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Research Article

Keywords: Advanced oxidation process improvement, Iron complexation, Microcontaminants elimination, Municipal wastewater treatment, Natural enhancers

Posted Date: March 22nd, 2021

DOI: <https://doi.org/10.21203/rs.3.rs-318961/v1>

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Version of Record: A version of this preprint was published at Environmental Science and Pollution Research on July 14th, 2021. See the published version at <https://doi.org/10.1007/s11356-021-15377-1>.

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21 **Abstract**

22 Extracts of copoazu (*Theobroma grandiflorum*), canangucha (*Maurita Flexuosa*),
23 and coffee (*coffea arabica*) were explored as enhancers of the solar photo-Fenton
24 process to eliminate acetaminophen, sulfamethoxazole, carbamazepine, and
25 diclofenac. The process performance, at pH 6.2 and 5 mg L⁻¹ of iron without the
26 extracts, has a very limited action (~35% of the pollutants degradation at 90 min of
27 treatment) due to the iron precipitation. Interestingly, the extracts addition increased
28 the soluble iron forms but only copoazu extract improved the pollutants degradation
29 (~95% of elimination at 90 min of the process action). The copoazu extract acted as
30 a natural complexing agent, maintaining the soluble iron up to 2 mg L⁻¹ even after 90
31 min, and consequently enhancing the pollutants degradation. The effect of copoazu
32 extract dose on the process performance was also assessed, finding that an iron:
33 copoazu extract molar ratio equal to 1:0.16 was the most favorable condition. Then,
34 the process improved by copoazu extract was applied to municipal wastewater.
35 Remarkably, the process led to ~90% of total pharmaceuticals degradation at 20 min
36 of treatment. This work evidences the feasibility of amazonian fruits extracts to
37 improve the solar photo-Fenton process to degrade pharmaceuticals in aqueous
38 matrices at near-neutral pH.

39

40 **Keywords:** Advanced oxidation process improvement; Iron complexation;
41 Microcontaminants elimination; Municipal wastewater treatment; Natural enhancers.

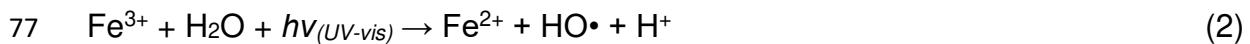
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43 **1. Introduction**

44 Untreated municipal wastewater (MWW) is recognized as the most hazardous
45 pollution source to environmental water due to the high presence of nutrients (e.g.,
46 N and P) and organic contaminants (Preisner 2020). Pharmaceuticals are among
47 organic pollutants commonly found in MWW (Liu et al. 2017; Shraim et al. 2017;
48 Botero-Coy et al. 2018). Although the discharge of pharmaceuticals into the
49 environment is not yet regulated, there is a growing interest in their final disposal,
50 thus they are considered as contaminants of emerging concern (CECs) (Küster and
51 Adler 2014; Sánchez Pérez et al. 2020). Indeed, pharmaceuticals can induce
52 toxicity, mortality, and developmental abnormalities in aquatic species (Bielen et al.
53 2017). Thus, their input into the environment should be limited/avoided. An option to
54 eliminate pharmaceuticals in MWW is the application of advanced oxidation
55 processes (AOP), which are based on the generation and utilization of short-lived,
56 strong oxidant, and highly reactive radical species (mainly hydroxyl radical, $\text{HO}\cdot$, $E^\circ: 2.80 \text{ V}$ (Armstrong et al. 2013)). To produce the hydroxyl radical, catalytic AOPs
57 such as photo-Fenton are widely used (Miklos et al. 2018).

59 Photo-Fenton comprises the reaction of ferrous ions with hydrogen peroxide initially
60 (Eq. 1). Then, the ferric species can be reduced in the aqueous medium by the action
61 of UV-Vis light, producing extra hydroxyl radicals (Eq. 2) and making the system a
62 catalytic process (Pignatello et al. 2006). Photo-Fenton process has shown to be
63 very efficient in the degradation of pollutants in MWW (Papoutsakis et al. 2016). This
64 process has gained the attention of researchers due to its operational easiness and
65 the possibility of solar irradiation utilization (Pouran et al. 2015; Clarizia et al. 2017).

66 However, the photo-Fenton process at pH higher than 3.0 is limited by Fe^{3+}
67 precipitation (Eq. 3) (Villegas-Guzman et al. 2017b). Interestingly, some organic
68 compounds can complex iron (III) keeping it in soluble forms, allowing it to work at
69 near-neutral pH conditions (Clarizia et al. 2017). Furthermore, many Fe^{3+} -organic
70 ligand complexes can absorb light in the near-UV and visible regions more efficiently
71 than aqua-complexes and undergo photo-reduction through a ligand-to-metal
72 charge transfer (LMCT). Then, the ferrous ion is regenerated (Eq. 4) and other
73 radicals are produced (Eqs. 4-5, which could contribute to the oxidation of organic
74 substances) (Miralles-Cuevas et al. 2014; Clarizia et al. 2017; Villegas-Guzman et
75 al. 2017a).



81 Several poly-carboxylate acids that act as chelating agents of iron such as oxalic
82 acid, citric acid, ethylenediamine-N,N-disuccinic acid (EDDS), ethylene-diamine-
83 tetra-acetic acid (EDTA), and nitrile-acetic acid (NTA), have been successfully
84 utilized to improve the solar photo-Fenton process performance at near-neutral pH
85 (Arzate et al. 2020). Nonetheless, EDTA or NTA have high toxicity and low
86 biodegradability. Meanwhile, chelation with citric or oxalic acid requires reduced pH
87 for their best performance (for instance, the optimal behavior for the Fe^{3+} -citrate and

88 Fe³⁺-oxalate complexes, occurs at pH 5.0 and 4.3, respectively) (Papoutsakis et al.
89 2015). In turn, iron complexation with EDDS represents additional costs to the
90 process by the purchasing of such reagent. Thus, to overcome these issues, natural
91 substances polyphenols-rich have been recently evaluated (Ruiz-Delgado et al.
92 2019; Prada-Vásquez et al. 2020). Also, current works are focused on the search for
93 non-toxic and biodegradable chelating compounds, which must be environmentally
94 sustainable and inexpensive to propose a large-scale photo-Fenton process at near-
95 neutral pH (Klamerth et al. 2013). Indeed, juices of orange or lime and coffee extract
96 (organic acid-rich compounds) have been evaluated as SPF enhancers with
97 promising results for the elimination of CECs and bacteria (Villegas-Guzman et al.
98 2017b).

99 In this context, natural products rich in polyphenols and/or organic acids have
100 interesting potential in the application of solar photo-Fenton (SPF) at natural pH.
101 Thus, chelating compounds that are biodegradable, do not generate toxicity after
102 carrying out the treatment, and inexpensive could lead to an SPF environmentally
103 sustainable with the potential to propose a large-scale (Klamerth et al. 2013). Some
104 of the abundant native products from Colombian Amazonia contain polyphenols
105 (Genaro-Mattos et al. 2015; Mussatto 2015; Geremu et al. 2016; Heeger et al. 2017),
106 such products have potential as iron complexing agents. Therefore, in our research,
107 three typical products of the region: copoazu (also named cupuassu, *Theobroma*
108 *gramdiflorum*), canangucha (*Maurita Flexuosa*), and husk coffee (*Coffea arabica*)
109 were considered.

110 Copoazu and canangucha are typical fruits of the Amazon region (Rogez et al. 2004;
111 Contreras-Calderón et al. 2011; Galeano 2011; Koolen et al. 2013; Nobre et al.
112 2018), these have been used in the food industry in Brazil and Peru. In the
113 Colombian Amazon region, copoazu and canangucha are part of the nascent agro-
114 industry, specifically the production of candies where just a part of pulp is used as
115 raw material and the rest is a waste. Likewise, coffee is an emblem of the national
116 economy of Colombia, and it is also cultivated in its Amazonian region.

117 This study aimed to evaluate the feasibility of extracts of three amazonian fruits to
118 enhance the SPF process at near-neutral pH to simultaneously degrade four
119 representative pharmaceuticals (acetaminophen, diclofenac, sulfamethoxazole, and
120 carbamazepine) in municipal wastewater. Under authors' knowledge, this work is the
121 first study about the use of amazonian fruits for enhancing the degradation of the
122 pharmaceuticals in MWW, which pretends to provide a starting point about the
123 utilization of amazonian agro-industrial products and wastes as iron complexing
124 agents in the photo-Fenton process. In our work is covered the following topics: i)
125 limitation of conventional SPF at near-neutral pH to eliminate the target pollutants,
126 ii) performance of the SPF system in the presence of extracts of the amazonian
127 fruits, iii) influence of the iron: extract ratio on the pollutants degradation and iron
128 availability, and iv) enhancement of the pharmaceuticals degradation in raw MWW
129 by the best extract.

130

131 **2. Materials and methods**

132 *2.1. Reagents*

133 Iron (III) sulfate heptahydrate and hydrogen peroxide were purchased from
134 Panreac®. Sulfuric acid, acetonitrile (HPLC grade), sodium bisulfite, and sodium
135 hydroxide were supplied by Merck®. All solutions were prepared using deionized
136 water or municipal wastewater from Florencia-Colombia (which characterization is
137 provided in Table 1).

138 *2.2. Analytical measurements*

139 The concentration of the four pharmaceuticals was analyzed using a Shimadzu
140 HPLC with a UV detector at 254 and 267 nm with a C18 column (5 µm, 4.6 x 150
141 mm), with a flow of 0.5 mL min⁻¹. The injection volume was 100 µL. The method
142 consisted of a gradient flow with phase A: 25 mmol L⁻¹ formic acid and phase B:
143 acetonitrile, as follows, at 0 min 90%A, 3 min 90%A, and 13 min 20%A, 13.1 min
144 0%A, 20 min 0%A, 20.1 min 90%A, and 25 min 90%A.

145 The removal percentage of each pharmaceutical was calculated using Eq. 6

$$146 \quad x_i = (C_{\text{initial}} - C_{\text{final}} / C_{\text{initial}}) * 100 \quad (6)$$

147 Where C_{initial} is the initial concentration and C_{final} means the final concentration of the
148 pharmaceutical (in µmol L⁻¹).

149 The pondered removal percentage (X) was determined by Eq. 7.

$$150 \quad X = \frac{\sum_{i=1}^n C_{\text{initial},i} * x_i}{\sum_{i=1}^n C_{\text{initial},i}} \quad (7)$$

151 Where the subscript i refers to each pharmaceutical (i.e., 1 ≤ i ≤ n; n = 4) and x_i
152 represents the removal percentage of each pharmaceutical.

153 The organic matter as total organic carbon (TOC) was monitored by the combustion
154 catalytic oxidation method using a Shimadzu TOC-L instrument. Dissolved iron was
155 determined by complexation with 1,10-phenanthroline with filtered and unfiltered
156 samples according to Standard Methods for the Examination of Water and
157 Wastewater (3500-Fe B) (APHA et al. 2012). The hydrogen peroxide concentration
158 was followed by the meta-vanadate method based on the reaction of H₂O₂ with
159 ammonium metavanadate in an acidic medium, which results in the formation of a
160 red-orange color from peroxovanadium cation, with maximum absorbance at 450 nm
161 (Nogueira et al. 2005).

162 The concentration of total polyphenols in fruit extracts was quantified using the Folin-
163 Ciocalteu (Horszwald and Andlauer 2011) with some modifications. The result is
164 given as mg of gallic acid equivalent per L (mg GAE L⁻¹) (Heeger et al. 2017).
165 Municipal wastewater (MWW) from Florencia-Caquetá Colombia was used. Physical
166 and chemical characterization of MWW was determined by analytical standardized
167 methods (Table 1) (APHA et al. 2012).

168

169 **Table 1.** Physicochemical characteristics of the MWW from Florencia-Caquetá
170 Colombia

Parameter	Method	Minimum	Maximum
		Value	Value
Nitrates (NO ₃ ⁻) (mg L ⁻¹)	HCl-UV 4500 –NO ₃ ⁻	4.62	14.38

Ammonia Nitrogen (NH_4^+), (mg L^{-1})	4500-NH ₃	17.73	43.34
Total Nitrogen (TN), (mg L^{-1})	Persulfate oxidation (4500 -N)	124	210.3
Orto-phosphates (PO_4^{3-}), (mg L^{-1})	Ascorbic acid (4500-P)	1.46	3.21
Sulfates (SO_4^{2-}), (mg L^{-1})	Titrimetric (4500-SO ₄ ⁻²)	45.60	188.3
Chlorides (Cl^-), (mg L^{-1})	Argentometric (4500-Cl ⁻ -B)	62.05	103.34
Alkalinity (mg L^{-1})	Titrimetric (2320-B)	53.49	146.76
Hardness, (mg L^{-1})	Titrimetric (2340-C)	57.1	119
pH	Potentiometric (4500-pH)	6.45	6.75
Total organic carbon (TOC, mg C L^{-1})	Combustion-IR (5310-B)	28	89.32
Chemical oxygen demand (COD, $\text{mg O}_2 \text{ L}^{-1}$)	Closed reflux (5220 C)	256	580
Biochemical oxygen demand (BOD_5 , $\text{mg O}_2 \text{ L}^{-1}$)	Respirometry-Method (5210 B)	126	302

Total suspended solids (TSS, mg L ⁻¹)	Gravimetric (2540 D)	54.12	147.67
Total solids (TS, mg L ⁻¹)	Gravimetric (2540 B)	128.56	258.78

171 *All methods were taken from Standard Methods for examination of water and
 172 wastewater (APHA et al. 2012).

173
 174 *2.3. Reaction system for the solar photo-Fenton process*

175 SPF experiments were carried out at near-neutral pH (6.2) using 0.0066 mM of each
 176 pharmaceutical (SMX, DCF, CBZ, and ACT) spiked simultaneously in 1 L of
 177 deionized water or actual raw municipal wastewater (MWW) in a 1L beaker. 5 mg L⁻¹
 178 of Fe (III) and 68 to 120 mg L⁻¹ of H₂O₂ were added to the reactor. The degradation
 179 was done with a solar simulator (Suntest CPS+, Atlas) equipped with an air-cooled
 180 xenon lamp with an illumination surface of 560 cm² and 250 - 750 W m⁻² of solar
 181 intensity (global irradiance).

182

183 *2.4. Extracts characterization*

184 Three amazonian fruits (canangucha “*Maurita Flexuosa*”, coffee “*Coffea arábica*”
 185 and copoazu “*Theobroma grandiflorum*”) were considered. Copoazu fruits were
 186 collected in a medium green state, and 10 mL of the liquid extract was obtained by
 187 squeezing 5 g of the pulp from inside the fruits. The ripe fruit of canangucha was
 188 taken from a palm, 30 g of the pulp fruit were squeezed, obtaining 8 mL of the liquid
 189 extract. For the coffee extract, the ripe fruits of coffee trees were taken, the seed

190 was extracted and the external shell (30 g) was used to obtain its aqueous extract
191 by maceration using 20 mL of distilled water. The content of total phenols (by Folin-
192 Ciocalteu methodology) and total organic carbon in the extracts were assessed
193 (Table 2). Extracts of the amazonian fruit were added to the aqueous samples to be
194 treated by the SPF systems considering the molar ratio between the iron and total
195 polyphenols (i.e., Fe: total polyphenols in the extract).

196

197 **3. Results and discussion**

198 **3.1. Treatment of the representative pharmaceuticals by solar photo-Fenton**
199 The capability of the SPF process to degrade a mix of the four representative
200 pharmaceuticals in deionized water (at pH 6.2) was initially tested, the results are
201 presented in Fig. 1A. It was found that the SPF process exhibited a low degrading
202 action on three of the considered pharmaceuticals. After 90 min of treatment, it was
203 obtained removals of 18% for both CBZ and ACT, 35% for SMX, and 68% for DCF.
204 In addition to the pollutants elimination, the evolution of iron and the consumption of
205 hydrogen peroxide were also followed during the SPF process (Fig. 1A). The
206 analysis of dissolved iron in the SPF system indicated that iron is not available in
207 solution, at 20 min of treatment there was no soluble iron, and at this time only ~5
208 mg L⁻¹ of H₂O₂ were consumed. These results suggest that the production of
209 hydroxyl radicals by the reaction between the iron and the hydrogen peroxide (Eq.
210 1) is limited due to the iron is not in soluble forms (Pignatello et al. 2006; Serna-
211 Galvis et al. 2020). Consequently, the elimination of pollutants is low.

212 To understand the observed degradation by the process, some sub-
213 systems/components were also tested. It was considered the direct action of the
214 solar irradiation on the pharmaceuticals (i.e., photolysis) and the combination of light
215 with hydrogen peroxide (i.e., H₂O₂/light). Fig. 1B shows the results for such control
216 experiments. The sunlight alone induced a high degradation of DFC (~ 75% of
217 elimination at 90 min of irradiation), whereas SMX and CBZ experienced moderate
218 photolysis (removals between 10 and 15% after 90 min of treatment). In contrast,
219 ACT showed no elimination by the sole action of the solar light.

220 The differences among the photolysis of the four pollutants can be understood
221 considering the UV-Vis spectra of them (see Fig. S1 in the Supplementary material).
222 From the spectra can be noted that DCF, SMX, and CBZ have light absorption above
223 300 nm, while ACE only absorbs wavelengths below 300 nm. In turn, the used solar
224 simulator emits light from 300 to 800 nm. Thus, there are intersections of the
225 absorption and emission spectra more significant for DCF, SMX, and CBZ than for
226 ACT, which can explain the behavior observed in Fig. 1B.

227 According to the literature, photolysis of DCF promoted by solar light yields
228 carbazole structures from the elimination of a chlorine-substituent with the
229 subsequent cyclization (Salaeh et al. 2016; Leydy Katherine Ardila et al. 2019).
230 Meanwhile, SMX begins its photolytic degradation with the breaking of the
231 sulfonamide bond and rearranging the isoxazole ring (Alharbi et al. 2017; Martínez-
232 costa et al. 2018). In the case of CBZ, the solar light can induce transformations,
233 such as hydration of central double bond, ring expansions, and complementary
234 hydroxylation to yield acridone and acridine (Yazdanbakhsh et al. 2019).

235 On the other hand, it can be noted that for CBZ and ACT, which were more resistant
236 to the photolysis, their degradations are improved by the hydrogen peroxide addition
237 (i.e., the H₂O₂/light subsystem, Fig. 1B). H₂O₂ enhances the pharmaceuticals
238 elimination probably due to its direct oxidant effect (Eq. 8, (Alharbi et al. 2017)) and
239 the formation of some radicals by its homolysis by the UVB component of the solar
240 light (Eq. 9, (Mohapatra et al. 2014; Alharbi et al. 2017)).

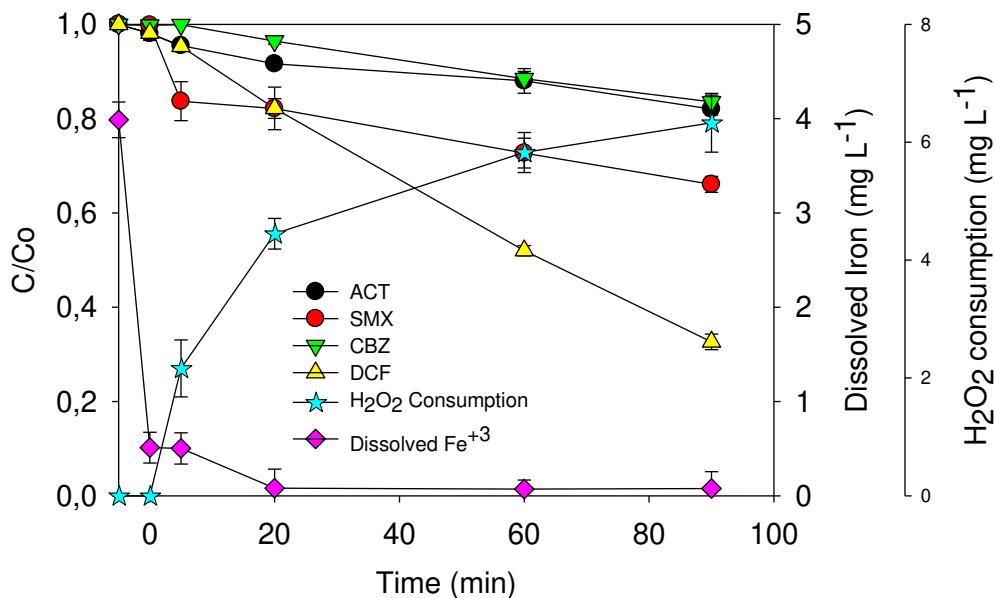


243 The results from the subsystems (control experiments) suggest that, under the
244 tested experimental conditions, the main routes of the SPF action on the
245 pharmaceuticals are photolysis and attack of some radicals generated by both the
246 Fenton reaction and the hydrogen peroxide homolysis by the solar light.
247 Consequently, under work conditions (pH 6.2), iron precipitation was probably
248 responsible for the poor pharmaceutical degradation by the SPF system (Eq. 3,
249 (Pignatello et al. 2006; Serna-Galvis et al. 2020)). In fact, the pondered degradation
250 of pharmaceuticals was ~ 35%. Particles as insoluble iron can avoid light penetration
251 (Christensen and Linden 2003; Mamane et al. 2006), then, this decrease the
252 efficiency of direct (photolysis) and indirect (H₂O₂/light) irradiation action on the
253 pharmaceuticals. Therefore, to obtain a higher removal of the target pollutants,
254 strategies to maintain the iron as soluble species and increase the hydroxyl radical
255 production should be applied. Such topic is developed in the next section.

256

257

A

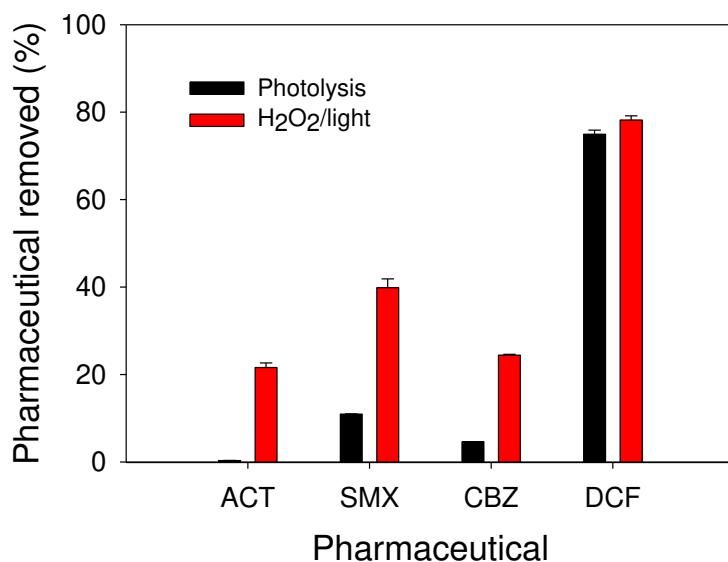


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B



261

262 **Fig. 1. A.** Treatment of pharmaceuticals mixture in deionized water by SPF at
263 near-neutral pH. **B.** removal of pharmaceuticals by control systems. Experimental

264 conditions $[Fe^{3+}]_0 = 5 \text{ mg L}^{-1}$ ($90 \mu\text{mol L}^{-1}$). $[\text{Pharm}]_0 = 0.0066 \text{ mM}$. $[\text{H}_2\text{O}_2]_0 = 120$
265 mg L^{-1} , and $\text{pH}_{\text{initial}} = 6.2$.

266

267 **3.2. Performance of SPF in presence of COFE, CANE, and COPE**

268 The effect of the addition of extracts of three amazonian fruits to the SPF system (at
269 pH around 6) on both soluble iron keeping and degradation of ACT, CBZ, DCF, and
270 SMX was evaluated. The considered extracts were CANE (canangucha extract),
271 COFE (coffee extract), and COPE (copoazu extract). Fig. 2 shows the treatment of
272 the mix of the target pharmaceuticals in the presence of COFE, CANE, or COPE.

273 Remarkably, in the presence of all extracts (Fig. 2A-C), the amount of dissolved iron
274 was higher than in their absence (Fig. 1). Furthermore, the addition of COPE to the
275 photo-Fenton system enhanced significantly the degradation of the four
276 pharmaceuticals (Fig. 2D). From Fig. 2D, it can be also noted that despite COFE
277 and CANE keep iron in soluble forms, these two extracts had a detrimental effect on
278 the process performance for the pollutants elimination (with an exception for ACT in
279 CANE presence).

280 Table 2 contains relevant characteristics of the tested extracts, which provide
281 information to interpret the different results for CANE, COFE, and COPE. It must be
282 mentioned that the COFE addition is able to provide a coloration more intense to the
283 water (Fig. S2), this is probably due to the content of tannins and chlorogenic acid,
284 (Genaro-Mattos et al. 2015)), which can affect the light penetration, avoiding the
285 photolysis of pharmaceuticals and photo-regeneration of ferrous ions from iron-
286 organic-complexes (Eq. 4). Also, the COFE has the highest content of TOC (this is

287 twice that contributed by COPE and five times that of CANE, Table 2), such organic
288 matter can also compete by the radicals, inhibiting the pharmaceuticals elimination.

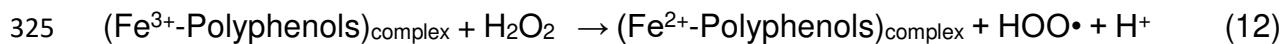
289 In the case of CANE, a particular behavior was found. Only the elimination of ACT
290 was improved concerning the conventional SPF. The results for the elimination of
291 SMX, CBZ, and DFC with CANE can be rationalized analogously (competence by
292 organic matter) to the indicated in the COFE case. Meanwhile, the selective
293 elimination of ACT requires another view. The non-removal of ACT by CANE alone
294 (Fig. S3) allowed us to discard a direct interaction between this pharmaceutical and
295 polyphenols of CANE. Then, it could be suggested that ACT is transformed by a
296 ferric-organic complex (or ferric ions under special complexing environments, Eq.
297 10), which favor redox reactions for hydroquinone-type or analogous structures as
298 presented by ACT (Issa et al. 2013; Jiang et al. 2015; Peng et al. 2017). Although,
299 such interaction allows to rationalize the elimination of ACT from the water by SPF
300 in the CANE presence. This particular result should be studied in depth in future
301 investigations.



303 On the other hand, when COPE was added, an acceleration of the pharmaceuticals
304 degradation and a significant increase in the H_2O_2 consumption was observed (Figs.
305 2C and 2D). To investigate the possible action of COPE as a photosensitizer, the
306 treatment of the pharmaceuticals in presence of the fruit extract and solar light was
307 carried out. When the sunlight is combined with COPE, after 90 min of treatment,
308 the pharmaceutical removal was the same as the obtained by photolysis with

309 sunlight alone. This suggests that COPE did not act as a photosensitizer for
310 pharmaceuticals degradation (Porras et al. 2016).

311 It can be mentioned that COPE is rich in polyphenols (Hernández Londoño 2010;
312 Carlos Carmona-Hernandez et al. 2018; Pereira et al. 2018), as demonstrated in
313 Table 2. The polyphenols complex the ferric ions easily (Eq. 11, (Ruíz-Delgado et
314 al. 2019)), maintaining the iron in soluble forms as supported by Fig. 2C. Such
315 complexes can interact with the hydrogen peroxide to produce other degrading
316 species (e.g., hydroperoxyl radical, Eq. 12, (Ruíz-Delgado et al. 2019)). Besides, it
317 is well-known that polyphenols can reduce ferric ions to ferrous ions (Fig. S4, (Oakes
318 2013)), which are required for the Fenton reaction (Eq. 1); thus, favoring the catalytic
319 cycle of iron. Indeed, in COPE presence, there was a very high H₂O₂ consumption
320 (Fig. 2C), which confirms the interactions indicated by Eqs. 1 and 12. Hence, the
321 role of COPE as a complexing agent of iron and regenerator of ferrous ions favors
322 the production of radicals in the system; then, an enhancement of the degradation
323 of target pollutants occurs.

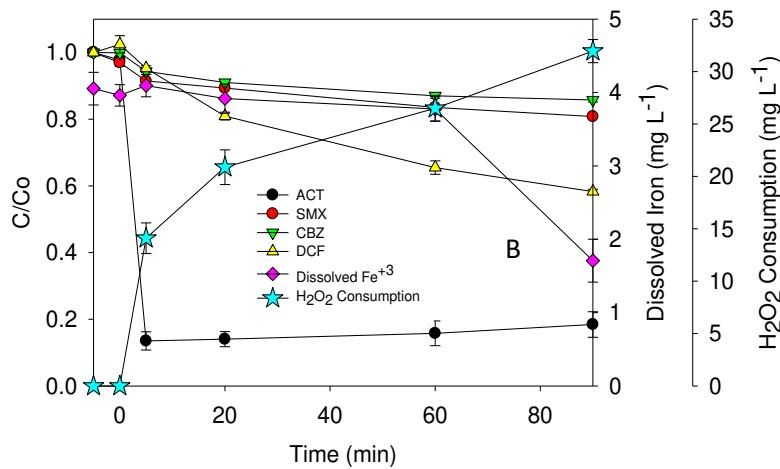
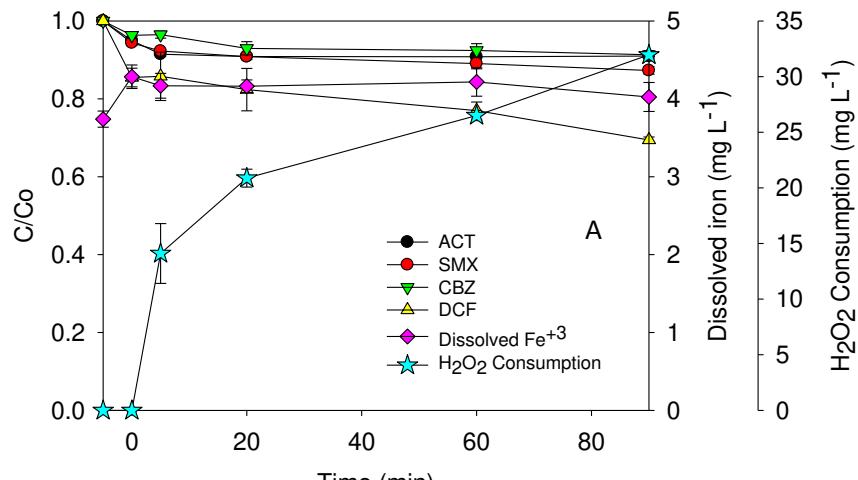


326 It can be mentioned that the dissolved iron in the presence of COPE remained lower
327 than in the cases of CANE and COFE (see Figs. 2A-C), suggesting that CANE and
328 COFE exert an iron complexation stronger than COPE. Then, the complex of iron
329 with COPE seems less stable, making the iron more available to interact with the
330 hydrogen peroxide (Eqs. 1 and 12). Additionally, as HO[·] is non-selective, this can

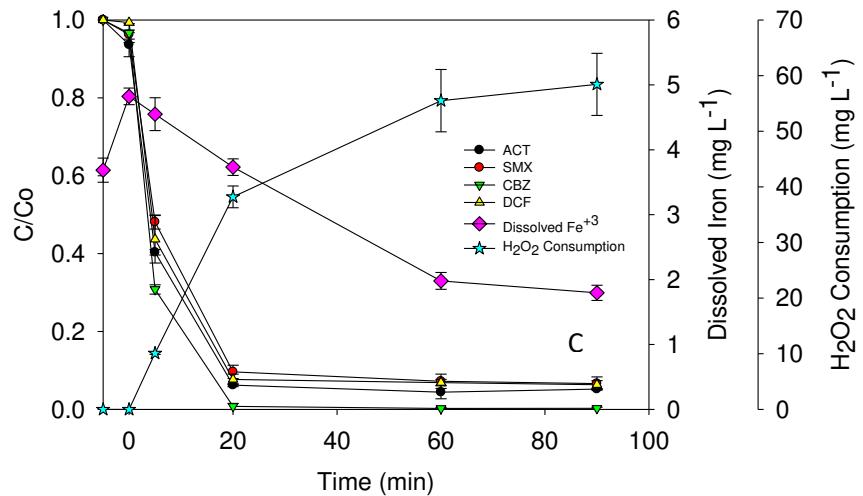
331 also react with the COPE components. As a consequence, the concentration of
332 COPE decreases, and the soluble iron may diminish.

333 In the COPE presence, up to 95% of the pondered elimination of pharmaceuticals
334 was reached, which highlights the high potential of COPE as an enhancer of the SPF
335 process at near-neutral pH. To better understand its enhancing effect, experiments
336 about variations of COPE: iron ratio were performed and their outcomes are
337 presented and discussed in the next subsection.

338

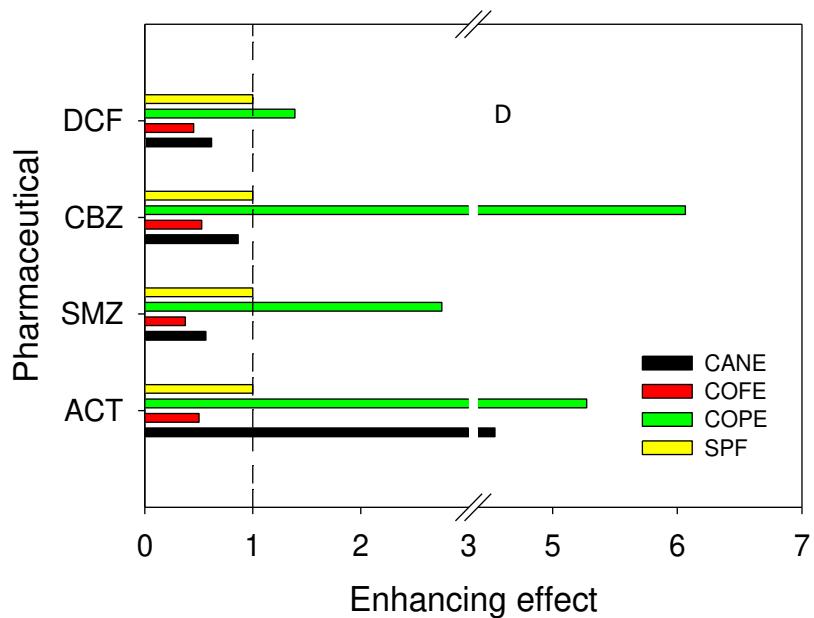


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340

341



342

343 **Fig. 2.** Treatment of the pharmaceuticals mix in deionized water by SPF at natural
 344 pH in presence of extracts. **A.** COFE. **B.** CANE. **C.** COPE. **D.** Enhancing effect of
 345 three extracts. Molar ratio Fe:extract = 1:0.16. $[Fe^{3+}]_0 = 5 \text{ mg L}^{-1}$. $[Pharm]_0 = 0.0066$
 346 mM. $[H_2O_2]_0 = 120 \text{ mg L}^{-1}$, and $pH_{initial} = 6.2$.

347

Table 2. Content of total phenols and TOC in the extracts from three amazonian fruits.

Extract	Extraction technique	Total phenols (mg GAE L ⁻¹)	TOC content in the extract
<i>Canangucha</i> <i>extract (CANE)</i>	Squeezed fruit pulp	297±10	32±3
<i>Coffee extract</i> <i>(COFE)</i>	Maceration of husk	454±32	162±11
<i>Copoazu extract</i> <i>(COPE)</i>	Squeezed fruit pulp	481±21	92±5

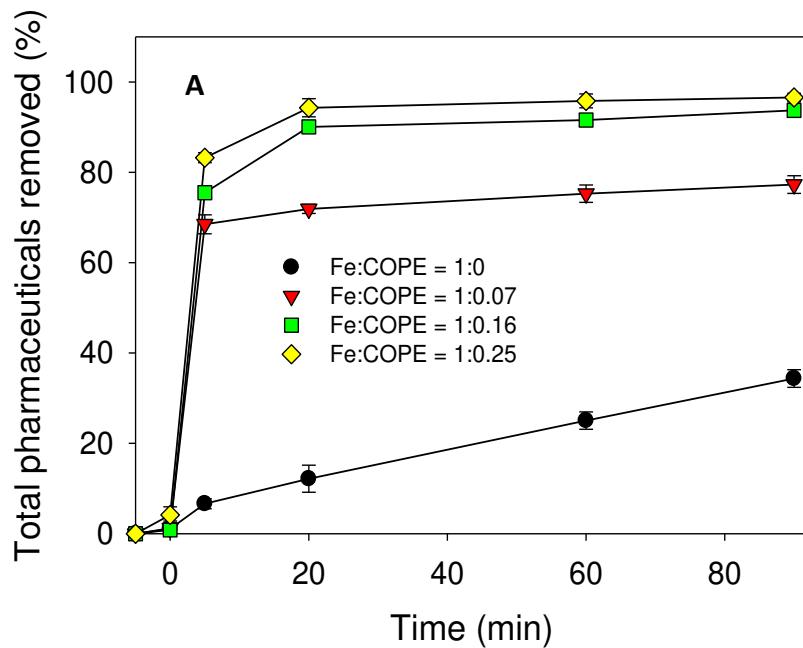
350

3.3. Influence of Fe:COPE ratio on the degradation of pollutants

After evaluation of the different amazonian fruits and considering the best performance of COPE, the pharmaceuticals mix in deionized water was treated by SPF at pH 6.2, using different amounts of the extract (Fig. 3a). It can be noted that the efficiency of SPF for the degradation of the pharmaceuticals significantly increased as the Fe:COPE ratio augmented from 1:0.07 to 1:0.16, and a slight improvement of the degradation was found for changes of the ratio from 1:0.16 to 1:0.25.

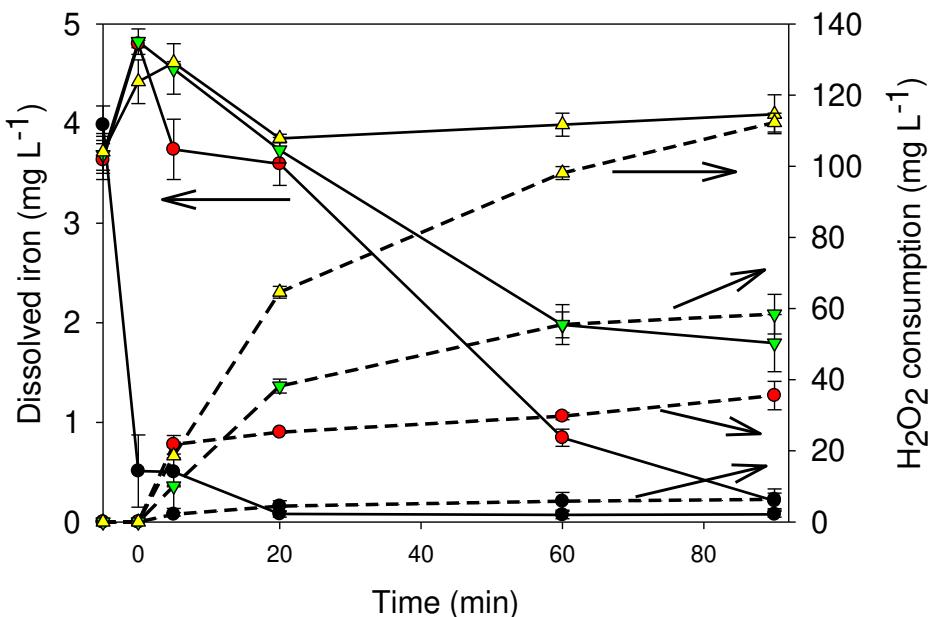
359 The positive effect on the degradation of the pollutants caused by the increment of
360 COPE concentration was related to the ability of COPE to act as a complexing agent

361 of iron. Hence, the dissolved iron and H₂O₂ consumption were measured during the
362 SPF process at the different Fe:COPE ratios (Fig. 3b). These measurements
363 showed that both the iron in solution and hydrogen peroxide consumption were
364 higher as COPE amount increased. In fact, in presence of the highest COPE
365 presence, 80% of the initial iron remained in the solution during the whole experiment
366 (on the contrary, in absence of COPE more than 95% of iron was quickly
367 precipitated).



368

369



370

371 **Fig. 3.** A. SPF at near-neutral pH applied to the elimination of pharmaceuticals in
 372 deionized water. $[Fe^{3+}]_0 = 5 \text{ mg L}^{-1}$. $[Pharm]_0 = 0.0066 \text{ mM}$. $[H_2O_2]_0 = 120 \text{ mg L}^{-1}$,
 373 and $pH_{initial} = 6.2$. B. Behavior of dissolved iron and H_2O_2 consumption during SPF
 374 at near-neutral pH (6.2) with COPE in different molar relations Fe: COPE.

375 (—): Dissolved iron and (----): H_2O_2 consumption.

376

377 As the COPE amount was higher, more quantity of complexing agents (i.e.,
 378 polyphenols) are available to keep the iron as soluble forms that can react with H_2O_2 ,
 379 thus favoring the radicals generation (Eqs. 1-5), and consequently, the elimination
 380 of the pharmaceuticals is enhanced. Nevertheless, at the Fe:COPE ratio 1:0.25,
 381 despite the iron remains in solution by interaction with the COPE (Fig. 3), there was
 382 only a little improvement of the pollutants removal concerning 1:0.16. The highest
 383 addition of the amazonian fruit extract may represent a very elevated load of organic

384 matter. Thus, the organic matter in excess could compete by the radicals ((Klammerth
385 et al. 2013; Papoutsakis et al. 2015; De la Obra et al. 2017), and therefore a
386 significant increase in the process efficiency does not occur. Indeed, it would be
387 expected that a Fe: COPE ratio higher than 1: 0.25 will induce a detrimental effect
388 on the pharmaceuticals removal by the SPF process. Therefore, the above results
389 indicated that 1:0.16 is the suitable Fe: COPE ratio for the subsequent experiments.

390

391 **3.4. Pharmaceuticals degradation in raw municipal wastewater by SPF in
392 COPE presence**

393 After demonstrating the enhancing role of COPE on the SPF to degrade pollutants
394 in deionized water, the process in presence of the copoazu extract was evaluated
395 for the treatment of raw municipal wastewater (MWW) spiked with the target
396 pharmaceuticals. The improving character of COPE in the complex matrix was also
397 proven. Fig. 4A shows that similar to the observed in deionized water (Fig. 2), in the
398 MWW, the presence of the extract led to a high pollutants degradation (~80% in 90
399 min of treatment). Again, the COPE maintains the iron in solution in the MWW matrix
400 and increases the production of the radicals (evidenced indirectly by a higher H₂O₂
401 consumption) (Fig. 4B), which favors the pharmaceuticals removal.

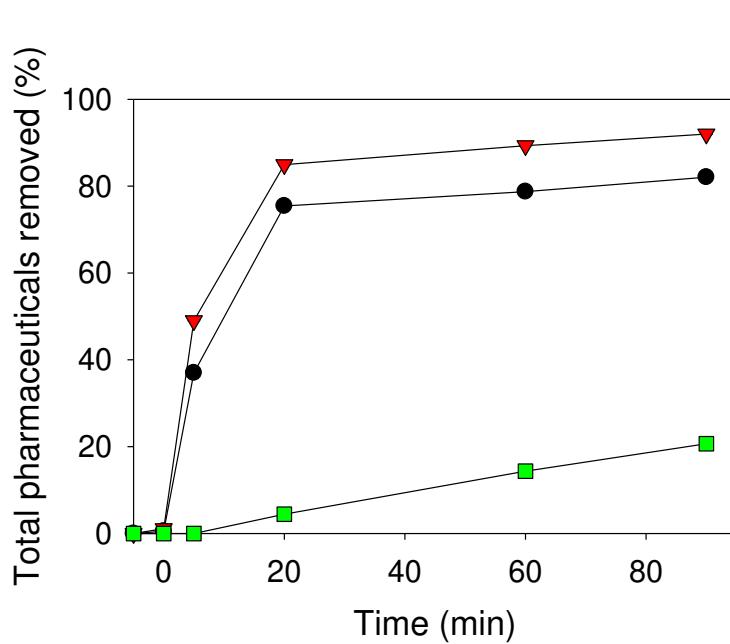
402 Naturally, the presence of intrinsic organic and inorganic components in MWW
403 competed by the radicals, decreasing a few pharmaceuticals elimination compared
404 to the degradation in deionized water (Fig. S5). Organic matter from MWW (e.g.,
405 humic substances) can behave as radical scavengers or can induce a radiation
406 attenuation, affecting the removal efficiencies (Mirzaei et al. 2017). Likewise, at the

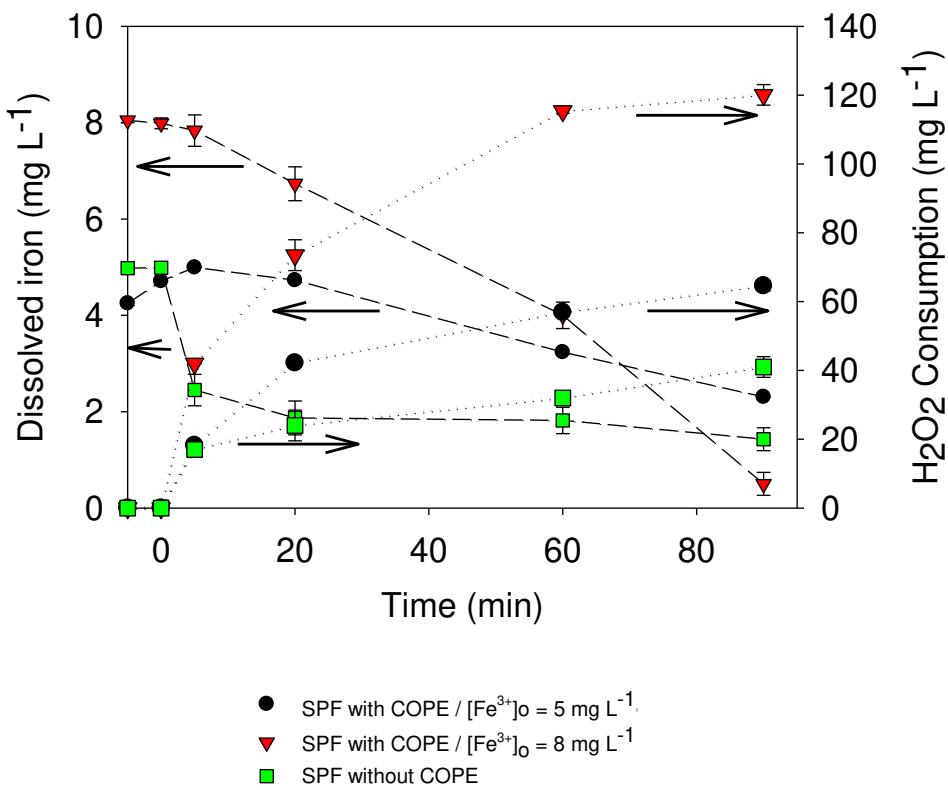
407 experimental pH (pH: 6.2), the HCO_3^- ion is the main inorganic carbon species in the
408 MWW (alkalinity of the matrix under study ranges from 53.49 to 146.76 mg L⁻¹, Table
409 1). HCO_3^- has high reactivity with $\text{HO}\cdot$ ($8.5 \times 10^6 \text{ M}^{-1} \text{ s}^{-1}$) (Klameth et al. 2012; De
410 la Obra et al. 2017) and consequently, also competes with the pollutants for the
411 available radicals. Furthermore, in MWW, the presence of chloride, sulfate, and
412 phosphate ions (see Table 1) may reduce the efficiency of the SPF process, since
413 these anions react with hydroxyl radicals and generate other species that are lesser
414 powerful and lesser reactive toward the pollutants. Besides, the presence of
415 phosphate limits the pharmaceuticals degradation (see SPF without COPE in Fig.
416 4A) and decreases the soluble iron (see SPF without COPE in Fig. 4B) by
417 precipitation as iron phosphate (Eq. 13) (Mirzaei et al. 2017).



419 In addition to the evaluation of COPE addition for treating MWW, the initial
420 concentration of iron (III) was varied from 5 to 8 mg L⁻¹ (consequently, the amount of
421 added COPE was increased, keeping the Fe:COPE ratio in the 1: 0.16 proportion).
422 Total removal of pharmaceuticals and consumption of hydrogen peroxide plus iron
423 evolution were also followed (Fig. 4). As the initial concentration of iron increased,
424 the pondered removal percentage of pharmaceuticals, as well as the peroxide
425 consumption, increased. Remarkably, high pollutants removal (>90%) was achieved
426 during 90 min of treatment, at 8 mg L⁻¹ of iron. The increasing of the initial
427 concentration of iron can promote the formation of more radicals, leading to a higher
428 degradation of the pollutants (Eqs. 1-5), as indirectly evidenced by the highest H₂O₂
429 consumption at 8 mg L⁻¹ of iron. Additionally, the presence of COPE favors the

430 catalytic cycle of iron, and its maintenance in the solution (Fig. 4B). However, the
431 SPF system using 8 mg L⁻¹ of iron in the COPE presence, can generate a high
432 amount of non-selective radicals as HO•, which can also attack the COPE,
433 decreasing the extract availability. For this reason, at a long treatment time (e.g., 90
434 min) the soluble iron diminishes significantly (Fig. 4B).





448

449 **Fig. 4.** Degradation of the four pharmaceuticals in MWW by SPF at near-neutral pH
 450 (6.2) without COPE and with COPE. (—) Total pharmaceuticals removal; (---)
 451 Dissolved iron, and (....) H₂O₂ consumption. $[Pharm]_0 = 0.0066 \text{ mM}$. Fe:COPE =
 452 1:0.16 and $[H_2O_2]_0 = 120 \text{ mg L}^{-1}$, for the experiments of SPF without COPE the
 453 concentration of Fe^{3+} was 5 mg L^{-1} .

454 It should be mentioned that the high removal of the pollutants in the tested MWW by
 455 SPF in presence of amazonian fruit extracts, evidenced the great feasibility of this
 456 treatment to eliminate pharmaceuticals in actual complex matrices. This kind of AOP
 457 could receive special attention in developing countries in the amazonian region (e.g.,
 458 Colombia, Peru, or Brazil), which have abundant solar light and agro-industrial

459 wastes coming from the processing of fruits as copoazu. Nonetheless, additional
460 tests on optimization (using a design of experiments) of operational parameters
461 (such as COPE amount, concentrations of iron and hydrogen peroxide, and radiation
462 intensity) to treat MWW can be developed in future work to obtain a deeper
463 exploration of the intensification of the SPF system by COPE.

464

465 **4. Conclusions**

466 After the development of the present research, we can conclude that:

- 467 • The application of SPF without extracts induced a low removal of the four
468 representative pharmaceuticals due to limitations in the iron availability in the
469 solution.
- 470 • The addition of COFE, CANE, and COPE increased the dissolved iron but
471 only COPE improved significantly the degradation of the target pollutants due
472 to its ability to complex iron with the polyphenols present in this extract.
- 473 • The increment of Fe: COPE ratio from 1:0 to 1:0.16 augmented significantly
474 the pollutants removal but an excess of COPE has little enhancer effect. Even
475 a detrimental effect could be found at very high amounts of COPE due to
476 competition by radicals and/or the increasing of water turbidity (light filtering).
- 477 • The treatment of MWW reached degradations > 75% after 90 min in the
478 process improved with COPE. Furthermore, the increase of iron from 5 to 8
479 mg L⁻¹ (maintaining the Fe: COPE molar ratio at 1: 0.16) augmented the
480 pharmaceuticals elimination until 90% at only 20 min of treatment.

481 • This work showed the high feasibility of amazonian fruit extracts to improve
482 the efficiency of photo-Fenton at near-neutral pH and high iron concentrations
483 to remove pharmaceuticals in aqueous media. Additionally, our research also
484 evidenced the viability of this process to be applied in developing countries,
485 valorizing agro-industrial materials (considered mainly as raw matters or just
486 as wastes) for the wastewater treatment.

487

488 ***Declarations***

489 **- Ethical approval and consent to participate**

490 Not applicable.

491 **- Consent to publish**

492 Not applicable.

493 **- Availability of data and materials**

494 The datasets used and/or analyzed during the current study are available from the
495 corresponding author on reasonable request.

496 **- Competing interests**

497 The authors declare that they have no competing interests

498 **- Funding**

499 This research was funded by Universidad de la Amazonia through the project
500 "Treatment of Florencia-Caquetá municipal wastewater by the combination of
501 biological processes and Fenton-type advanced oxidation processes".

502 **- Acknowledgments**

503 The authors from UdeA also acknowledge the support provided by MINCIENCIAS
504 COLOMBIA (before named COLCIENCIAS) through project No. 111577757323; as
505 well as Universidad de Antioquia UdeA by means "Programa de Sostenibilidad". E.
506 A. Serna-Galvis thanks MINCIENCIAS COLOMBIA for his Ph.D. fellowship during
507 July 2015-June 2019 (Convocation 647-2014), and the Postdoctoral scholarship
508 (Convocation 848-2019).

509 **- Authors' contributions**

510 **Lis Manrique-Losada:** Investigation, Methodology; Formal analysis, Writing-original
511 draft, Resources, Funding acquisition. **Heidy L. Santanilla-Calderón:** Investigation,
512 Methodology; Writing-original draft. **Efraím A. Serna-Galvis:** Conceptualization,
513 Formal analysis, Writing-review & editing. **Ricardo A. Torres-Palma:**
514 Conceptualization, Writing-review & editing, Resources, Funding acquisition.

515

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Figures

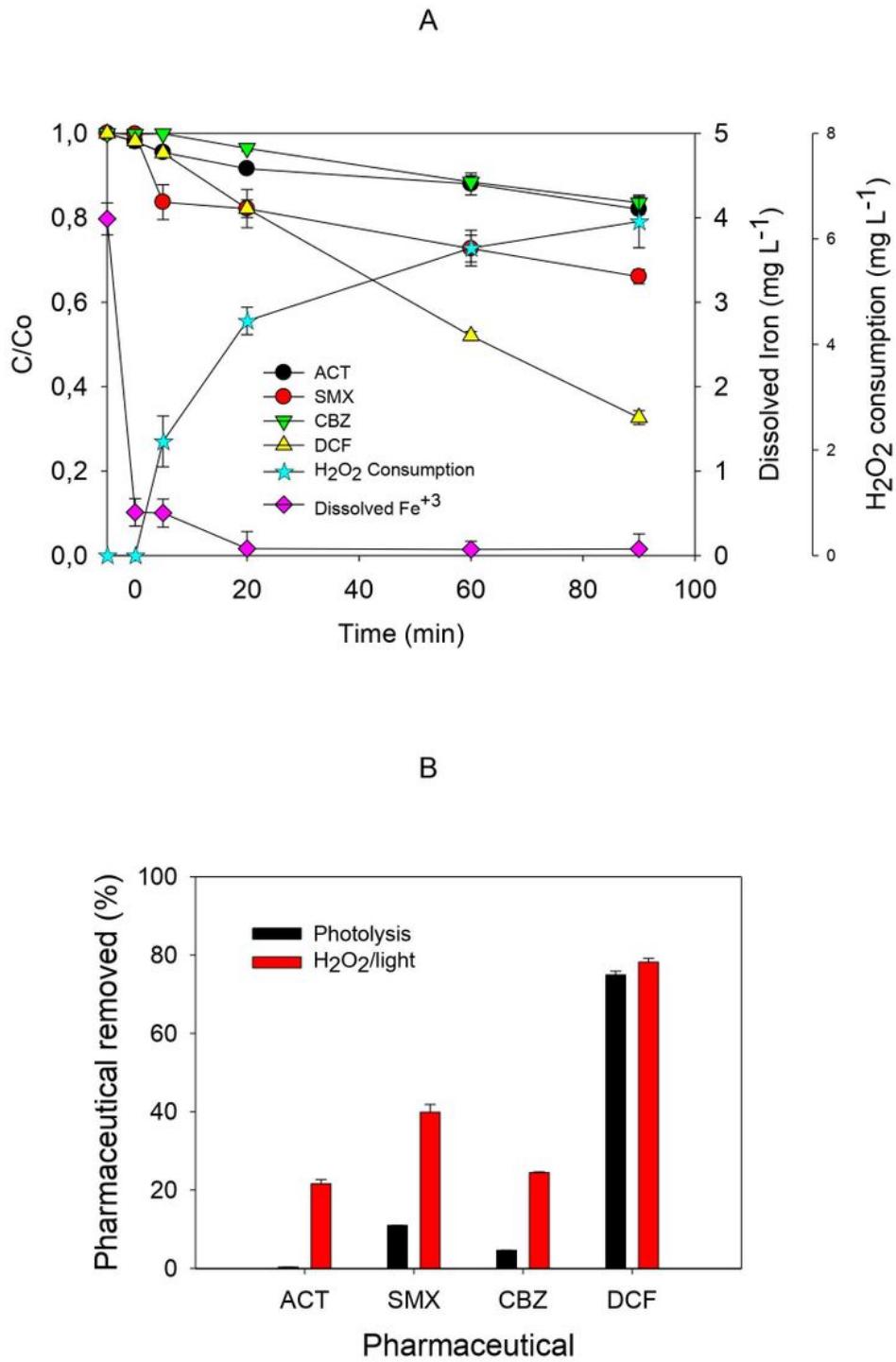


Figure 1

Treatment of pharmaceuticals mixture in deionized water by SPF at near-neutral pH. B. removal of pharmaceuticals by control systems. Experimental conditions $[Fe^{3+}]_0 = 5 \text{ mg L}^{-1}$ ($90 \mu\text{mol L}^{-1}$). $[Pharm]_0 = 0.0066 \text{ mM}$. $[H_2O_2]_0 = 120 \text{ mg L}^{-1}$, and $pH_{initial} = 6.2$.

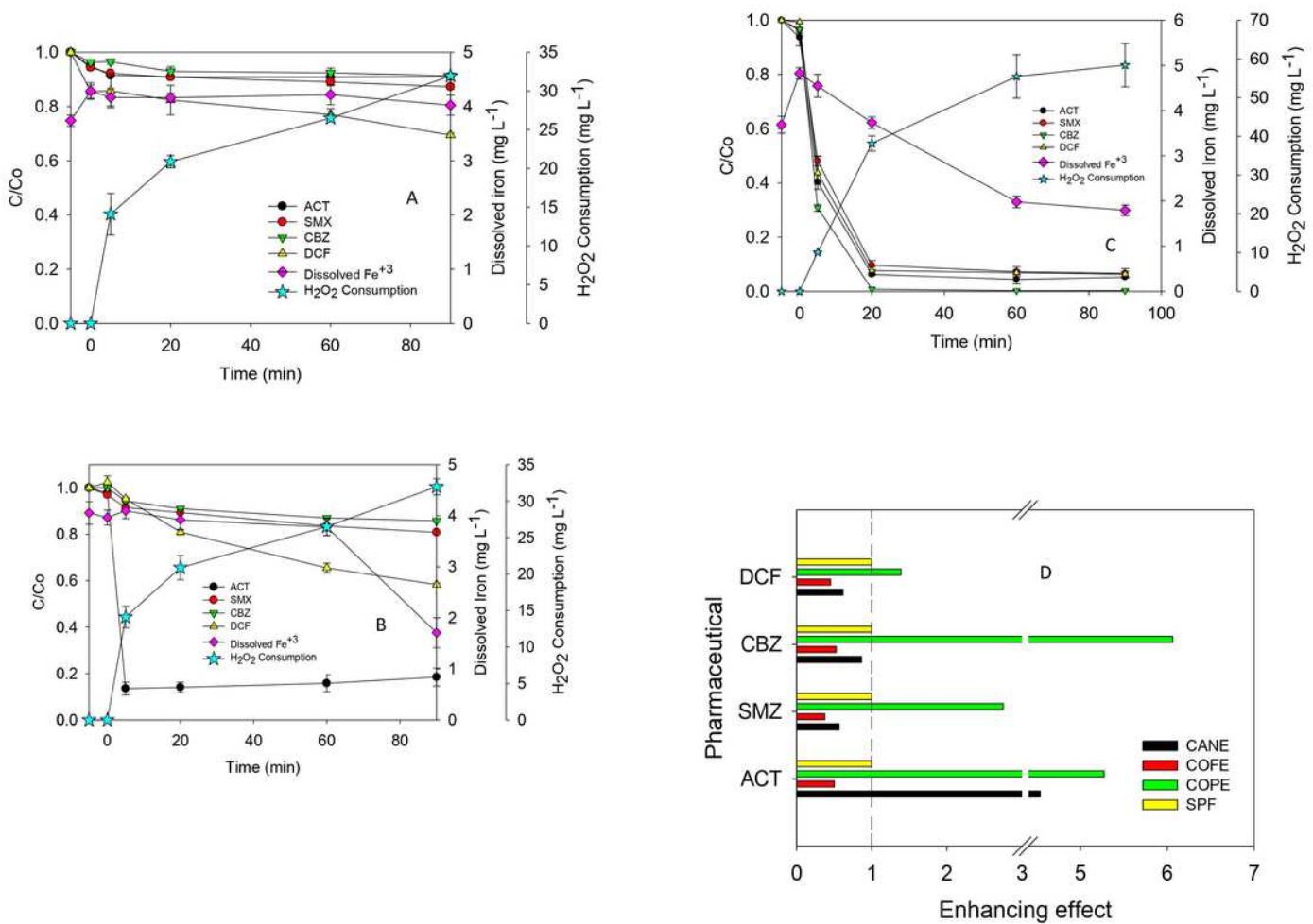


Figure 2

Treatment of the pharmaceuticals mix in deionized water by SPF at natural pH in presence of extracts. A. COFE. B. CANE. C. COPE. D. Enhancing effect of three extracts. Molar ratio Fe:extract = 1:0.16. $[\text{Fe}^{3+}]_0 = 5 \text{ mg L}^{-1}$. $[\text{Pharm}]_0 = 0.0066 \text{ mM}$. $[\text{H}_2\text{O}_2]_0 = 120 \text{ mg L}^{-1}$, and $\text{pH}_{\text{initial}} = 6.2$.

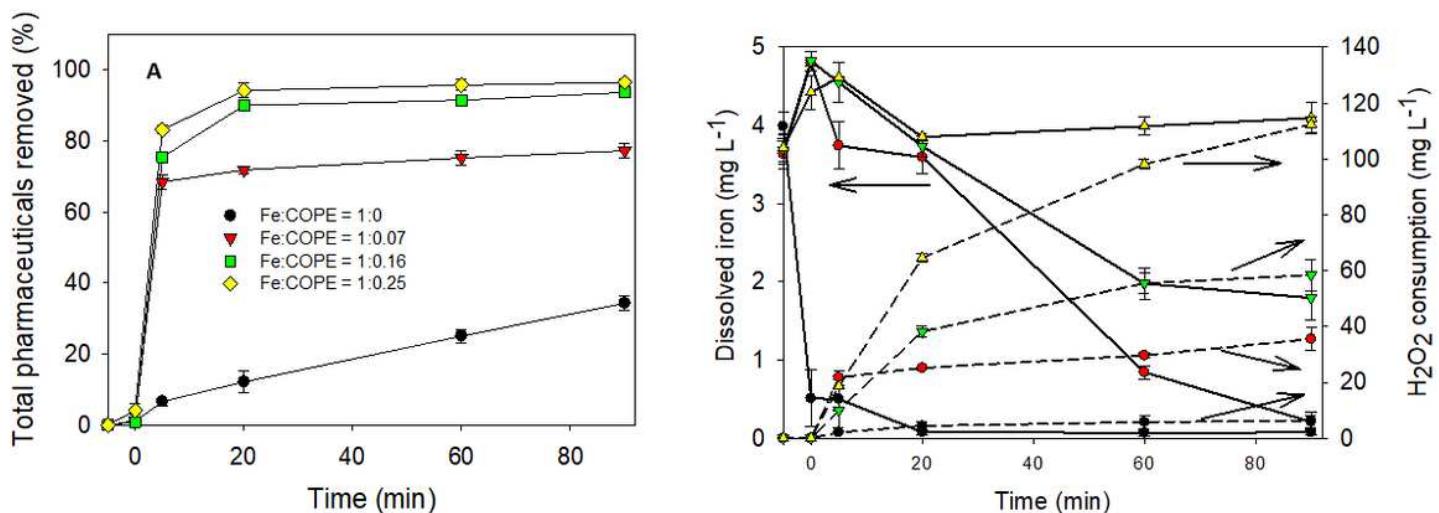


Figure 3

A. SPF at near-neutral pH applied to the elimination of pharmaceuticals in deionized water. $[Fe^{3+}]_0 = 5 \text{ mg L}^{-1}$. $[Pharm]_0 = 0.0066 \text{ mM}$. $[H_2O_2]_0 = 120 \text{ mg L}^{-1}$, and $pH_{initial} = 6.2$. B. Behavior of dissolved iron and H_2O_2 consumption during SPF at near-neutral pH (6.2) with COPE in different molar relations Fe:COPE. (●): Dissolved iron and (—): H_2O_2 consumption.

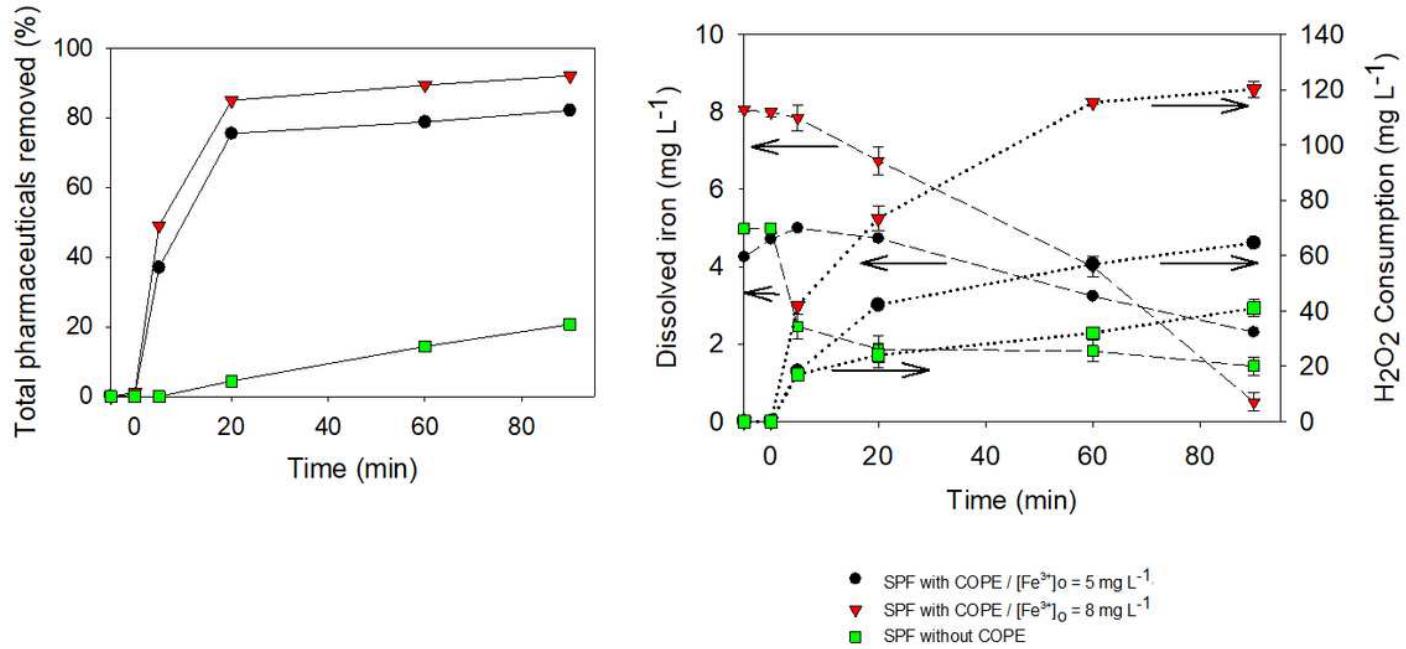


Figure 4

Degradation of the four pharmaceuticals in MWW by SPF at near-neutral pH (6.2) without COPE and with COPE. (●) Total pharmaceuticals removal; (—) Dissolved iron, and (....) H_2O_2 consumption. $[Pharm]_0 = 0.0066 \text{ mM}$. Fe:COPE = 1:0.16 and $[H_2O_2]_0 = 120 \text{ mg L}^{-1}$, for the experiments of SPF without COPE the concentration of Fe^{3+} was 5 mg L^{-1} .

Supplementary Files

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