

Quantification of economic and environmental benefits of faecal sludge derived compost and briquettes using a REVAMP tool: The case of Dar es Salaam, Tanzania

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Research

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1 **Quantification of economic and environmental benefits of**
2 **faecal sludge derived compost and briquettes using a**
3 **REVAMP tool: *The case of Dar es Salaam, Tanzania***

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20 **Abstract**

21 In recent years, the concept of resources recovery from waste particularly faecal sludge
22 (FS) has gained much attention and popularity. The aim of this study is to
23 quantify/estimate the economic and environmental benefits of nutrients and energy that
24 could be recovered from FS. The empirical data came from three unplanned settlements
25 of Dar es Salaam City; Keko, Kipawa and Manzese. Two different FS resource recovery
26 scenarios were assessed, namely nutrients and energy. The resource value mapping
27 (REVAMP) tool was used to quantify benefits of nutrients and energy/solid fuels
28 recovered from FS. The results indicate a daily economic benefit from FS recovery
29 ranging between 680-950 USD for energy and up to 7,000 USD for nutrients recovered,
30 depending on the composition. The co-composting of FS with organic waste to recover
31 nutrients was found more profitable than FS derived briquettes. The environmental
32 benefits, beyond the disposal of untreated FS into the local environment, include a
33 potential saving of up to 5 hectares of forest area when substituting the use of wood
34 charcoal with FS-derived briquettes for domestic cooking energy needs. Even here, co-
35 composting FS is estimated to be more profitable than FS derived briquettes. The study
36 concludes that to obtain FS derived briquettes with high calorific value, FS should be co-
37 fuelled with other feedstock materials. Guidelines and standards for safe re-use of FS
38 derived products should be developed, adapted and information campaigns and
39 demonstration sites to illustrate the economic and environmental benefits of resource
40 recovery from FS for energy and agricultural purposes should be made available to
41 relevant stakeholders.

42 **Keywords:** *Faecal sludge; Resource recovery; Resource value mapping; REVAMP;*
43 *Environmental; Economic, briquettes; compost; energy; nutrients*

44 **1. Introduction**

45 In the world of scarce resources, several alternatives have been explored to meet the
46 growing resource demand. Resources recovery and reuse from waste is one of the
47 alternatives under considerations. Resource recovery and reuse (RRR) is a new dimension
48 in sanitation management that focuses on optimizing resource use and connect to water,
49 energy and food systems [1]. Contrary to linear waste management solutions focus on
50 treatment for disposal, the RRR concept is circular and attempts to ‘close the sanitation
51 loop’ and turn human waste into a valuable resource [2]. RRR is particularly applicable
52 in areas dominated by Onsite Sanitation Systems (OSS). In Dar es Salaam (DSM),
53 Tanzania, about 90% of its inhabitants of 5 million, depend on onsite sanitation systems
54 (OSSs) like pit latrines and septic tanks for their sanitation services and needs. A study
55 from 2015 estimated that only about 43% of the collected faecal sludge (FS) from the
56 city’s OSSs are safely managed by two designated Waste stabilization ponds (WSPs)
57 located at Vingunguti and Buguruni areas [3]. A piled-up semi-solid faecal sludge,
58 partially treated in WSPs needs proper and adequate management to ensure safe disposal
59 without posing any environment and human health risks [4]. Various options for
60 sustainable Faecal Sludge Management (FSM) have been developed in the last 10 years
61 that makes it possible to safely and sustainably recover FS resources to be used as plant
62 nutrients, water for irrigation, energy, and protein for animal feed [5].

63 The positive impacts of proper FSM through resources recovery have been widely
64 documented. Reducing and averting nutrient loads causing eutrophication due to disposal
65 of FS into water bodies is a major environmental benefit from increased FS-RRR [6].
66 RRR also when mainstreamed in FSM strategies, has the potential to contribute to
67 achieving several sustainable development goals (SDGs) such as improving soil fertility

68 and animal feed options to ensure food security (SDG2), provide renewable energy
69 sources to contribute to energy security (SDG8), reduce risks of waterborne diseases to
70 improve human health and safe management of sanitation (SDG3 and SDG6) as well as
71 reduce environmental degradation (SDG15) [7]. Moreover, RRR could also have
72 economic value by creating job opportunities and generated incomes from sales of end
73 products [8]. When practised in the back-end of the sanitation value chain, FS-RRR has
74 been proved to increase the sustainability of FSM by off-setting a portion of the treatment
75 and disposal costs [4]. Even though several products can be obtained from RRR, it is
76 known that potential market values of the same end-products vary significantly in
77 different countries where FS-RRR has been practiced [8]. For example, in Dakar, Senegal
78 protein recovery appears to be more financially viable than other end-products; while in
79 Kampala, Uganda, FS derived solid fuel was found to be more profitable than other
80 products and in Accra, Ghana biogas currently has more biggest market value [5]. To
81 ensure the most viable and sustainable end-product it is therefore important to assess not
82 only the FS amount and nutrient content within the sanitation value chain but also the
83 economic and environmental values of the anticipated end-product [9].

84 While the practice of FS-RRR in DSM is still at infancy stage, lacking information on
85 economic and environmental benefits of produced FS-based products, therefore research
86 on the potential and possibilities of FS-RRR has begun. In a study by Mkude et al., [10]
87 the FS characteristics and available volume was estimated to determine the resource
88 recovery potential from OSSs in three unplanned settlements of DSM. Drawing on the
89 same data, the objective of this article is to assess the potential economic and
90 environmental benefits of future FS-derived products using the Resource Value Mapping
91 (REVAMP) tool developed by the Sustainable Sanitation Initiative at the Stockholm
92 Environment Institute (SEI, 2016).

93 **2. Methodology**

94 **2.1 Study design**

95 The study employed a cross-sectional research design combining both qualitative and
96 quantitative methods during data collection and analysis. The quantitative method
97 included a household survey and observation of available resource recovery practices and
98 products in study areas. Qualitative methods involved key informant interviews and focus
99 group discussions (FGD) with key stakeholders involved in faecal sludge management in
100 Dar es Salaam. Furthermore, as a continuation study, results from preliminary analysis on
101 faecal sludge amount and nutrients content from study areas documented by Mkude and
102 others (2019) [10] were used as main inputs for quantification of recovery products.

103 **2.1.1 Description of the study site**

104 The study was conducted in three unplanned settlements in Dar es Salaam city, namely
105 Keko, Kipawa and Manzese, located in Temeke, Ilala and Ubungo Municipality
106 respectively. The areas are densely populated, estimated to have 224,140 people in total
107 occupied in 19 km² area with average household occupancy of 6 people [11]. The areas
108 were considered due to almost all population (99.2%) depend on OSSs including pit
109 latrines and septic tanks for their sanitation services and needs [10]. To fulfil the
110 populations' water and sanitation needs, two main water sources serve domestic needs
111 through communal water standpoints for surface water and private and public boreholes
112 for groundwater sources. The reported average per capita water use is still low,
113 approximately to 22.4 ± 9.1 l/cap/d for bathing and 46.5 ± 27.9 l/cap/d for other purposes.

114 **2.2 Study approach**

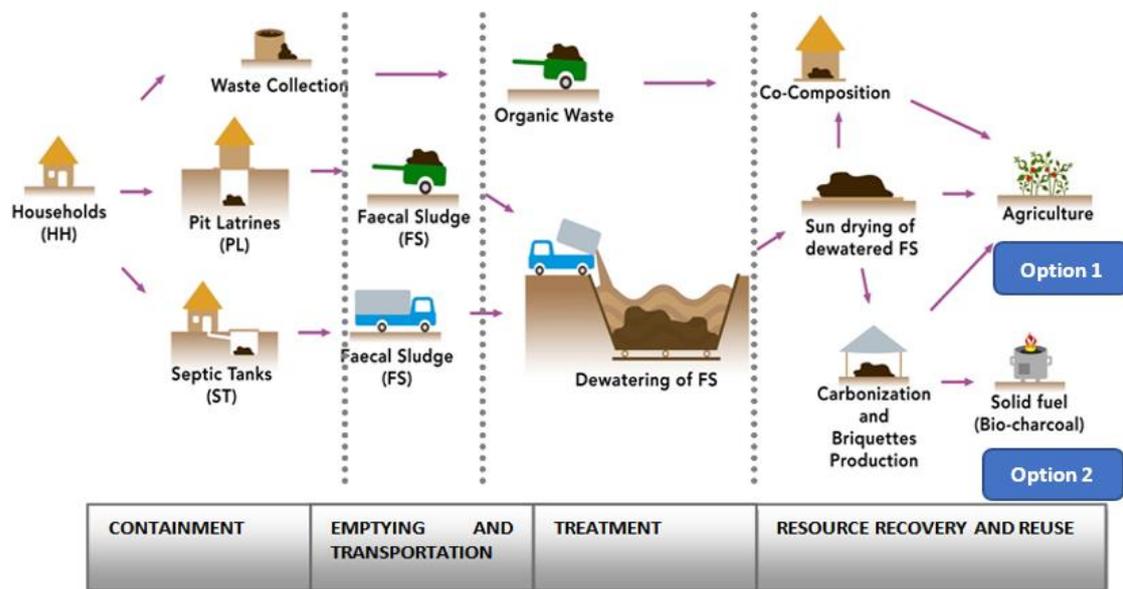
115 **2.2.1 Resource Recovery Scenarios Analysis**

116 **i. Zero or Baseline Scenario**

117 The status quo of FSM in DSM does not consider resource recovery. The same situation
 118 was used in MFA to present amount of FS and nutrients contained from collection to
 119 disposal.

120 **ii. A scenario for a Change**

121 This article considers two scenarios to represent hypothetical future resources recovery in
 122 DSM. The two scenarios include first, energy recovery (*EnRec*) via production of FS-
 123 derived briquettes to be used as domestic cooking fuel and second, nutrient recovery
 124 (*NutRec*) via compost production to be used as fertilizer in agriculture (see Figure 1).



125

126 **Figure 1.** Schematic presentation of proposed resource recovery scenarios; co-
 127 composting and solid fuel production

128 The *EnRec* scenario considers the possible use of solid fuels as an alternative cooking fuel
 129 to the dominant wood charcoal currently used by the majority of residents in the study
 130 areas. The *NutRec* scenario considers the possible use of composted FS as an alternative

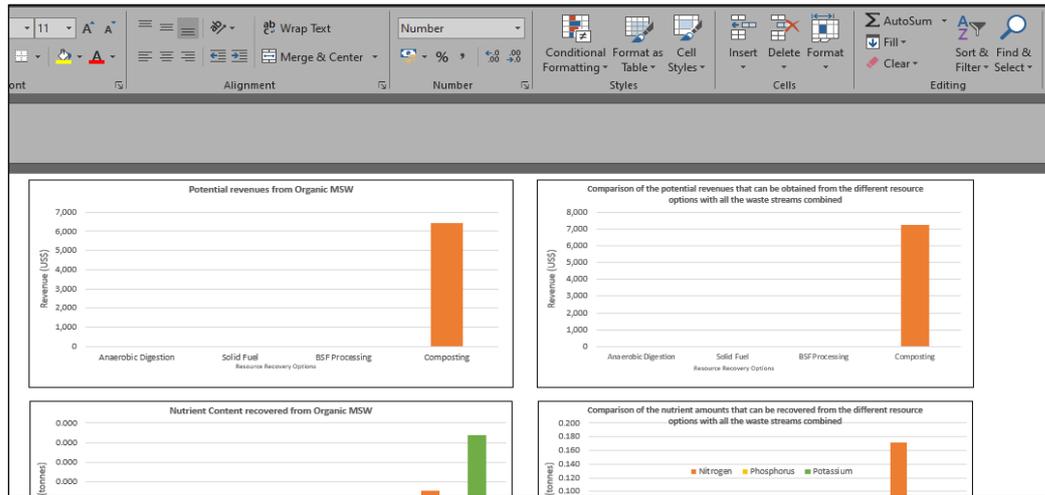
131 or supplement to UREA fertilizers and other organic and animal waste currently applied
 132 by local farmers to improve soil fertility [12]. To obtain a higher nutrient content level,
 133 FS was co-composted as recommended by Tayler, [13]. Accordingly, in the NutRec
 134 scenario, FS was co-composted with organic waste, sorted from collected municipal solid
 135 waste, due to its availability throughout the year. The end products were analyzed and
 136 presented in the REVAMP tool and compared based on mass volume and revenue.

137 2.2.2 Resource Value Mapping (REVAMP) tool

138 The resource value mapping (REVAMP) is a tool developed as a Microsoft Excel
 139 spreadsheet with built-in formulae that consists of six (6) working sheets. The tool was
 140 developed by Stockholm Environmental Institute (SEI) for rapid estimation, visualization
 141 and valuation of the potential resources to be recovered from organic waste streams in
 142 urban areas. The tool includes both sewage sludge, faecal sludge, as well as food and other
 143 organic solid waste [14]. The current available recovery technologies for evaluation in
 144 REVAMP tool are biogas production, solid fuel for combustion, insect larvae for livestock
 145 feeds, and compost or soil conditioner materials for agriculture [15]. Figures 2 (a) and (b)
 146 displays examples of REVAMP tool computer interfaces.

Parameter	Units	Faecal Sludge	Reference(s)
Volatile Solids Degradation rate, VS_D	%	70.00	Estimate based on Von Sperling & de Lemos Chernicharo
Dry Mass Reduction rate for anaerobic digestion (AD) residue, DMR_{AD}	% of initial TS		Alfa et al. (2014)
Biomass conversion rate for black soldier fly (BSF) larvae, BCR	%		Banks (2014)
Dry Mass Reduction rate for BSF residue, DMR_{BSF}	% of initial TS		Banks (2014)
Total nitrogen (TN) reduction in BSF residue, TNR_{BSF}	% of initial TN		Assumed based on Sheppard, Newton and Burtle (2008)
Total phosphorus (TP) reduction in BSF residue, TPR_{BSF}	% of initial TP		Assumed based on Sheppard, Newton and Burtle (2008)
Total potassium (TK) reduction in BSF residue, TKR_{BSF}	% of initial TK		Assumed based on Sheppard, Newton and Burtle (2008)
Dry Mass Reduction in compost, DMR_C	% of initial mass	19.40	Average based on Breitenbeck & Schellinger (2004)
Dry nitrogen (TN) reduction during composting, TNR_C	% of initial TN	34.30	Average based on Galvin (2013)
Total phosphorus (TP) reduction during composting, TPR_C	% of initial TP	1.77	Average based on Eghball et al. (1997)

147 **Figure 2(a).** The REVAMP tool interface estimating waste streams treatment processes
 148



149 **Figure 2(b).** The REVAMP tool interface displaying final results
 150

151 **2.2.3 Steps followed in the application of REVAMP analysis**

152 In the first step, all the required data according to the waste streams used for analysis were
 153 filled in the data input sheet separately for each study area. Composting and solid fuel
 154 production were selected among treatment processes based on two resource recovery
 155 scenarios. Prices of products were identified from the literature review and interviews
 156 with Key Informants. Table 1 presents benchmark information including the fraction of
 157 faecal sludge amount emptied and delivered to treatment plants and estimated nutrients
 158 flows calculated by MFA [10].

159 **Table 1.** The general information used to estimate input data in developed scenarios

	Study location			Reference
	Keko	Kipawa	Manzese	
Population	79,453	74,180	70,507	[11]
Fraction of pit latrines out of OSSs, %	27.1	35.9	61.8	[9]

Fraction of septic tanks out of OSSs, %	72.1	63.4	38.2	[9]
Volume of FS delivered to treatment plants (l/ca/yr)	408.80	434.35	646.05	[9]
Nitrogen (tons/yr)	293.91	318.70	359.48	[9]
Phosphorus (tons/yr)	29.95	23.03	28.59	[9]
	9,048.	9,891.3		
N (mg-N/l)	80	0	7,891.81	This study
P (mg-P/l)	922.1	714.77	627.65	This study

160 **2.3 Data collection method**

161 Data collection took place between September 2017 and November 2018. A total of 395
 162 respondents participated in the household survey, 19 key informants were interviewed and
 163 six focus group discussions and observations were conducted in all three study areas. A
 164 literature review and field observations were carried out to obtain an overview of OSS
 165 facilities and population served as well as the faecal sludge contents potential for recovery.
 166 Focus group discussions, key informant interviews and household surveys identified
 167 availability of and prices for different energy and nutrient products.

168 The amount of FS generated and transported along the sanitation service chain was then
 169 quantified in each locality, drawn from the Material flow analysis (MFA) technique and
 170 using the Software for Substance Flow (STAN). The MFA was used to trace and quantify
 171 the annual FS generated, stored and deposited together with nutrients (nitrogen and
 172 phosphorus) contained in FS from three study areas [10].

173 **2.4 Data analysis using the REVAMP tool**

174 The input data were obtained from different data collection methods representing a status
175 quo (no changes) scenario. The tool incorporates different built-in mathematical
176 expressions in the quantification of the desirable end-products according to the waste
177 streams used. Apart from the general information that cut across each study area, data
178 analysis for each recovery option was conducted based on eight (8) equations by Ddiba,
179 2016 [15] as expressed accordingly.

180 **2.4.1 Potential amount of FS-derived briquettes**

181 Energy recovery was sought to be achieved through the production of dried FS briquettes
182 to be used for combustion. It is recommended that for the faecal sludge to be used as an
183 energy source, needs 90% dryness [16]. Estimation of solid fuel amount by REVAMP
184 was first given in mass (tonnes) then evaluated for its calorific value (energy content). The
185 mass of briquettes (B_m) was obtained from the known volume of collected faecal sludge
186 (FS_v) as shown in table 1 and volume of total solids (TS_v) in mg/l by using equation 1.

$$187 \quad B_m(\text{tonnes}) = FS_v \times 1000 \times \frac{TS_v}{10^9} \times \frac{100}{90} \quad (\text{Eq. 1})$$

188 Then, the fuel energy content, E_F (MJ) was calculated based on calorific value (CV) of
189 dry matter in MJ/kg TS, as recommended by Diener et al. (2014) using the Eq. 2

$$190 \quad E_F(\text{MJ}) = FS_v \times 1000 \times \frac{TS_v}{10^9} \times \frac{100}{90} \times CV \quad (\text{Eq. 2})$$

191 The revenue that can potentially be generated from the briquettes used was calculated
192 based on a mass of briquettes in tonnes (B_m) and the price in US\$/tonne (F_p) as;

$$193 \quad \text{Potential revenue from briquettes} = B_m(\text{tonnes}) \times F_p(\text{US\$/tonne}) \quad (\text{Eq.3})$$

194 *Input data for EnRec scenario*

195 Since the production and use of solid fuel from FS is still not common in Tanzania,
 196 particularly in Dar es Salaam, most of the input data related to faecal sludge-based
 197 briquettes was obtained from the literature. A Total Solids (TS) of 30,000 mg/l for dried
 198 FS was adopted [17]. The price of FS briquettes was lacking. Instead, the available and
 199 comparable solid energy source which is either wood charcoal or other biomass briquettes
 200 could be used. Since the wood charcoal is the dominant source used by more than 90% of
 201 the population, its cost was taken as alternative. Table 2 summarizes input data used for
 202 *EnRec* scenario.

203 **Table 2.** Input data to REVAMP tool specific for *EnRec* scenario

S/N	Item	Value adopted and used	Reference
1.	Faecal sludge total solids (TS)	30,000 mg/l	Schoebitz et al., [17]
2.	Faecal sludge calorific value	16.2 MJ/kg	Muspratt et al., [5]
3.	Cost of charcoal	0.60 US\$/kg	Msuya et al., [18]
4.	Cost of biomass briquettes	0.26 US\$/kg	Lohri et al [19]

204 **2.4.2 Potential amount of FS-derived compost**

205 The designed *NutRec* scenario was achieved through the co-composting of faecal sludge
 206 and organic waste fraction from municipal solid waste (MSW). Data of two waste streams;
 207 faecal sludge and organic waste were needed in the REVAMP analysis. Faecal sludge
 208 dryness of 60% is recommended to achieve composting and through the process, reduction

209 of mass of combined waste streams is expected [15]. The value of dry mass reduction
 210 during the composting (Compost Mass Reduction-CMR) expressed in percentage (%) was
 211 obtained from the literature. The amount of compost, C_m (in tonnes) that can be obtained
 212 at 60% dryness, was calculated using equation 4.

$$213 \quad C_m(\text{tonnes}) = FS_v \times 1000 \times \frac{TS_v}{10^9} \times \frac{100-CMR}{100} \times \frac{100}{60} \quad (\text{Eq.4})$$

214 The revenue that can potentially be generated from the compost if put on sale was
 215 calculated based on the mass of compost (C_m) and the price of compost fertilizer (C_p)
 216 according to Eq. 5.

$$217 \quad \text{Potential revenue from compost} = C_m(\text{tonnes}) \times C_p(\text{US\$/tonne}) \quad (\text{Eq.5})$$

218 When the sanitary waste stream is treated through composting, results in nutrients
 219 reduction. It was therefore, necessary to quantify nutrient content and consider the
 220 percentage of nutrient reduction during the co-composting (NR_C) of faecal sludge and
 221 organic waste (OW). The expected nutrient content in the compost, NUT_C (tonnes), was
 222 therefore calculated using Eq.6 for faecal sludge and Eq. 7 for organic MSW. Nitrogen
 223 and Phosphorus were the two nutrients analyzed in the analysis.

$$224 \quad NUT_C(\text{tonnes}) = FS_v \times 1000 \times \frac{NUT_v}{10^9} \times \frac{100-NR_C}{100} \quad (\text{Eq.6})$$

$$225 \quad NUT_C(\text{tonnes}) = OW_w \times \frac{TS_m}{100} \times \frac{NUT_w}{10^6} \times \frac{100-NR_C}{100} \quad (\text{Eq.7})$$

226 Where; NUT_C expresses the nutrient content, either Nitrogen or Phosphorus and OW_w is
 227 the weight of organic waste sorted from municipal solid waste.

228 The nutrient content in terms of percentages was calculated by the amount of each nutrient
 229 in the compost and amount of compost itself in tonnes according to Eq. 8.

230 Nutrient content in the compost (%) = $\frac{NUT_C (\text{tonnes})}{C_m(\text{tonnes})} \times 100$ (Eq.8)

231 *Input data for NutRec scenario*

232 In DSM, the average per capita solid waste generation rate has been reported to be 0.815
 233 kg/cap/d out of 78% is composed of organic waste [20–22]. A market price approach was
 234 used to obtain the cost of organic compost through key informant interview. Other
 235 technical values were adopted from the literature. Such values included reduction of FS
 236 and organic solid waste dry mass and nutrients (NR) for nitrogen and phosphorus during
 237 composting was customised to fit the study situation. Table 3 presents a summary of input
 238 data specific for *NutRec* scenario applied to the REVAMP tool.

239 **Table 3.** Input data to REVAMP tool specific for *NutRec* scenario analysis

S/N	Item	Available	Reference
1.	Organic waste (78% of total waste generation)	0.09-3.0 kg/ca/d	[20–22]
2.	Price of packaged compost	0.43 US\$/kg	Interview, 2018 (UMAWA)**
3.	Total Nitrogen reduction from FS during composting, (NR _{N-FS})	34.3% of the initial mass	Schoebitz et al., [17]
4.	Total Nitrogen reduction from OW during composting, (NR _{N-OW})	50% of the initial mass	Galvin, [23]

5.	Total Phosphorus reduction during composting, (NR _P)	1.8% of the initial mass	Schoebitz et al., [17]
6.	Dry mass reduction in compost	19.4% of the initial mass	Breinenbeck and Schellinger [24]

240 ^{**} UMAWA (*Usafi wa Mazingira na Watu*), a local NGO, manage the DEWATS at Kigamboni,
 241 Dar es Salaam.

242 **Other necessary assumptions made**

- 243 • The total number of population in specific areas were adopted from 2017,
 244 forecasted from the last census of 2012 used in the calculation [11].
- 245 • Faecal sludge considered for resource recovery comes from both septic tanks and
 246 pit latrines. No separation of OSS types was taken into considerations.
- 247 • Amount of FS used was calculated from percentages of per capita generation rate
 248 of faeces, urine and greywater goes to OSSs, flowing along the sanitation service
 249 chain as reported in the sheet flow diagram of DSM [3].

250 **2.5 Results implication and use from REVAMP**

251 **2.5.1 Economic benefits indicators of FS resource recovery**

252 The economic values of resource recovery from faecal sludge were determined in two
 253 ways. One, the revenues calculated from potential direct sale based on known amount and
 254 price of products and second value was obtained from potential saved costs from FS
 255 desludging, transport, treatment and disposal [14]. The market value price was used to
 256 estimate the economic value. The currency exchange rate of 1 US\$ = 2,277 Tshs from

257 15thNovember, 2019 was used in revenue calculations (www.bot.go.tz). Second, the
258 economic benefits from diverting FS amount needed for recovery units were obtained
259 from potential Total Cost (TC) to be saved as expressed in equation 9 modified from
260 Ranga et al., [25].

$$261 \quad TC = \sum [(FS \text{ volume, (m}^3/\text{yr)} \times \text{desludged frequency, (month/year)} \times \text{Desludging fee, (USD /m}^3)) + (FS \text{ volume disposed, (m}^3/\text{yr)} \times \text{Disposal fee, (USD /m}^3))] \quad (\text{Eq. 9})$$

263 **where**, TC = total cost saved

264 FS = Feacal sludge

265 USD = United States Dollar

266 **2.5.2 Environmental benefits indicators from resource recovery implementation**

267 The environmental benefits of applying RRR was analysed as one among contributions in
268 combating deforestation. It was evaluated by estimating the forest area saved. The area
269 that would be saved from wood charcoal production by substituting with solid fuel was
270 obtained as explained by Mwampamba [26].

$$271 \quad \text{Forest area (hectares)} = (M_s \times E_k) / S \quad (\text{Eq. 10})$$

272 **Where;**

273 M_s is the mass of single sack contains charcoal (kg of charcoal/sack),

274 E_k is the charcoal kiln efficiency (tones of wood/tones of charcoal) and

275 S is the stock density (tonnes of wood/hectare of the forest).

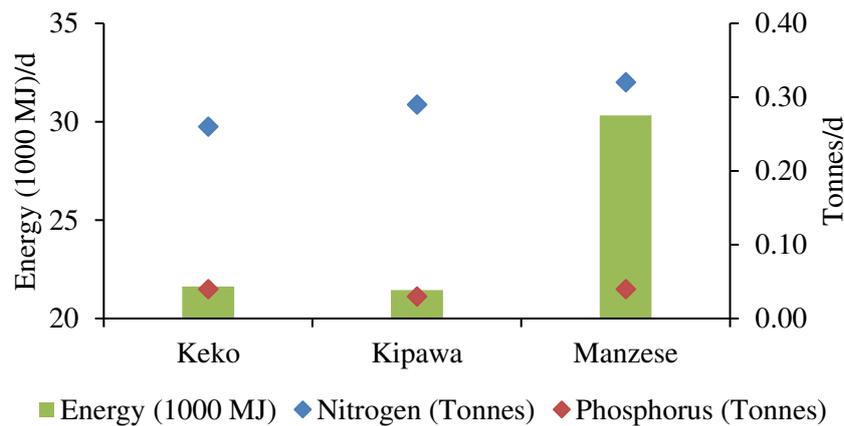
276 In this case, a common single sack of charcoal packaged was taken as 50 kg. The
277 efficiency of traditional unimproved earth kilns commonly used in Tanzania perform at
278 10% efficiency [18]. The average stock density is 66 t/ha [27]. Not all 100% of the forest

279 is harvested for wood charcoal, hence 93% of stock density is suggested for calculation
 280 purposes [26]. Using the formula expressed as Eq.10 and adopted parameters, the forest
 281 area needed to produce 1 tonne of charcoal is 0.00162 ha.

282 3. RESULTS AND DISCUSSION

283 3.1 Nutrients and Energy contents in FS from study areas

284 Results displayed from the REVAMP tool present daily nitrogen and phosphorus as well
 285 as energy amount available in FS for recovery. Based on the amount of FS collected and
 286 delivered to treatment plants from each study area, the daily potential nutrients and energy
 287 content is estimated and presented in a combined bar and scatter plot chart (Figure 3).



288
 289 **Figure 3.** Nitrogen, phosphorus and energy content estimated from the available daily amount of
 290 FS

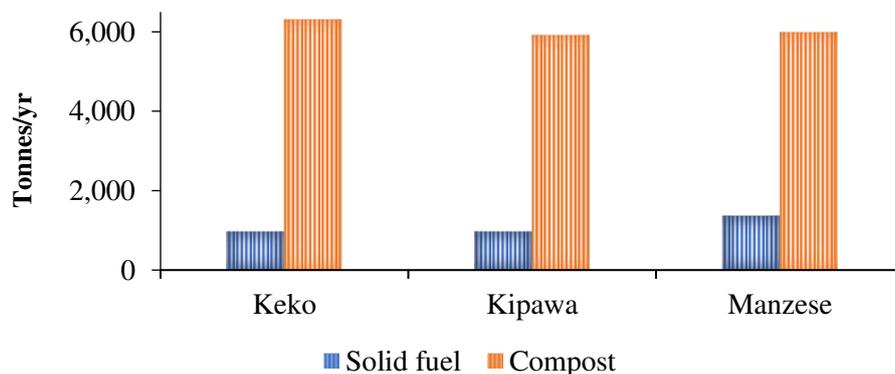
291 The findings show that the largest amount of nitrogen could be obtained from Manzese,
 292 about 29% more than Keko and Kipawa. This could be explained by the fact that Manzese
 293 is predominantly served by pit latrines. Pit latrines contain undiluted and relatively slowly
 294 digested FS as compared to septic tanks [28–30]. The accumulation and dispersion
 295 mechanisms which determine the nitrogen and phosphorus concentrations are slow in pit

296 latrines as compared to septic tanks [31]. Similarly, energy value from raw FS varied with
297 location, with the highest amount coming from Manzese, again, likely linked to the
298 undigested nature of FS found in pit latrines.

299 3.3 Analysis of resource recovery scenarios

300 3.3.1 Potential amount of composted FS and solid fuel generated

301 In the first analysis composted FS and solid fuel production was based on 100% of faecal
302 sludge, without any other organic waste (OW). Here the results indicates that Keko has
303 the possibility of producing significantly larger amounts of compost compared to Manzese
304 and Kipawa. In the second analysis, co-composting was performed by mixing FS with
305 OW, calculated from the per capita generation rate. As expected, the largest amount of
306 co-composted FS thus comes from the most populated area, Keko, here a possible total
307 daily amount of 49.9 tonnes could be produced, with a nutrient composition of 9%, 6%
308 and 0.4% for nitrogen, phosphorus and potassium respectively.



309

310 **Figure 5.** Amount of compost and solid fuel produced per year in the study area

311 The total annual amount of solid fuel generated across the study areas are 3,310.55 tonnes
312 equivalent to 9.07 tonnes per day. The lower amount of solid fuel compared to the
313 composted FS was probably attributable to the pyrolysis process which reduces the initial

314 weight of dried FS up to 50% or more [19]. There were no additives which were added as
315 the case for composting which could also enhance the addition in mass for the final
316 product obtained.

317 Generally, the findings show that the availability of nutrients (nitrogen and phosphorus)
318 contents from FS which can be recovered for agricultural purposes is high. About 0.26 –
319 0.32 tonnes of nitrogen and up to 0.04 tonnes of phosphorus contents in raw FS could be
320 recovered annually. The source of phosphorus content in OSSs could be from soap and
321 detergents used in cleaning and personal health care products [6]. Moreover, the majority
322 of houses in this study were observed to divert greywater away from the OSSs, possibly
323 diluting phosphorus contents [10].

324 The collective estimated FS-based compost from produced in this study, resulted in 49.9
325 tonnes/d. The recommended National fertilizer application rate in agricultural land of
326 Tanzania is 7 kg/ha [12]. Based on that, the obtained amount is equivalent to and could
327 be enough for application in 7,128 hectares.

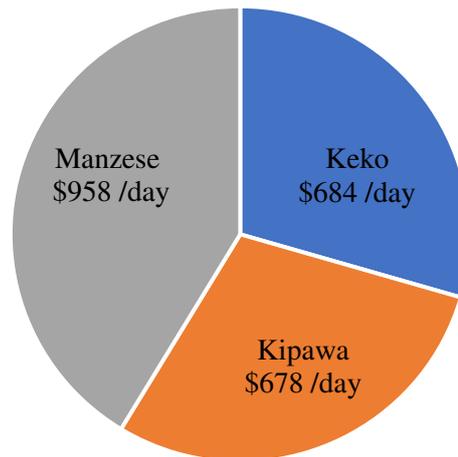
328 The *NutRec* scenario analysis indicates that in terms of nutrients content, the obtained
329 amount of FS-based compost analyzed in the REVAMP tool has 9% nitrogen content.
330 However, the most common chemical fertilizer consumed by local farmers in Tanzania is
331 UREA which contains 46% of nitrogen (460 kg of nitrogen in 1 tonne of fertilizer) [32].
332 Therefore, the faecal sludge-based compost from this study could only substitute 19% of
333 the nitrogen compared to UREA. Based solely on the nitrogen content, FS-based compost
334 is therefore not a viable alternative to UREA to improve direct plant growth. However, it
335 could be used as a soil conditioner to improve soil texture, permeability and porosity [33].

336 Results from *EnRec* scenario analysis were compared to the 32 MJ/kg energy value of
337 wood charcoal [26,34]. The calculations include the hypothetical FS energy value of 16.2
338 MJ/kg and the maximum energy content of 30,000 MJ obtained in Manzese. The results
339 indicate that the energy potential of FS-based briquettes would only substitute an average
340 of 15% of wood charcoal by mass and 50% of the energy value. In other words, to reach
341 the energy value required, two portions of FS briquettes would be needed to substitute a
342 single portion of wood charcoal (e.g. 2kg briquettes equivalent to 1 kg wood charcoal).

343 **3.3.2 Potential Economic Benefits from Recovered Products**

344 *Potential revenue from products sales*

345 It is important to note that the revenue obtained in this study considers the direct sales of
346 the FS derived products at local market prices during the study period. There were no
347 additional revenue adjustments done linked to crop yields increase from application of
348 FS-derived soil conditioners on agricultural land. For the FS-derived compost the average
349 price was based on estimates obtained during the interviews while the price of FS-derived
350 briquettes was estimated in comparison to wood charcoal. The estimated daily revenue
351 generated from FS based products from possible direct sales was then analyzed based on
352 equations 3 and 5 with summarized results in figures 6 and 7.



353

354

Fig 6. Daily revenue generated from potential sales of FS-based solid fuel

355

As shown in figure 6, Manzese shows the potential for the highest daily revenue generated from direct sales of solid fuel, about 28% more than other areas. Again, this might be attributed to the fact that the largest amount of solid fuel with high calorific value is generated FS from pit latrines which is more biologically unstable than that of septic tanks.

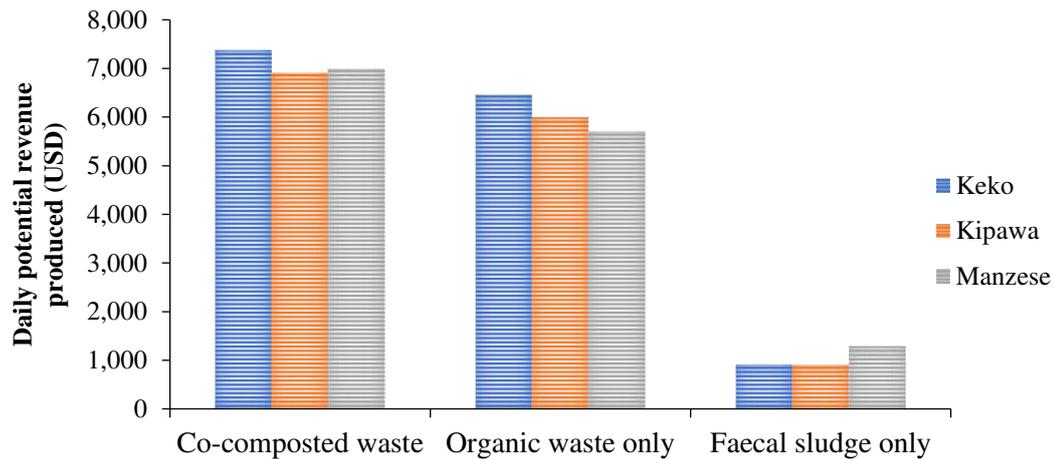
359

Furthermore, the economic value of compost was compared from composted FS-alone, OW-alone, and co-composted FS and OW. As can be seen from Figure 7, the products from the co-compost waste generates more profit across all study areas than other compost materials. The products from individual waste material were the least profitable.

360

361

362



363

364 **Figure 7.** Potential revenue generated from sales of compost products from different waste
 365 materials

366 The revenue analysis shows the variation of daily profit generated from compost material.
 367 The minimum revenue ranges 900 -1300 USD can potentially be generated by composting
 368 faecal sludge only while the revenue increased to the maximum of 7,000 USD as different
 369 waste streams co-composted together. The revenue obtained from compost product is
 370 slightly higher than from FS derived briquettes which its highest daily profit generated
 371 was below 1,000 USD. The findings from this study are in line with the results reported
 372 in other studies that composting of FS offers a business and development opportunity that
 373 could benefit millions of poor farmers [2]. Kampala, Uganda, for example, FS has been
 374 commonly sold for agriculture purposes from 5 USD to 10 USD per tonne [15]. Research
 375 by Gold et al., [4] also show that the interest in using FS as a soil conditioner in urban
 376 areas of Sub-Saharan Africa is higher than other FS-based products. Indeed, it has been
 377 estimated that if all FS collected in the city of Dakar, Senegal and Kampala, Uganda was
 378 sold as soil conditioners it could generate a gross value of USD 12,480/year in Dakar and
 379 USD 81,120/year in Kampala. When compared to the revenues of FS-derived briquettes
 380 indicating and even higher profitability compared to soil conditioners [33]. In this

381 particular case this was possible because the primary consumers of the briquettes were
382 private sector companies buying them in bulk as an energy source to fuel industrial
383 processes

384 When considering composting it has been shown that when FS is mixed with other OW it
385 has the highest revenue possibilities [35]. This is because co-composting can ensure a
386 steady volume and quality end-product [36]. This finding needs to be taken into
387 considerations when planning for resource recovery through composting.

388 *Potential saved costs from desludging and disposal*

389 In the studied sites and Dar es Salaam at large, costs for FS desludging and transport
390 remains a heavy financial burden for tenants and land lords at the household level and the
391 discharge fees at the disposal points are incurred by service delivery entities [37]. With
392 improved resource recovery of FS throughout the sanitation value chain part of these costs
393 could be reduced or avoided altogether [38]. Currently households in the study areas
394 reported an average desludging rate of once every three months or four times annually
395 [10]. Today the minimum desludging costs averages 100,000 Tsh (44 USD) per pit plus a
396 minimum FS disposal fee of 10,000 Tsh (4.4 USD) for a vacuum tanker of 6000 litres
397 capacity (0.73 USD per m³ of FS) paid by emptiers to the city's WSP facilities.
398 Calculations of the potential costs saved from FS resource recovery in the study areas (See
399 Eq. 9) indicate an estimated daily saving of 152 USD, 150.8 USD and 213 USD from
400 Keko, Kipawa and Manzese respectively. These funds could then possibly be used to
401 offset operations and maintenance (O&M) costs linked with running future resource
402 recovery treatment plants that would benefit both individual households and the city as a
403 whole

404 **3.3.3 Environmental Benefits**

405 Besides possible economic benefits from FS resource recovery and improving FSM,
406 additional environmental benefits could be linked beyond the sanitation system itself. The
407 use of FS-based briquettes as an alternative to wood charcoal has opportunity to contribute
408 to mitigate climate change, by combating deforestation. In this study, the total forest area
409 that could be saved by substituting FS-derived briquettes for wood charcoal has been
410 calculated in equation 10 and the results are summarized in Table 4.

411 **Table 4.** Forest area that would be saved by using FS based briquettes produced

Study area	Annual Mass of briquettes produced (tonnes/yr)	Equivalent Forest area saved (ha)
Keko	974.55	1.58
Kipawa	967.25	1.57
Manzese	1,368.75	2.22
TOTAL	3,310.55	5.37

412 Collectively, the annual quantity of FS-based briquettes produced from the three study
413 areas could substitute an average of 15% of the wood charcoal currently used for cooking
414 in the settlements. This substitution suggests an estimated 5 hectares of forest land that
415 could potentially be conserved from being cut down per year.

416

417 **4. CONCLUSIONS**

418 Two scenarios analyzing the recovery of nutrients and energy from FS for agricultural and
419 domestic uses respectively were quantified for economic and environmental benefits.
420 Analysis from REVAMP indicates the possibilities of recovering nutrients as well as
421 energy from collected FS. Results further revealed the possibility of recovering nutrients
422 with a daily total production of 49.9 tonnes of composted fertilizer. This amount is enough
423 to substitute 19% of nitrogen obtained from UREA fertilizer. Along with organic matter,
424 the FS based compost could improve soil texture, porosity and permeability. These are
425 benefits which cannot be realized using chemical fertilizers. The co-composting of FS
426 with organic waste to recover nutrients was found to be more profitable than solid fuels.
427 The produced 9.07 tonnes of solid fuel could be able to substitute 15% of wood charcoal
428 consumed, covers 50% of energy amount. However, the revenue generated from both
429 recovery options varied. Compost is more profitable than energy source. The study also
430 found that to obtain briquettes with high calorific value, FS should be co-fuelled with
431 other feedstock materials and co-composting should be advocated when planning for
432 resource recovery through composting.

433 So far however, FS resource recovery practices across the African continent, and in
434 Tanzania especially, is still rare, but case study research, including this one, indicates that
435 there are both economic and environmental benefits to be had, beyond the likely positive
436 impacts for FSM with significant reduction of untreated FS volumes entering and
437 polluting local waterways and soils and posing human health risks. Yet, in order to start
438 this process towards FS resource recovery and reuse the safety and health aspects of
439 handling and manufacturing FS based products needs more attention and guidelines and

440 standards for safe re-use of FS derived products should therefore be developed and
441 adopted.

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449 **Authors' contributions**

450 All authors read and approved the final manuscript.

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457 **Availability of data and materials**

458 All data generated or analysed during this study are included in submitted manuscript

459 **Competing interests**

460 The authors declare they have no competing interests.

461 **REFERENCES**

- 462 1. Esrey SA, Andersson I, Hillers A, Sawyer R. Ecological sanitation for food
463 security. Sawyer R, editor. Tepoztlán, Mexico: Swedish International Development

- 464 Cooperation Agency; 2000. 107 p.
- 465 2. Rao KC, Otoo M. Resource Recovery and Reuse as an Incentive for a More Viable
466 Sanitation Service Chain. *Water Altern.* 2017;10(2):493–512.
- 467 3. Brandes K, Schoebitz L, Kimwaga R, Strande L. Shit Flow Diagram (SFD)
468 Promotion Initiative Dar es Salaam Tanzania. EAWAG/SANDEC- Water and
469 Sanitation in Developing Countries. Dar es Salaam, Tanzania; 2015.
- 470 4. Gold M, Niang S, Niwagaba CB, Eder G, Muspratt AM, Diop PS, et al. Results
471 from FaME (Faecal Management Enterprises) – can dried faecal sludge fuel the
472 sanitation service chain? In: 37th WEDC International Conference-
473 SUSTAINABLE WATER AND SANITATION SERVICES FOR ALL IN A
474 FAST CHANGING WORLD, Hanoi, Vietnam, 2014. pp. 1–6.
- 475 5. Diener S, Semiyaga S, Niwagaba CB, Muspratt AM, Gning JB, Mbéguéré M, et al.
476 A value proposition: Resource recovery from faecal sludge - Can it be the driver
477 for improved sanitation? *Resour Conserv Recycl.* 2014;88:32–8.
- 478 6. Harder R, Wielemaker R, Larsen TA, Zeeman G, Öberg G. Recycling nutrients
479 contained in human excreta to agriculture: Pathways, processes, and products. *Crit
480 Rev Environ Sci Technol.* 2019;49(8):695–743.
- 481 7. Gwenzi W, Chaukura N, Mukome FND, Machado S, Nyamasoka B. Biochar
482 production and applications in sub-Saharan Africa: Opportunities, constraints, risks
483 and uncertainties. *J Environ Manage.* 2015;150:250–61.
- 484 8. Harada H, Strande L, Fujii S. Challenges and Opportunities of Faecal Sludge
485 Management for Global Sanitation. In: Katsumi T, Hashimoto S, editors. *Towards*

- 486 Future Earth: Challenges and Progress of Global Environmental Studies. Kaisei
487 Publishing, Tokyo; 2016. p 81–100.
- 488 9. Loetscher T, Keller J. A decision support system for selecting sanitation systems
489 in developing countries. *Socioecon Plann Sci.* 2002;36:267–90.
- 490 10. Mkude IT, Mbwette T, Kimwaga R, Gabrielsson S. Material Flow Analysis as a
491 Decision Supporting tool for Faecal Sludge Resource Recovery : Mathematical
492 formulation and quantification. *Tanzania J Eng Technol.* 2019;38(1):97–115.
- 493 11. NBS. Annual Agriculture Sample Survey Crop and Livestock Report. United
494 Repub Tanzania. 2017;1–181.
- 495 12. Benson T, Kirama SL, Selejio O. The Supply of Inorganic Fertilizers to
496 Smallholder Farmers in Tanzania Evidence for Fertilizer Policy Development.
497 International Food Policy Research Institute, Addis Ababa, Ethiopia. IFPRI
498 Discussion paper 01230. 2012.
- 499 13. Tayler K. Faecal Sludge and Septage Treatment: A guide for low and middle
500 income countries. Rugby, Warwickshire: Practical Action Publishing; 2018. 370 p.
- 501 14. SEI. Resource Value Mapping (REVAMP): A tool for evaluating the resource
502 recovery potential of urban waste streams. DISCUSSION BRIEF Resource,
503 Stockholm, Sweden. 2016.
- 504 15. Ddiba D. Estimating the potential for resource recovery from productive sanitation
505 in urban areas. TRITA-LWR Degree Project, Royal Institute of Technology
506 (KTH); 2016.
- 507 16. Diener S, Reiser JC, Murray A, Mbéguéré M, Strande L. Recovery of industrial

- 508 waste heat for faecal sludge drying. Sandec News. 2012;(13):16.
- 509 17. Schoebitz L, Niwagaba C, Francis O, Bischoff F, Strande L. FAQ : Faecal Sludge
510 Quantification and Characterization – Kampala. Sandec News.
511 2014;15(January):12–3.
- 512 18. Msuya N, Masanja E, Temu AK. Environmental Burden of Charcoal Production
513 and Use in Dar es Salaam, Tanzania. J Environ Prot (Irvine, Calif).
514 2011;02(10):1364–9.
- 515 19. Lohri CR, Sweeney D, Rajabu HM. Carbonizing urban biowaste for low-cost char
516 production in developing countries - A review of knowledge, practices and
517 technologies. A Rev Knowledge, Pract Technol Jt Rep by Eawag, MIT D-Lab
518 UDSM. 2015;58.
- 519 20. Hoornweg D, Bhada-Tata P. What a Waste: A Global Review of Solid Waste
520 Management. Urban Development Series; knowledge papers no.15, World Bank.
521 Washington D.C: world Bank; 2012. 116 p.
- 522 21. Kaseva ME, Mbuligwe SE. Appraisal of solid waste collection following private
523 sector involvement in Dar es Salaam city, Tanzania. Habitat Int. 2005;29(2):353–
524 66.
- 525 22. Kirama A, Mayo AW. Challenges and prospects of private sector participation in
526 solid waste management in Dar es Salaam City, Tanzania. Habitat Int.
527 2016;53:195–205.
- 528 23. Galvin C. Embodied Energy and Carbon Footprint of Household Latrines in Rural
529 Peru: The Impact of Integrating Resource Recovery. University of South Florida;

- 530 2013.
- 531 24. Breitenbeck GA, Schellinger D. Calculating the Reduction in Material Mass And
532 Volume during Composting. *Compost Sci Util.* 2004;12(4):365–71.
- 533 25. Uding Ranga J, Syed Ismail S, Rasdi I, Karuppiah K, Ikmal Irozi M.
534 Environmental Impact, Health Risk, and Management Cost of Landfilling Practice :
535 A Case Study in Klang, Selangor, Malaysia. *J Waste Manag Dispos.* 2019;2(1):1–
536 12.
- 537 26. Mwampamba TH. Has the woodfuel crisis returned ? Urban charcoal consumption
538 in Tanzania and its implications to present and future forest availability. *Energy*
539 *Policy.* 2017;35:4221–34.
- 540 27. Lupala ZJ, Lusambo LP, Ngaga YM. Management, Growth, and Carbon Storage
541 in Miombo Woodlands of Tanzania. *Int J For Res.* 2014;2014:1–11.
- 542 28. Strande L, Schoebitz L, Bischoff F, Ddiba D, Okello F, Englund M, et al. Methods
543 to reliably estimate faecal sludge quantities and qualities for the design of treatment
544 technologies and management solutions. *J Environ Manage.*
545 2018;223(February):898–907.
- 546 29. Nzouebet L, Kengne IM, A. Rechenburg. Does Depth And Sanitation Type Affect
547 The Quality Of Faecal Sludge In The Tropics? The Case of Yaoundé, Cameroon.
548 *Open Water.* 2015;3(1):15.
- 549 30. Bassan M, Tchonda T, Yiougo L, Zoellig H, Mahamane I, Mbéguéré M, et al.
550 Characterization of faecal sludge during dry and rainy seasons in Ouagadougou,
551 Burkina Faso. In: 36th WEDC International Conference: Delivering Water,

- 552 Sanitation and Hygiene Services in an Uncertain Environment. Nakuru, Kenya;
553 2013. p. 1–6.
- 554 31. Montangero A. Material Flow Analysis for Environmental Sanitation Planning in
555 Developing Countries-An approach to assess assess material flows with limited
556 data availability. [Austria]: Leopold-Franzens-University Innsbruck; 2006.
- 557 32. Chianu J, Chianu JM. Mineral fertilizers in the farming systems of sub-Saharan
558 Africa. A review. *Agron Sustain Dev Springer Verlag/EDP Sci.*
559 2012;32(2):pp.545-566.
- 560 33. Impraim R, Nikiema J, Cofie O, Rao K. Value from faecal sludge and municipal
561 organic waste : fertilizer cum soil conditioner in Ghana. In: 37th WEDC
562 International Conference SUSTAINABLE WATER AND SANITATION
563 SERVICES FOR ALL IN A FAST CHANGING WORLD. Hanoi, Vietnam; 2014.
564 p. 3–8.
- 565 34. Muspratt AMA, Nakato T, Niwagaba C, Dione H, Kang J, Stupin LJ, et al. Fuel
566 potential of faecal sludge: calorific value results from Uganda, Ghana and Senegal.
567 *J Water, Sanit Hyg Dev.* 2014;4(2):223.
- 568 35. Drechsel P, Kunze D. *Waste Composting for Urban and Peri-Urban Agriculture.*
569 CABI Publishing; 2001. 229 p.
- 570 36. Keener HM, Dick WA, Hoitink HAJ. Composting and Beneficial Utilization of
571 Composted By-Product Materials. In: *Land application of Agricultural, Industrial
572 and Municipal By-Products.* 2000. p. 315–41.
- 573 37. Seleman A, Gabrielsson S, Mbwette TSA, Kimwaga R. Drivers of unhygienic

574 desludging practices in unplanned settlements of Dar es Salaam, Tanzania. *J Water*
575 *Sanit Hyg Dev.* 2020;10(3):512–26.

576 38. Tilley E, Dodane P. Financial Transfers and Responsibility in Faecal Sludge
577 Management Chains. In: Strande L, Mariska R, Damir B, editors. *Faecal Sludge*
578 *Management Systems Approach for Implementation and Operation.* IWA
579 Publishing; 2014. p. 273–91.

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Figures

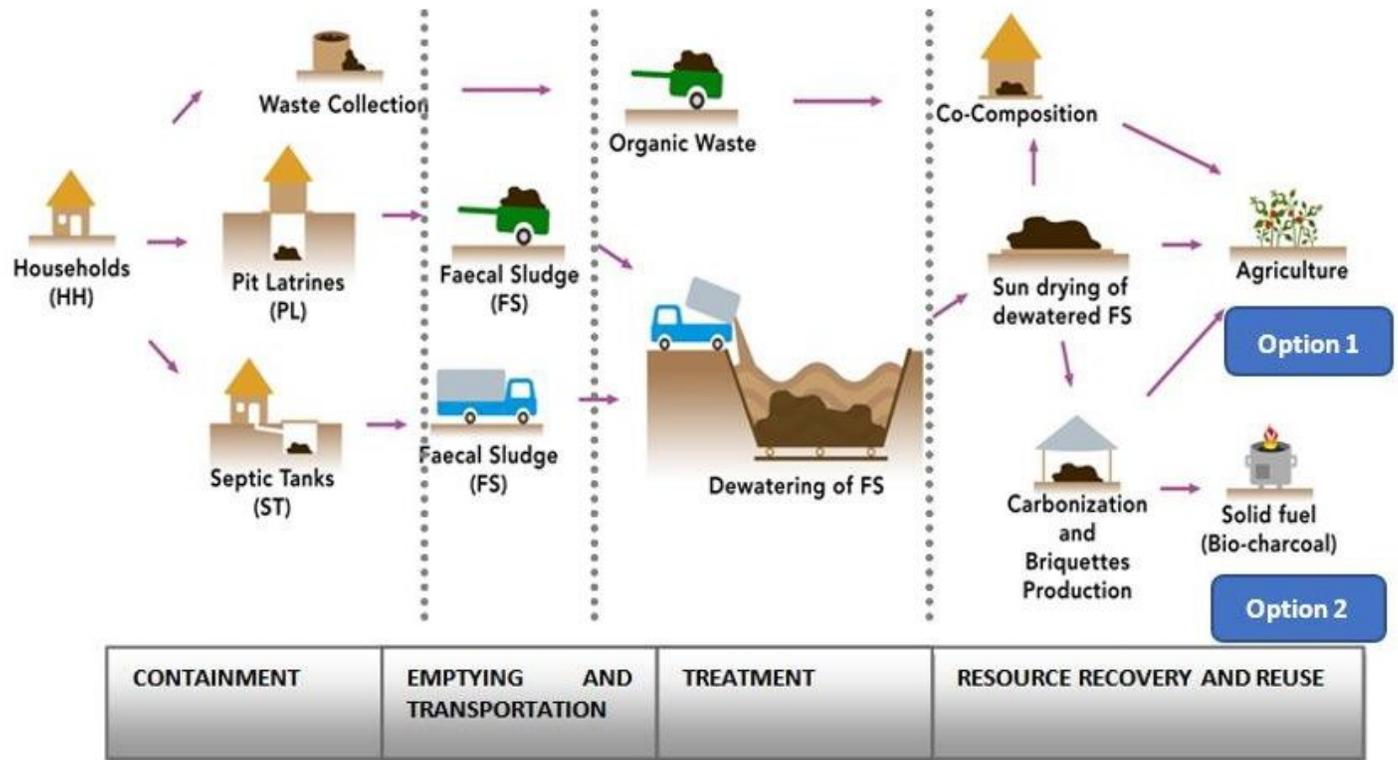
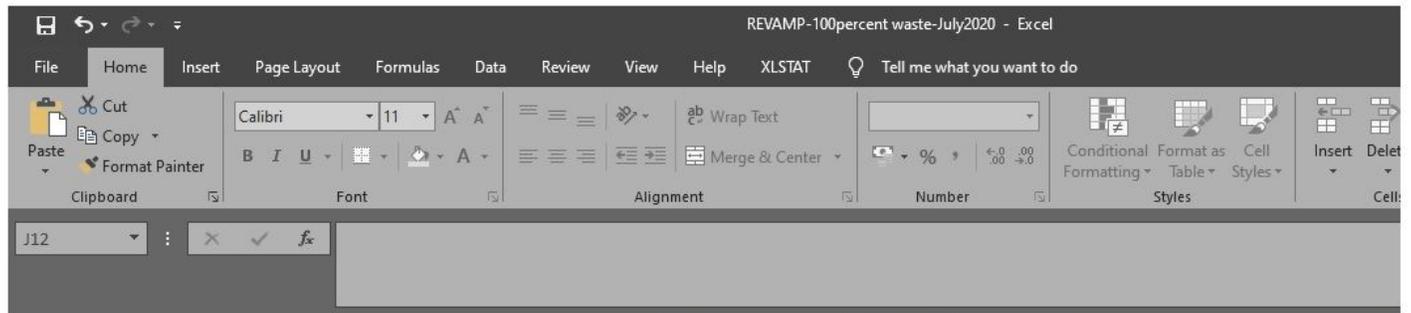


Figure 1

Schematic presentation of proposed resource recovery scenarios; co-composting and solid fuel production



Parameter	Units	Faecal Sludge	Reference(s)
Volatile Solids Degradation rate, VS_D	%	70.00	Estimate based on Von Sperling & de Lemos Chernicharo
Dry Mass Reduction rate for anaerobic digestion (AD) residue, DMR_{AD}	% of initial TS		Alfa et al. (2014)
Biomass conversion rate for black soldier fly (BSF) larvae, BCR	%		Banks (2014)
Dry Mass Reduction rate for BSF residue, DMR_{BSF}	% of initial TS		Banks (2014)
Total nitrogen (TN) reduction in BSF residue, TNR_{BSF}	% of initial TN		Assumed based on Sheppard, Newton and Burtle (2008)
Total phosphorus (TP) reduction in BSF residue, TPR_{BSF}	% of initial TP		Assumed based on Sheppard, Newton and Burtle (2008)
Total potassium (TK) reduction in BSF residue, TKR_{BSF}	% of initial TK		Assumed based on Sheppard, Newton and Burtle (2008)
Dry Mass Reduction in compost, DMR_C	% of initial mass	19.40	Average based on Breitenbeck & Schellinger (2004)
Total nitrogen (TN) reduction during composting, TNR_C	% of initial TN	34.30	Average based on Galvin (2013)
Total phosphorus (TP) reduction during composting, TPR_C	% of initial TP	1.77	Average based on Eghball et al. (1997)

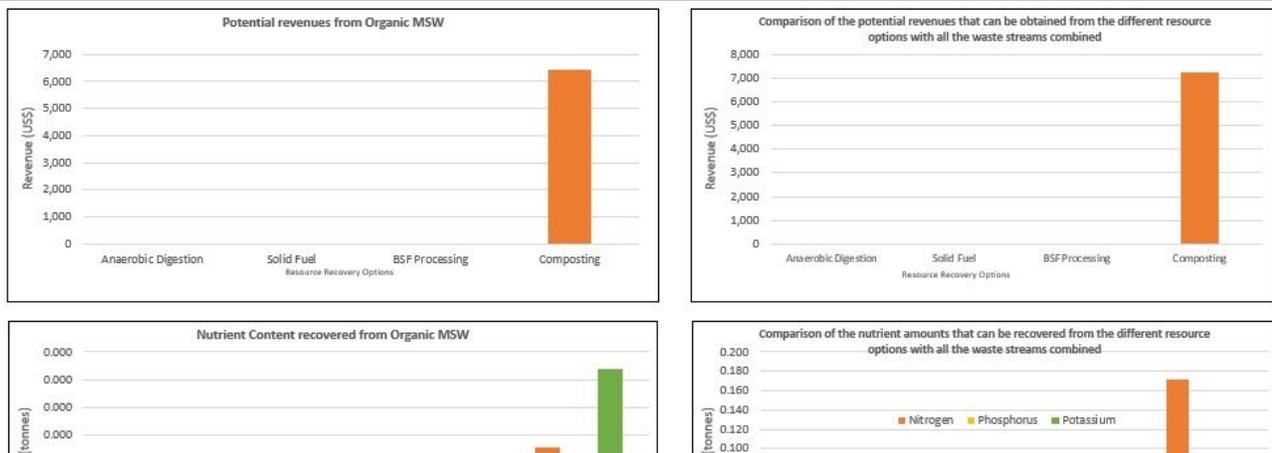
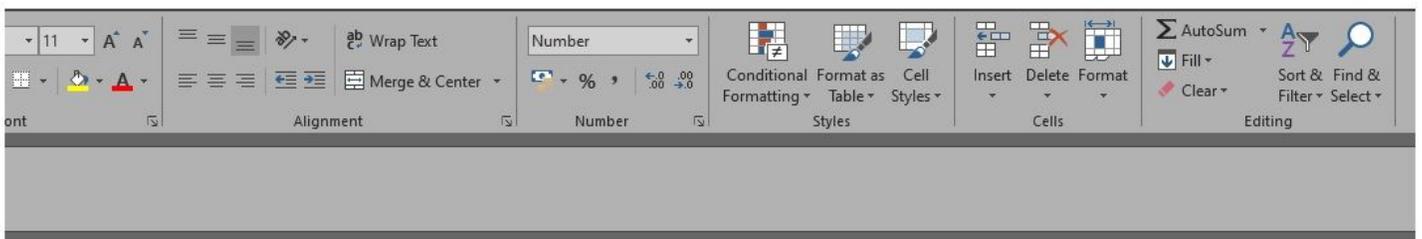


Figure 2

(a). The REVAMP tool interface estimating waste streams treatment processes (b). The REVAMP tool interface displaying final results

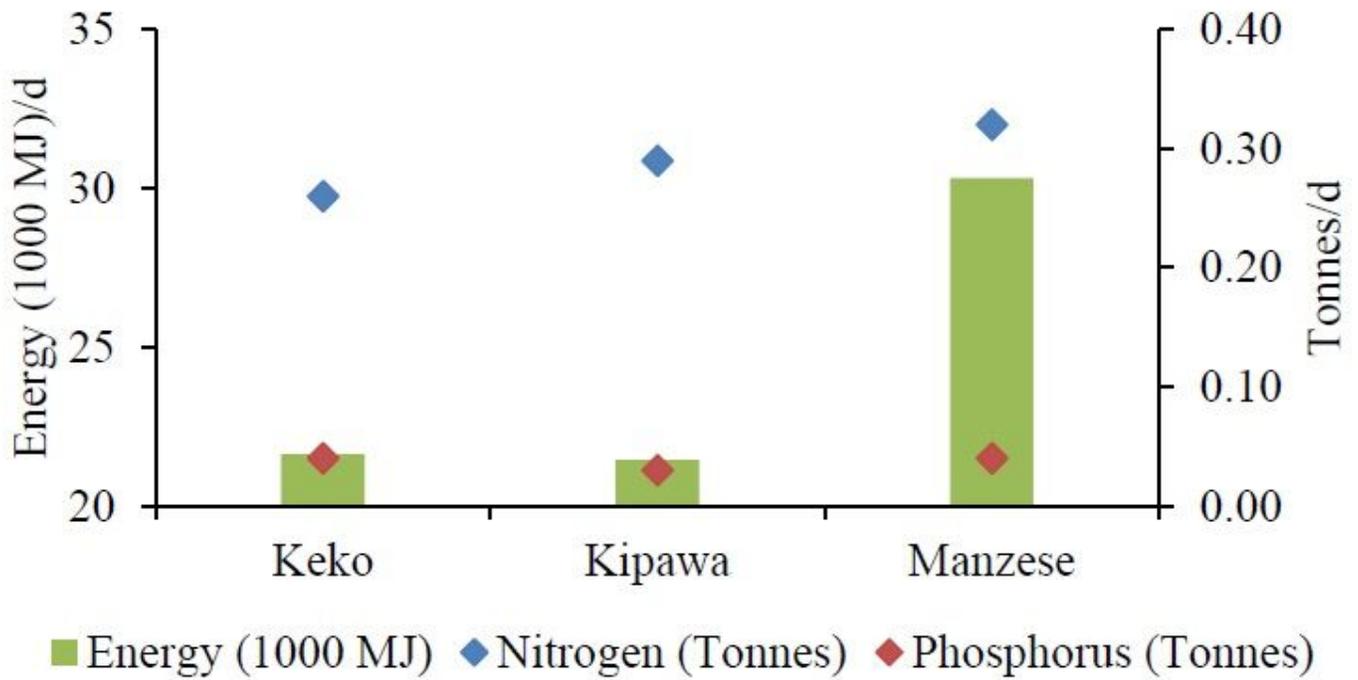


Figure 3

Nitrogen, phosphorus and energy content estimated from the available daily amount of FS

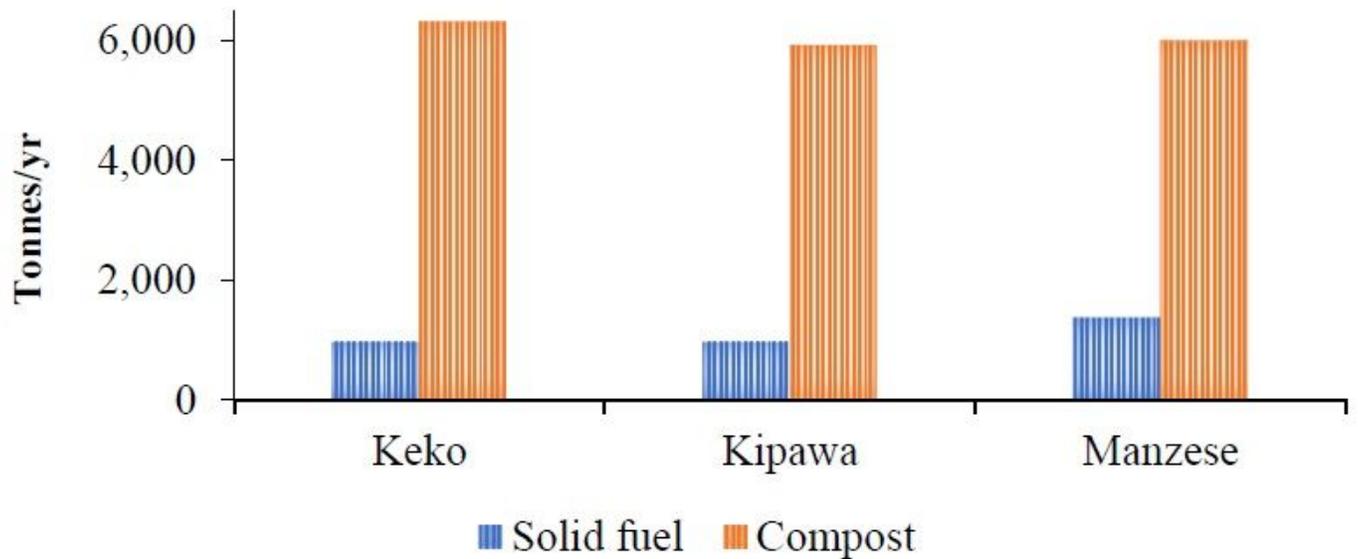


Figure 4

Amount of compost and solid fuel produced per year in the study area

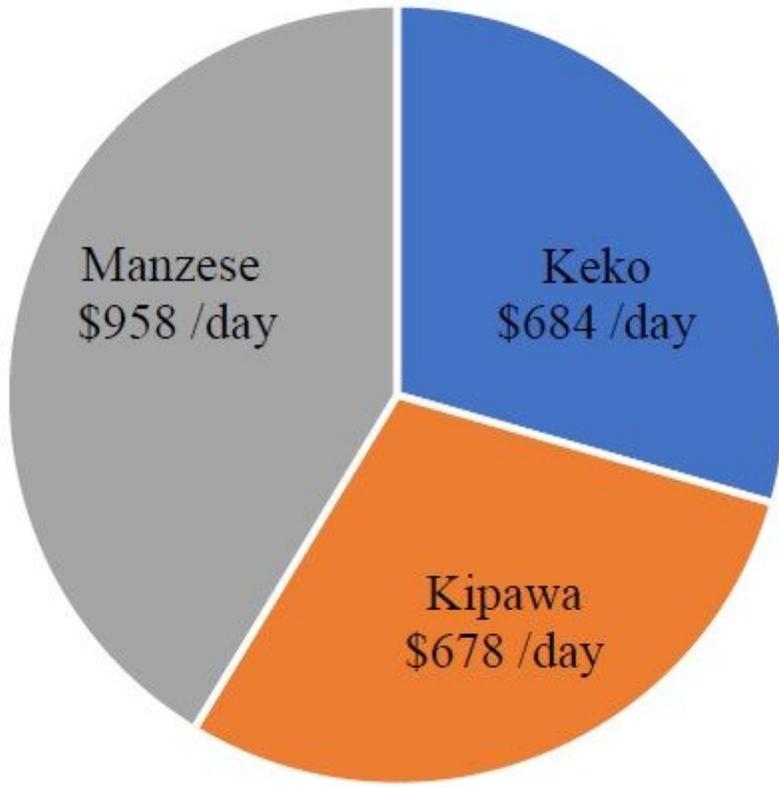


Figure 5

Daily revenue generated from potential sales of FS-based solid fuel

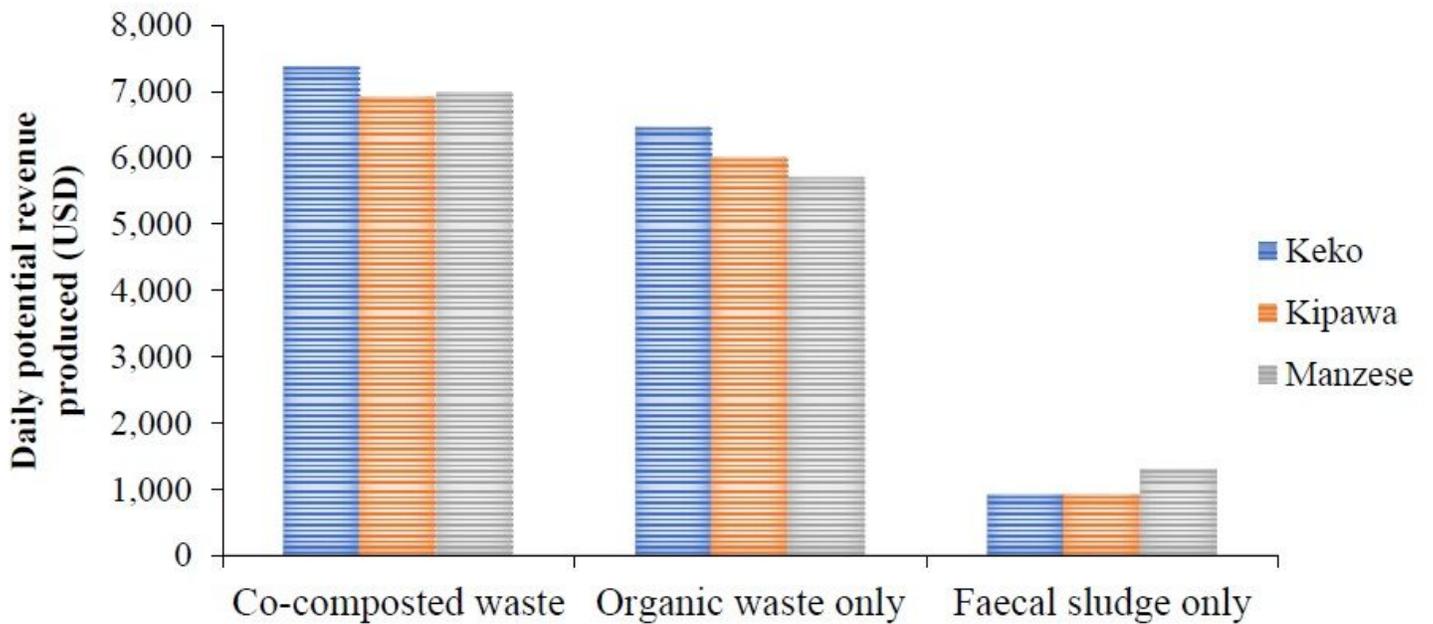


Figure 6

Potential revenue generated from sales of compost products from different waste materials