoter of the Ar sensor was replaced with two weak promoters to maintain the receptor (arsR) concentration encoded through that promoter. As a result, sensitivity and output amplitude were significantly increased.

Additionally, the output expression was further improved up to 440-fold (without a signal noise) and a limit of detection of up to 100-fold only by incorporating two layers of a transcriptional amplifier cascade (Amp30E-Amp31E11A) in between the sensor and the reporter module. However, to decrease the metabolic load on the host system, this amplifier cascade was connected with a tight input sensor (J117-arsR-ParsR) in a low-copy-number plasmid. The incorporation of the third amplifier (Amp33RinA) in a cascade consequently exhibited an inadequate dose response. Therefore, there was a shift to a high-copy-number plasmid along with its promoter and reporter, and as a result, the three-layered amplifiers were able to enhance the sensitivity and output amplitude irrespective of the type of the used amplifier. The engineered sensor exhibited a high output range for other contaminants, e.g., Hg. The sensor's background leakage was also be minimised by using protein-based post-translational degradation control (Wan et al. 2019).

### ***Recent advances for biosensors specific to heavy metals***

#### ***Arsenic***

Accelerated industrialisation leads to uncontrolled anthropogenic activities, contributing to the accumulation of heavy metals in the environment. For the detection of heavy metals, various biosensors have been developed (Table 1). According to the WHO, arsenic, which can be present in groundwater, is hazardous in its inorganic form as its long-term use can cause tumours and cardiovascular diseases ([www.who.int](http://www.who.int)). It is also a well-known harmful metalloid that occurs in natural environments mainly as a result of geologic processes (e.g., volcanic activity) and from artificial sources (e.g., heavy industry, pesticides). Small-scale analytical protocols require costly instrumentation and highly skilled personnel, thereby limiting their applicability and on-site detection ability; these assay protocols also have low detection limits, which further restrains their application (Komorowicz et al. 2016). Therefore, novel and more sensitive sensors for the detection of heavy metals have been created by designing reporter gene systems along with *Pseudomonas fluorescens* as host cells. The host range of two arsenites (pTPT21 and pTPT31) sensor plasmids that show metal availability through luminescence character was created and assigned into *Escherichia coli* DH5 $\alpha$  and *Pseudomonas fluorescens* OS8. These biosensors are useful for online observations of contaminants, which is not possible via conventional methods (Petänen et al. 2001).

In another report, the generation of WCBMB using two different forms of arsenic (As<sup>3+</sup> and As<sup>5+</sup>) contaminants by joining an arsenic-resistant promoter and the regulatory gene *arsR* to *Escherichia coli* DH5 $\alpha$ , and *phiYFP* and inserted into *E. coli* DH5 $\alpha$  to create an arsenic-resistant biosensor (WCB-11), has been described. The experimental results suggested that the biosensor has an excellent activity to arsenic and the expression of *phiYFP*. When strain WCB-11 was exposed, yellow fluorescence occurred in the presence of As<sup>3+</sup> and As<sup>5+</sup>, which was time- and dose-dependent. The newly synthesised construct is an inexpensive and accessible method for the detection of arsenic in contaminated sites (Qing Hu et al. 2010).

Alternatively, a GFP-merged *ArsR* protein, attached to a cis-element, was used, which loses the binding capability to arsenic at the As-binding conformational states. The GFP-tagged *ArsR* and cis-element complex immobilised on a solid surface in the presence of samples containing toxic metals. The quantification of metal ions is directly proportional to the fluorescence intensity; where the light intensity significantly increases with an increase in the concentrations of toxic metals. The detection limit was 5  $\mu\text{g/L}$  for As(III) in purified water, tap water and bottled mineral water. This biosensor was stable at 4°C and responded to 5  $\mu\text{g/L}$  As(III). Genetically produced biosensors are sensitive, less time-consuming, portable, and could offer benefits such as on-site detection of toxic metals in drinking water (Jia et al. 2019). In a pioneering study, the signal-to-noise ratio was enhanced and the detection limit was significantly decreased by creating plasmids bearing three tandem copies of an *ars* promoter/operator and a *gfp* gene placed in trans in another plasmid bearing a *lac* promoter/operator or a T7 promoter, which is under the control of IPTG. The following construct was incorporated into recombinant *E. coli* cells that were effectively able to sense as at the lowest detection limit of 7.5  $\mu\text{g/mL}$  reported to date (Tani et al. 2009).

The use of whole cells as biosensors is more convenient as they are self-sustained and can detect multiple analytes simultaneously, with a straightforward and quantifiable output. The bacterial operon, which can either populate on the plasmid or on the chromosome, is responsive against arsenic (Fekih et al. 2018) and can further be exploited for generating sophisticated sensors. Various bacteria, such as *Bacillus subtilis*, can survive in tropical environments and can therefore be used at high temperatures for the detection for arsenate in water samples.

The baculosensor was fabricated that composed of *ars* promoter, repressor *arsR* of *B. subtilis* fused with a *xylE* reporter on a pVK168 plasmid with an enhanced detection limit as suggested by the WHO. This biosensor is more sophisticated for arsenic recognition, as the output was reported simply by analysing the yellow-coloured dye within 4 h of exposure with arsenic (Wicke et al. 2018). Most of the bacteria with an operon on a plasmid show resistance against arsenate by establishing a robust efflux system. In one study, *E. coli*-based WCBMB used an *ars* operon fused with mutant GFP (due to the high stability and quantum yield) as a reporter gene and incorporated into non-pathogenic *E. coli* MG1655. This constructed biosensor was able to detect bioavailable arsenic ions as well as arsenate content in water within 2 h of induction, unlike other analytical methods that measure total arsenic content (Elcin et al. 2018).

Most whole-cell-based biosensors rely on the inducible operon that, whenever it senses a particular metal, activates transcription. A genuine arsenic bioreporter was created using the *nikA* promoter of the repressible *nik* operon from *E. coli*, which represses in the presence of sufficient nickel fused with *egfp* as a reporter domain to generate pNik-eGFP plasmid (Yoon et al. 2016).

## ***Cadmium***

Widespread development, economic growth and water scarcity contribute to the increased accumulation of Cadmium (Cd) and, thus, disrupt the ecological balance. The massive accumulation of Cd in the environment is hazardous for organisms, especially in the case of chronic exposure. Inhalation and ingestion of any remarkable amount of Cd lead to respiratory tract infections and may also cause liver and kidney poisoning (Mahmood et al. 2019). Although the accumulation of Cd in the environment can cause toxicity, it is still widely used in industries due to its broad applications. Cadmium is released as a by-product from various sectors, mixes with waterbodies and reaches the food chain (Hayat et al. 2019). In this context, it is crucial to develop a method that can monitor and detect the presence of hazardous Cd in the environment with high sensitivity efficiency in a cost-effective manner.

To create a reporter gene construct with high expression of GFP under the control of the cadR gene that confers resistance for Cd, the recombinant strain *E. coli* cadR30, carrying the cadR gene from *Pseudomonas aeruginosa* BC15 in pET30b expression vector, was used and cloned. Clones were detected by increased expression of green fluorescence protein under UV illumination and separated on SDS-PAGE. Thus, a GFP-based *E. coli* bacterial biosensor for the detection of heavy metals in soil and water was constructed (Edward and Selvam 2011).

The profiling of gene expression associated with Cd treatment identified novel responsive genes in the methylotrophic yeast *Hansenula polymorpha*. A set of genes showed more than a 6-fold increase in their expression on treatment with 300  $\mu$ M Cd for 2 h. Promoters with approximately 1,000 bp upstream were merged with yeast-enhanced GFP gene. Promoter strength and specificity were assessed by inserting the reporter gene construct into *H. polymorpha*; the *H. polymorpha* SE01 gene (HpSE01)-derived promoter most strongly expressed the GFP gene on treatment with Cd. Thus, the HpSE01 promoter can potentially be used as a bio-element in constructing a whole-cell biosensor to mitigate heavy metal contamination, especially for Cd (Jeong et al. 2007).

A GFP-based metal-binding biosensor with cis-element-bound GFP-tagged CadC proteins was produced for the on-site determination of toxic metal ions. This biosensor, along different water samples containing metal pollutants, was incubated and immobilised on a solid surface. Fluorescence intensity was measured for metal quantification and increased with increasing metal levels. In purified water and tap water, as well as bottled mineral water, detection limits of 1 and 10  $\mu$ g/L for Cd(II), respectively, were achieved by the use of a fluorometer after 15 mins. The biosensor was stable at 40° C and responded to Cd (II). The solid phase-based biosensor for Cd is highly sensitive, portable, less time-consuming and can be used for online monitoring of xenobiotics in drinking water (Siddiki et al. 2011).

Most of the previously developed biosensors rely on the presence of metal resistance operons with active promoters and regulators, which are controlled by transcriptional regulators that can bind with the different promoters even in the absence of Cd. Moreover, most of the promising biosensors use the single output mechanism, where the signal output is fluorescence or luminescence. Multiplex microbial biosensors have unprecedented capabilities to monitor the health and metabolic activity of the cell as well as the environmental fluctuations that cause variability in the signal output pattern. Bereza and co-authors (2017) constructed a multiplex Cd biosensor by assembling a single-output Cd biosensor built along with its promoter/operator and constitutively expressing mrp1. This construct was transferred into three different Gram-negative bacterial species (*Pseudomonas aeruginosa*, *Shewanella oneidensis*, *Enterobacter spp.*) isolated from Cd-contaminated soil, resulting in the production of a dual output green and red signal that facilitates the simultaneous measurement of bacterial growth and viability (Bereza et al. 2017).

## ***Chromium***

Chromium is the most common element found in the earth's crust and is widely used in industrial applications such as leather tanning, textile dyeing, ink pigmentation and pharmaceutical drug production, resulting in its heavy discharge into waterbodies (Bharagava and Mishra 2018). Such uncontrolled release puts pressure on aquatic systems and, consequently, affects marine life (Handa and Jindal

2019). Because of the accumulation of chromium in the food chain, human health is at risk. Unlike other contaminants, heavy metal persist longer in the environment and are toxic even at lower concentrations (Li et al. 2019).

*Salmonella typhimurium* TA1535 was genetically modified with plasmid pPLS-1, containing the promoterless lux operon of *Photobacterium leiognathi* as reporter element under the guidance of a DNA damage-dependent SOS promoter from ColD as sensing element, known as SOS-LUX-TEST. This system depends on the level of xenobiotics, stimulating DNA damage in the microbial cell and emitting light that can easily be quantified.

Similarly, the LAC-FLUORO-TEST, used for the identification of physiological responses of cells to different cytotoxic materials, applies the bacterial protein expression vector pGFPuv, integrated with GFP. The presence of xenobiotics results in a dose-dependent decrease in GFP fluorescence. Comparative studies have been performed in the presence of heavy metals through  $K_2Cr_2O_7$  and  $CrCl_3$  simultaneously. Luminescence induction and fluorescence reduction respectively have been measured to show the genotoxic and cytotoxic ability of Cr metal ions. Such tests are easy to conduct, inexpensive and environmentally friendly (Elke et al. 2002).

One of the studies described the construction of a *E. coli* bioreporter system by creating a transcriptional fusion between the GFP reporter gene, the genetic unit of the promoter (chr) and the chromate regulator (chrB) obtained from *Ochrobactrum tritici* 5bv11. The construct was then cloned into the broad host range vector pPROBE –NT, incorporated into *E. coli* DH5 $\alpha$  and designated as pCHRGFP1. Another construct, pCHRGFP2, was designed by inserting the mob gene of pCHRGFP1, which was introduced in the *O. tritici* strain. In this manner, two bioreporter systems were created to assess Cr(VI) and Cr(III) levels (Branco et al. 2013). The bioluminescent whole-cell-based bioreporter was developed using *E. coli* recA:luxCDABE, providing a more precise evaluation of bioavailable Cr. It was also tested in a real-time scenario where the biosensor was applied to Cr-contaminated soil and seawater collected from oil spills (Jiang et al. 2017).

Bacterial whole-cell bioreporters have attracted considerable attention because of their powerful detection properties and their ability to detect bioavailable levels of contaminants in the sample. Nevertheless, most of the respective studies were performed in aqueous samples with the use of *E. coli* instead some other host organisms. Therefore, Jiang et al. (2015) used the native soil microorganism *Acinetobacter baylyi* ADP1 as a bioreporter. This biosensor consists of the promoter less luxCDABE cassette from *Photobacterium luminescens* in its chromosome that is controlled by the SOS-inducible recA promoter. *Acinetobacter baylyi* ADPWH\_recA with fused recA regulon, when exposed to different concentrations of Cr(VI) in soil samples, produced easily quantifiable bioluminescence (Jiang et al. 2015).

## **Copper**

The ubiquitous metal ions such as Cu, Zn and Mg play an essential role as cofactors for various enzymes that take part in metabolic processes. These ions can maintain cell integrity by regulating osmotic pressure. They are also crucial trace elements but can act as a toxin if present in higher concentrations (Maret and Wedd 2014; Rasheed et al. 2018). The microorganisms that inhabit the extreme environment loaded with high levels of heavy metals can sense even small fluctuations of metal concentrations. Many microbial-based biosensors have been developed for Cu identification but failed to detect the bioavailability of an analyte (Chandrasekhar et al. 2012). When studying the copper-responsive promoter in *Achromobacter* sp. AO22, the two regulator genes copR and copA (multicopper oxidase gene), along with a palindromic cop box sequence, were determined. The identified new strain was engineered using a lacZ reporter construct, pCOPRP, in *E. coli*; after modifying the host, expression was enhanced significantly (250–400 fold) for copper, but minimally to other metals. Thus, *Achromobacter* sp. AO22 is highly efficient in the detection of copper bioavailability; for *E. coli*, experiments are still laboratory-based (Ping et al. 2012).

In another approach, engineered *Saccharomyces cerevisiae* strains containing plasmids with the  $Cu^{2+}$  inducible promoter of the CUP1-gene from *S. cerevisiae* were combined with the lacZ-gene from *E. coli*. In this construct, the promoter was first stimulated by  $Cu^{2+}$ , followed by the use of lactose, which acts as a substrate for quantification. The carbon source, lactose, is used if  $Cu^{2+}$  is present and leads to

alterations in oxygen consumption; the sensor measures the copper level, which ranges between 0.5 and 2mM CuSO<sub>4</sub> (Matthias et al. 2000). One hypothesis describes that few microorganisms employ a mechanism to resist metal toxicity, maintaining metal homeostasis. Kang et al. (2018) exploited the gene operon present in *E. coli* and responsible for metal transport to prepare a biosensor against Cu. Upon Cu binding, the cop operon is regulated by the interaction between the promoter and its regulator in the presence of reporters, but the copA promoter was not as active as other promoters; this limitation was overcome by deleting copA and cueR from *E. coli*, further incorporated with a plasmid-harboring pCopAp and GFP reporter.

The dual detection of Cu, as well as gold (Au), was performed with engineered *E. coli* and *Synechococcus* sp. 7002-based biosensors by incorporating the transcriptional activators GoISA113T and GoISCL with a replaced metal-binding loop from CueR (part of a copper operon). This bacterial sensor produces fluorescence signals when it comes in contact with both Au and Cu. The GoISCL-based *E. coli* biosensor responded to even low concentrations of Au and Cu. Finally, the Au and Cu sensors were expressed in the cyanobacterium *Synechococcus* sp. 7002, producing signals in response to Au only (Lacey et al. 2019).

In another study, Chen and co-workers (2017) designed a robust whole-cell biosensor for monitoring the bioavailable levels of Cu ions in tap as well as freshwater samples. The authors used several vital parts such as microbial chassis, promoters and output signal to develop a novel biosensor for Cu. They successfully minimized the response time by examining betaxanthin, a plant pigment for fluorescence signals (Chen et al. 2017).

### **Lead/Zinc**

Lead (Pb) is a naturally occurring heavy metal in the earth's crust. Numerous industries are heavily relying on Pb because of its density, malleability, resistance against corrosion and low degradability. The significant use of Pb and its incorrect disposal results in uncontrolled environmental contamination. According to WHO guidelines, the allowed limit of Pb in drinking water is about 10 µg/L (WHO, Geneva, 2011).

The fusion of a reporter gene to a sensitive metal ion gene, such as zinc and lead, produces biosensors. After many efforts, the limit of detection of a *Pseudomonas putida* KT2440-based Zn/Cd/Pb-biosensor could be increased by a factor of up to 45 by interrupting four primary efflux transporters for Zn/Cd/Pb, causing the metals to concentrate inside the cell. Changes in sensor elements lead to a higher specificity of bioreporter genes. From *P. putida*, a Zn-specific bioreporter was achieved by using the promoter of the cadA1 gene as a sensor element. The engineered *P. putida* reporter strain can detect Zn<sup>2+</sup> concentrations about 50 times lower when compared to other available Zn bioreporters. The achieved limit is much lower than the permitted limit values for Zn and Pd in different soil and water samples, allowing the identification of minute xenobiotic concentrations in the environment (Anu et al. 2010).

The luxCDABE genes of *Vibrio fischeri* and the metal-responsive smt operator/promoter region of *Synechococcus* PCC7942 were fused in plasmid DNA (pJLE23); this plasmid showed metal ion-inducible luminescence to transformed cyanobacteria. *Synechococcus* PCC7942 (pJLE23) was adequate in ZnCl<sub>2</sub> within a range of 0.5–4/~M, as constituted by the stimulus of luminescence (Erbe et al. 1996).

Bereza and co-workers (2016) assessed the ecotoxicological level of Pb in environmental samples with the help of a biosensor which was created by using the host inhabiting the heavy metal-contaminated soil. The plasmid-based whole-cell bioreporter was constructed by fusing pbrR, a regulatory protein, and a reporter gene (GFP) that produces fluorescence in the presence of Pb. This construct was further incorporated into a low-copy-number broad host range plasmid and, subsequently, Gram-negative bacteria, including *Pseudomonas*, *Shewanella* and *Enterobacter* species (Bereza et al. 2016).

## Mercury

The bioaccumulation of heavy metals results in altered food chains, causing ecological imbalance (Wepener et al. 2001). The identification, quantification and monitoring of heavy metals typically involve expensive equipment and frequently require crucial sample pre-treatments. Recombinant bacterial biosensors have been produced, containing zinc regulatory gene *zntR* and *zntA* promoters from the *znt* operon of *E. coli* to induce the expression of GFP reporter protein. The sensor can report up to 3–800 ppm of  $Zn^{2+}$ , 0.005–4 ppm of  $Cd^{2+}$  and 0.001–0.12 ppm of  $Hg^{2+}$  ions. Studies with the liquid medium have shown that the maximum stimulation for  $Zn^{2+}$ ,  $Cd^{2+}$  and  $Hg^{2+}$  was at 20, 0.005 and 0.002 ppm, respectively. The application of the pPROBE-*zntR-zntA* biosensor can therefore be the primary screening technique for aquatic heavy metal pollution (Gireesh and Chaudhari 2012).

The generated sensor is used for peculiar detection and studies on the bioavailability of metals by modified *Pseudomonas fluorescens* with reporter gene systems. One broad host is mercury (pTPT11) plasmids that emphasise metal occurrence by luminescence were established and inserted into *E. coli* DH5 $\alpha$  and *Pseudomonas fluorescens* OS8. The capability of the strain OS8 to measure specific heavy metals in spiked soil and soil extracts is higher than that of conventional chemical analysis. The strain *Pseudomonas* was isolated from pine rhizosphere in oil and heavy metal-contaminated soil. It is a common soil bacteria and, therefore, an excellent agent for quantifying soil heavy metal bioavailability (Petänen et al. 2001). This environmentally relevant bacterial species offer a definite advantage in detecting environmental contaminants as they can mobilise the bioavailable portion of compounds to be used by other organisms coexisting with them in the ecosystem. Therefore, they can play an important role as biosensors. For instance, Rijavec et al. (2017) considered two bacteria, *Pseudomonas putida* and *Allivibrio fischeri*, in biosensor preparation. The Hg-sensing construct consists of the *mer-lux* element incorporated into *A. fischeri* as well as *P. putida*. The constructed biosensor has a lower limit of detection (100 times increased sensitivity) and can detect even the lowest concentration of Hg in the environmental samples (Rijavec et al. 2017). The potential colorimetric biosensor was prepared by Wang et al. by modifying the biosynthesis of soluble pigment pyocyanin under the regulation of MeR, the signal-sensing transcriptional activator of  $Hg^{2+}$  that provides the visual detection signals. The construct exhibited extraordinary sensitivity for  $Hg^{2+}$ , with high selectivity and sensitivity, showing a working potential within a wide pH range (pH 4–10) (Wang et al. 2020).

## Recent advances of biosensors for xenobiotic compounds: hydrocarbons and Pesticides

A xenobiotic is a man-made chemical substance toxic to the ecosystem. To detect the accumulation of these compounds, several strategies are currently being used. One is the use of a luciferase-based whole-cell biosensor, which was constructed and characterised for measuring toluene, benzene, xylene and similar molecules. The reporter TOL plasmid from *Pseudomonas putida* mt2 was constructed by incorporating *xylR*, a transcriptional activator that binds explicitly to toluene-like compounds, and a *luc* gene under the control of promoter Pu. The luminescence response of transformed *E. coli* proved to be a capable biosensor for measuring environmental contaminants. Furthermore, other than the biotechnological workhorse *E. coli*, different soil-borne bacteria, such as *S. aureus*, *P. fluorescens* and *P. putida*, have been used in the engineering of whole-cell biosensors (Renella and Giagnoni 2016), providing the additional advantage of monitoring soil-based contaminants.

The immediate use of industrial and agricultural chemicals results in elevated levels of toxic compounds in the environment (Bilal et al. 2019; Hernandez et al. 2018). These contaminants are potential carcinogens, endocrine disruptors and mutagenic in nature (Barrios et al. 2018; Bilal et al. 2017). The amelioration of these pollutants will have a profound impact on contributing significant efforts towards the success of mitigation and pollution management programs. The traditional strategies to monitor pollutants involve long detection periods and labour-intensive and costly equipment. In this direction, the whole-cell-based biosensor seems to be pivotal for the active sensing of environmental contaminants because of its improved sensitivity, selectivity and lower detection limits (Maduraiveeran and Jin 2017). Until now, various biosensors have been introduced to monitor an array of analytes including xylene, toluene, heavy metals, polychlorinated biphenyls (PCB), polycyclic aromatic hydrocarbons (PAH), alkyl sulphonates, among others (Chang et al. 2017).

## Hydrocarbons

The lipophilic, inert and toxic properties of aromatic hydrocarbons classify them as severe environmental pollutants. Although chlorinated aromatics are under control in many countries, the major problem of their distribution in the environment still prevails. The potential development of recombinant microorganisms capable of bioremediation has increased with the growing knowledge of xenobiotics metabolism. Genetically modified microorganisms with a high degrading capacity for a wide range of chemical contaminants and bioremediation of soil, water and activated sludge can potentially be used as bioreporters.

*Pseudomonas fluorescens* HK44 was the first genetically modified whole-cell bioreporter containing the bioluminescent gene luxCDABE; the phenotype is linked to naphthalene catabolic pathways (Trögl et al. 2012). Soil contamination due to oil spills and petroleum-based products is the leading cause of environmental pollution and a great matter of concern (Cocârță et al. 2017). The naphthalene types of PAH occur in the subsurface environment; they have high solubility and can cause cancer in humans. Sun et al. (2017) prepared a bioreporter which consists of a vector pWH1274\_Nah incorporated with the naphthalene-degrading gene nahAD, cloned into a potent host *Acinetobacter* ADPWH\_lux. The salicylate-inducible gene was introduced into the vector along with the luxCDABE operon; the constructed reporter was highly sensitive towards naphthalene (Sun et al. 2017).

Three reporter gene fusions were constructed through the use of a pcbC promoter of *Pseudomonas* sp. DJ-12 to identify and remediate xenobiotics such as biphenyl and 4-chlorobiphenyl through meta-cleavage dioxygenase. The genes used for production were luc, luxCDABE and GFPuv, inserted into *E. coli* XL1-Blue as a transformant host which emits light upon exposure to aromatic xenobiotics, e.g. 4-chlorobiphenyl, 4-hydroxybiphenyl, 2,3-dihydroxybiphenyl, catechol and 4-chlorocatechol. The cells containing the engineered gene fusions showed a concentration of 0.1 mM for 10 min (Sang et al. 2004).

In another approach the construct was performed by fusing the dnaK promoter of *Pseudomonas* sp. DJ-12 or *E. coli* to the luc marker gene. The fusion product of dnaKp-DJ::luxCDABE in *Pseudomonas* sp. DJ-12 dnaK showed a nearly 5-fold higher response to ethanol compared to dnaKp-EC::luxCDABE in *E. coli*. Along with luc, the response was considerably weaker than that of lux. A biosensor containing dnaKp-DJ::luxCDABE fusion was tested for bioluminescence production upon exposure to other aromatic xenobiotics, such as 4-hydroxybenzoate, 4-chlorobiphenyl and catechol. Specifically, light emitted by dnaKp-DJ::luxCDABE gene fusion showed sensitivity to 1mM biphenyl and 4-chlorobiphenyl when subjected to identification for 80 min; there are similar results for other aromatic substances. This suggests that biosensors produced via dnaKp-DJ::luxCDABE gene fusion have a high capacity to detect many aromatic compounds and different kinds of xenobiotics (Park et al. 2002).

The significant problems with the detection of hydrocarbons through bioreporters are based on their low solubility in water, leading to an enhanced response time and a weaker signal-to-noise ratio. To overcome this issue, the whole cell acts as a host, as it is better adapted to the metabolism of hydrocarbons. Sevilla et al. (2015) compared biosensors based on two different host systems: *E. coli*, which is not optimised for alkane degradation, and *Alcanivorax borkumensis*, which is a marine hydrocarbonoclastic bacterium. These marine bacteria are better adapted to oil and can use hydrocarbons as a sole energy and carbon source (Naether et al. 2013). The construct was made by fusing the transcriptional activator AlkS from *Pseudomonas putida* OCT plasmid under the control of promoter PalkB with GFP as a reporter gene; AlkS activates the transcription in the presence of alkanes. This construct was cloned into the broad-host-range plasmid pSEVA431, resulting in pKSB1. The hydrocarbonoclastic *Alcanivorax borkumensis* is more sensitive towards low-hydrocarbon-chain compounds such as alkanes, with significantly lower noise during fluorescence and four-fold lower threshold levels in comparison to *E. coli*, which was compromised by its higher sensitivity towards salt (Sevilla et al. 2015).

## **Pesticides**

The extensive use of pesticides in modern agricultural practices leads to various pollution issues and, ultimately, ecological imbalance. Organophosphate pesticides or organothiophosphate (OP) compounds share a global insecticide market and are potent neurotoxins; however, they are widely used as pesticides. These compounds disrupt the action of acetylcholinesterase, which is required for the proper functioning of nerve cells. Therefore, it becomes imperative to assess the detrimental effects of this pollutant. In this sense, biosensors

provide the definite advantage of sustainable production and stability as compared to other traditional assays. One of the broad-spectrum widely used organophosphate insecticide is chlorpyrifos (CPF). Its low water solubility, high soil absorption capacity and high mobility result in increased accumulation in the soil during application, causing various health hazards (Lester et al. 2017). By taking advantage of the distinctive properties of microbes, they can easily be manipulated to enhance their use in practical applications. A cheaper alternative to the traditional approach is the monitoring of CPF, which is dependent on the genetic-based whole-cell biosensor. On specific exposure to the pesticide, the target promoter gets activated and fuses with the suitable reporter gene, which in turn produces a higher exposure of the reporter gene and, ultimately, quantifiable signals. Whangsuk and co-workers (2016) prepared a CPF-sensing biosensor based on the CPF-responsive transcription factor (chpR) and the chpA promoter, which gets activated in the presence of CPF and produces fluorescence during positive response due to the presence of lacZ. Another plasmid consists of the *atsBA* gene that encodes the formylglycine-generating enzyme (FGE) and sulfatase; here, sulfatase was used as the reporter system. Both fusion plasmids were incorporated into the *E. coli* host system, and upon exposure to CPF, chpR triggers the expression of the sulfatase gene that converts 4-methylumbelliferyl sulphate to the fluorescent product 4-methylumbelliferyl. Thus, the sulfatase response was more robust and sensitive than that of the  $\beta$ -galactosidase reporter (Whangsuk et al. 2016).

In one study, a DNA fragment comprising the open reading frame of *mpd* (methyl parathion hydrolase encoding gene) and a cognate regulator from *Pseudomonas putida* DLL-1, having methyl-parathion (MP) degrading capacity, was constructed. It was cloned by the gene shotgun method in the pBR1MCS-2 vector to develop a recombinant plasmid pBBR-*mpd*. A genetically engineered microorganism, CDS-pBBR-*mpd*, degrading both carbofuran and MP, was built by transforming the above DNA fragment into the carbofuran-degrading strain *Sphingomonas* sp. CDS-1. It preserved stability and maintained high MP hydrolase activities, making it a promising genetically engineered microorganism for bioremediation (Zhi Liu et al. 2006).

Another biosensor was developed using constitutively expressed redox-sensitive GFP (roGFP2) in *E. coli*, which is induced by chemicals. This work shows that just a few minutes are sufficient to detect oxidation by using *E. coli*-roGFP2, compared to the traditional oxidative microbial biosensor. The minimum quantity (in ppm) detected was  $1.0 \times 10^{-7}$  (arsenite),  $1.0 \times 10^{-4}$  (naphthalene),  $1.0 \times 10^{-4}$  ( $\text{Cu}^{2+}$ ),  $3.8 \times 10^{-4}$  ( $\text{H}_2\text{O}_2$ ),  $1.0 \times 10^{-3}$  ( $\text{Cd}^{2+}$ ),  $1.0 \times 10^{-3}$  ( $\text{Zn}^{2+}$ ),  $1.0 \times 10^{-2}$  (menadione), 1.0 (triphenyltin), 1.56 (zinc pyrithione), 3.1 (selenite) and 6.3 ( $\text{Pb}^{2+}$ ), respectively. The induction of oxidation occurred through heavy metals, while the sigmoid curve was identified for another compound, which is concentration-dependent. A study demonstrates that *in vitro* roGFP2 and *in vivo* GSH concentrations with *E. coli*-roGFP2 justified that roGFP2 is recognised as sensitive to redox potential and thiol modification stimulated by xenobiotics (Carlos et al. 2010).

Biosensors generated with the use of cyanobacteria might be able to detect herbicides and other xenobiotic compounds. The *Synechocystis* sp. strain PCC6803 was chromosomally labelled with the *luc* gene to develop a new strain of a bioluminescent cyanobacterium. Successful expression showed that it produces bioluminescence, and bioassays revealed that a new luminescent biosensor was established, which responded to a variation of xenobiotics and herbicides (Shao et al. 2002). This review indicates that the fusions of a reporter gene into microbial biosensors for the identification of xenobiotics in the environment are valuable for examining the bioavailability of the inducing contaminants. Biosensors are considered to provide new chances for the rapid screening of contaminated sites or the detection of potential environmental damage (Shao et al. 2002; Sang et al. 2004).

### **Potential environmental effects and biosafety concerns**

With advancements in synthetic biology, newer biosensors with advanced features and characteristics can be developed and applied for *in situ* monitoring of environmental contaminants. While considering synthetic biology-based biosensors, it is imperative to assess the possible risks in terms of their interaction with the environment and how these can be mitigated. The whole-cell bacterial biosensors are classified as genetically modified organisms (GMO's), thus raised pertinent questions about potential interaction with commensals, non-target organisms and other biomolecules. Unlike other forms of polluting substances such as chemical spills, it is not easy to contain or control the self-replicating organism released into the environment. The engineered microorganism with even known biology and mechanism would be difficult to predict the unexpected and unwanted environmental effects in advance.

Moreover, engineered organisms would not fit into the natural environment, leading to finding them suitable ecological niche and modifying accordingly by swapping genes with other organisms. Horizontal Gene Transfer (HGT) mechanism is a common and uncontrollable trait among bacterial systems (Thomas et al., 2005) that relies on acquiring foreign genes. The primary issue with HGT is concerned with transferring genetic sequences from the genetically modified bacteria to the indigenous microorganism resulting in altering their genome and, ultimately, their ecological niche (Heuer and Smalla, 2007). Even if the engineered microorganism does not survive in the natural environment, the free synthesised DNA in the modified microorganism can pose an environmental risk as it can be actively assimilated by many other bacteria (Thomas and Nielsen, 2005), unicellular and multicellular eukaryotes (Boschetti et al., 2012). In this context, it is crucial to prevent the transfer of modified (recombinant and synthetic) DNA to the environment and other organisms via HGT (Wright et al., 2013). This prevents the overburdening of polluting the environment with engineered microorganisms. The antibiotic resistance genes have been utilised in the plasmid for selecting modified organisms, but this can pose a significant challenge in generating antibiotic-resistant superbugs. Keeping these points in view, it is critical to test the WCBMB extensively on a laboratory or pilot scale before releasing it in the environment.

### ***The future of whole-cell biosensors***

In recent years, increasing pollution has diverted the attention of researchers to mitigate the problem through innovative state-of-the-art techniques. The bacterial community has the smartest system to quickly adapt to new environmental conditions by directly evolving coping mechanisms. Easy adaptation and high robustness make them an ideal platform for scalable biosensing devices. The ongoing industrialisation and the resulting emergence of new pollutants lead the way to a better understanding of reporter genes and to an improved detection and degradation of these pollutants. Furthermore, miniaturized bioreporter assays have potential for the monitoring of toxicants in environmental samples. This, however, calls for the creation of a metagenomics library consisting of large insert fragments, generating a wealth of possibilities for screening novel genes to construct genetically modified microorganisms (Xing et al. 2020) that will additionally provide better specificity to the concerned biosensor. Further, the comprehensive study of bacterial whole genome sequences (Ali et al., 2018; Behare et al., 2020) and pan proteome investigations (Pragya et al., 2017; Nataraj et al., 2020) allows the determination of crucial genes for the specific niches (Kaur et al., 2017; Ali et al., 2020). Therefore, we suggest two crucial concerns that demand attention, namely response period and detection limit.

Although the whole-cell-based microbial biosensor can outperform the conventional chemical-based sensor, continuous efforts are still being made to increase the applicability of these sensors. Another concern is the increase in pollution levels in the environment, increasing the possibility of co-contaminated sites with multiple pollutants. It could be envisaged that fabricating multifunctional biosensors by incorporating a combination of numerous regulatory proteins provides the explicit advantage of the simultaneous sensing of various contaminants (Hou et al. 2015). Another prime focus for the use of whole-cell microbial biosensors relies on the construction of multifunctional biosensors that will be able to sustain harsh environmental conditions with high temperatures, acidity, alkalinity and salinity. Thus, different microbial species, such as extremophiles, thermophiles, alkalophiles or halophiles, could be useful as host cells for constructing whole-cell microbial biosensors. The presence of different toxic compounds in the samples either inhibit or decrease the activity of microbial cells and can be mitigated by self-powered biosensors.

Although microbial biosensors have tremendous potential, only a few applications are being implemented within ecosystems, primarily because of legislative restrictions that are generating ethical issues as the components of the whole cell-based biosensor may interact, and thus interfere, with the host metabolism. With the recent advancement in sequencing methods, robust characterisations can now be performed to better elucidate and understand the fate, activities and interactions of this microbial biosensor strain under real-world scenarios to stimulate and enable future full-scale field applications. Mainly the transcriptomic approaches can be used to identify the promoter/operator pair better to be rearranged into the bacterial system for the desired functions.

The complex biological systems can pose a significant hurdle in engineering biosystems with desired features, as exemplified by the presence of interconnected transcriptional regulatory networks that often produce unpredictable results. Transcriptional biosensing dictates the microbial cells to act following the surrounding environmental signals. The introduction of synthetic biology provides novel avenues for creating a manipulated sensing system based on the re-programming genomes and genetic circuits, in addition to tuning

environment-responsive transcriptional regulators in a bacterial order. Continuous efforts are being made to meet the advanced sensitivity, threshold, dynamic and optical range of the microbial modulating system, thus introducing biocomputational tools, including Boolean logic gates, and programmable cellular computation with scalable signal-processing capacity, benefiting the revolutionary biosensing technique.

However, the new scenario is changing as scientists begin to incorporate decision tree models to identify numerous xenobiotics and pattern-learning algorithms via specific light emission profiles. Moreover, machine learning tools become an intricate tool to design biosystems with optimised performance and desired properties because of their power of better predictability (Volk et al. 2020). Despite all the advantages of microbial biosensors, it is crucial to investigate the portability and miniaturisation of signal detection systems to increase the use in the field. It is a diverse, fast-growing field that can be coupled to other high-throughput technologies such as microfluidics, which provides an opportunity to detect multiple analytes at the same time with high sensitivity and long-term preservation (Roggo and Meer 2017). Another hardware platform, such as smartphone-based whole-cell microbial biosensors harbouring engineered bacteria, allows induced response to quantifiable signals.

Moreover, the integration of microfluidic devices for sample attainment and smartphones for imaging and data calculation provides a laboratory-independent sensing micro-machined platform (Yang et al. 2016). Portable, lightweight device systems, such as a paper strips or a card, are promising tools and could be implied in on-site environmental monitoring. Such technologies will evolve rapidly into more programmed, compressed and wireless devices for the real-time detection of a wide range of environmentally relevant contaminants, indicating a keener perception of genetically engineered whole-cell microbial-derived biosensors for broader performance. This review summarises the vast range of analyses using various genetic elements of an operon, such as promoter/operator, reporter genes and metal sensitiveness, and their on-site degradation for the establishment of biosensors. This simultaneous detection and deterioration of environmental pollutants resulted in an innovative and multidisciplinary approach.

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## Declarations

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### Consent for publication

The authors gave their consent for publication of the research results.

## Tables

**Table 1.** Comprehensive list of biosensors used in recent studies for the detection of heavy metals and xenobiotics.

Target analyte	Host strain	Expression vector/ sensor	Detection limit	Response time	Sample type	References
<b>Heavy metals</b>						
Arsenic	<i>E. coli</i> BL21	pET-3a	5 µg/ml	15 min	Water	Shohel et al., 2011
	<i>E. coli</i>	pIRC140	1 ppm	10-12 h	Spiked sample	Roberto et al., 2002
	<i>E. coli</i> (AW3110)	pTO031	1-10 µM	90 min	Spiked water samples	Tauriainen et al., 1999
	<i>E. coli</i> DH5α	ars-puc18-yfp	8 µmol/l	2 h	Spiked sample	Qing Hu et al., 2010
	<i>Pseudomonas fluorescens</i>	pTPT21 and pTPT31	10 µM	2 h	Spiked sample	Petänen et al., 2001
	<i>E. coli</i> JM109	pSB1A2	<5 ppb	5 h	Water	Aleksic et al., 2007
	<i>E. coli</i> K-12	POLA (VtP)-gfp	5-140 µg/l	1 h	Water	Pola et al., 2018
	<i>E. coli</i>	pHg-lux, pAs-lux, pHg-gfp, pVLAS1	0.025-0.1 µM	2 h	Sewage water	Huang et al., 2015
	<i>Thermus thermophilus</i> HB27	TtArsXpbgaA-nqo TtSmtB	5 µM, 10 µM	16 h	Spiked sample	Antonucci et al., 2018
	<i>E. coli</i> MG1655	pBR-arsR773	10 µg/l	5 h	Spiked sample	Elcin and Öktem, 2020
Cadmium	<i>E. coli</i> BL21	pI258	10 µg/l	30 min	Water	Shohel et al., 2011
	<i>S. cerevisiae</i> Y2805	YEpSSEO-gfp	300 µM	2 h	Spiked sample	Jeong et al., 2007
	<i>Pseudomonas, Shewanella</i> and <i>Enterobacter</i>	pBBcadRgfp-rfp	0.01-10 µg/ml	20-40 min	Spiked water	Bereza et al., 2017
	<i>E. coli</i> BL21	pegfp-N1	5 µM	1-2 h	Spiked sample	Lee et al., 2019
	<i>Thermus thermophilus</i> HB27	TtArsXpbgaA-nqo TtSmtB	5 µM, 10 µM	16 h	Spiked sample	Antonucci et al., 2018
Chromium	<i>Salmonella typhimurium</i> TA1535	pPLS-1	0.13 mmol/l	5 h	Spiked sample	Elke et al., 2002
	<i>E. coli, Ochrobactrum tritici</i>	pCHR-gfp1, pCHR-gfp2	0.5-2 µM, 2-10 µM	<14 days	Artificial and natural soil	Coelho et al., 2015

	<i>Acinetobacter baylyi</i>	ADP1-RecA-lux	520 mg/kg	6 h	Contaminated soil	Song et al., 2014	
Copper	<i>E. coli</i> DH5α	YEp352-FUSI; pYe(lacZ)1-CUP1	0.5-2 mM	15 min	Spiked sample	Lehmann et al., 2000	
	<i>E. coli</i> CB454 and <i>Achromobacter</i> sp. AO22	pCOPRP	0.8 and 2 nM	2 h ( <i>E. coli</i> ) and 4.5 h ( <i>Achromobacter</i> )	Spiked sample	Ng et al., 2012	
	<i>E. coli</i> J53	pK19mobsacGFP	20-600 μM	4 h	Spiked sample	Martinez et al., 2019	
	<i>Synechococcus</i> sp. PCC 7002	GolS-ypet	10 μM	1 h	Spiked sample	Lacey et al., 2019	
Lead	<i>Pseudomonas. Putida</i> KT2440	pDN-PcadA1lux and pDNPzc1lux	0.90 μM		Spiked sample	Anu et al., 2010	
	<i>E. coli</i> DH5α <i>P. aeruginosa</i> PAO1 <i>S. oneidensis</i> MR-1 <i>Enterobacter</i> sp. NCR3 <i>Enterobacter</i> sp. LCR17	pBBbrR-gfp	0.2–1 μg/mL	30-100 min	Spiked tap water, groundwater	Bereza-Malcom et al., 2016	
	<i>E. coli</i> DH5α	kan ΔpbrA-PpbrRT-pbrR-luxRΔ2–162	5 μM	8 h	Contaminated and Spiked sample	Jia et al., 2018	
	<i>E. coli</i>	pPpbr-RFP-lacZ.	3 μM	4 h	Spiked sample	Hui et al., 2020	
	Mercury	<i>Pseudomonas fluorescens</i>	pTPT11	0.003 μg kg <sup>-1</sup>	24 h	Spiked soil	Petänen et al., 2002
		<i>E. coli</i> DH5α	pPROBE-KT	0.002 ppm	16 h	Spiked sample	Gireesh and Chaudhari, 2012
<i>Pseudomonas putida</i>		merR-egfp	<40 μg/kg	6 days	Soil	Wei et al., 2014	
<i>Sphingobium</i> SA2		merA-gfp	0-40 nM	5 h	Water	Mahbub et al., 2017	
<i>E. coli</i>		pHg-gfp, pVLAS1, pHg-lux	0.025-0.1 μM	2 h	Sewage water	Huang et al., 2015	
<i>E. coli</i>		pRLucMer19, pRGfpMer19	1 ppb	1 h	Spiked sample	Gupta et al., 2019	
Zinc	<i>Synechococcus</i> PCC7942	pJLE23	2 μM	4 h	Spiked sample	Erbe et al., 1996	
	<i>E. coli</i>	zntA-PzntR-gfp	0-25 mg/l	3 h	Liquid media, soil	Hurdebise et al., 2015	

Gold	<i>Synechococcus</i> sp. PCC 7002	GoIS-ypet	100 nM	1 h	Spiked sample	Lacey et al., 2019
	<i>E. coli</i> DH5a	pgolB-rfp	0.1 µM	10 h	Spiked sample	Yan et al., 2018
<b>Other xenobiotics</b>						
p-nitrophenol (PNP)	<i>Arthrobacter</i> . JS 443.	pUTGFP2	5 nM	5 min		Lei et al., 2004
	<i>Moraxella</i> sp.	pPNC033	20 nM	5 min	Spiked sample	Mulchandani et al., 2005
Organophosphate insecticides	<i>Pseudomonas putida</i>	CDS-pBBR-mpd	0.2 g/l	24 h	Spiked sample	Liu et al., 2006
	<i>E. coli</i>	pET22b	4 units/mL	8 h	Spiked sample	Kang et al., 2006
	<i>E. coli</i> BL21 (DE3)	XL1-Blue/pETLG/pINCM	5 ppm	8 h	Spiked sample	Yang et al., 2012
Naphthalene	<i>E. coli</i> DH5a	pRSET-roGFP2	1.0 × 10 <sup>-4</sup> ppm	2-6 min	Spiked sample	Carlos et al., 2010
	<i>Acinetobacter</i>	pWH1274-lux	0.01 µM	<6 h	Groundwater, soil	Sun et al., 2017
Hydrogen peroxide	<i>Saccharomyces cerevisiae</i>	roGFP1-R12	10 mM	—	Spiked sample	Yu et al., 2009
Polychlorinated biphenyls (PCB)	<i>Pseudomonas fluorescens</i> F113L::1180-GFP	pJBA26	0.09-1.48 nmol/mg of protein/min	60 min	Spiked sample	Villacieros et al., 2005
Aromatic pollutant	<i>Pseudomonas</i> sp.	dnaKp-DJ::luxCDABE	1 mM	80 min	Spiked sample	Park et al., 2002
	<i>Pseudomonas</i> sp. DJ-12	pcbCp::luxCDABE	0.1 mM	10 min	Spiked sample	Park et al., 2004
Herbicides	<i>Synechocystis</i> sp. strain PCC6803	pTLUC	7.47 mg/l (urea-based)	30 min	Spiked sample	Shao et al., 2002
2,4-dinitrotoluene (DNT)	<i>E. coli</i>	yqjF-lux	100 µM	60 min	Soil	Kroll et al., 2014

	<i>E. coli</i> DH5α	yqjf-gfp	4 mg/L	∅4 h	Soil	Kabessa et al., 2016
2,4,6-trinitrotoluene(TNT)	<i>E. coli</i>	ybiJ-lux	148 μM	60 min	Soil	Kroll et al., 2014
Phenanthrene	<i>P. putida</i>	phnS-lux	<10 mg/kg	6 days	Soil	Wei et al., 2014
Atrazine, cyanuric acid	<i>E. coli</i> SM004, <i>E. coli</i> SM003	pBBluxPatzD	7.83, 1.08 μM	<4 h	Spiked sample	Hua et al., 2015
2-nitrobenzoate	<i>Cupriavidus sp.</i> ST-14	onb-egfp	0.5 nM-10 mM	1 h	Spiked sample	Deb et al., 2018
n-alkanes	<i>E. coli</i>	PakB-gfp	5 μM	1-3 h	Spiked sample	Sevilla et al., 2015
Toxic substances	<i>Vibrio aquamarines</i> VPKM B-11245, <i>E. coli</i> MG 1655	pRecA-lux, pKatG-lux, pSoxS-lux, plbpA-lux, pFabA-lux			Contaminated wastewater	Sazykin et al., 2016
Mitomycin C, benzo[a]pyrene	<i>Acinetobacter baylyi</i>	ADP1_recA_lux	0.4 mg/kg, 0.5 mg/kg	6 h	Contaminated soil	Song et al., 2014

## Figures

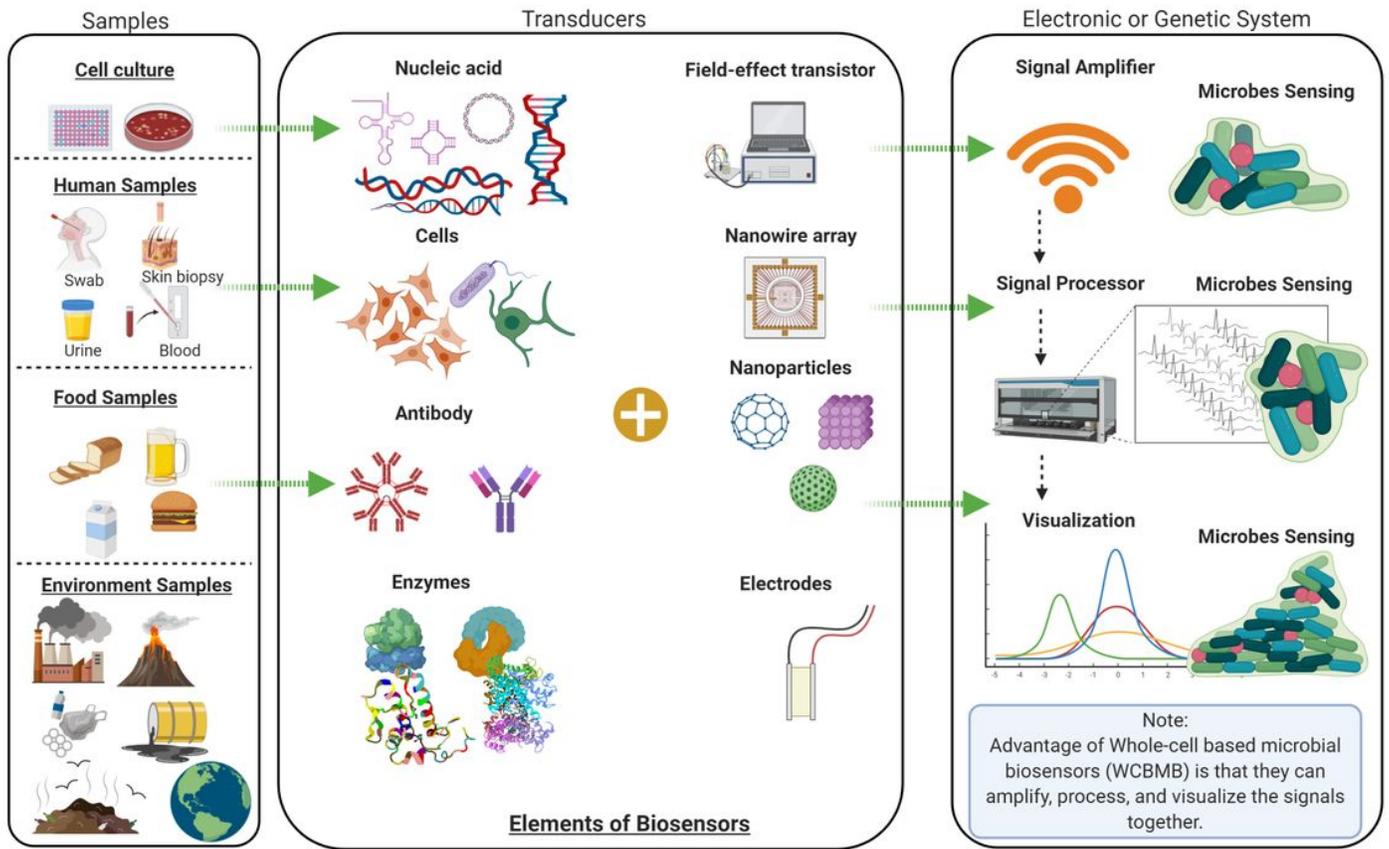


Figure 1

Comprehensive overview of elements of biosensors.