

Effects of Indoor Air Pollution on Household Health: Evidence from Turkey

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1 **Effects of Indoor Air Pollution on Household Health: Evidence from Turkey**

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13 **Abstract**

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17 Indoor air pollution caused by use of biomass energy in heating and cooking affects the health status of household
18 members adversely. In Turkey, despite with the rapid economic growth in the last decade, biomass has been among
19 the most preferred type of energy by households for heating and cooking due to inadequate infrastructure, dependence
20 on foreign energy, and high energy prices. This study aims to provide empirical evidence from Turkey to the literature
21 on indoor air pollution caused by households' energy choice and health status. This study is analyzed these effects
22 with the random effects panel discrete ordered models using the Income Living Conditions Micro Longitudinal Data
23 Set for the period 2014-2017. As a result of the analysis, we found that age, being a woman, having dependent children,
24 and indoor air pollution have adverse effects on the health status. However, education level, and income level affect
25 the health status positively. The most important result obtained from this study is that even if households have high-
26 income, the inability to access clean energy resources affects their health adversely.

27 **Keywords:** Indoor Air Pollution, Health, Panel Discrete Ordered Models, Turkey

1. Introduction

In developing countries, indoor air pollution has been a public health hazard because of using solid biofuels such as manure, wood, crop residues, and coal for daily cooking and residential space heating. According to the World Health Organization, the use of this fuel by poor households, whose number is more than three billion, causes risk on households' health by creating more air pollution than levels allowed by international ambient air quality standards (WHO, 2018). Indoor air pollution causes to an increase in the risk of some important diseases such as chronic respiratory diseases and lung cancer, low birth weight, pneumonia, stroke, asthma, and cataracts in adults and children (Bruce et al., 2002).

Energy choice theory is often based on "the energy stack" and the "energy ladder" model. The Energy stack model assumes that households tend to choose mix energy sources among all alternatives and households can switch to energy types. On the other hand, energy ladder model was designed as a hierarchical relationship between household' income and socio-economic status and energy type of households for cooking and heating. The polluting effect, efficiency, and costs of fuel are generally identified by the "energy ladder" model. Cheap, inefficient, and most polluting dry animal manure, dropped branches, and grasses are at the bottom of the stairs in the energy ladder. In the second stair, households use coal, kerosene, and charcoal. In the third stair, high-income households tend to use modern fuels such as electricity and LPG. Moving through the stairs of this ladder is based on the income growth of households (Barnes and Floor, 1999). There is a transition from biofuels to petroleum products (kerosene, LPG) and electricity in developed countries. On the other hand, in developing countries, even cleaner and more advanced fuels are available, households generally continue to use biomass (Smith, 1987). Poverty is one of the main constraints to the adoption of cleaner fuels, and unfortunately, the slow growth rate in many countries indicates that biofuels will continue to be used by the poor. Hence, the possible dangerous effects of indoor pollution on human health caused by energy choices, which are closely related to household income and socioeconomic characteristics, should be considered.

In the literature, there are various studies examined the disease risks of indoor air pollution caused by using the fuel in cooking for women and children, who are spending a significant amount of time near the cooking stove (Cerqueiro et al. 1990; Armstrong and Campbell 1991; Johnson and Aderele 1992; Collings et al. 1990; Shah et al. 1994; Albalak et al. 1999; Commodore et al. 2013). The first study analyzing the effect of indoor cooking smoke on children's respiratory diseases was conducted by Sofoluwe (1968). Mishra et al. (2004) investigated the relationship between relying on highly polluting biomass fuels such as wood, manure, or straw preferred by households and the prevalence of acute respiratory infections in children using the cross-sectional logistic regression method. The results show that children living in households using biomass fuels are more than twice as likely to be exposed to acute respiratory diseases. Agrawal and Yamamoto (2015) analyzed the effect of cooking smoke produced by biomass and solid fuel combustion on asthma disease reported among adult men and women in India using multivariate logistic regression on cross-sectional data. The results show that adult women who live in a household using biomass and solid fuels more likely to have asthma. Again, Mishra (2003) analyzed the effect of cooking smoke on asthma using logistic regression for elderly adults. Study results provide that exposure to food smoke is strongly associated with asthma regardless of some demographic factors such as age, education, and standard of living. Duflo et al. 2008; Bruce et al. 2000; De Francisco 1993; Khalequzzaman et al. 2007 are also other studies investigating the effect of cooking smoke on health.

On the other hand, Boy et al. (2002) examined the relationship between exposure to indoor air pollution caused by heating energy choice during pregnancy and low birth weight in rural Guatemala. Lakshimi et al. (2013) investigated the relationship between indoor pollution and adverse pregnancy outcomes such as miscarriage and postnatal infant mortality using the poisson regression. Both studies found a strong association between stillbirths and low birth weight and indoor pollution. Morris et al. (1990) analyzed the effects of fuels used in heating on respiratory tract diseases using a multiple logistic regression, using the self-reported health data. They concluded that households living in houses heated with biomass have a high risk of respiratory disease. Besides, there are also various studies investigating the relationship between health status and indoor air pollution (Ezzati and Kammen 2001;2002; Demjen et al. 2000; Bruce et al. 2002; Smith et al. 2000).

Previous studies have also showed that there are strong associations between indoor air pollution and acute lower respiratory tract infections in young children, chronic obstructive pulmonary disease, and lung cancer in adult women (Fullerton et al.2008; Salvi and Barnes, 2010; Dionisio et al., 2008; Ezzati 2005; Kim et al.2011; Bruce et al.2000; 2004). Moreover, some studies determined the nexus between exposure to coal and biomass smoke and lung cancer in men and women (Subramanian and Govindan 2007; Shrestha and Shrestha 2005), asthma in school-age children

81 (Smith et al., 2000; Mishra 2003; Schei et al.2002) Kovesi et al. 2006), and cataracts and tuberculosis in adults (Ezzati
82 and Kammen, 2001;2002; Mishra et al. 1999; Saha et al. 2005).

83 Although there are various studies investigating the effect of indoor pollution on health in less developed countries,
84 there are relatively less studies on developing countries in the literature (Qui et al. 2019; Kim et al. 2011; Perez-Padilla
85 et al. 2010). Despite the rapid economic growth in Turkey since the early 2000s, biomass fuels such as manure, straw,
86 and wood have still preferred by poorest and most vulnerable households for cooking and heating due to some reasons
87 such as inequalities in income distribution, high energy prices, lack of infrastructure.

88 On the other hand, generally, air pollution is considered only as outdoor air pollution. However, we spend most of our
89 time indoors and indoors are significant microenvironments considering the impact of air pollution on health.
90 Nowadays, the whole world continues to be affected by the rapid spread of the COVID-19 epidemic, which has been
91 expanding globally since the beginning of 2020. Especially during the Covid-19 pandemic, quarantine implications
92 and lockdowns have also increased the time spent indoors. Therefore, determining the effects of indoor air pollution
93 on human health have become an important issue again for researchers and politicians. Although indoor air pollution
94 is generally caused by wet or damp walls, cigarette smoke, house dust, dwelling properties, energy fuel choice in the
95 dwelling is the most critical factor that causes indoor pollution. Especially in developing countries, like Turkey, even
96 though income, economic growth, or some socio-economic factors trigger the transition from dirty fuels to clean fuels,
97 many households rely on biomass fuels for cooking and heating. Therefore, the main motivation of this study is to
98 provide important clues to policies targeting to reduce indoor air pollution and its effect on health by determining
99 these effects of indoor pollution caused by energy choice in Turkey. With this motivation, in this study, we investigated
100 the impacts of indoor air pollution caused by energy choice on the health status applying the random effects panel
101 discrete ordered models on the Income and Living Conditions Research Micro Longitudinal Data Set (2014-2017).
102 To our best of knowledge, this study is the first study analyzing the potential effects of indoor air pollution based on
103 energy choice on physical health in Turkey. Thus, it is expected to fill gap in the relevant literature by providing
104 empirical results of the effect of indoor air pollution on physical health in Turkey.

105 The remainder of the paper is organized as follows. Section 2 provides the data sets used for the empirical analysis.
106 Section 3 details the empirical strategy. Section 4 presents the econometric results. Section 5 presents the discussion
107 of the results.

108 **2. Data and Methodology**

109 **2.1. Data**

110 In this study, the Income and Living Conditions Micro Longitudinal Data Set (ILC) was used in determining the nexus
111 between household health status and indoor air pollution in Turkey. ILC contains 4-waves panel microdata including
112 the overlapping records in the years of 2017-2016-2015-2014. The design of ILC is a two-stage stratified cluster
113 sampling. Household is described as the final sampling unit in the survey. It is possible to produce country-wide
114 estimates from the annual panel research results. In the study, observations without available data for the basic
115 variables, and observations for household members except household head were excluded from the data set. The data
116 set is a four-year balanced panel, and the data set includes a total of 4881 households and 19524 observations.

117 In the analysis, the health status of the household head was used as the dependent variable. Health status was measured
118 as a 5-Likert scale (from 1 Very bad to 5 Very good) based on the self-reported health status. In the survey, the
119 household answered the question of "How is your health?". Since self-reported health status is an ordered categorical
120 variable, we preferred the random effects panel discrete ordered models in the study. Indoor air pollution refers to
121 chemical, biological, and physical contamination of indoor air. It may result in adverse health effects (OECD,2021).
122 However, indoor air pollution is not directly measurable and the ILC data set includes only the primary energy source
123 used by households. Therefore, in the analysis the energy choices of households are made under the assumption of the
124 energy ladder hypothesis. According to this hypothesis, pollution decreases as the primitive fuels switch to modern
125 fuels. For this reason, the energy type preferred by the households is considered as indoor air pollution in the study.

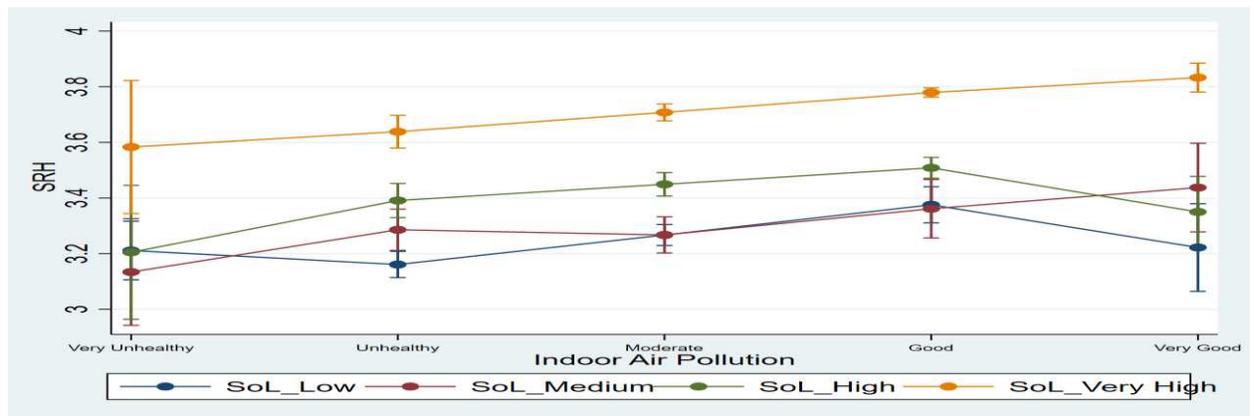
126 Besides indoor pollution, demographic and economic variables that are related to the health status of the individuals
127 are also included in the model. Educational status, gender, age, marital status, and dependent child dummy variables
128 were taken as demographic factors. Household economic status variables which are homeownership, log income, and
129 SoL index were also added to the model as exogenous variables. Table 1 provides definitions and summary statistics
130 of the variables used in the study.

131 **Table 1:** Description of variables.

	Variable	Definition	Mean	S.D.	Min	Max	
Dependent Variable	Self-Reported Health	(1) Very Bad	0.01	0.10	0	1	
		(2) Bad	0.11	0.32	0	1	
		(3) Fair	0.26	0.44	0	1	
		(4) Good	0.55	0.50	0	1	
		(5) Very Good	0.05	0.23	0	1	
Indoor Air Pollution (Caused by type of main energy sources of households)	Level of Indoor Air Pollution	Very Unhealthy	0.02	0.15	0	1	
		Unhealthy	0.16	0.37	0	1	
		Moderate	0.33	0.47	0	1	
		Good	0.41	0.49	0	1	
		Very Good	0.05	0.22	0	1	
Demographics Characteristics of Head of Household	Gender	if female=1 otherwise=0	0.18	0.39	0	1	
	Marital status	if married =1 otherwise=0	0.80	0.40	0	1	
	Age	Age is numerically measured.	50.54	15.11	18	95	
	Dependent child	if there is a dependent child= 1	0.57	0.50	0	1	
	Education	No literacy		0.07	0.26	0	1
		Literate		0.06	0.23	0	1
		Primary school		0.41	0.49	0	1
		Secondary school		0.12	0.32	0	1
		High school		0.18	0.38	0	1
Higher education		0.17	0.37	0	1		
Economic Characteristics	Home ownership	homeownership=1 otherwise=0	0.66	0.47	0	1	
	Logincome	Log of household income	7.96	0.55	0	1	
	Standard of Living	(1) Low		0.24	0.43	0	1
		(2) Medium		0.08	0.27	0	1
(3) High			0.20	0.40	0	1	
(4) Very High			0.46	0.49	0	1	

132 It is widely accepted that there is a strong relationship between individuals' health and housing conditions. Therefore,
 133 using the principal components analysis, we created the standard of the living index over physical properties of the
 134 house associated with health, such as a bathroom, toilet, kitchen, water system, hot water system, and economic assets
 135 such as telephone, internet, TV, and computer. After that, we divided the SoL index into four subgroups as low,
 136 medium, high, and very high, and after we clustered the households.

137 **Graph 1:** Estimated Means of SRH by Level of Indoor Air Pollution and SoL with %95 CIs



138
 139 The energy ladder hypothesis assumes that households follow a certain energy path depending on household income
 140 levels or living standards. Graph1 clearly shows that as one goes up the steps of energy type used as a basis in the
 141 household, his health status is improving. Also, it presents a positive relationship between SoL and individuals' health

142 status. It also shows that there is a decrease in health status of the households using electricity as the primary energy
 143 source. Because even if households have sufficient income levels, electrical energy cannot be used efficiently for
 144 heating due to insufficient infrastructure. Moreover, the positive impact of electrical energy on health due to low
 145 emissions is significantly reduced because of insufficient heating.

146 2.2. Methodology

147 The health status of the household head was used as a dependent variable in the study. Because this variable is ordered
 148 and categorical, we preferred the ordered discrete choice models considering the unobserved individual heterogeneity.
 149 Random effect discrete ordered models are fit via maximum likelihood the random-effects model.

$$150 \Pr(y_{it} > j | \mathbf{J}, x_{it}, v_i) = \Omega(x_{it}\beta + v_i - J_j) \quad (1)$$

151 for $i = 1, \dots, n$ panels, where $t = 1, \dots, n_i$, y is the observed ordinal responses, \mathbf{J} is a set of cut points j_1, j_2, \dots, j_{J-1} ,
 152 where J is the number of possible outcomes, and v_i are independent and identically distributed $N(0, \sigma_v^2)$. If $\Omega(\cdot)$ is the
 153 logistic cumulative distribution function it is random effect ordered logit model. On the other hand, if $\Omega(\cdot)$ is the
 154 standard normal cumulative distribution function, it is random effect ordered probit model.

155 From the above, we can derive the probability of observing outcome j for response y_{it} as

$$156 p_{itj} \equiv \Pr(y_{it} = j | \mathbf{J}, x_{it}, v_i) = \Pr(j_{J-1} < X_{it}\beta + v_i + \varepsilon_{it} \leq J_j) \quad (2)$$

$$157 = \Pr(j_{J-1} - X_{it}\beta - v_i < \varepsilon_{it} \leq J_j - X_{it}\beta - v_i)$$

$$158 = \Omega(J_j - X_{it}\beta - v_i) - \Omega(j_{J-1} - X_{it}\beta - v_i)$$

159 where j_0 is taken as $-\infty$, and j_j is taken as $+\infty$. Here X_{it} does not contain a constant term, because its effect is absorbed
 160 into the cut points.

161 We may also express this model in terms of a latent linear response, where observed ordinal responses y_{it} are generated
 162 from the latent continuous responses, such that.

$$163 SRH_{it}^* = X_{it}\beta + v_i + \varepsilon_{it} \quad (3)$$

$$164 SRH_{it} = \begin{cases} 1 & \text{if } SRH_{it}^* \leq j_1 \\ 2 & \text{if } j_1 < SRH_{it}^* \leq j_2 \\ \vdots & \\ J & \text{if } j_{J-1} < SRH_{it}^* \end{cases} \quad (4)$$

165 SRH_{it}^* indicates the health status of the i^{th} household head at time t , is a latent variable and random effect discrete
 166 ordered model is given above. J is the number of possible outcomes; and j_i is a set of cut points. X_{it} is a matrix of
 167 explanatory variables, representing indoor pollution caused by energy type of household (dry dung, firewood,
 168 charcoal, natural gas, and electricity) and socioeconomic variables (such as age, gender, education, income and SoL)
 169 for the household. β is predicting parameter vector of the explanatory variables. ε_{it} time-varying unobservable
 170 heterogeneity. Depending on the distribution property of the error term ε_{it} , random effect discrete ordered model is
 171 divided into two as random effect ordered logit and random effect ordered probit model. If the errors term
 172 ε_{it} distributed as standard normal with mean zero and variance one, it is called a random effect probit model. However,
 173 if the errors term ε_{it} is distributed as logistic with mean zero and variance $\pi^2/3$, it is called random effect ordered
 174 logit model. Except for the assumption of the distribution of error terms, there is no significant difference between the
 175 two models. For this reason, in the study, the model results are estimated by both methods and the findings are
 176 interpreted based on the goodness of fit model results.

177 2.3. Empirical Results

178 In analyses, if there is a serial correlation in the error term or the panel data do not have an identical distribution,
 179 clustering over the panel variable allows us to obtain consistent estimators (Wooldridge, 2002; Baltagi, 2001). For
 180 this reason, we applied the Wooldridge autocorrelation test for serial correlation. The null hypothesis, which claims

181 there is no serial correlation, was rejected. Therefore, we obtained robust clustered standard errors by
 182 Huber/White/sandwich variance-covariance matrix estimators (Wooldridge, 2020).

183 Clustering on the panel variable produces a consistent VCE estimator when the disturbances are not identically
 184 distributed over the panels or existence serial correlation in the error term. The cluster robust VCE estimator requires
 185 existence many clusters and the disturbances are uncorrelated across the clusters. The panel variable must be nested
 186 within the cluster variable because of the within-panel correlation, which is generally induced by the random-effects
 187 transform when heteroskedasticity or within-panel serial correlation in the idiosyncratic errors exists.

188 Equation 2 shows that parameters β do not have a subscript j . It implies that the estimated coefficients in discrete
 189 ordered models are assumed to be the same regardless of the category of the output variable. This assumption is called
 190 the parallel line assumption or the proportional-odds assumption. In our model, log income (p-value 0.55), the
 191 homeowner (p-value 0.23), and marital status (p-value 0.11) violate the parallel line assumption at 0.10 level. Although
 192 there are many evidence that the assumptions of the ordered models are frequently violated (Long & Freese, 2014) we
 193 applied the parallel lines restriction for these three variables. The AIC and BIC statistics can be used to evaluate the
 194 trade-off between the better fit of the restricted model and the loss of parsimony from having a $J - 1$ coefficient for
 195 each independent variable instead of just one (Long & Freese, 2014). The smaller values of both the AIC and BIC
 196 statistics were obtained for the unrestricted model in which the parallel regression assumption is relaxed compared
 197 with the restricted model.¹

198 Table 2 contains both random effects ordered logit and random effects ordered probit model results. the estimation
 199 results of both models are parallel to each other. However, random effects ordered logit model fit the data set better
 200 when the AIC, BIC, and Log pseudo-likelihood values are compared. Moreover, Mc Fadden, Cox-Snell, and Cragg-
 201 Uhler Pseudo R² values are calculated for the model goodness of fit. All pseudo R² values show that the random effect
 202 logit model has higher goodness of fit than the random effect probit model. The panel-level variance component of
 203 the random effect (sigma2_u) is both large and significant. This result supports that our empirical model captures
 204 unobserved heterogeneity between the households. Cut points represent the values of j_j in Equation 2 and are not
 205 expected to be statistically equal. If the ordered cut points are identical to each other, the relevant cut point must be
 206 eliminated. The null hypothesis is tested by the Wald test and the null hypothesis is strongly rejected. It implies that
 207 cut points are not statistically equal.

208 **Table 2:** Estimation Results of REOPROBIT and REOLOGIT Models

		REOPROBIT		REOLOGIT	
	Variables	Coefficient	Cluster Robust Std. Err.	Coefficient	Cluster Robust Std. Err.
Indoor Air Pollution	<i>Base category</i>				
	Unhealthy	0.2432**	0.0989	0.4281**	0.1813
	Moderate	0.2196**	0.0966	0.3877**	0.1773
	Good	0.2828***	0.0988	0.5163***	0.1813
	Very good	0.3170***	0.1133	0.5679***	0.2082
Demographics Characteristics of Head of Household	Female	-0.3656***	0.0475	-0.6727***	0.0869
	Married	0.0004	0.0433	-0.0210	0.0788
	Age	-0.0460***	0.0014	-0.0842***	0.0025
	Dependent child: Yes	-0.0814**	0.0331	-0.1355**	0.0604
	<i>Base category</i>				
	Literate	0.2886***	0.0804	0.5317***	0.1465
	Primary school	0.5940***	0.0656	1.0647***	0.1204
	Secondary school	0.6565***	0.0803	1.1826***	0.1478
	High school	0.7923***	0.0776	1.4306***	0.1427
	Higher education	1.0630***	0.0805	1.9293***	0.1486
Economic Characteristics	Homeowner: Yes	0.1257***	0.0322	0.2278***	0.0592
	logincome	0.0409	0.0274	0.0662	0.0500

¹ AIC and BIC values for restricted model are 35857.17, 35809.85 respectively.

	<i>SoL_Low</i>	<i>Base category</i>			
	SoL_Medium	0.0956**	0.0474	0.1813**	0.0855
	SoL_High	0.1533***	0.0409	0.2728***	0.0740
	SoL_Very High	0.2729***	0.0408	0.4875***	0.0745
Cut points of Outcomes	/cut1	-4.7132***	0.2585	-8.8271***	0.4734
	/cut2	-2.8533***	0.2534	-5.2832***	0.4631
	/cut3	-1.3550***	0.2514	-2.5791***	0.4596
	/cut4	1.4582***	0.2492	2.5570***	0.4554
	/sigma2_u	0.7174***	0.0306	2.4447***	0.1034
Model	Number of Obs		19524		19524
Diagnostics	Number of Groups		4881		4881
	Wald chi2(18)		3130.24		3078.06
	Wooldridge F(1,4880)		36.394***		36.394***
Goodness of Fits	Log pseudolikelihood		-17842.454		-17756.499
	AIC		35730.91		35559.00
	BIC		35912.13		35740.22
	Pseudo R ²		0.034		0.039
	Mc Fadden R ²		0.034		0.039
	Cox-Snell R ²		0.064		0.072
	Cragg-Uhler R ²		0.075		0.084

209 ***P <0.01, **P <0.05, *P < 0.10.

210 Since both model estimation coefficients are not practical for interpretation directly, marginal effects are estimated to
 211 understand better the analysis results. Marginal effects are estimated when other estimators assumed constant at a
 212 certain level (usually for mean or median values), and they represent the relationship between their predicted
 213 probabilities. In other words, marginal effects measure how the probability of the output changes when the value of
 214 the estimator changes by one unit. Marginal effects of the REOLOGIT model are given in Table 3.

215 **Table 3:** Marginal effects at the sample mean.

Variable	SRH=1	SRH=2	SRH=3	SRH=4	SRH=5
Unhealthy	-0.0016**	-0.0282**	-0.0426**	0.0610**	0.0113***
Moderate	-0.0015*	-0.0257**	-0.0383**	0.0554**	0.0101***
Good	-0.0018**	-0.0321**	-0.0499***	0.0702***	0.0136***
Very good	-0.0020**	-0.0354***	-0.0563***	0.0780***	0.0156***
Female	0.0024***	0.0418***	0.0642***	-0.0909***	-0.0175***
Age	0.0002***	0.0047***	0.0085***	-0.0108***	-0.0026***
Dependent Child	0.0004**	0.0082**	0.0150**	-0.0190**	-0.0046**
Literate	-0.0041***	-0.0461***	-0.0371***	0.0800***	0.0073***
Primary school	-0.0066***	-0.0845***	-0.0881***	0.1594***	0.0198***
Secondary school	-0.0069***	-0.0911***	-0.0995***	0.1743***	0.0232***
High school	-0.0075***	-0.1039***	-0.1249***	0.2048***	0.0315***
Higher education	-0.0082***	-0.1242***	-0.1760***	0.2552***	0.0532***
Homeowner	-0.0007***	-0.0131***	-0.0230***	0.0299***	0.0068***
SoL_Medium	-0.0006**	-0.0110**	-0.0169**	0.0239**	0.0045*
Sol_High	-0.0009***	-0.0171***	-0.0273***	0.0378***	0.0075***
SoL_Very High	-0.0015***	-0.0287***	-0.0496***	0.0650***	0.0147***

216 ***P <0.01, **P <0.05, *P < 0.10.

217 Note: Only the statistically significant effects are listed. Marginal effects of the REOPROBIT are available upon
 218 request.

219 Table 3 shows the effect of the factor affecting the health status compared to the reference group based on each health
 220 level when the other variables are constant at their average levels. As level of indoor pollution is getting better in
 221 households with Very Bad, Poor, and Moderate health levels, the probability of being at their current health level

222 decreases. On the other hand, in households with good and very good health levels, as the level of indoor pollution is
223 getting better, the probability of being at these health levels increases significantly. Households with well health levels
224 increase their chances of having this health level by 6%, 5.5%, 7%, and 7.8%, respectively, depending on the indoor
225 air pollution.

226 Moreover, age, having dependent children, and being a woman increase the probability of being at fair and lower
227 health status but decreases the possibility of having a good and very good health status level. These results are
228 consistent with existing literature and the realities of the country. Because women are responsible for cooking and
229 they spend a long time at home due to the gender role women have in Turkey. Furthermore, income level, education
230 level, and homeownership decrease the probability of being at a fair and below health status and increase the possibility
231 of being at a good and very good health status level.

232 **Conclusion**

233 Approximately half of the world's population prefers biomass fuels and coal such as wood, animal manure, and crop
234 residues for their energy needs (WHO, 2018). There is evidence that the transition from biomass to cleaner fuels in
235 heating and cooking, especially in low-income groups, has been decreased significantly over the years (Bruce et al.
236 2002). It causes to increase indoor air pollution and health problems.

237 Although the health impacts of air pollution have been examined by many studies in Turkey, the number of studies
238 on the health effects of indoor air pollution has been limited. In this paper we determined the effects of indoor air
239 pollution caused by households' energy choice on household health applying the random effects panel discrete ordered
240 models on the Micro Longitudinal Data Set in Turkey. Differently to previous studies we consider unobserved
241 heterogeneity between the households. We found that age, gender, and having dependent children in the household
242 affect household health negatively. However, income, education, and homeownership affect positively. Education is
243 an important variable that positively affects a person's health status. We also measured the standard of living of
244 households and showed the effect on indoor air pollution. We noted that although the socioeconomic variables of the
245 household affect health, these effects are relatively limited. On the other hand, indoor air pollution is an important
246 variable that affects the health status of individuals. Moreover, the impact of indoor air pollution on health is higher
247 than that of income level.

248 These are significant because indoor air pollution, which is caused by relying on biomass energy in heating and
249 cooking, has serious effects on the health status of individuals living in the household. Therefore, determining this
250 effect empirically will provide some clues to decision-makers in energy and environmental policies. As a result of the
251 study, we report that even if households have high-income levels, their inability to access clean energy resources has
252 serious negative effects on health. Especially in developing countries such as Turkey, only economic development
253 and increases in the national income are not good enough to increase public health. In such countries, access to clean
254 and safe energy sources should be increased, as well. For this reason, policymakers should pay attention to
255 infrastructure services to increase households' access to clean and reliable energy sources.
256

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365

366 **DECLARATIONS**

367 **Ethics approval and consent to participate**

368 Not applicable

369 **Consent for publication**

370 Not applicable

371 **Availability of data and materials**

372 The data that support the findings of this study are available from Turkish Statistical Institute.

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380 **Authors' contributions**

381 All authors contributed to the study conception and design. Material preparation, data collection
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