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Elevated CO₂ levels promote both carbon and nitrogen cycling in global forests

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Forests provide vital ecosystem services, particularly as carbon sinks for nature-based climate solutions. However, the global impact of elevated atmospheric carbon dioxide (CO₂) levels on carbon and nitrogen interactions of forests remains poorly quantified. We integrate elevated CO₂ experimental observations and biogeochemical cycle model to elucidate the synergies between enhanced nitrogen and carbon cycling in global forests under elevated CO₂. Elevated CO₂ levels alone increase net primary productivity by 26% (95% CI, 21-30%) and leaf C:N ratio by 32% (18-46%), while stimulating biological nitrogen fixation by 72% (27-136%) and nitrogen use efficiency by 22% (8-38%). Under the elevated CO₂ middle road scenario for 2050, forest carbon sink is projected to increase by 0.32 billion tonnes (PgC), with forest products increasing by 4 million tonnes (Tg) nitrogen, reactive nitrogen loss to the environment decreasing by 8 Tg, and fertilizer input decreasing by 4 Tg nitrogen relative to the baseline scenario. The monetary impact assessment of the direct elevated CO₂ impact on forests represents a social value of US\$292 billion. These findings should inform the development of forest management strategies for future climate change adaptation and mitigation.

Forests cover approximately 31% of the Earth's land area and serve as habitats for a diverse range of wildlife¹. They play a crucial role as natural assets that support the livelihoods of 1.6 billion people, particularly those vulnerable segments of society residing in or near forested regions². Forests provide essential ecosystem services to humanity, including forest production, water and soil conservation, as well as carbon capture and storage^{3,4}. Specifically, forest ecosystems have the potential to act as carbon sinks, contributing to nature-based decarbonization solutions for combating climate change, and helping offset anthropogenic carbon emissions from agriculture and industrial sectors to achieve Net-Zero emissions^{5,6}.

The unprecedentedly high levels of CO₂ in the atmosphere have driven anthropogenic climate change while also influencing the biophysiological and biogeochemical processes of forest ecosystems, such as stimulating plant growth and productivity through the CO₂ fertilization effect⁷. The terrestrial carbon sink has more than doubled in the past five decades largely attributing to the CO₂ fertilization^{8,9}. The altered carbon stock capacity highly depends on nitrogen availability, and thus, the responses of the nitrogen cycle in the context of climate

50 change might determine whether forests act as carbon sinks or sources^{10,11}. The alteration of
51 carbon and nitrogen interactions in forests under elevated CO₂ is subject to debate. Some
52 forests could be nitrogen-limited ecosystems, leading to progressive nitrogen limitation under
53 CO₂ enrichment^{12,13}. Whereas, a recent long-term field study indicates that nitrogen limitation
54 may not occur due to increased litterfall turnover and nitrogen resorption, which sustain the
55 CO₂ fertilization effect in an alpine forest.¹⁴

56
57 Globally, nearly one-third of forests are managed primarily for the production of wood and
58 non-wood products¹. In recent years, increasing nitrogen deposition, combined with human
59 application of synthetic fertilizer, has led to higher nitrogen inputs and associated reactive
60 nitrogen (N_r) loss in some forests¹⁵. However, the specific impact of elevated CO₂ as a key
61 driver of climate change on forest nitrogen cycling and N_r loss, is still not well understood and
62 quantified in global forests. The representation of the nitrogen cycle and nitrogen loss (N_r loss)
63 in current Earth System Models has been insufficient, particularly in relation to accounting for
64 the responses of the nitrogen cycle to climate change¹⁶. It is essential to incorporate the
65 feedback of carbon and nitrogen cycles, along with their interactive processes, into forest
66 management policy-making, for both adapting to and mitigating the impacts of future climate
67 change¹⁷.

68
69 In this study, we aim to identify the impacts of elevated CO₂ on carbon and nitrogen cycles
70 using a global dataset of elevated CO₂ experiments conducted in forests. Subsequently, we
71 propose a modelling framework by integrating the impact of elevated CO₂ experiments on
72 carbon and nitrogen cycles with the global forest carbon and nitrogen budgets simulated by the
73 Dynamic Land Ecosystem Model (DLEM)¹⁸ and Coupled Human and Natural Systems
74 (CHANS) model¹⁹. The integration allows us to project the spatial-temporal variations in forest
75 carbon and nitrogen budgets in response to elevated CO₂ under multiple future scenarios.
76 Finally, we conduct a monetary impact assessment of the elevated CO₂ on the carbon and
77 nitrogen cycles in global forests, evaluating its economic implications for human society.

78 79 **Impact of elevated CO₂ on forest carbon and nitrogen cycling**

80 The impact of elevated CO₂ on forest carbon and nitrogen cycles was investigated using a
81 global forest dataset of elevated CO₂ experiments. The elevated CO₂ experiments, including
82 Free-Air CO₂ Enrichment (FACE), Open-Top Chambers (OTC), and Greenhouse Chambers
83 (GC), have been conducted at various forest sites across North America, Central America,
84 Europe, Asia, and Oceania (Fig. 1a). A total of 1,059 response ratios of site-based observations
85 were analyzed to form this global dataset. Globally, elevated CO₂ profoundly enhances the
86 carbon cycle, as reflected in promoted plant productivity, plant biomass, soil respiration, and
87 carbon content. Plant net primary productivity (NPP) shows an overall increase of 26% (95%
88 CI: 21-30%, hereinafter) under elevated CO₂ (Fig. 1b), with the response sensitivity decreasing
89 as the squared mean annual precipitation (MAP) increases (Fig. S2a). This might be attributable
90 to the reduced stomatal conductance and transpiration under elevated CO₂, resulting in higher
91 water use efficiency (+114%, 73-149%) (Fig. S3a) that could ameliorate drought stress and
92 further promote photosynthesis, especially in arid or semi-arid areas²⁰. The biomass of different
93 plant components also exhibits distinct increases due to elevated CO₂, including leaf biomass
94 (+24%, 16-33%), stem biomass (+24%, 16-33%), and root biomass (+46%, 38-55%) (Fig. 1b).
95 Simultaneously, soil respiration (R_s, soil CO₂ emissions from plant roots and microbes)
96 increases by 28% (23-33%). Furthermore, elevated CO₂ stimulates soil organic carbon (SOC)
97 (+5%, 1-8%), dissolved organic carbon (DOC) (+16%, 3-34%), and soil microbial biomass
98 carbon (MBC) (+19%, 12-26%).

99

100 Meanwhile, elevated CO₂ levels induce a 72% increase in rates of biological nitrogen fixation
101 (BNF) (27-136%), and a 24% increase in denitrification rates (4-53%) (Fig. 1b), suggesting a
102 higher microbial capability to transform inert N₂ into plant-available nitrogen and to reduce
103 nitrate to N₂ under CO₂ enrichment. These increases likely result from the stimulated activities
104 of nitrogen-cycling relevant microorganisms, induced by the greater availability of carbon²¹.
105 The improved nitrogen use efficiency (NUE) (+22%, 8-38%) under CO₂ enrichment is
106 associated with reduced loss of N_r, as nitric oxide (NO_x) emissions decrease by 28% (4% to
107 46%), and leaching and runoff nitrate (NO₃⁻) decrease by 39% (9% to 60%). Moreover,
108 elevated CO₂ leads to decreases in nitrogen concentration of vegetation organisms, such as
109 leaves by 13% (10% to 15%), stems by 7% (2% to 13%), and roots by 8% (1% to 15%) (Fig.
110 1b). Generally, the accelerated nitrogen cycle, including higher nitrogen input and nitrogen
111 transformation would sustain the CO₂ fertilization effect on plant productivity.

112
113 Overall, our findings indicate a synergistic enhancement in both the carbon and nitrogen cycles
114 in global forests under elevated CO₂ levels, accompanied by shifts in carbon-to-nitrogen (C:N)
115 stoichiometry. The C:N ratios increase in leaves by 32% (18-46%) and in soil by 5% (1-9%)
116 due to elevated CO₂ (Fig. 1b). Elevated carbon inputs facilitate nitrogen cycling, while
117 accelerated nitrogen cycling and alleviation of nitrogen limitation, in turn, benefits carbon
118 cycling.

119 **Global variations of carbon and nitrogen budgets under elevated CO₂**

120 We utilize the DLEM¹⁸ and CHANS¹⁹ models to deliver a plausible global gridded model of
121 forest carbon and nitrogen budgets (Fig. S1). By incorporating the impacts of elevated CO₂
122 experiments on carbon and nitrogen cycles into the parameterization optimization of model
123 simulation, we project the carbon sinks and nitrogen budgets in global forests under multiple
124 future scenarios. Different levels of socioeconomic development and climate change are
125 hypothesized based on the Shared Socioeconomic Pathways (SSPs) and Representative
126 Concentration Pathways (RCPs). The future atmospheric CO₂ concentrations are derived from
127 CMIP6 models, leading to the formulation of the eCO₂ SSP1-2.6 (SSP1-RCP2.6, “Sustainable
128 society” under elevated CO₂ levels) and eCO₂ SSP2-4.5 (SSP2-RCP4.5, “Middle road” under
129 elevated CO₂ levels), along with the baseline scenarios (SSP1, SPP2, no-climate-change under
130 fixed CO₂ levels)²². Our results indicate that, in the eCO₂ SSP2-4.5 scenario, by the year 2050,
131 forest carbon sink (net biome productivity) is projected to increase by 0.32 billion tonnes (Pg
132 C yr⁻¹), while the total nitrogen inputs are projected to increase by 13 million tonnes (Tg N yr⁻¹)
133 (Fig. 2). The increased nitrogen input deriving from promoted BNF under elevated CO₂ could
134 help sustain the nitrogen demand of the enhanced carbon sink in forests. Additionally, nitrogen
135 in global forest products is expected to increase by 4 Tg N yr⁻¹, accumulation nitrogen in
136 biomass and soil is estimated to increase by 19 Tg N yr⁻¹, N_r losses are projected to decrease
137 by 8 Tg N yr⁻¹, and NUE is projected to increase from 65% to 79% in global forests.

138
139
140 Notably, the global forest carbon sink is projected to rise from 1.05 ± 0.32 Pg C yr⁻¹ in the
141 baseline SSP2 scenario to 1.37 ± 0.41 Pg C yr⁻¹ in the eCO₂ middle road (SSP2-4.5) scenario
142 by the year 2050. This enhanced carbon sink indicates a greater potential for future carbon
143 sequestration and decarbonization capabilities within forest ecosystems, particularly under
144 higher atmospheric CO₂ levels compared to current levels. In specific geographical contexts,
145 substantial enhancements in carbon sinks are foreseen in pivotal zones like the tropical forests
146 of the Amazon, the Congo Basin, Southeast Asia, and certain regions of northern Australia (Fig.
147 2a-c), renowned for their designation as land carbon sink hotspots^{23,24}.

148
149 Under eCO₂ middle road scenario, the projected increase of 13 Tg N yr⁻¹ in total nitrogen input
150 changes is the sum of the altered various input sources, including BNF (+19 Tg N yr⁻¹),

151 deposition (-1 Tg N yr^{-1}), and fertilizer (-4 Tg N yr^{-1}) in 2050 (Fig. 3). The changes in natural
152 sources of annual nitrogen input include BNF increases from $66 \text{ Tg} \pm 23 \text{ N}$ to $84 \pm 38 \text{ Tg N}$,
153 while nitrogen deposition slightly decreases from $21 \pm 7 \text{ Tg N}$ to $19 \pm 9 \text{ Tg N}$. Regionally, the
154 largest increases of BNF occur in tropical and subtropical forest, particularly the rainforests in
155 Amazon, the Congo Basin, Southeast Asia, and parts of northern Australia (Fig. S4a-c). The
156 rest of the areas, mainly the temperate and boreal forests, experience slight increments of BNF
157 under elevated CO_2 . The reductions of nitrogen deposition are dominant in the vast global area,
158 except for occasional minor increases in some regions (Fig. S4d-f). The application of nitrogen
159 fertilizer occurs mainly in some managed forests in the United States, Europe, Asia, and
160 Oceania (Fig. S4g-i). Due to the significant increases in BNF and NUE, which could meet the
161 nitrogen demands of ecosystems, the human source of fertilizer is proposed to reduce from $4 \pm$
162 1 Tg N yr^{-1} to zero. In sum, the largest increases of total nitrogen input occur in tropical and
163 subtropical forests (Fig. 2d-f). Boreal forests experience slightly increased nitrogen input at the
164 high latitude in North America and Eurasia, relative to the slightly decreased nitrogen input in
165 some temperate forests in western and eastern parts of North America, western and central
166 Europe, and East Asia. The distinct pattern of variations in total nitrogen inputs across regions
167 depends on the trade-offs among changing BNF, deposition and fertilizer in different types of
168 forest by climatic domains. For instance, the increased nitrogen in tropical forests is dominated
169 by the profoundly increased BNF, much higher than the summed reduction of nitrogen
170 deposition and fertilizer.

171
172 The global aggregated nitrogen in forest products increases from $22 \pm 4 \text{ Tg N yr}^{-1}$ in the baseline
173 SSP2 scenario to $26 \pm 6 \text{ Tg N yr}^{-1}$ in the e CO_2 middle road scenario for 2050. The forest
174 products, including wood and non-wood products, originate from some forests in all the
175 continents, apart from the intact forests without any human interventions (Fig. 2g-i). Increases
176 in forest products are mainly projected in wood production hotspots such as Europe, Eastern
177 Asia, North America, southeastern Latin America, and parts of Sub-Saharan Africa¹.
178 Additionally, nitrogen accumulation in the living biomass and soil stock increases from $37 \pm$
179 14 Tg N yr^{-1} to $56 \pm 17 \text{ Tg N yr}^{-1}$ due to elevated CO_2 (Fig. S5a-c). The majority of increases
180 in accumulation occur in tropical and subtropical forests, followed by boreal forests, suggesting
181 that increased nitrogen input dominates these regions. On the other hand, minor decreases in
182 accumulation take place in temperate forests in Europe, western North America, and
183 Northeastern Asia, where intensive production activities might be responsible for depleting the
184 nitrogen pool²⁵.

185
186 The N_r losses all exhibit a decreasing trend in the e CO_2 middle road scenario for 2050, with
187 reductions in NH_3 ($-0.8 \text{ Tg N yr}^{-1}$), N_2O ($-0.9 \text{ Tg N yr}^{-1}$), NO_x ($-1.2 \text{ Tg N yr}^{-1}$), and NO_3^- (-5 Tg
188 N yr^{-1}), respectively (Fig. 3). Regionally, the most significant reductions in aggregated N_r losses
189 are projected in northern South America, central Africa, South and Southeast Asia, and parts of
190 northern Australia (Fig. 2j-l). For the specific N_r component, reductions in NH_3 emissions are
191 dominant in most areas, with profound reductions in parts of North America, Europe, and East
192 Asia (Fig. S6a-c). The spatial reductions in N_2O emissions are similar to that of NO_x emissions
193 (Fig. S6d-i). Slight increases in N_2O and NH_3 emissions occur in certain intact tropical and
194 boreal forests due to the significant increase in nitrogen input from higher BNF. The reductions
195 in NO_3^- leaching and runoff to water bodies are most substantial, especially in the tropical and
196 subtropical forests in Amazon, Congo Basin, Asia, and North Australia, followed by the
197 temperate forests in eastern North America, Europe, and Northeast Asia (Fig. S6j-l).

198 199 **Multiple scenario analysis and impact assessment**

200 Monte Carlo simulations were employed to estimate the averages and uncertainty ranges of

201 nitrogen budget during the study period from 2000 to 2050 (Fig. 4). Historical data indicates
202 that the global forest area has remained relatively stable with a slight downward trend in recent
203 two decades. In the projection beyond 2020, the global forest area is expected to expand over
204 time due to anticipated afforestation and reforestation in the baseline SSP1 “Sustainable society”
205 scenario; conversely, the forest area is projected to shrink due to potential deforestation
206 resulting from human interventions in the baseline SSP2 “Middle road” scenario²⁶. The forest
207 area in SSP1 is greater than that in SSP2, and the forest area is correlated with the size of the
208 nitrogen budget, thus leading to higher baseline carbon sinks, nitrogen inputs, forest products,
209 and N_r losses in SSP1 compared to SSP2 at the same time point. Both elevated CO_2 SSP1-2.6
210 and SSP2-4.5 scenarios show consistent effects on forest carbon and nitrogen budgets. These
211 effects include increased carbon sinks, enhanced total nitrogen inputs, higher forest product
212 yields, and reduced N_r losses under the e CO_2 scenarios in comparison to their corresponding
213 baseline scenarios. This suggests that the positive impacts of elevated CO_2 on carbon-nitrogen
214 interactions remain robust across diverse socioeconomic and climate scenarios spanning from
215 2030 to 2050. Using the SSP1-2.6 scenario as an illustration, elevated CO_2 is projected to
216 increase carbon sink from $1.58 \pm 0.24 \text{ Pg C yr}^{-1}$ to $2.06 \pm 0.31 \text{ Pg C yr}^{-1}$, boost nitrogen input
217 from $144 \pm 19 \text{ Tg N yr}^{-1}$ to $162 \pm 20 \text{ Tg N yr}^{-1}$, raise forest products from $29 \pm 4 \text{ Tg N yr}^{-1}$ to
218 $34 \pm 4 \text{ Tg N yr}^{-1}$, and decrease N_r losses from $32 \pm 4 \text{ Tg N yr}^{-1}$ to $18 \pm 7 \text{ Tg N yr}^{-1}$ by the year
219 of 2030.

220

221 Subsequently, we undertake a comprehensive economic assessment to gauge the direct impacts
222 of elevated CO_2 on the forest carbon and nitrogen cycles under e CO_2 middle road scenario.
223 Our evaluation isolates elevated CO_2 as the sole driver of climate change, bypassing the
224 consideration of other climate change drivers like warming and altered precipitation regimes.
225 The economic valuation of the societal benefits amounts to US\$292 billion, encompassing
226 diverse aspects of human health benefits (US\$17 billion), ecosystem benefits (US\$44 billion),
227 climate impacts (US\$17 billion), and forest production (US\$213 billion) (Fig. 5). Foremost
228 among these benefits is forest production, making a substantial contribution to the overall
229 benefit with a noteworthy surge of US\$211 billion in forest product revenues, along with a
230 US\$2 billion reduction in fertilizer input expenses. The second-highest benefit stems from
231 climate impact, encompassing a US\$10 billion benefit arising from carbon sequestration,
232 coupled with an additional US\$7 billion derived from N_r -induced climate impact (i.e., the
233 reduction of N_2O emissions and its subsequent decrease in global warming potential). Human
234 health and ecosystem benefits primarily result from the reduction of N_r emissions, thereby
235 avoiding harm to both humans and the ecosystem health. Among various geographical regions,
236 North America, including Canada and the United States, stands out as the primary beneficiary,
237 accumulating benefits totaling US\$56 billion. Europe closely follows with US\$51 billion, and
238 China secures the third position with benefits amounting to US\$48 billion.

239

240 **Future perspective**

241 Accelerated carbon and nitrogen cycles in forests create potential synergies that offer
242 opportunities to enhance forest production, carbon sequestration, and mitigate nitrogen
243 pollution. To harness and optimize these benefits in global forests, it is essential to recognize
244 the changes in the coupled carbon-nitrogen relationship and develop sustainable forest
245 production in the context of future climate change.

246

247 Our study projects that elevated CO_2 will benefit forest carbon cycling by promoting forest
248 productivity and increasing the living biomass stock. Increased nitrogen input from natural
249 sources, aided by BNF, helps sustain CO_2 fertilization effects on forest growth and productivity.
250 This reduces the need for synthetic fertilizer in nitrogen-limited natural forests and plantations.

251 The net incomes from forest production are likely to increase, given the higher yields of forest
252 products and the lower cost of fertilizer. As forest production provides livelihoods for the
253 population residing in impoverished mountainous regions², the increased incomes of producers
254 could directly contribute to alleviating poverty and reducing regional inequality.

255

256 Moreover, the CO₂ fertilization presents an opportunity to expand forest carbon sinks, making
257 a greater contribution to carbon neutrality and the goal of limiting global warming below 2°C
258 or even 1.5°C. Enhancing forest conservation, restoration, and afforestation efforts can
259 facilitate and maximize the realization of this potential. This involves implementing ecological
260 restoration projects to expand forested areas and selecting tree species adapted to high
261 atmospheric CO₂ concentrations and possessing high carbon density^{17,27}. Meanwhile, global
262 warming and altered precipitation regimes along with eCO₂ would also change the future
263 distribution of forest, and global forest carbon and nitrogen cycles²⁸⁻³⁰. This would require
264 further efforts from integrated studies in the context of climate change.

265

266 Additionally, the altered carbon-nitrogen interactions under elevated CO₂ demand optimized
267 forest nutrient management. Nitrogen fertilizer usage should be adjusted, and fertilization
268 strategies optimized based on soil nutrient dynamics to improve NUE and reduce production
269 costs. This practice helps mitigate excess N_r losses to the environment and the associated
270 nitrogen pollution, ultimately yielding ecosystem and health benefits for society as a whole³¹.
271 Furthermore, our study is subject to uncertainties arising from data sources and modeling
272 procedures. Future efforts are required to narrow down these ranges of uncertainty. Additional
273 field manipulation experiments in tropical forests are essential to gather firsthand observations
274 on how climate change modulates biogeochemical cycles, which can help fill the data gaps in
275 tropical regions. Further research is needed to examine the responses of other nutrient element
276 balances, aside from nitrogen, to climate change. Field studies have indicated that phosphorus
277 limitation can influence plant productivity responses to elevated CO₂ and warming in natural
278 ecosystems^{32,33}. Maintaining a balanced stoichiometry of nutrient elements, including
279 phosphorus (P) and potassium (K), is crucial for preserving the health and service functions of
280 forest ecosystems.

281

282

283 **Methods**

284 **Database of elevated CO₂ experiments in forests and global synthesis**

285 Data from elevated CO₂ experiments (listed in Supplementary Information [Table S1](#)) and
286 additional sources were compiled to create a comprehensive global dataset on elevated CO₂ in
287 forest ecosystems. Our selection criteria ensured the inclusion of qualified studies, which met
288 the following criteria: (1) experiments with control (ambient CO₂) and treatment (elevated CO₂)
289 groups; (2) regular measurements of variables related to nitrogen and carbon cycles; (3) studies
290 published in peer-reviewed journals and indexed in authoritative databases such as Web of
291 Science, Google Scholar, and Scopus. Site locations, experimental settings, and variable
292 information were extracted from text, tables, and figures. Data from figures were extracted
293 using WebPlotDigitizer 4.4 (<https://apps.automeris.io/wpd/>). In addition, climate data, soil
294 texture, and climate zones were compiled from external sources. Climate data for study sites
295 were obtained from the WorldClim database using site coordinates
296 (<https://worldclim.org/data/index.html#>). Soil texture information was derived from the Global
297 Land Data Assimilation System (GLDAS) (<https://ldas.gsfc.nasa.gov/gldas/soils>). Climate
298 zones were assigned based on the Köppen-Geiger climate classification³⁴.

299

300 For the global synthesis of response mechanisms to elevated CO₂ levels, we employed multi-
 301 level meta-analyses across four levels: (i) individual observations, (ii) combinations by climate
 302 zones (e.g., cold, temperate, arid, tropical), (iii) global scale.

303
 304 To calculate the response ratio of an individual observation, we used the natural logarithm of
 305 the response ratio (*lnR*) formula³⁵, as follows:

$$306 \quad \ln R = \ln \frac{x_{eCO_2}}{x_{aCO_2}} \quad (1)$$

307 where x_{eCO_2} and x_{aCO_2} represent the means of parameters at elevated CO₂ levels and ambient
 308 CO₂, respectively.

309
 310 The weight of each individual observation was determined based on the number of
 311 experimental replications:

$$312 \quad Weight = \frac{n_{eCO_2} \times n_{aCO_2}}{n_{eCO_2} + n_{aCO_2}} \quad (2)$$

313 where n_{eCO_2} and n_{aCO_2} denote the number of replications at elevated CO₂ levels and ambient
 314 CO₂ levels, respectively.

315
 316 Subsequently, we acquired weighted mean response ratios (*RR*) at various levels,
 317 accompanied by 95% confidence intervals, utilizing a randomized resampling procedure
 318 through bootstrapping over 4,999 iterations with *MetaWin 2.0*³⁶. A significant response ratio
 319 ($P < 0.05$) was indicated by non-overlapping 95% confidence intervals with zero.

320
 321 The results were reported as percentage changes of response ratios (*RR%*) for clarity.

$$322 \quad RR\% = (e^{RR} - 1) \times 100\% \quad (3)$$

323 To explore the spatial heterogeneity of response patterns, we conducted meta-regressions for
 324 each variable with potential moderators using the *metafor* package in the *R* platform (version
 325 4.1.3)³⁷. The moderators considered include manipulation magnitude (ΔCO_2), mean annual
 326 temperature (*MAT*), mean annual precipitation (*MAP*), maximum temperature, minimum
 327 temperature, soil texture, and others (Fig. S2).

328 329 **Global forest carbon and nitrogen budgets**

330 The global gridded carbon and nitrogen budgets for forests were estimated using the
 331 DLEM¹⁸ and CHANS¹⁹ models at a spatial resolution of $0.5^\circ \times 0.5^\circ$ (Fig. S1).

332
 333 DLEM is a dynamic global vegetation model that simulates the daily cycles of carbon,
 334 water, and nitrogen driven by atmospheric chemistry, climate, land-use changes, and
 335 disturbances¹⁸.

336
 337 The calculation of plant net primary productivity (*NPP*) at the grid *i* level is based on the
 338 following equations:

$$339 \quad NPP_i = GPP_{sun,i} + GPP_{shade,i} - G_{r,i} - M_{r,i} \quad (4)$$

340
$$G_{r,i} = 0.125(GPP_{sun,i} + GPP_{shade,i}) \quad (5)$$

341
$$GPP_{sun,i} = 12.01 \times 10^{-6} \times A_{sun,i} \times LAI_{sun,i} \times dayl_i \times 3600 \quad (6)$$

342
$$GPP_{shade,i} = 12.01 \times 10^{-6} \times A_{shade,i} \times LAI_{shade,i} \times dayl_i \times 3600 \quad (7)$$

343 where GPP_{sun} and GPP_{shade} represent gross primary productivity (GPP) of sunlit and shaded
 344 canopy, respectively ($\text{g C/m}^2/\text{d}$); Gr denotes the growth respiration of plants ($\text{g C/m}^2/\text{d}$);
 345 A_{sun} and A_{shade} are leaf level assimilation rates of sunlit and shaded canopy, respectively
 346 ($\mu\text{mol CO}_2/\text{m}^2/\text{s}$); LAI_{sun} and LAI_{shade} are projected leaf area index of sunlit and shaded canopy,
 347 respectively (fraction); $dayl$ is daytime length (hour) in a day; 12.01×10^{-6} is a constant to
 348 change the unit from $\mu\text{mol CO}_2$ to g C .

349

350 Maintenance respiration (Mr) of plants ($\text{g C/m}^2/\text{d}$) is a function of surface air temperature
 351 and biomass carbon content, including carbon pools in different plant parts (i.e., leaf, sapwood,
 352 fine root, and coarse root). The calculation of Mr is performed by summarizing all plant parts
 353 as follows:

354
$$M_{r,i} = \sum (\min(Rsep_{coef} \times f(T), r_{max}) \times CV_m) \quad (8)$$

355 where $Rsep_{coef}$ is a plant functional type specific respiration coefficient; $f(T)$ is the temperature
 356 factor, calculated as a function of daily average air temperature; r_{max} is the maximum
 357 respiration rate of different carbon pools; CV_i is the carbon content (g C/m^2) of vegetation pool
 358 m .

359

360 DLEM has 6 vegetation pools, 8 soil pools, 6 litter pools, and 3 product pools. The
 361 calculation of annual net biome productivity at time t is based on the following equation:

362
$$NBP_t = (CV_t + CS_t + CL_t + CP_t) - (CV_{t-1} + CS_{t-1} + CL_{t-1} + CP_{t-1}) \quad (9)$$

363 where NBP_t is net biome productivity at time t ; CV_t , CS_t , CL_t , CP_t are the carbon content (g
 364 C/m^2) of vegetation pool, soil pool, litter pool, and product pool at time t , respectively; CV_{t-1} ,
 365 CS_{t-1} , CL_{t-1} , CP_{t-1} are the carbon content (g C/m^2) of vegetation pool, soil pool, litter pool, and
 366 product pool at time $t-1$, respectively.

367

368 CHANS stands as a nitrogen cycle model that simulates nitrogen flows within diverse
 369 interlinked subsystems of the natural-human interface^{19,38}. These subsystems encompass
 370 cropland, grassland, forest, atmosphere, surface water, and groundwater. Our study
 371 concentrates specifically on the forest subsystem.

372

373 The calculation of forest nitrogen budget is carried out at the grid i level based on the N
 374 mass balance principle. The key nitrogen parameters, including nitrogen input ($N_{input,i}$), N_r
 375 ($N_{r,i}$), forest products ($N_{products,i}$), forest accumulation ($N_{accumulation,i}$), and NUE (NUE_i),
 376 are identified using the following equations:

377
$$\sum_I^k N_{input,i} = \sum_I^k N_{r,i} + \sum_I^k N_{2,i} + \sum_I^k N_{products,i} + \sum_I^k N_{accumulation,i} \quad (10)$$

$$\sum_I^k N_{input,i} = \sum_I^k N_{BNF,i} + \sum_I^k N_{deposition,i} + \sum_I^k N_{fertilizer,i} \quad (11)$$

$$\sum_I^k N_{r,i} = \sum_I^k NH_{3,i} + \sum_I^k N_2O_i + \sum_I^k NO_{x,i} + \sum_I^k NO_{3,i}^- \quad (12)$$

$$NUE_i = \frac{N_{products,i} + N_{accumulation,i}}{N_{input,i}} \quad (13)$$

where $N_{input,i}$ represents the total N input, consisting of BNF ($N_{BNF,i}$), N deposition including both wet deposition (rainfall and snow) and dry deposition (direct settling of particles and gases) ($N_{dep,i}$), and synthetic fertilizer ($N_{fer,i}$); reactive nitrogen ($N_{r,i}$) includes NH_3 emissions ($NH_{3,i}$), N_2O emissions (N_2O_i), NO_x emissions ($NO_{x,i}$), and N leaching and runoff to water ($NO_{3,i}^-$); $N_{2,i}$ denotes N_2 emissions; $N_{products,i}$ represents the quantity of N in both wood and non-wood forest products; $N_{accumulation,i}$ denotes the N increment in living biomass, litterfall, and soil stock.

388

389 The emission factor ($F_{emit,i}$) are defined as:

$$F_{emit,i} = \frac{N_{emit-component,i}}{N_{surplus,i}} \quad (14)$$

391 where $N_{emit-component,i}$ could be any component of $N_{r,i}$, such as $N_{NH_{3,i}}$, $N_{N_2O,i}$, $N_{NO_{x,i}}$, and
 392 $NO_{3,i}^-$.

393

394 In this study, we adopt a multi-model simulation approach to establish robust global forest
 395 carbon and nitrogen budgets, effectively mitigating uncertainties. The gridded data generated
 396 by the DLEM is systematically integrated into the CHANS model. This integration involves a
 397 comparison and calibration process with the nationally-scaled data embedded within the
 398 CHANS model, generating a finely-detailed and accurate gridded dataset.

399

400 Scenario design and model simulation

401 In this study, we developed two sets of scenarios: (i) Baseline scenarios, representing no-
 402 climate-change conditions, consist of SSP1 ('Sustainable society') and SSP2 ('Middle road');
 403 (ii) eCO₂ scenarios, encompassing SSP1-2.6 ('Sustainable society') and SSP2-4.5 ('Middle
 404 road'), consider elevated CO₂ levels as the sole driver of climate change. The future
 405 atmospheric CO₂ concentrations were derived from CMIP6 models²². Additionally, the future
 406 forest areas under different socio-economic pathways (SSP1 and SSP2) were projected based
 407 on a Global Change Analysis Model for future land use²⁶.

408

409 Next, we conducted a multi-model simulation under various scenarios. The impact of
 410 elevated CO₂ on plant NPP and stem nitrogen content is integrated into the forest products
 411 within grid i as below:

$$N_{products,i}^{eCO_2} = N_{products,i}^{base} \times (1 + RR\%_{NPP,i}) \times (1 + RR\%_{stem[N],i}) \quad (15)$$

413 where $N_{products,i}^{eCO_2}$ and $N_{products,i}^{base}$ represent the N in the forest products under the elevated
 414 CO₂ scenario and baseline scenario, respectively; $RR\%_{NPP,i}$ denotes the response ratio of
 415 NPP to elevated CO₂; $RR\%_{stem[N],i}$ denotes the response ratio of stem N content to elevated
 416 CO₂.

417
 418 The effects of elevated CO₂ on NUE are incorporated into the base NUE within grid i as
 419 follows:

$$420 \quad NUE_i^{eCO_2} = NUE_i^{base} \times (I + RR\%_{NUE,i}) \quad (16)$$

421 where $NUE_i^{eCO_2}$ and NUE_i^{base} represent the NUE under the elevated CO₂ and baseline
 422 scenarios, respectively; $RR\%_{NUE,i}$ denotes the response ratio of NUE to elevated CO₂.

423
 424 As for the calculation of N_r emissions, the effects of elevated CO₂ on N_r are incorporated
 425 into the emission factors within grid i as follows:

$$426 \quad F_{emit,i}^{eCO_2} = F_{emit,i}^{base} \times (I + RR\%_{Nrcomponent,i}) \quad (17)$$

427 where $F_{emit,i}^{eCO_2}$ and $F_{emit,i}^{base}$ represent the emission factors under the elevated CO₂ and baseline
 428 scenarios, respectively; $RR\%_{Nrcomponent,i}$ denotes the response ratio of any N_r component to
 429 elevated CO₂.

430
 431 The total N input under elevated CO₂ ($N_{input}^{eCO_2}$) is obtained by summing up all the N output
 432 components within grid i as follows:

$$433 \quad N_{input,i}^{eCO_2} = N_{r,i}^{eCO_2} + N_{2,i}^{eCO_2} + N_{products,i}^{eCO_2} + N_{accumulation,i}^{eCO_2} \quad (18)$$

434 where $N_{r,i}^{eCO_2}$, $N_{2,i}^{eCO_2}$, $N_{products,i}^{eCO_2}$, and $N_{accumulation,i}^{eCO_2}$ represent the N_r, N₂, and the N in the
 435 forest products and ecosystem accumulation under the elevated CO₂ scenario, respectively.
 436

437 The BNF under the elevated CO₂ scenario ($N_{BNF,i}^{eCO_2}$) is attained by integrating the effects
 438 of elevated CO₂ on the base BNF rates as follows:

$$439 \quad N_{BNF,i}^{eCO_2} = N_{BNF,i}^{base} \times (I + RR\%_{BNF,i}) \quad (19)$$

440 where $N_{BNF,i}^{base}$ represents the BNF under the baseline scenario; $RR\%_{BNF,i}$ denotes the response
 441 ratio of BNF to elevated CO₂.

442
 443 The N deposition under the elevated CO₂ scenario ($N_{deposition,i}^{eCO_2}$) is attained by integrating
 444 the effects of elevated CO₂ on the base N deposition as follows:

445
$$N_{deposition,i}^{eCO_2} = N_{deposition,i}^{base} + \Delta NH_{3,i} + \Delta NO_{x,i} \quad (20)$$

446 where $N_{deposition,i}^{base}$ represents the deposition under the baseline scenario; $\Delta NH_{3,i}$ and
 447 $\Delta NO_{x,i}$ denote the changes of NH₃ emissions and NO_x emissions due to elevated CO₂,
 448 respectively.

449
 450 The allocation of human-source fertilizer under the elevated CO₂ scenario is conducted
 451 based on the disparity between the total N input and the natural-source N input, including BNF
 452 and N deposition.

453
 454 The effects of elevated CO₂ on carbon contents are incorporated into the NBP as follows
 455 within grid i :

456
$$NBP_i^{eCO_2} = \sum (NBP_{i,m}^{base} \times (1 + RR\%_{[C],i,m})) \quad (21)$$

457 where $NBP_i^{eCO_2}$ represent the NBP under the elevated CO₂ scenario; $NBP_{i,m}^{base}$ denotes the
 458 NBP of carbon pool k under the baseline scenario; $RR\%_{[C],i,m}$ denotes the response ratio of
 459 carbon content to elevated CO₂ for carbon pool m .

461 Impact assessment

462 The monetary impact analysis of elevated CO₂ (I_{eCO_2}) as a single climate change driver is
 463 conducted at the grid level in global forests, considering its potential impacts on human health
 464 (I_{health}), ecosystem (I_{eco}), climate change ($I_{climate}$), and forest production (I_{pro}) within grid i
 465 as follows:

466
$$I_{eCO_2,i} = I_{health,i} + I_{eco,i} + I_{climate,i} + I_{pro,i} \quad (22)$$

467 The human health impact is determined by the altered health damage resulting from
 468 varying N_r emissions under elevated CO₂ levels within grid i as follows³⁹:

469
$$I_{health,i} = \Delta N_{r,i} \times d_{health,i} \quad (23)$$

470 where $\Delta N_{r,i}$ represents the change in N_r elevated CO₂ at grid i ; $d_{health,i}$ denotes the human
 471 health damage cost of N_r, which is calculated based on the metric of N-share to PM_{2.5} pollution.
 472 The ecosystem impact is quantified as the altered damage cost of N_r effects on the ecosystem
 473 within grid i using the following equation^{40,41}:

474
$$I_{eco,i} = \Delta N_{r,i} \times d_{eco,EU} \times \frac{WTP_j}{WTP_{EU}} \times \frac{PPP_j}{PPP_{EU}} \quad (24)$$

475 where $\Delta N_{r,i}$ represents the change in N_r elevated CO₂ scenario relative to the baseline,
 476 including NH₃, N₂O, NO_x, and NO₃⁻ losses at grid i ; $d_{eco,EU}$ stands for the estimated
 477 ecosystem damage cost of N_r emission in the European Union (EU) based on the European N
 478 Assessment; WTP_j and WTP_{EU} denote the values of the willingness to pay for ecosystem
 479 service in the country/ area j and the EU, respectively; PPP_j and PPP_{EU} denote the
 480 purchasing power parity of the country/ area j and the EU. We extend the ecosystem damage

481 cost of N_r emission in the EU to other countries by applying willingness to pay and purchasing
482 power parity adjustments for comparable ecosystem benefits worldwide, due to data limitations.

483 The assessment of climate impact is conducted considering the influence of carbon
484 sequestration and the N_r losses on climate change within grid *i* as follows⁴²:

$$485 \quad I_{climate,i} = \Delta C_i \times p_{C,i} + \Delta N_{r,i} \times d_{climate,i} \quad (25)$$

486 where the change in C sequestration under elevated CO₂ is estimated by multiplying change of
487 carbon sequestration (ΔC_i) and the carbon price ($p_{C,i}$); we use the national carbon prices for
488 calculation⁴³, and the missing values for some countries are supplemented with means of the
489 income groups; the influence of N_r losses on climate change is estimated by multiplying change
490 of N_r losses ($\Delta N_{r,i}$) and climate damage cost of N_r. The effects of N_r on climate change can be
491 positive or negative, i.e., N₂O contributes to climate warming as a potent greenhouse gas, while
492 NO_x and NH₃ exert climate cooling give that they are precursors of aerosol reflecting long-
493 wave solar radiation.

494 The monetary evaluation of forest production is conducted in terms of production cost
495 (i.e., fertilizer application) and incomes from forest products within grid *i*, as shown in the
496 following equation:

$$497 \quad I_{pro,i} = \Delta N_{fertilizer,i} \times p_{fertilizer,i} + \Delta N_{pro,i} \times p_{pro,i} \quad (26)$$

498 where $\Delta N_{fertilizer,i}$ denotes the changes in N fertilizer under elevated CO₂; the N fertilizer
499 price ($p_{fertilizer,i}$) is estimated by dividing the value of fertilizers traded by the quantity based
500 on the UN Comtrade Database (<https://comtrade.un.org/>); where $\Delta N_{pro,i}$ denotes the changes
501 in forest products under elevated CO₂; the price of forest products ($p_{pro,i}$) is estimated by
502 dividing the value of forest products traded by the quantity based on the FAO database
503 (<https://www.fao.org/faostat/en/#data/FO>). For some countries with missing values of price,
504 the global mean value is used as a substitute.

505

506 **Uncertainty analysis**

507 To evaluate the uncertainty of our model outputs, we conducted an uncertainty analysis using
508 the CHANS model with Monte Carlo simulations. We performed 1,000 iterations to generate
509 projection ensembles and calculate the average and variability of nitrogen budgets. Coefficients
510 of variation (CV) were utilized to quantify the relative ranges of uncertainty for nitrogen budget
511 data and the effects of warming on nitrogen dynamics (Table S2).

512

513 **Correspondence and requests for materials** should be addressed to Baojing Gu.

514

515 **Data availability**

516 Data on the main findings can be found in the Supplementary Information.

517

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522

523 **Author contributions**

524 B.G. and J.C. designed the study. J.C. analyzed the data and wrote the first draft of the paper.

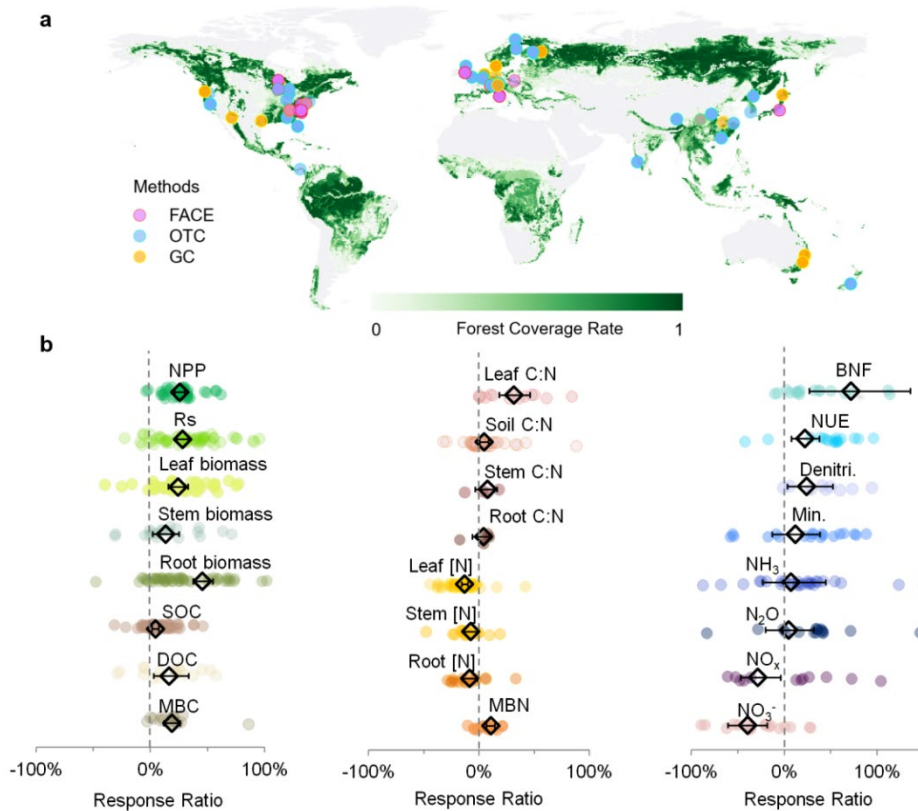
525 All authors contributed to the discussion and revision of the paper. M.Z. provided support for
526 data collection and visualization. X.Z. provided support for CHANS model and impact
527 assessment. Z.B., N.P., and H.T. provided modelling support for DLEM. Z.Q. collected data
528 from climate change experiments. J.X. contributed to the discussion of the study.

529

530 **Competing interests**

531 The authors declare no competing interests.

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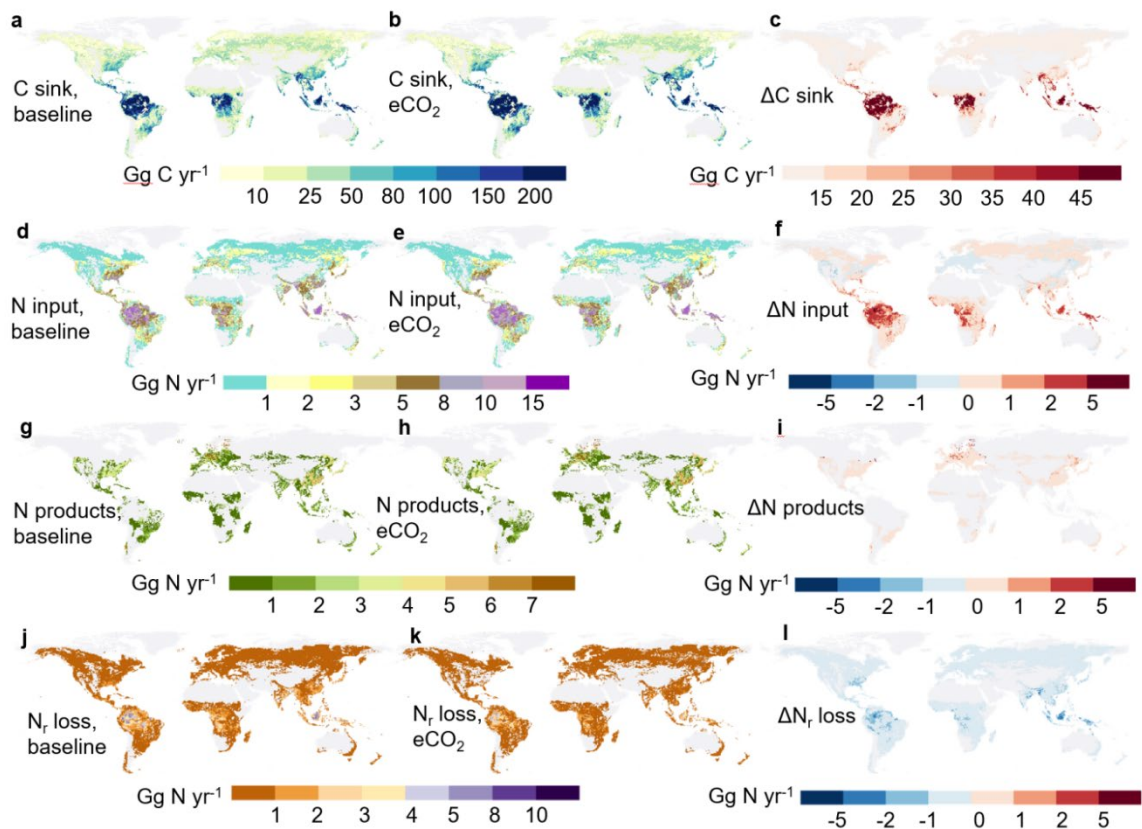
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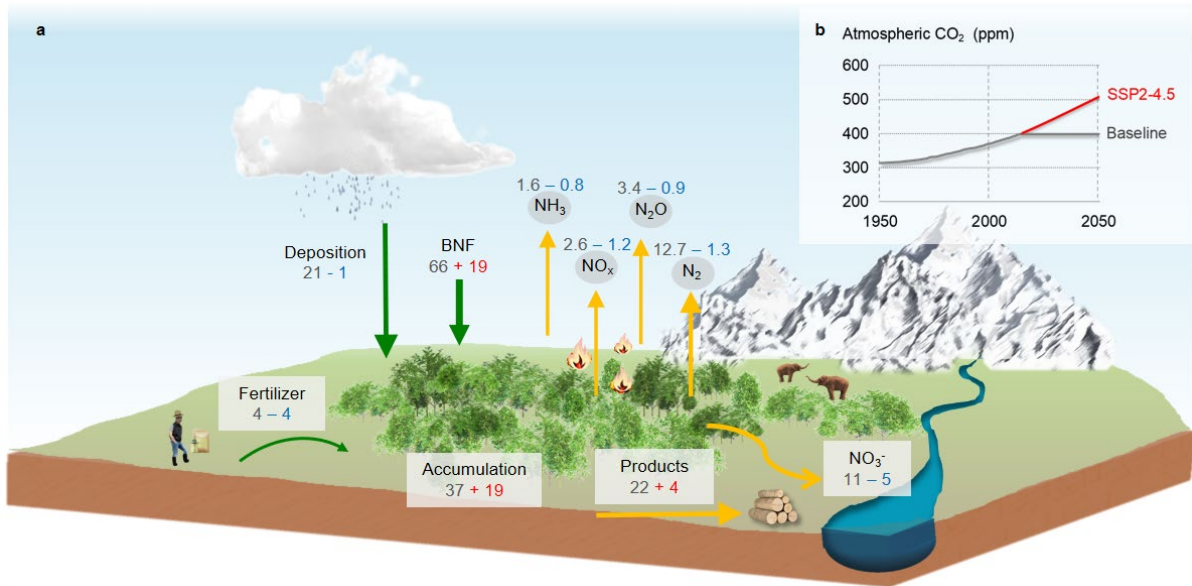
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Fig. 1 | Global elevated CO₂ (eCO₂) experimental sites and eCO₂ impact on carbon and nitrogen variables in global forests. (a) eCO₂ experiments sites with diverse manipulation methods. FACE, Free-Air CO₂ Enrichment; OTC, Open-Top Chamber; GC, such as Greenhouse and Growth Chamber. **(b)** Response ratios of key carbon and nitrogen cycling parameters to eCO₂ from meta-analysis. Scatter plots represent response ratios of observations, and diamonds with error bars indicate mean values of response ratios with 95% confidence intervals. The value of response ratio is significant if the 95% confidence interval does not overlap zero. NPP, net primary productivity; Rs, soil respiration; SOC, soil organic carbon; DOC, dissolved organic carbon; MBC, microbial biomass carbon; [N], nitrogen content; MBN, microbial biomass nitrogen; BNF, biological nitrogen fixation; NUE, nitrogen use efficiency; Denitri., Denitrification; Min., nitrogen mineralization; NO₃⁻, leaching and runoff nitrate to water. The base map is applied without endorsement from GADM data (<https://gadm.org/>).



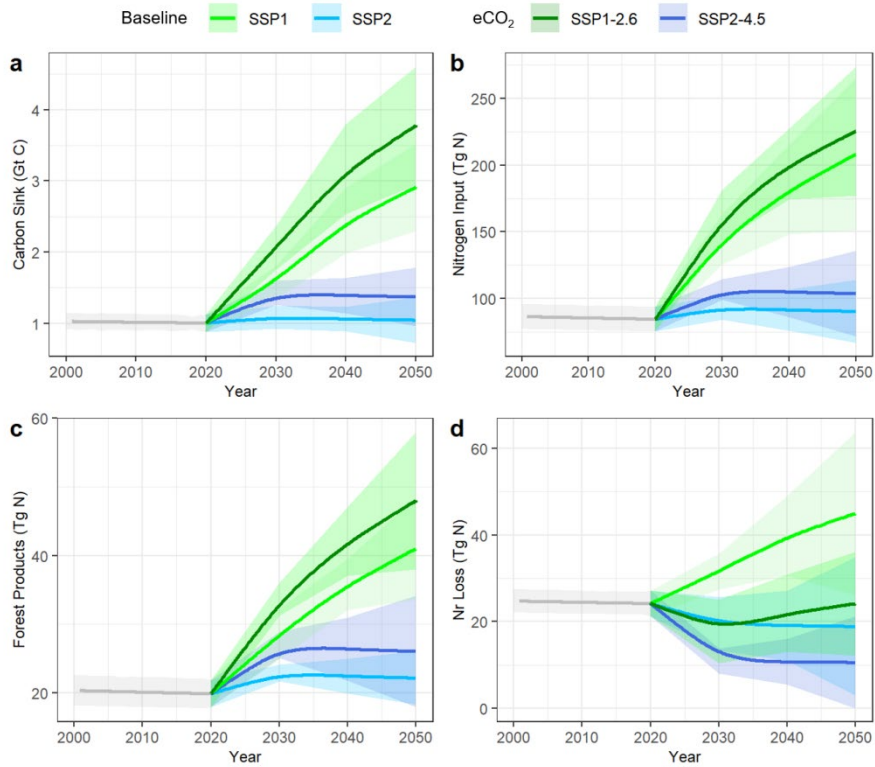
547

548 **Fig. 2 | Carbon sink and nitrogen budgets of global forests and their changes between**
 549 **elevated CO₂ middle road scenario (SSP2-4.5) and baseline scenario in 2050.** Carbon
 550 sink (net biome productivity) in baseline scenario (a), eCO₂ scenario (b), and ΔC sink (eCO₂-
 551 induced change) (c); N input in baseline scenario (d), eCO₂ scenario (e), and ΔN input (eCO₂-
 552 induced change) (f); N products in baseline scenario (g), eCO₂ scenario (h), and ΔN products
 553 (i); N_r loss in baseline scenario (j), eCO₂ scenario (k), and ΔN_r loss (l). Values in the legend
 554 reflect the average annual budgets from forest within a grid cell (0.5° × 0.5°). The base map
 555 is applied without endorsement from GADM data (<https://gadm.org/>).



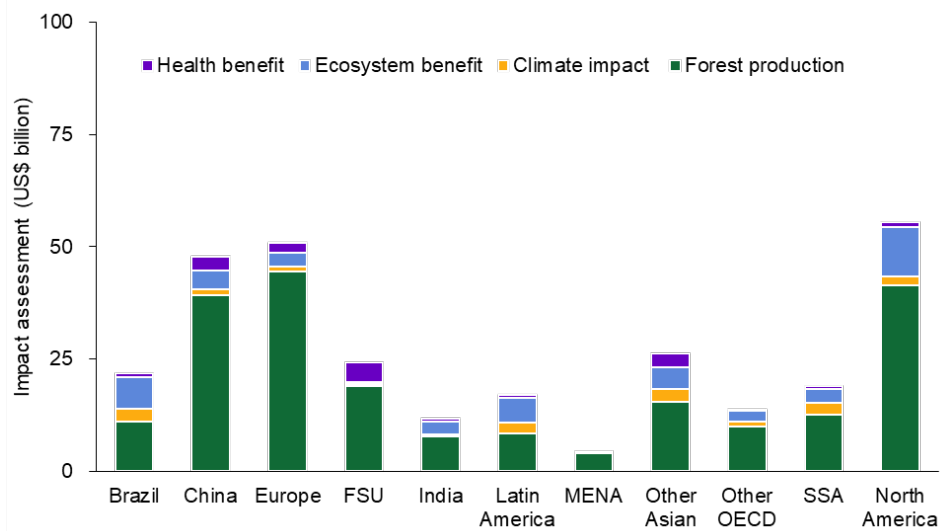
556

557 **Fig. 3 | Nitrogen flows in global forests under elevated CO₂ middle road scenario (SSP2-**
 558 **4.5) for 2050. (a)** Schematic representation of nitrogen budgets in global forests. Green and
 559 yellow arrows represent nitrogen input and output, respectively. Accumulation denotes the
 560 nitrogen residue in living trees and soil stock. Values of nitrogen flows in dark grey represent
 561 flows in the baseline scenario, with the red or blue values indicating increases or decreases in
 562 nitrogen flows due to elevated CO₂. The unit is Tg N yr⁻¹. **(b)** Historical and future
 563 atmospheric CO₂ levels under the baseline and eCO₂ SSP2-4.5 scenarios during 1950–2050.



564

565 **Fig. 4 | Time series of carbon and nitrogen budgets in global forests over the period**
 566 **2000-2050 under baseline and elevated CO₂ scenarios.** Solid lines represent mean values
 567 of carbon sink (a), total nitrogen input (b), forest products (c), and N_r loss (d). Shadings
 568 represent standard deviations of the model ensembles.



569

570 **Fig. 5 | Impact assessment of elevated atmospheric CO₂ levels as a single climate change**
 571 **factor on forests under the elevated CO₂ middle road scenario (SSP2-4.5) relative to the**
 572 **baseline in 2050.** The positive values indicate benefit. FSU, Former Soviet Union; MENA,
 573 Middle East and North Africa; OECD, Organization for Economic Cooperation and
 574 Development; SSA, Sub-Saharan Africa.

575 **References**

- 576 1. FAO, Global forest resources assessment 2020. Rome (2020).
- 577 2. DESA, The global forest goals report 2021. New York (2021).
- 578 3. Bonan, G. B. Forests and climate change: Forcings, feedbacks, and the climate benefits
579 of forests. *Science*. **320**, 1444–1449 (2008).
- 580 4. IPCC. *Global Warming of 1.5°C*. (2018) doi:10.1017/9781009157940.
- 581 5. Harris, N. L. *et al.* Global maps of twenty-first century forest carbon fluxes. *Nat. Clim.*
582 *Chang.* **11**, 234–240 (2021).
- 583 6. Walker, W. S. *et al.* The global potential for increased storage of carbon on land. *Proc.*
584 *Natl. Acad. Sci. U. S. A.* **119**, 1–12 (2022).
- 585 7. Chen, C., Riley, W. J., Prentice, I. C. & Keenan, T. F. CO₂ fertilization of terrestrial
586 photosynthesis inferred from site to global scales. *Proc. Natl. Acad. Sci. U. S. A.* **119**,
587 (2022).
- 588 8. Ruehr, S. *et al.* Evidence and attribution of the enhanced land carbon sink. *Nat. Rev.*
589 *Earth Environ.* (2023) doi:10.1038/s43017-023-00456-3.
- 590 9. Fernández-Martínez, M. *et al.* Global trends in carbon sinks and their relationships
591 with CO₂ and temperature. *Nat. Clim. Chang.* **9**, 73–79 (2019).
- 592 10. Lu, X. *et al.* Nitrogen deposition accelerates soil carbon sequestration in tropical
593 forests. *Proc. Natl. Acad. Sci. U. S. A.* **118**, 1–7 (2021).
- 594 11. Hong, S. *et al.* Asymmetry of carbon sequestrations by plant and soil after forestation
595 regulated by soil nitrogen. *Nat. Commun.* **14**, 1–10 (2023).
- 596 12. Luo, Y. *et al.* Progressive nitrogen limitation of ecosystem responses to rising
597 atmospheric carbon dioxide. *Bioscience* **54**, 731–739 (2004).
- 598 13. Kou, D. *et al.* Progressive nitrogen limitation across the Tibetan alpine permafrost
599 region. *Nat. Commun.* **11**, (2020).
- 600 14. Guo, Y. *et al.* Enhanced leaf turnover and nitrogen recycling sustain CO₂ fertilization
601 effect on tree-ring growth. *Nat. Ecol. Evol.* **6**, 1271–1278 (2022).
- 602 15. Du, E., Fenn, M. E., De Vries, W. & Ok, Y. S. Atmospheric nitrogen deposition to
603 global forests: Status, impacts and management options. *Environ. Pollut.* **250**, 1044–
604 1048 (2019).
- 605 16. Feng, M. *et al.* Overestimated nitrogen loss from denitrification for natural terrestrial
606 ecosystems in CMIP6 Earth System Models. *Nat. Commun.* **14**, 1–9 (2023).
- 607 17. Koch, A. & Kaplan, J. O. Tropical forest restoration under future climate change. *Nat.*
608 *Clim. Chang.* **12**, 279–283 (2022).
- 609 18. Tian, H. *et al.* Model estimates of net primary productivity, evapotranspiration, and
610 water use efficiency in the terrestrial ecosystems of the southern United States during
611 1895–2007. *For. Ecol. Manage.* **259**, 1311–1327 (2010).
- 612 19. Gu, B. *et al.* Cost-effective mitigation of nitrogen pollution from global croplands.
613 *Nature* **613**, 77–84 (2023).
- 614 20. Ainsworth, E. A. & Long, S. P. What have we learned from 15 years of free-air CO₂
615 enrichment (FACE)? A meta-analytic review of the responses of photosynthesis,
616 canopy properties and plant production to rising CO₂. *New Phytol.* **165**, 351–372
617 (2005).
- 618 21. Terrer, C., Vicca, S., Hungate, B. A., Phillips, R. P. & Prentice, I. C. Mycorrhizal
619 association as a primary control of the CO₂ fertilization effect. *Science* vol. 353 72–74

- (2016).
- 621 22. Cheng, W. *et al.* Global monthly gridded atmospheric carbon dioxide concentrations
622 under the historical and future scenarios. *Sci. Data* **9**, 1–13 (2022).
- 623 23. Yu, Z. *et al.* Forest expansion dominates China’s land carbon sink since 1980. *Nat.*
624 *Commun.* **13**, 5374 (2022).
- 625 24. Roebroek, C. T. J., Duveiller, G., Seneviratne, S. I., Davin, E. L. & Cescatti, A.
626 Releasing global forests from human management: How much more carbon could be
627 stored? *Science.* **380**, 749–753 (2023).
- 628 25. Hume, A. M., Chen, H. Y. H. & Taylor, A. R. Intensive forest harvesting increases
629 susceptibility of northern forest soils to carbon, nitrogen and phosphorus loss. *J. Appl.*
630 *Ecol.* **55**, 246–255 (2018).
- 631 26. Chen, M. *et al.* Global land use for 2015–2100 at 0.05° resolution under diverse
632 socioeconomic and climate scenarios. *Sci. Data* **7**, 1–11 (2020).
- 633 27. Zhang, J., Fu, B., Stafford-smith, M., Wang, S. & Zhao, W. Improve forest restoration
634 initiatives to meet sustainable development goal 15. *Nat. Ecol. Evol.* (2015)
635 doi:10.1038/s41559-020-01332-9.
- 636 28. Reich, P. B. *et al.* Even modest climate change may lead to major transitions in boreal
637 forests. *Nature* **608**, 540–545 (2022).
- 638 29. Wu, Q. *et al.* Contrasting effects of altered precipitation regimes on soil nitrogen
639 cycling at the global scale. *Glob. Chang. Biol.* **28**, 6679–6695 (2022).
- 640 30. Wang, J. *et al.* Precipitation manipulation and terrestrial carbon cycling: The roles of
641 treatment magnitude, experimental duration and local climate. *Glob. Ecol. Biogeogr.*
642 **30**, 1909–1921 (2021).
- 643 31. Shah, N. W. *et al.* The effects of forest management on water quality. *For. Ecol.*
644 *Manage.* **522**, (2022).
- 645 32. Ben Keane, J. *et al.* Grassland responses to elevated CO₂ determined by plant–
646 microbe competition for phosphorus. *Nat. Geosci.* (2023) doi:10.1038/s41561-023-
647 01225-z.
- 648 33. Tian, Y. *et al.* Long-term soil warming decreases microbial phosphorus utilization by
649 increasing abiotic phosphorus sorption and phosphorus losses. *Nat. Commun.* **14**, 864
650 (2023).
- 651 34. Beck, H. E. *et al.* Present and future köppen-geiger climate classification maps at 1-km
652 resolution. *Scientific Data* vol. 5 180214 (2018).
- 653 35. Hedges, L. V., Gurevitch, J. & Curtis, P. S. The meta-analysis of response ratios in
654 experimental ecology. *Ecology* **80**, 1150 (1999).
- 655 36. Rosenberg, M., Adams, D. & Gurevitch, J. MetaWin: Statistical software for meta-
656 analysis. Version 2.0. *Sinauer Assoc.* (2000).
- 657 37. Viechtbauer, W. Conducting meta-analyses in R with the metafor. *J. Stat. Softw.* **36**, 1–
658 48 (2010).
- 659 38. Gu, B. *et al.* Toward a generic analytical framework for sustainable nitrogen
660 management: application for China. *Environ. Sci. Technol.* **53**, 1109–1118 (2019).
- 661 39. Gu, B. *et al.* Abating ammonia is more cost-effective than nitrogen oxides for
662 mitigating PM_{2.5} air pollution. *Science.* **374**, 758–762 (2021).
- 663 40. Kristal, S. L., Randall-Kristal, K. A. & Thompson, B. M. *The society for academic*
664 *emergency medicine’s 2004-2005 emergency medicine faculty salary and benefit*

- 665 *survey. Academic Emergency Medicine* vol. 13 (2006).
- 666 41. Sobota, D. J., Compton, J. E., McCrackin, M. L. & Singh, S. Cost of reactive nitrogen
667 release from human activities to the environment in the United States. *Environ. Res.*
668 *Lett.* **10**, 025006 (2015).
- 669 42. Zhang, X. *et al.* Societal benefits of halving agricultural ammonia emissions in China
670 far exceed the abatement costs. *Nat. Commun.* **11**, 1–10 (2020).
- 671 43. Dolphin, G. & Xiahou, Q. World carbon pricing database: sources and methods. *Sci.*
672 *Data* **9**, 1–7 (2022).
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