

Removal of NO With Fe(II)NTA Solution Catalyzed By The Carbon Treated With Ethylenediamine

Xin-wei Dou

East China University of Science and Technology

Pei-yun Chen

East China University of Science and Technology

Ruo-chuan Zhang

East China University of Science and Technology

Xiangli Long (✉ longdragon@ecust.edu.cn)

East China University of Science and Technology

Research Article

Keywords: activated carbon, catalysis, Fe(II)NTA, nitric oxide, ethylenediamine

Posted Date: May 13th, 2021

DOI: <https://doi.org/10.21203/rs.3.rs-360602/v1>

License: © ⓘ This work is licensed under a Creative Commons Attribution 4.0 International License.

[Read Full License](#)

Removal of NO with Fe(II)NTA solution catalyzed by the carbon treated with ethylenediamine

Xin-wei Dou, Pei-yun Chen, Ruo-chuan Zhang and Xiang-li Long*
State Key Laboratory of Chemical Engineering, School of Chemical Engineering,

East China University of Science and Technology, Meilong Road 130, Shanghai 200237, People's Republic of China

Abstract: Fe(II)NTA solution manifests a good performance in the simultaneous removal of sulfur dioxide and nitric oxide. Activated carbon is used to catalyze the reduction of Fe(III)NTA to Fe(II)NTA to retain the ability of absorbing NO. Ethylenediamine(EDA) solution is capable of changing the physical structure and chemical characteristics on the carbon surface to improve the catalytic capability of activated carbon. The experiments suggest that the best treatment condition be immersing the carbon in 5.0 mol l⁻¹ EDA solution for 6 h followed by being heated at 700 °C in N₂ for 4 h. The modification with EDA increases the surface area and alkalinity on the carbon. The experiments also indicate that the removal efficiency of nitric oxide catalyzed by the modified carbon is significantly improved compared with that of the original one.

Keywords: activated carbon, catalysis, Fe(II)NTA, nitric oxide, ethylenediamine

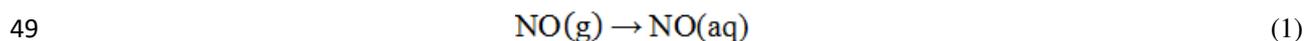
1 Introduction

The emissions of sulfur dioxide and nitric oxide are causing a series of environmental problems such as acid rain, ozone layer destruction, photochemical smog, and even endangering human health(Toro et al. 2014; Saikawa et al. 2017; Ma et al. 2017). Faced with the severe impact of these pollutants, and more and more stringent national emission requirements, various technologies have been developed to control the emission of SO₂ and NO_x. Wet flue gas desulfurization (FGD) is the most popular technology used for SO₂ removal due to its excellent performance in commercial application[4-5](Dou et al. 2009; Zheng et al. 2014). However, this approach is incapable of eliminating NO_x since 90–95% of the NO_x in typical flue gases is the water-insoluble nitric oxide (NO). In order to solve this problem, some oxidants have been used to transform NO into soluble NO₂ (Pan et al. 2015; Yan et al. 2018; Kang et al. 2020; Guo 2018; Liémans and Thomas 2013; Mondal and Chelluboyana 2013; Khan and Adewuyi 2010) However, the oxidation of NO by chemical agents has not been applied commercially yet due to their high costs and the production of large amounts of waste water. The SCR technology is most widely used in coal-fired power plants to complete the reduction of nitric oxide in the flue gases. But this technology suffers the disadvantage of high capital and operating costs(Tang et al. 2020). Hence there is an urgent need to develop a low-cost, easily- industrialized method of denitrification.

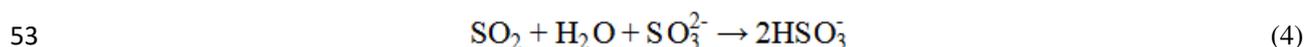
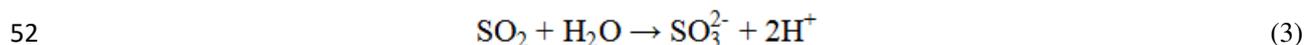
*Corresponding author, Professor Xiang-li Long, Tel.: +86 21 6425 3267; fax: +86 21 6425 3528; E-mail: longdragon@ecust.edu.cn (X. L. Long), State-Key Laboratory of Chemical Engineering, East China University of Science and Technology, Shanghai, 200237, P. R. China.

34 The low solubility of NO in water can be avoided effectively by binding NO with complexants. Such
 35 liquid-phase complex method has a broad industrial application prospect because it holds the advantages of
 36 high denitrification efficiency, fast reaction rate and small equipment input. The approach that introduces
 37 Fe(II)(NTA) (NTA, nitrilotriacetic acid) or Fe(II)(EDTA)(EDTA, ethylenediaminetetraacetate) into the
 38 scrubbing liquor to enhance the solubility of NO via the formation of Fe(II)(NTA)NO or
 39 Fe(II)(EDTA)NO has been studied extensively(Hofeleet al.1996; Chandrashekhar et al.2015;
 40 Chandrashekhar et al. 2013; Zhu et al. 2010). Compared with EDTA, NTA holds several advantages
 41 such as smaller molecular weight and less toxicity. Furthermore, the complex Fe(II)(NTA)NO, is less
 42 stable than Fe(II)(EDTA)NO(Wolak and van Eldik 2002), which is helpful to the reduction of NO to N₂.
 43 Although Fe(II)(NTA) can obtain a high NO removal efficiency, it is easily oxidized to Fe(III)(NTA) that
 44 is not capable of binding NO. Activated carbon can speed up the regeneration of Fe(II)(NTA) with
 45 sulfite/bisulfite as a reductant to maintain the NO removal efficiency so as to realize the simultaneous
 46 removal of SO₂ and NO from flue gas streams for a long period. The combined elimination of NO and
 47 SO₂ can be described briefly as follow.

48 Fe(II)(NTA) may react with dissolved NO according to the following equations:



51 In the meantime, the SO₂ existing in the gas stream also dissolves into the aqueous solution:



54 However, the oxygen coexisting in the flue gases may oxidize Fe(II)(NTA)⁻ to Fe(III)(NTA)(Eq.(5)) during
 55 the gas scrubbing. NO removal efficiency will decrease quickly as the operation proceeds due to the
 56 decline of Fe(II)(NTA) concentration.



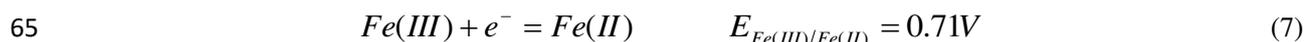
58 To retain the NO removal efficiency, activated carbon can be used as a catalyst and the sulfite/bisulfite
 59 ions produced by SO₂ absorption into the aqueous solution act as reductants to regenerate Fe(II)(NTA)⁻.

60 The mechanism of Fe(III)(NTA) catalytic reduction can be expressed as follows:

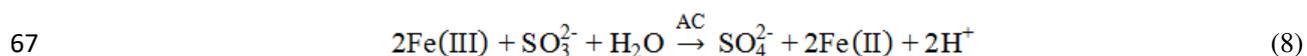
61 The Fe(III)NTA adsorbed on the carbon surface disintegrates into Fe(III) and NTA:



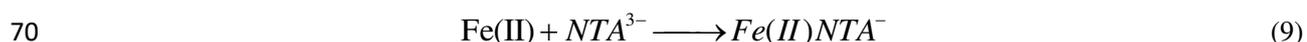
63 Electrochemical half-cell reduction potential of Fe(III)/Fe(II)(Eq(7)) exhibits that Fe(III) is a strong
 64 oxidant and can be reduced to Fe(II) easily.



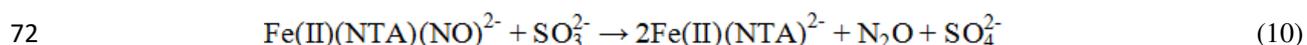
66 The net reaction for the regeneration of Fe(II) ions can be written as follow:



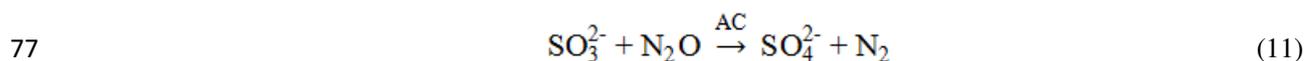
68 Fe(II)(NTA) is regenerated by Fe(II) coordinating with NTA(Eq. (9)). Therefore, the NO removal
69 efficiency can be sustained for a long time.



71 Besides, the NO coordinated with Fe(II)(NTA) may be reduced to N₂O by sulfite.



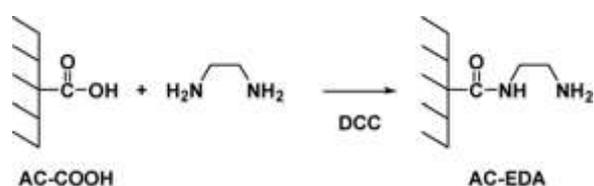
73 Nitrous oxide is also a harmful pollutant that contributes to the depletion of the ozone layer(Ravishankara
74 et al. 2009). It has been reported that N₂O can be reduced to N₂ by CO under the catalytic action of
75 carbon(Chen et al. 2019). In the process discussed in this paper, N₂O can be further reduced to N₂ by
76 sulfite/bisulfite under the catalysis of activated carbon.



78 According to the discussion above, this technology realizes not only the absorption and reduction of
79 nitric oxide but also the absorption and oxidation of sulfur dioxide.

80 Activated carbon plays an important role in the process of regeneration of Fe(II)(NTA). The catalytic
81 activity of activated carbon is dependent on its surface characteristics. By transforming the physical
82 structure and chemistry characteristics on the carbon surface, the capabilities of activated carbon may be
83 improved to a certain extent. Wang et al.(Wang et al. 2019) treated the activated carbon to upgrade
84 isophthalic acid yield from the oxidation of m-xylene under the catalysis of phosphotungstic acid loaded on
85 the activated carbon. Zhang et al.(Zhang et al. 2016) found that the thermal modification can enhance
86 phenol adsorption on carbon samples. Fu et al.(Fu et al, 2016) reported the introduction of amino groups
87 enhanced the adsorption capacity of heavy metals. Zhang et al.(Zhang et al. 2016) prepared a kind of
88 activated carbon with well-developed micropore structure and abundant basic nitrogen-containing
89 functionalities by KOH activation and ammonia modification. Zheng et al.(Zheng et al. 2019) used
90 dicyandiamide to dope activated carbon with nitrogen.

91 Ethylenediamine is a strong alkaline with nitrogen content of 46.6%. It may combine with the acidic
92 groups, such as carboxyl groups, as follow(Li et al. 2009):



94 The N-containing groups on the surface increases the electronic density and the basicity of the activated

95 carbon(Messele et al. 2014). Li et al(Li et al. 2009) performed a study on the modification of activated
 96 carbon with ethylenediamine for selective solid-phase extraction and preconcentration of metal ions.
 97 Messele et al.(Messele et al. 2014) found that the carbon modified with ethylenediamine significantly
 98 enhanced phenol removal efficiency. Fu et al.(Fu et al. 2016) treated the activated carbon with
 99 ethylenediamine to enhance the adsorption capability for Cr(VI). In this paper, ethylenediamine solution
 100 has been tried to treat the activated carbon of coconut to promote its catalytic capacity in the regeneration of
 101 Fe(II)(NTA) in order to obtain higher NO removal efficiency.

102

103 **2 Experimental**

104 **2.1 Materials**

105 The coconut activated carbon purchased from Shanghai Activated Carbon Co., Ltd, was purged with
 106 deionized water and then dried at 110 °C in a vacuum for 24 h. Carbon samples of 100 ~ 120 mesh were
 107 prepared before being used as catalysts or modified with EDA solution. The carbon was dealt with EDA
 108 solution in the following steps: Firstly, 5 g activated carbon was impregnated in 250 mL EDA solution for
 109 several h at room temperature; secondly, after being filtrated, the carbon sample was put in a vacuum
 110 drying oven at 110 °C for 12 h; thirdly, the samples obtained were heated under N₂ atmosphere in a
 111 furnace at set temperature for a few h.

112 **2.2 Reduction of Fe(III)NTA**

113 The experiments to test the catalytic capability of the carbon samples in the reduction of Fe(III)NTA
 114 were carried out in a stirred glass flask of 250 ml with a turbine impeller of diameter 3 cm mounted on the
 115 bottom of the stirring rod. The stirring speed was 300 rpm. When the temperature of the solution
 116 reached 70 °C, 1.0 g activated carbon and 3.15 g Na₂SO₃ were added sequentially into the glass flask filled
 117 with 200 ml 0.01 mol L⁻¹ Fe(III)NTA solution. In the course of the experiments, 1.0 mL liquid sample
 118 was taken from the flask every few min into a 100 mL volumetric flask containing 5 mL 0.025 mol L⁻¹
 119 phenanthroline solution, 5 mL pH 2.9 glycine solution and 1 mL 0.1mol L⁻¹ NTA solution. And then the
 120 liquid in the volumetric flask was raised to 100 ml by adding deionized water. The Fe²⁺ concentration was
 121 measured with a spectrophotometer at 25°C from the absorbance at 510 nm. The Fe(II) calibration curve
 122 was obtained using standard FeSO₄·7H₂O solutions ranged from 0.00 to 0.12 mmol L⁻¹. Least-squares fits
 123 to the data yield Eq. (13) with a correlation coefficient (r²) 0.9998.

$$124 \quad C = 10.888 A - 0.0006 \quad (13)$$

125 where A stands for absorbency and C for Fe(II) concentration (10⁻³ mol L⁻¹).

126 Fe(III) concentration was computed from the difference between total iron and Fe(II). Fe(III)NTA]
 127 conversion ($X_{Fe(III)NTA} \%$) is determined as follow:

$$128 \quad X_{Fe(III)NTA} \% = \frac{C_{Fe_{total}} - C_{Fe(III)}}{C_{Fe_{total}}} = \frac{C_{Fe(II)}}{C_{Fe_{total}}} \times 100 \quad (14)$$

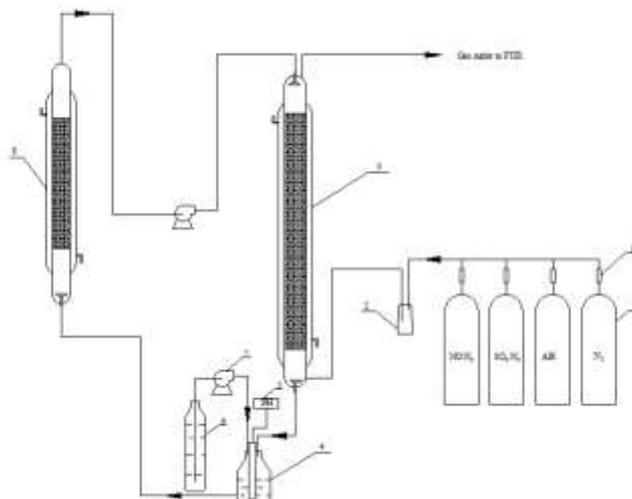
129 Where $C_{Fe_{total}}$, $C_{Fe(III)}$ and $C_{Fe(II)}$ are the total iron, Fe(III) concentration and Fe(II) concentration at t

130 time in the solution, respectively.

131

132 2.3 Combined removal of NO and SO₂

133 The schematic diagram of the experimental apparatus for the simultaneous removal of NO and SO₂ is
134 shown in Fig. 1 The absorption was performed in a packed tower (18 mm i.d., 1000 mm long) and the
135 Fe(II)NTA⁻ regeneration was carried out in a fixed-bed (20 mm i.d.) reactor packed with 20 g activated
136 carbon of 20-40 mesh. The temperature of the absorber and regeneration tower was controlled at 50 °C
137 by the jackets through which water from thermostatic baths was recycled. Five hundred milliliter
138 Fe(II)NTA⁻ solution together with measured amount of Na₂SO₃ was added into the circulation tank. The
139 pH was controlled at 5.5 using NaOH (1.0 mol/L) solution by a THORNTON M300 pH/ORP transmitters
140 as well as a pH-electrode in the course of the experiment. The absorber was operated with a continuous
141 gas stream feeding at 270 ml min⁻¹ from the bottom and a continuous scrubbing solution feeding at 25 ml
142 min⁻¹ from the top. The absorbent discharging from the packed tower was fed into the circulation tank.
143 When the regeneration of Fe(II)NTA⁻ started, the absorbent in the circulation tank ran through the
144 regeneration reactor upwardly and into the packed tower to scrub NO and SO₂ directly after it left the
145 regeneration reactor. The experiment was performed under atmospheric pressure.



146

147 **Fig.1** Flowchart of absorption and regeneration reactor system

148 1-Cylinder; 2-buffer tank; 3-packed column; 4-circulation tanker; 5-pH meter; 6-NaOH solution;

149 7-pump; 8-regeneration reactor; 9-massmeter

150 The quantitative analysis of gas compositions was achieved by an on-line Fourier transform infrared
151 spectrometer (Nicolet E.S.P. 460 FT-IR) equipped with a gas cell and a quantitative software package,
152 named Quant Pad. The length of the gas cell in the FTIR is 2 m. The peaks in the region 1150-1200
153 cm⁻¹, 2850-2935 cm⁻¹, 2150-2225 cm⁻¹ and 1875-1960 cm⁻¹ were applied to identify SO₂, NO₂, N₂O and
154 NO, respectively. The inlet and outlet gases moved directly into the gas cell of the FTIR to obtain the
155 transient N₂O, NO, NO₂ and SO₂ concentrations in both the inlet and outlet gases, as well as the transient
156 NO conversion. The heights of the peaks are proportional to the concentrations of the components to be

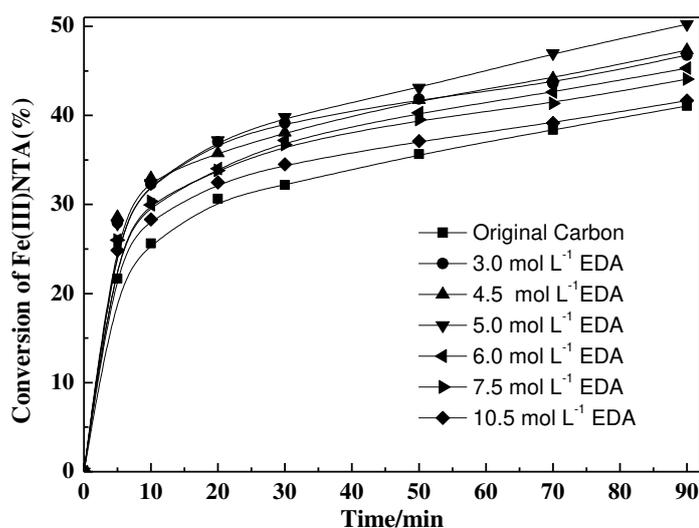
157 detected and the relative standard deviations of the measurements were determined to be within 1%.
158 Hence this set-up is feasibly and conveniently operated to monitor the NO and SO₂ removal efficiency.

159 2.4 Characterization of carbon samples

160 FT-IR was used to analyze the functional groups on the surface of activated carbon by KBr
161 compression method(O'reilly and Mosher 1983) and the point of zero charge(pH_{pzc}) was determined by
162 mass titration.. An ASAP2020 surface Analyzer(Micromeritics Co. USA) was used to measure the
163 specific surface area of activated carbon with nitrogen as adsorption medium at 77K. The total surfaces of
164 the carbon samples were calculated by BET method. The surface area and volume of mesopores were
165 obtained by BJH method, and the micropores of which were computed by t-plot method. The content of
166 acidic and basic functional groups on the surface of activated carbon were measured by Boehm
167 titration(Boehm 1994). XPS was characterized by an ESCALAB 250 electron spectrometer from Thermo
168 Corporation with 300 W AlK α radiation at the base pressure of 3 \times 10⁻⁹ mbar.

170 3 Results and discussion

171 3.1 Effect of ethylenediamine concentration



172
173 **Fig. 2 Effect of the EDA concentration on the catalytic performance of carbon samples**

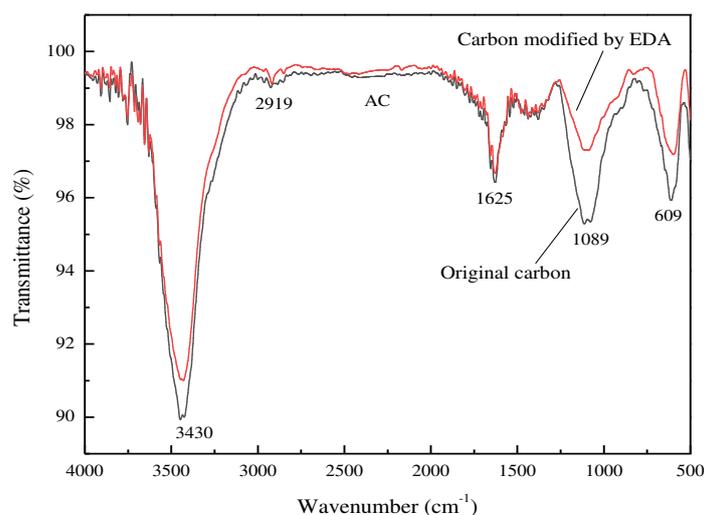
174 [Fe(III)NTA]=0.01 mol L⁻¹, SO₃²⁻=0.1 mol L⁻¹, pH=6.17, 300 rpm, 70 °C

175 To explore the effect of EDA concentration on the catalytic capacity of the activated carbon, six
176 carbon samples were impregnated in EDA solutions with a concentration of 3.0 mol L⁻¹, 4.5 mol L⁻¹, 5.0
177 mol L⁻¹, 6.0 mol L⁻¹, 7.5 mol L⁻¹ and 10.5 mol L⁻¹, respectively, for 6 h at ambient temperature. And then
178 they were heated in N₂ at 800 °C for 4 h. The prepared samples were used to catalyze the reduction of
179 Fe(III)NTA at 70 °C. The conversions of Fe(III)NTA presented in Fig.2 prove that the catalytic
180 performance of the activated carbon is ameliorated after the carbon is treated with EDA solution. After 90
181 min's operation, the Fe(III)NTA conversion got by the original carbon is 41.07 % while those acquired by

182 the one treated with 3.0, 4.5, 5.0, 6.0, 7.5 and 10.5 mol L⁻¹ EDA solution are 46.76, 47.34, 50.24, 45.32,
183 44.06 and 41.07%, respectively. As a result, the best EDA concentration for the carbon modification is
184 5.0 mol L⁻¹.

185 The reason for the improvement in the catalytic ability of the carbon samples treated with EDA
186 solution may be given according to the change of its surface characteristics.

187 The carbon samples were detected by FTIR to analyze their surface chemistry. The FTIR
188 transmission spectrum of the original carbon and the one soaked in 5.0 mol L⁻¹ EDA solution are shown in
189 Fig.3. According to the literatures(Allwar 2012; Chen et al. 2014), the absorption peak observed at 3430
190 cm⁻¹ in the FTIR spectra is resulted from hydroxyl O-H and adsorbed H₂O on the carbon surface. The
191 band exhibited at 2919 cm⁻¹ is ascribed to the stretching vibration of the hydrocarbon single bond C-H.
192 The carbonyl absorption peak from lactonic and carboxyl is exhibited at 1625 cm⁻¹. The peak at 1089
193 cm⁻¹ is due to the phenolic -OH group and C-O group. It can be seen from Fig.3 that the overall shapes of
194 these two spectra are very similar, which suggests they hold the similar chemical characteristics. The
195 strength of the peaks at 3430 cm⁻¹ and 1089 cm⁻¹ become weaker after the carbon has been treated with
196 EDA solution. The summary that there is no new functional group produced and the amount of hydroxyl
197 as well as the content of ester group or ether bonds decrease can be made.



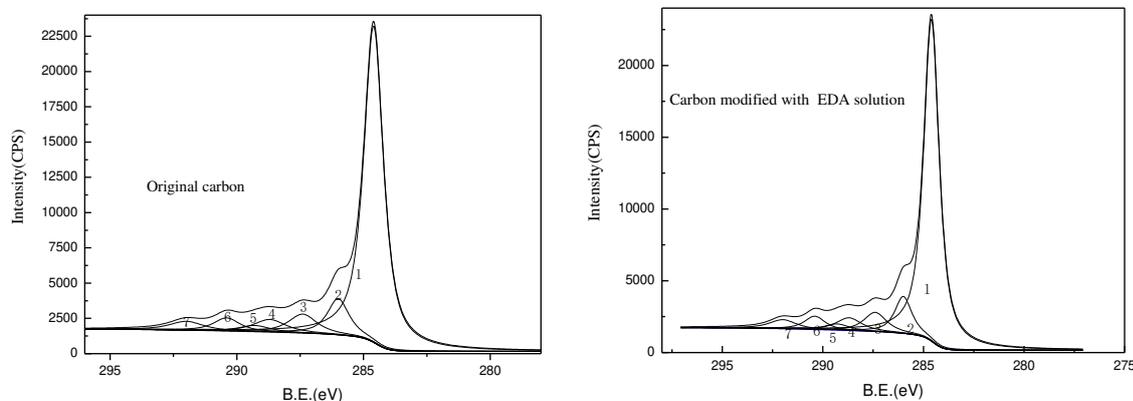
198

199

Fig.3 FTIR spectra of original carbon and modified carbon with EDA

200 XPS characterization has been made for the original one and the one immersed in 5.0 mol L⁻¹ EDA
201 solution. The XPS C 1s spectra shown in Fig.4 are resolved into several individual component peaks.
202 The molar percentages based on the peak resolution are illustrated in Table 1. The data listed in Table 1
203 indicates that the molar percentage of graphitic carbon on the carbon surface increases from 70.07% to
204 71.70% after being modified with EDA solution. The percentage of the carbon in phenolic, alcohol or
205 ether groups decreases slightly (from 8.97% to 8.47%) after the carbon treated by EDA solution. The
206 carbon in carbonyl or quinone groups reduces imperceptibly from 7.41% to 7.22%. The C in lactonic or

207 nitrogenous group drops from 5.16% to 4.75%. The molar percentage of carboxyl decreases sharply
 208 from 2.31% to 0.92%. XPS C 1s spectra depicts that the EDA treatment ameliorates the π structure on
 209 the carbon.



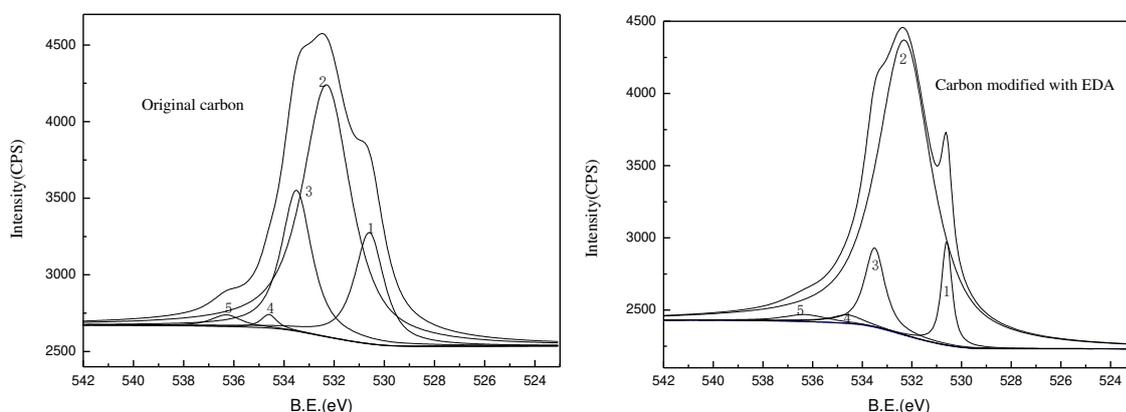
210
 211
 212

Fig. 4 C1s spectra of the original carbon and the modified carbon

Table 1 Fitted C1s peak parameters deduced from XPS for carbon samples

Peaks	B.E (eV)	Assignment	Original carbon (%)	Carbon modified (%)
1	284.6	C=C	70.07	71.70
2	286.0	C-OH, C-O-C, C-O-R	8.97	8.48
3	287.4	C-N, C=O	7.41	7.22
4	288.7	C=N-, -N=C-O-, COOC-	5.16	4.75
5	289.3	COOH	2.31	0.92
6	290.4	$\pi \rightarrow \pi^*$ shake-up satellite	3.54	3.97
7	292.0	Plasmon	2.54	2.96

213



214

Fig. 5 O1s spectra of the original carbon and the modified carbon

215
 216
 217
 218

The XPS O 1s spectra depicted in Fig. 8 are resolved into five individual component peaks. The molar percentages based on the peak resolution are listed in Table 2. The data illustrated in Table 2 reveals that C=O bond in the carboxyl group decreases sharply from 16.80% to 5.64% after the EDA

219 treatment. The molar percentage of C=O bond in amide groups on the carbon modified is 23.08% above
 220 that on the original one. The oxygen in ether group decreases from 21.07% to 8.31%. The functional
 221 groups that peak 4 stands for on B1 are 1.41% while those on AC are only 1.06%. In terms of XPS data, it
 222 can be concluded that acidic groups, such as carboxyl, carbonyl and lactonic group decrease sharply due to
 223 the EDA treatment. The amide groups increases obviously.

224 **Table 2 Fitted O1s peak parameters deduced from XPS for carbon samples**

225

Peaks	B.E (eV)	Assignment	Original carbon (%)	Carbon modified (%)
1	530.6	C=O(Carboxyl)	16.80	5.64
2	532.3	C=O(ester, amides)	59.96	83.04
3	533.5	C-O-C(ether oxygen)	21.07	8.31
4	534.6	C-OH, -COOH, N-O-C	1.06	1.41
5	536.3	H ₂ O _{ads} , O _{2ads}	1.71	1.60

226
 227 The concentrations of the surface functional groups determined by Boehm titration listed in Table 3
 228 may also account for the improvement of the carbon catalytic ability. It can be seen that the total acidic
 229 groups and basic groups have been changed greatly after being modified with EDA solution. The total
 230 basic groups on the modified carbon are significantly raised while the total acidic groups are reduced
 231 compared to those on the original one. For instance, the total basic groups on the original carbon are only
 232 4.02×10^{-4} mol g⁻¹ but those on the one immersed in 6.0 mol L⁻¹ EDA solution rise to 7.18×10^{-4} mol g⁻¹.
 233 The phenolic hydroxyl on this modified carbon is reduced by 81.25% compared with that on the
 234 unmodified one. The carboxylic also drops greatly from 1.50×10^{-4} mol g⁻¹ to 0.11×10^{-4} mol g⁻¹. And the
 235 amount of lactonic on this modified carbon is nearly one half times that on the original one. The physical
 236 characteristics of the carbon samples listed in Table 4 suggest that the modification with EDA solution
 237 gives rise to the increase in total surface area, mesopore area and micropore area on the carbon surface.
 238 For instance, the total surface area on the original carbon is 779 m² g⁻¹, but that on the one impregnated in
 239 5.0 mol L⁻¹ EDA solution is 813 m² g⁻¹. The micropore area on this modified carbon increases by 26 m²
 240 g⁻¹ compared with that on the original one. The reason for the magnification of both S_{BET} and S_{mic} is that
 241 ethylenediamine can etch the activated carbon and remove the ash in the pores. Furthermore,
 242 ethylenediamine as well as the acidic groups on the carbon surface decompose at high temperature, which
 243 may bring about an increase in the pore structure and surface area on the carbon surface. The
 244 modification with EDA solution not only increases the basic functional groups on the carbon surface but
 245 also amplifies the pore structure of the carbon, which helps to adsorb Fe(III)NTA and disintegrate it into
 246 Fe³⁺ and NTA³⁻, accelerating the reduction of Fe(III). Therefore, the carbon modified with EDA is
 247 superior to the original one as a catalyst in the regeneration of Fe(III)NTA.

248 Table 3 also manifests that the total basic groups goes up gradually while the total acid groups go
 249 down slightly as the EDA concentration increases from 3.0 mol L⁻¹ to 6.0 mol L⁻¹. This may be because
 250 more ethylenediamine is adsorbed onto the activated carbon and reacts with the acidic groups on the carbon
 251 surface as EDA concentration increases, which is beneficial to the formation of basic groups at high
 252 temperature.

253 **Table 3 Chemical characteristic of carbon samples (EDA concentration)**

EDA concentration	Phenolic hydroxyl 10 ⁻⁴ mol g ⁻¹	Carboxylic 10 ⁻⁴ mol g ⁻¹	Lactonic 10 ⁻⁴ mol g ⁻¹	Total acid 10 ⁻⁴ mol g ⁻¹	Total basic 10 ⁻⁴ mol g ⁻¹	pH _{pzc}
0 mol L ⁻¹	1.28	1.50	0.45	3.23	4.02	8.12
3.0 mol L ⁻¹	0.25	0.12	0.35	0.72	6.95	10.25
5.0 mol L ⁻¹	0.25	0.11	0.24	0.60	7.09	10.60
6.0 mol L ⁻¹	0.24	0.11	0.24	0.59	7.18	10.69

254 The data in Table 4 reveals that the total surface area and micropore area magnify by 10 m²/g and 17
 255 m²/g, respectively with the EDA concentration rising from from 3.0 mol L⁻¹ to 5.0 mol L⁻¹. However,
 256 when the EDA concentration increases from 5.0 mol L⁻¹ to 6.0 mol L⁻¹, the total surface area and micropore
 257 area decrease by 8 m²/g and 10 m²/g, respectively. Therefore, the S_{BET} of the carbon will increase with the
 258 EDA concentration because more pores will be produced due to the reaction bwtween EDA and carbon
 259 when the samples are calcined at high temperature. However, as the EDA concentration increases over 5
 260 mol L⁻¹, S_{BET} decreases because some micropores will be transformed into mesopores and macropores due
 261 to the violent reaction between carbon and EDA.

262 **Table 4 Physical characteristics of carbon samples (EDA concentration)**

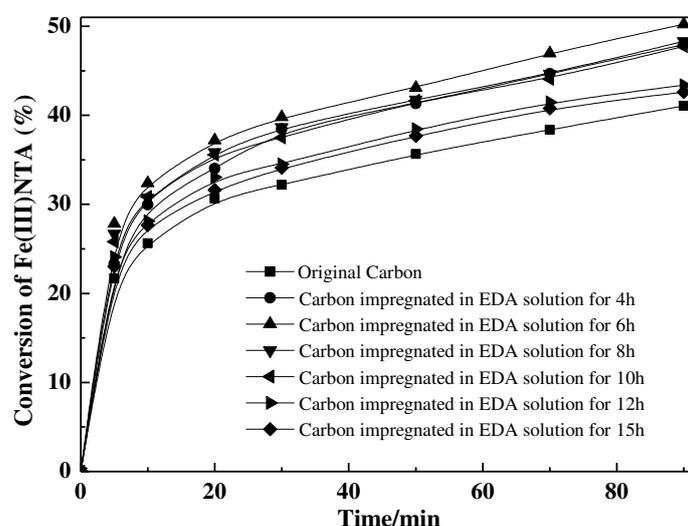
EDA concentration /(mol/L)	S _{BET} (m ² /g)	S _{mic} (m ² /g)	S _{ext} (m ² /g)	V _t (m ³ /g)	V _{mic} (m ³ /g)	D _{B,H} (nm)
0	779	679	100	0.376	0.061	2.143
3.0	803	688	115	0.431	0.072	2.146
5.0	813	705	108	0.440	0.070	2.142
6.0	805	695	110	0.430	0.069	2.165

263 S_{BET} and S_{mic} of the carbon soaked in 5.0 mol L⁻¹ EDA solution are bigger than those of the carbon
 264 soaked in 6.0 mol L⁻¹ EDA solution and the basicity of the former is slightly weaker than that of the latter.
 265 Greater S_{BET} and S_{mic} are conducive to the catalytic ability of the carbon in the regeneration of Fe(II)NTA.
 266 Therefore, the carbon soaked in 5.0 mol L⁻¹ EDA solution exhibits stronger catalytic ability than the carbon
 267 soaked in 6.0 mol L⁻¹ EDA solution. S_{BET} of the carbon soaked in 6.0 mol L⁻¹ EDA solution is slightly
 268 smaller than that of the carbon soaked in 3.0 mol L⁻¹ EDA solution. But the basicity of the carbon soaked
 269 in 6.0 mol L⁻¹ EDA solution is much stronger than that of the one soaked in 3.0 mol L⁻¹ EDA solution. At

270 a pH below the isoelectric point of the carbon, the carbon is positively charged and will adsorb
 271 preferentially anionic species(Rodriguez-Reinoso 1998). The higher the pH_{pzc} , the greater the positive
 272 charge density on the carbon surface, which is favorable for the adsorbability of the anionic NTA and
 273 sulfite on activated carbon. Thus, the reduction of Fe(III)NTA is benefited. The basic groups on the
 274 carbon samples play a more important role than their physical structure. Therefore, the carbon soaked in
 275 6.0 mol L⁻¹ EDA solution gets a higher Fe(III)NTA conversion than the one immersed in 3.0 mol L⁻¹ EDA
 276 solution .

277 3.2 Effect of impregnation time

278 The duration of the carbon impregnated in EDA solution is a vital factor influencing the effect of
 279 carbon modification. 2 g activated carbon of 100-120 mesh were immersed in 5.0 mol L⁻¹ EDA solution
 280 at room temperature for 4, 6, 8, 10, 12 and 15 h, respectively. Then the carbon samples were calcined at
 281 800 °C for 4 h in N₂. The Fe(III)NTA conversions catalyzed by these carbon samples are shown in Fig. 6.
 282 A conclusion can be drawn from Fig. 6 that the best impregnation time is 6 h. After 90 min's reaction, the
 283 Fe(III)NTA conversions obtained by the carbon impregnated in EDA solution for 4, 6, 8, 10, 12 and 15 h
 284 were increased by 3.47%, 9.17%, 7.24%, 6.66%, 2.31% and 1.54%, respectively, compared with that of
 285 41.07% obtained by the original carbon. The carbon soaked in EDA solution for 6 h exhibits the best
 286 catalytic activity.



287
 288 **Fig. 6 Effect of impregnation time on the catalytic performance of activated carbon**

289 [Fe(III)NTA]=0.01 mol l⁻¹, SO₃²⁻=0.1 mol l⁻¹, pH=6.17, 300 rpm, 70°C

290 The data in Table 5 reveals that the total acidic groups decrease from 0.80×10⁻⁴ mol g⁻¹ to 0.60×10⁻⁴
 291 mol g⁻¹ as the impregnation time prolongs from 4 h to 6 h. But if the impregnation time extends to 8 h, the
 292 amount of acidic functional groups is almost unchanged. The total basic groups increase with the
 293 impregnation time because the reaction between carbon and EDA is of benefit to the enhancement of the
 294 basicity of the carbon. The physical characteristics shown in Table 6 indicate that S_{BET} decreases from
 295 816 m² g⁻¹ to 705 m² g⁻¹ with the extension of the impregnation time from 4 h to 8 h.. This may be

296 because the long reaction time between carbon and EDA turns some micropores into mesopores and
 297 macropores.

298 Though the carbon soaked in the EDA solution for 4 h has the biggest S_{BET} , it exhibits the weakest
 299 catalytic ability in the regeneration of Fe(II)NTA^- because it holds the smallest basic groups on its surface.
 300 The basic groups on the carbon soaked in the EDA solution for 8 h are slightly more than those on the one
 301 immersed in the EDA solution for 6 h but the latter holds bigger S_{BET} than the former. Therefore, the
 302 carbon soaked in the EDA solution for 6 h can obtain a higher Fe(III)NTA conversion than the one soaked
 303 in the EDA solution for 8 h. Both the physical structure and surface chemistry of the activated carbon
 304 determine its catalytic ability in the reduction of Fe(III)NTA jointly.

305 **Table 5 Chemical characteristic of carbon samples (Impregnation time)**

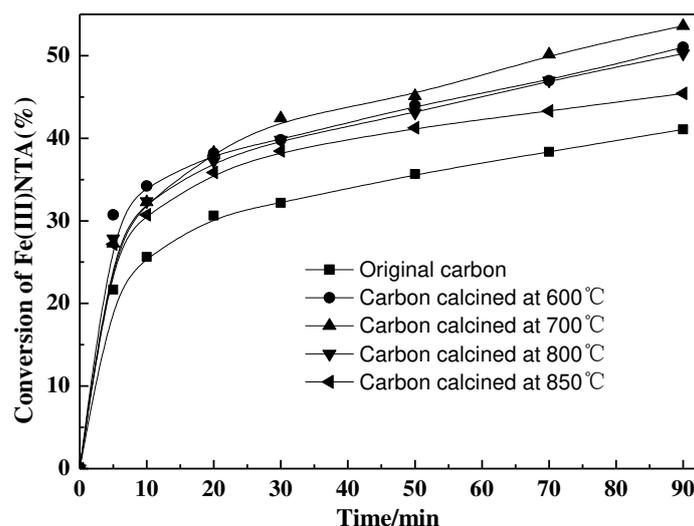
Impregnation time	Phenolic hydroxyl 10^{-4} mol/g	Carboxylic 10^{-4} mol/g	Lactonic 10^{-4} mol/g	Total acid 10^{-4} mol/g	Total basic 10^{-4} mol/g	pH_{pzc}
4 h	0.28	0.15	0.37	0.80	7.03	10.40
6 h	0.25	0.11	0.24	0.60	7.09	10.60
8 h	0.23	0.13	0.26	0.62	7.13	10.69

306 **Table 6 Physical characteristics of carbon samples (Impregnation time)**

Impregnation time	S_{BET} m^2/g	S_{mic} m^2/g	S_{ext} m^2/g	V_{t} m^3/g	V_{mic} m^3/g	D_{BJH} nm
4 h	816	703	113	0.438	0.071	2.146
6 h	813	705	108	0.440	0.070	2.165
8 h	791	682	109	0.427	0.070	2.159

307 3.3 Effect of calcination temperature

308 To explore the effect of calcination temperature on the catalytic capacity of the activated carbon, four
 309 carbon samples were heated in N_2 for 4 h at 600, 700, 800, and 850 $^\circ\text{C}$, respectively after having been
 310 impregnated in 5.0 mol L^{-1} EDA solution for 6 h at ambient temperature. The prepared samples were used
 311 to catalyze the reduction of Fe(III)NTA at 70 $^\circ\text{C}$. The conversions of Fe(III)NTA presented in Fig. 7 prove
 312 that the best calcination temperature for the carbon modification is 700 $^\circ\text{C}$. After 90 min's reaction, the
 313 Fe(III)NTA conversion obtained increases from 51.01 to 53.62% as the calcination temperature is raised
 314 from 600 to 700 $^\circ\text{C}$. However, when the temperature rises further to 800 $^\circ\text{C}$, the Fe(III)NTA conversion
 315 drops to 50.24%.



316
317 **Fig. 7 Effect of calcination temperature on the catalytic performance of activated carbon**

318 [Fe(III)NTA]=0.01 mol l⁻¹, SO₃²⁻=0.1 mol l⁻¹, pH=6.17, 300 rpm, 70 °C

319 Table 7 presents the chemical functional groups of the carbon samples impregnated in 5.0 mol L⁻¹
320 EDA solution for 6 h followed by being carbonized for 4 h at 600, 700, and 800 °C, respectively. It can
321 be seen that the total acidic groups and the total basic groups are almost unchanged when calcinating the
322 carbon samples at 600 and 700 °C. The total basic groups decrease from 7.20 ×10⁻⁴ mol g⁻¹ to 7.01×10⁻⁴
323 mol g⁻¹ when the calcining temperature rises to 800 °C. This may be because the acidic functional groups
324 on the carbon surface such as carboxyl, lactone and phenolic hydroxyl have been decomposed sufficiently
325 above 600 °C and if the calcination temperature increases to 800 °C, some basic functional groups begin
326 to decompose. The physical characteristics illustrated in Table 8 exhibit that the total surface area,
327 mesopore area and micropore area of the carbon samples increase gradually with the calcination
328 temperature rising from 600 to 800 °C because more micropores and mesopores are produced due to the
329 reaction between EDA and carbon proceeding more violently at higher temperature.

330 In spite of its biggest S_{BET}, the carbon calcined at 800 °C gets the lowest Fe(III)NTA conversion
331 because it holds less basic groups than the other two samples. The carbon calcined at 700 °C holds similar
332 chemistry characteristics with the carbon calcined at 600 °C, the former is superior to the latter as a catalyst
333 in the regeneration of Fe(II)NTA⁻ because the former owns larger surface area.

334 **Table 7 Chemical characteristics of carbon samples (calcination temperature)**

Activation temperature	Phenolic hydroxyl 10 ⁻⁴ mol g ⁻¹	Carboxylic 10 ⁻⁴ mol g ⁻¹	Lactonic 10 ⁻⁴ mol g ⁻¹	Total acid 10 ⁻⁴ mol g ⁻¹	Total basic 10 ⁻⁴ mol g ⁻¹	pH _{pzc}
600 °C	0.25	0.12	0.23	0.6 0	7.20	10.83
700 °C	0.22	0.11	0.25	0.58	7.20	10.78
800 °C	0.25	0.11	0.24		7.01	10.60

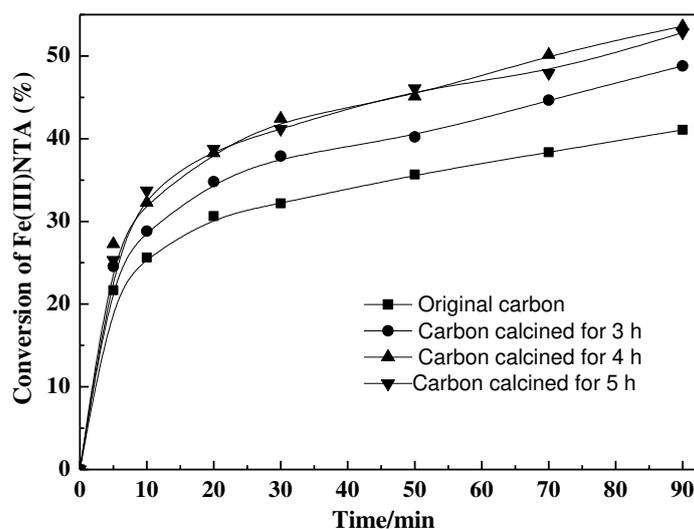
335

Table 8 Physical characteristics of carbon samples (calcination temperature)

Activation temperature	S _{BET} m ² /g g ⁻¹	S _{mic} m ² g ⁻¹	S _{ext} m ² g ⁻¹	V _t m ³ g ⁻¹	V _{mic} m ³ g ⁻¹	D _{B,H} nm
600 °C	789	696	93	0.426	0.064	2.157
700 °C	800	702	98	0.417	0.067	2.149
800 °C	813	705	108	0.440	0.070	2.165

336 3.4 Effect of calcination time

337 The effect of calcination time on the catalytic capability of activated carbon should also be
 338 investigated. Three carbon samples were calcined in N₂ at 700 °C for 3, 4, and 5 h, respectively after
 339 being immersed in 5.0 mol L⁻¹ EDA solution for 6 h at ambient temperature. Then the obtained samples
 340 were used to speed up the reduction of Fe(III)NTA at 70 °C. The conversions of Fe(III)NTA presented in
 341 Fig. 8 reveals that the optimal calcination time for the carbon modification is 4 h. After 90 min's operation,
 342 the Fe(III)NTA conversion catalyzed by the carbon carbonized for 4 h is 53.62% while those catalyzed by
 343 the samples carbonized for 3 and 5 h are 48.79 and 52.84%, respectively.



344

Fig. 8 Effect of calcination time on the catalytic performance of activated carbon

345

[Fe(III)NTA]=0.01 mol l⁻¹, SO₃²⁻=0.1 mol l⁻¹, pH=6.17, 300 rpm, 70 °C

346

347 The data listed in Table 9 depicts that the total acidic groups and the total basic groups on the surface
 348 of activated carbon change little as the calcination time prolongs from 3 h to 5 h. But the physical
 349 characteristics shown in Table 10 indicates that when the calcination time is extended from 3 h to 4 h, the
 350 total surface area and micropore area increase from 782 m² g⁻¹ and 674 m² g⁻¹ to 800 m² g⁻¹ and 702 m² g⁻¹,
 351 respectively. This is because appropriate extension of the calcination duration is favorable for the
 352 formation of micropores resulted from the reaction between carbon and EDA. The total surface area
 353 decreases to 791 m² g⁻¹ if the calcination time prolongs further to 5 h. The reason may be that excessive

354 calcination time leads to the transformation of some mesopores to macropores. The Fe(III)NTA
 355 conversions they have got are in accordance with the sequence of their physical structures. Therefore, the
 356 best calcination time is selected to be 4 h.

357

Table 9 Chemical characteristics of carbon samples (calcination time)

Activation time	Phenolic hydroxyl 10 ⁻⁴ mol/g	Carboxylic 10 ⁻⁴ mol/g	Lactonic 10 ⁻⁴ mol/g	Total acid 10 ⁻⁴ mol/g	Total basic 10 ⁻⁴ mol/g	pH _{pzc}
3h	0.27	0.10	0.22	0.59	7.21	10.80
4 h	0.22	0.11	0.25	0.58	7.20	10.78
5 h	0.21	0.11	0.25	0.57	7.18	10.69

358

Table 10 Physical characteristics of carbon samples (calcination time)

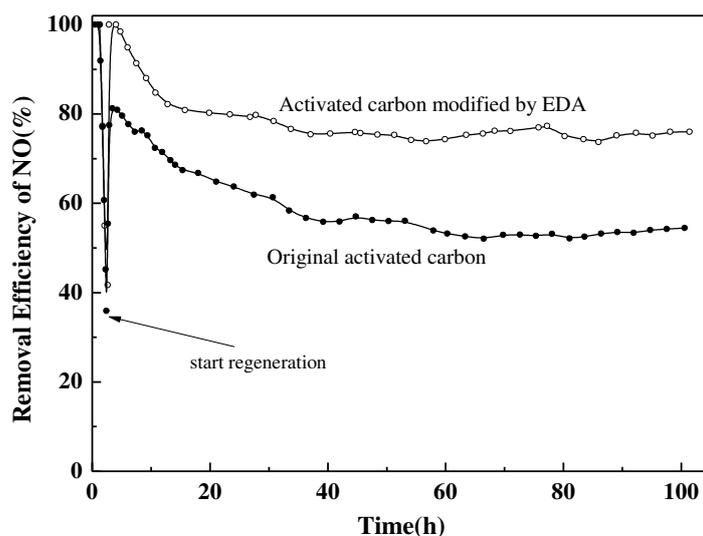
Activation time	S _{BET} (m ² /g)	S _{mic} (m ² /g)	S _{ext} (m ² /g)	V _t (m ³ /g)	V _{mic} (m ³ /g)	D _{BJH} (nm)
3 h	782	674	108	0.422	0.072	2.157
4 h	800	702	98	0.417	0.067	2.149
5 h	791	704	87	0.426	0.060	2.154

359 Simultaneous removal of NO and SO₂

360 The experimental schematic apparatus of the simultaneous removal of NO and SO₂ with Fe(II)NTA⁻
 361 solution as well as the Fe(II)NTA⁻ regeneration catalyzed by the raw carbon and the modified carbon is
 362 shown in Fig. 1. The modified carbon was treated with EDA solution under the optimal condition
 363 discussed previously. In the experiment, both of the absorber and the regeneration reactor are controlled
 364 at 50°C. The absorption solution is a mixture of 500 mL 0.05 mol L⁻¹ Fe(II)NTA⁻ and 0.04 mol L⁻¹
 365 Na₂SO₃. The concentrations of SO₂ and NO in the gas inlet are 1800 ppm and 580 ppm, respectively.
 366 O₂ in the simulated flue gas is 5.0% (volume). The gas flow rate is 270 ml min⁻¹ and that of the scrubbing
 367 solution is controlled by a peristaltic pump at 25 ml min⁻¹. The pH value of the solution in the circulation
 368 tank is controlled around 5.5. The experimental results are shown in Fig. 9.

369 It can be seen from Fig. 9 that the removal efficiency of NO decreases from 100% to about 35.9%
 370 after 2.4 h run due to the consumption of Fe(II)NTA⁻. At this time, the Fe(II)NTA⁻ regeneration is started.
 371 Obviously, the NO removal efficiency regain quickly in these two operations after the regeneration process
 372 begins. But there is evident difference between two operations. The NO removal efficiency rises to
 373 about 81.3% in 1 h and then begins to decrease gradually under the catalysis of the unmodified carbon.
 374 After 58 h operation, the NO removal efficiency drops to 57.9% and is maintained at 52.1-54.4%. As a
 375 contrast, when Fe(II)NTA⁻ regeneration is catalyzed by the modified carbon, the NO removal efficiency

376 reaches to 100% in about 15 min and is maintained at 100 % for 1.2 h. Thirty hours later, the NO removal
377 efficiency decreases to 78.4% and fluctuates between 73.7% and 77.3% in the run. Besides, there is no
378 SO₂ detected in the outlet gas by FTIR in the whole process. The explanation for such phenomenon is as
379 follows. Fe(II)NTA⁻ is gradually oxidized to Fe(III)NTA by oxygen, leading to the reduction in NO removal
380 efficiency. After the Fe(II)NTA⁻ regeneration starts, Fe(III)NTA is reduced by sulfite/bisulfite to
381 Fe(II)NTA⁻ under the catalysis of activated carbon and Fe(II)NTA(NO)⁻ also reacts with sulfite/bisulfite to
382 form Fe(II)NTA and N₂. As the operation goes on, the sulfite consumed is balanced with the
383 supplementary sulfite/bisulfite by SO₂ dissolving into the scrubbing liquor. Therefore, the NO removal
384 efficiency remains constant during the whole experiment. The experiments indicate that the modified
385 activated carbon can gain a higher NO removal efficiency than the original activated carbon. As a
386 consequence, the carbon modification with EDA solution is a cogent measures to ameliorate its catalytic
387 capability in the simultaneous elimination of NO and SO₂ with Fe(II)NTA⁻ solution.



388
389 **Fig. 8 NO removal coupled with Fe(II)NTA⁻ regeneration catalyzed by activated carbon**

391 Conclusions

392 Ethylenediamine solution has been used to modify activated carbon, and the following conclusions are
393 obtained from the experiments:

394 (1) The catalytic capability of the activated carbon in the regeneration of Fe(II)NTA⁻ is ameliorated
395 conspicuously by treating the carbon with EDA solution. The best modification conditions are
396 impregnating carbon in 5.0 mol L⁻¹ EDA solution for 6 h followed by carbonizing the sample at 700 °C for
397 4 h in N₂.

398 (2) The carbon surface characterization demonstrates that the treatment with EDA solution gives rise to an

399 evident increase in the basic groups and obvious decrease in acidic groups on the carbon surface. The
400 BET results prove that the modification also brings about a slight increase in the surface area. And these
401 changes are favorable for the improvement of the catalytic activity of the activated carbon in the generation
402 of Fe(II)NTA⁻. The catalytic ability of activated carbon in the Fe(II)NTA⁻ generation relies on its physical
403 structure and surface chemistry. The surface chemistry plays more important role than its physical
404 structure in determining the catalytic capability of carbon.

405 (3) The modified coconut activated carbon can achieve a much higher NO removal efficiency than the
406 unmodified coconut activated carbon. Therefore, this modification with EDA solution is an effective way
407 to enhance the catalytic ability of the activated carbon in the simultaneous removal of NO and SO₂ with
408 Fe(II)NTA⁻ solution.

409

410 **References**

411 Allwar A(2012) Characteristics of pore structures and surface chemistry of activated carbons by
412 physisorption, Ftir and Boehm methods. IOSR J Applied Chem 2(1): 9-15.

413 Boehm HP(1994) Some aspects of the surface chemistry of carbon blacks and other carbons. Carbon 32(5):
414 759-769.

415 Chandrashekhar B, Pai P, Morone A, Sahu N, Pandey RA(2013) Reduction of NO_x in Fe-EDTA and
416 Fe-NTA solutions by an enriched bacterial population. Bioresource technology 130: 644-651.

417 Chandrashekhar B, Sahu N, Tabassum H, Pai P, Morone A, Pandey RA(2015) Treatment of
418 ferrous-NTA-based NO_x scrubber solution by an up-flow anaerobic packed bed bioreactor. Applied
419 microbiology and biotechnology 99(12): 5281-5293.

420 Chen CJ, Li X, Tong ZF, Li Y, Li MF(2014) Modification process optimization, characterization and
421 adsorption property of granular fir-based activated carbon. Applied Surf Sci 315: 203-211.

422 Chen P, Gu MY, Chen G, Liu, FS, Lin YY(2019) DFT study on the reaction mechanism of N₂O reduction
423 with CO catalyzed by char. Fuel 254: 115666.

424 Dou BL, Pan WG, Jin Q, Wang WH, Li Y(2009) Prediction of SO₂ removal efficiency for wet Flue Gas
425 Desulfurization. Energy Conversion & Management 50(10):2547-2553.

426 Fu RQ, Liu Y, Lou ZM, Wang ZX, Baig SA, Xu XH (2016) Adsorptive removal of Pb (II) by magnetic
427 activated carbon incorporated with amino groups from aqueous solutions. Journal of the Taiwan Institute
428 of Chemical Engineers 62: 247-258.

429 Fu XF, Yang HP, Sun HH, Lu GH, Wu JM(2016) The multiple roles of ethylenediamine modification at
430 TiO₂/activated carbon in determining adsorption and visible-light-driven photoreduction of aqueous Cr (VI).
431 J Alloys and Compounds 662: 165-172.

432 Guo LN, Han CY, Zhang SL, Zhong Q, Ding J, Zhang BQ, Zeng YQ(2018) Enhancement effects of ·O₂⁻
433 and ·OH radicals on NO_x removal in the presence of SO₂ by using an O₃/H₂O₂ AOP system with inadequate
434 O₃ (O₃/NO molar ratio= 0.5). Fuel 233: 769-777.

435 Hofele J, van Velzen D, Langenkamp H, Schaber K(1996) Absorption of NO in aqueous solutions of
436 FeIINTA: determination of the equilibrium constant. *Chem Eng and Processing: Process Intensification*
437 35(4): 295-300.

438 Kang MS, Shin J, Tae UY, Hwang J(2020) Simultaneous removal of gaseous NO_x and SO₂ by gas-phase
439 oxidation with ozone and wet scrubbing with sodium hydroxide. *Chem Eng J* 381: 122601.

440 Khan NE, Adewuyi YG(2010) Absorption and Oxidation of Nitric Oxide (NO) by Aqueous Solutions of
441 Sodium Persulfate in a Bubble Column Reactor. *Ind Eng Chem Res* 49: 8749-8760.

442 Liémans I, Thomas D(2013) Simultaneous NO_x and SO_x reduction from oxyfuel exhaust gases using
443 acidic solutions containing hydrogen peroxide. *Energy Procedia* 37:1348-1356.

444 Li ZH, Chang XJ, Zou XJ, Zhu XB, Nie R, Hu Z, Li RJ (2009) Chemically-modified activated carbon with
445 ethylenediamine for selective solid-phase extraction and preconcentration of metal ions. *Analytica Chimica*
446 *Acta* 632: 272-277.

447 Ma SC, Chai J, Jiao KL, Ma L, Zhu SJ, Wu K(2017) Environmental influence and countermeasures for
448 high humidity flue gas discharging from power plants. *Renewable & Sustainable Energy Reviews*
449 73:225-235.

450 Messele A, Soares OSGP, Órfão JJM, Stüber F, Bengoa C, Fortuny A, Fabregat A, Font J(2014)
451 Zero-valent iron supported on nitrogen-containing activated carbon for catalytic wet peroxide oxidation of
452 phenols. *Appl. Catal. B: Environ.* 154-155: 329-338.

453 Mondal MK, Chelluboyana VR(2013) New experimental results of combined SO₂ and NO removal from
454 simulated gas stream by NaClO as low-cost absorbent. *Chem Eng J* 217: 48-53.

455 O'reilly JM, Mosher RA(1983) Functional groups in carbon black by FTIR spectroscopy. *Carbon* 21(1):
456 47-51.

457 Pan W, Zhang X, Guo R, Zhou Y, Jin Q, Ren J(2015) A thermodynamic study of simultaneous removal of
458 SO₂ and NO by a KMnO₄/ammonia solution. *Energy Sources, Part A: Recovery, Utilization, and*
459 *Environmental Effects* 37(7): 721-726.

460 Ravishankara AR, Daniel JS, Portmann RW(2009) Nitrous oxide (N₂O): the dominant ozone-depleting
461 substance emitted in the 21st century. *Science* 326(5949): 123-125.

462 Rodriduez-Reinoso F(1998) The role of carbon materials in heterogeneous catalysis. *Carbon* 36(3): 159
463 -175.

464 Saikawa E, Kim H, Zhong M, Avramov A, Zhao Y, Janssens-Maenhout G, Kurokawa J, Klimont Z, Wagner
465 F, Naik V(2017) Comparison of emissions inventories of anthropogenic air pollutants and greenhouse
466 gases in China. *Atmospheric Chemistry & Physics* 17(10): 6393-6421.

467 Tang MH, Tu SP, Sun WZ(2020) Research progress of flue gas desulfurization and denitrification
468 technology in coal-fired power plants. *Scientific Journal of Intelligent Systems Research* 2(10):71-79.

469 Toro RA, Donoso CS, Seguel RA, Morales RGES, Leiva MAG(2014) Photochemical ozone pollution in
470 the Valparaiso Region, Chile. *Air quality, atmosphere & health: An international journal* 7(1): 1-11.

471 Wang ZH, Liu, HJ, Fang ZW, Zhou, XZ, Long XL(2019) Production of isophthalic acid from M-Xylene
472 catalyzed by Co(II) and HPW@C modified with ZnCl₂ solution. *Canadian Chem Eng* 97(7): 2086-2096.

473 Wolak M, van Eldik R(2002) To be or not to be NO in coordination chemistry? A mechanistic approach.
474 Coordination chemistry reviews 230(1-2): 263-282.
475 Yan JR, Zhou FX, Zhou Y, Wu, XH, Zhu QL, Liu HY, Lu HF(2018) Wet oxidation and absorption
476 procedure for NO_x removal. Environ Tech Innovation 11: 41-48.
477 Zhang CM, Song W, Ma QL, Xie LJ, Zhang XC, Guo H(2016) Enhancement of CO₂ capture on
478 biomass-based carbon from black locust by KOH activation and ammonia modification. Energy & Fuels
479 30(5): 4181-4190.
480 Zhang D, Huo P, Liu W(2016) Behavior of phenol adsorption on thermal modified activated carbon.
481 Chinese J Chem Eng 24(4): 446-452.
482 Zheng CH, Xu, CR, Zhang YX, Zhang J, Gao X, Luo ZY, Cen KF(2014) Nitrogen oxide absorption and
483 nitrite/nitrate formation in limestone slurry for WFGD system. Applied Energy 129(15):187-194.
484 Zheng WL, Chen S, Liu HE, Ma, YD, Xu, WL(2019) Study of the modification mechanism of heavy
485 metal ions adsorbed by biomass-activated carbon doped with a solid nitrogen source. RSC Advances 9(64):
486 37440-37449.
487 Zhu HS, Mao YP, Yang XJ, Long XL, Yuan WK(2010) Simultaneous absorption of NO and SO₂ into
488 FeII–EDTA solution coupled with the FeII–EDTA regeneration catalyzed by activated carbon. Separation
489 and Purification Technology 74(1):1-6.

490
491

492 **Declarations**

493 **Ethics approval and consent to participate**

494 Not applicable

495 **Consent for publication**

496 Not applicable.

497 **Availability of data and materials**

498 Not applicable.

499 **Competing interests**

500 The authors declare that they have no competing interests.

501 **Funding**

502 Not applicable.

503 **Authors' contributions**

504 XD and PC analyzed the experimental data, and was a major contributor in writing the manuscript.

505 RZ did the experiments. XL made the experimental plan and finished the manuscript.

506 All authors read and approved the final manuscript.

507
508
509
510

Figures

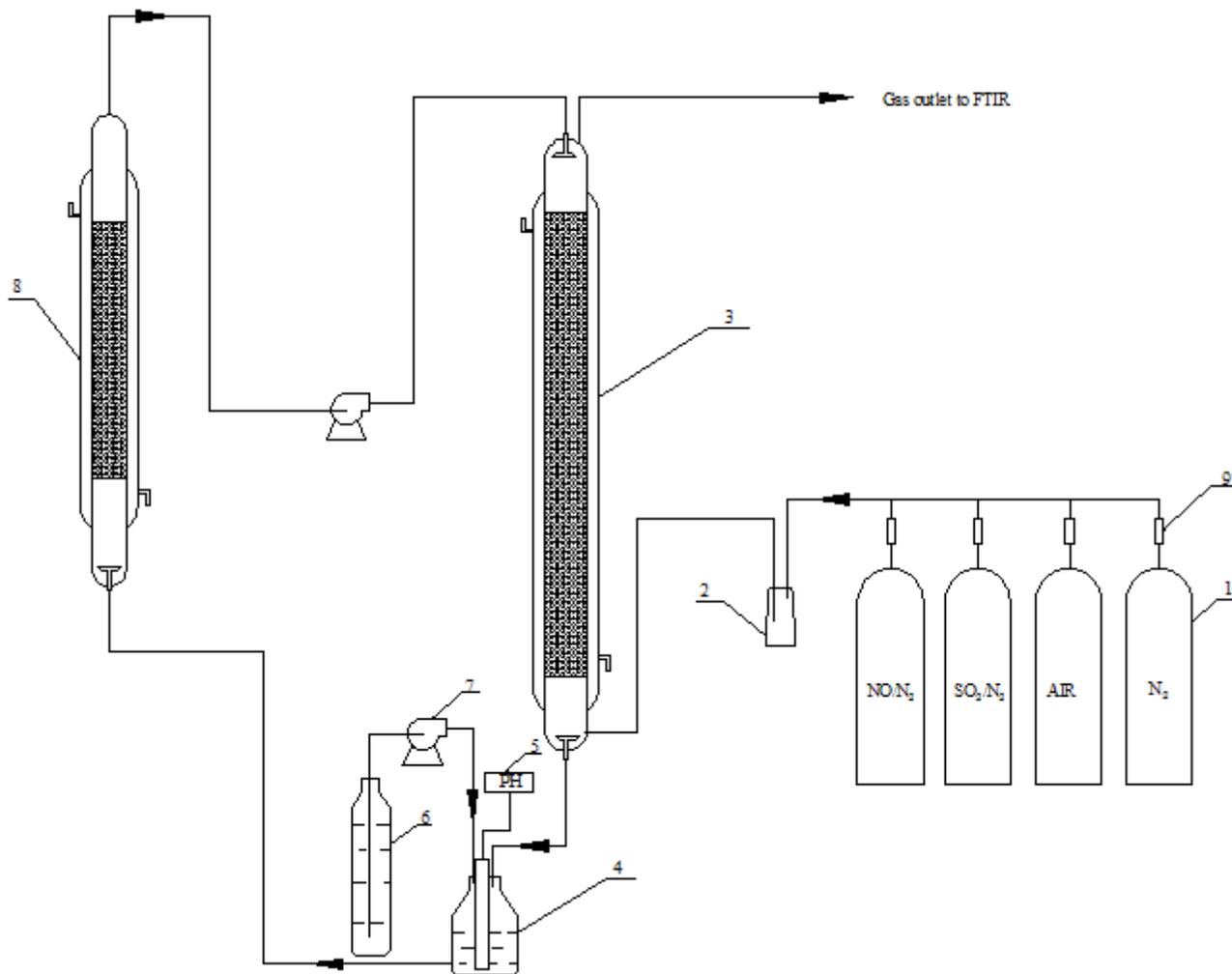


Figure 1

Flowchart of absorption and regeneration reactor system

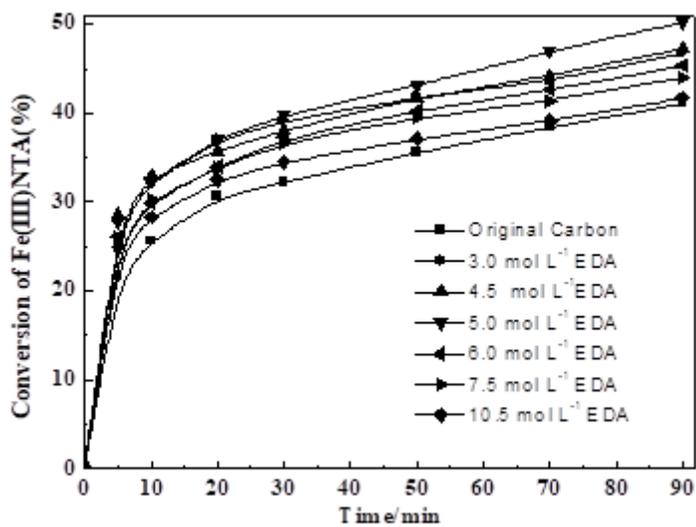


Figure 2

Effect of the EDA concentration on the catalytic performance of carbon samples

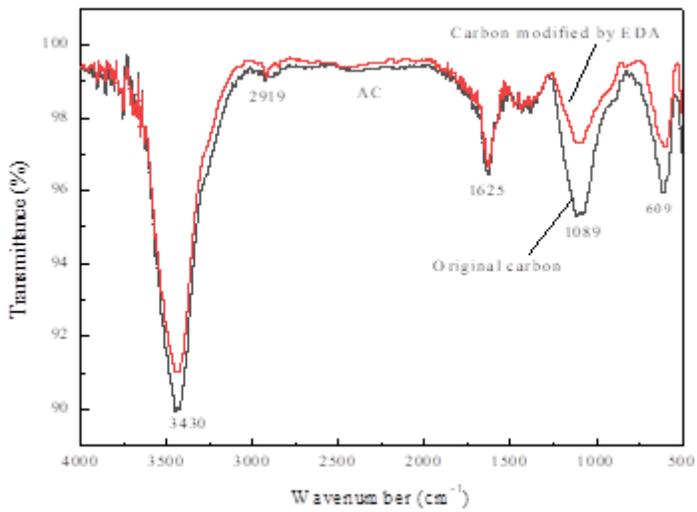


Figure 3

FTIR spectra of original carbon and modified carbon with EDA

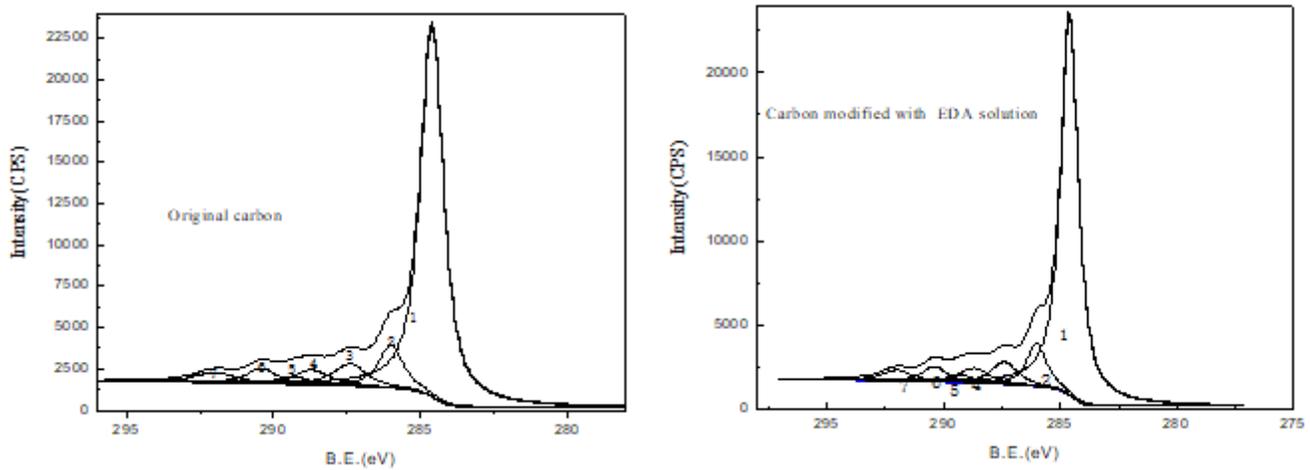


Figure 4

C1s spectra of the original carbon and the modified carbon

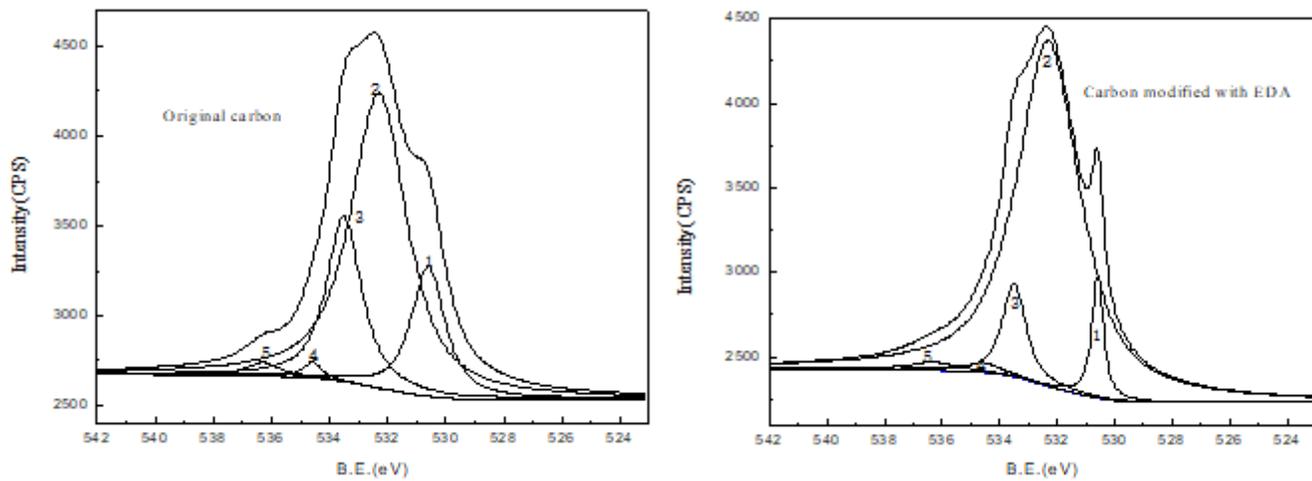


Figure 5

O1s spectra of the original carbon and the modified carbon

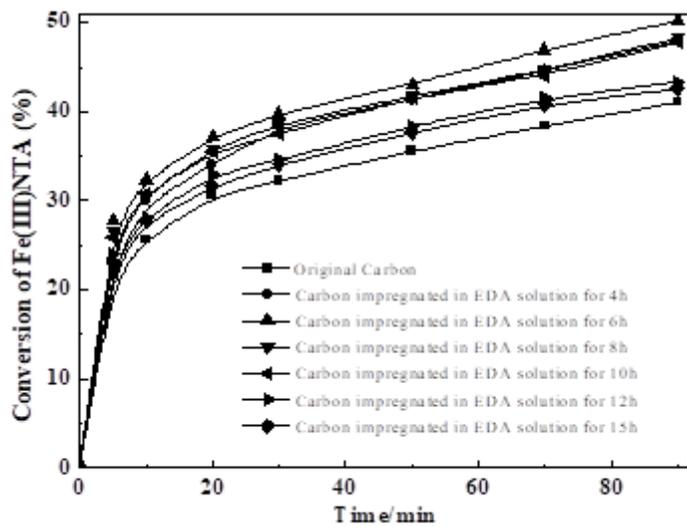


Figure 6

Effect of impregnation time on the catalytic performance of activated carbon

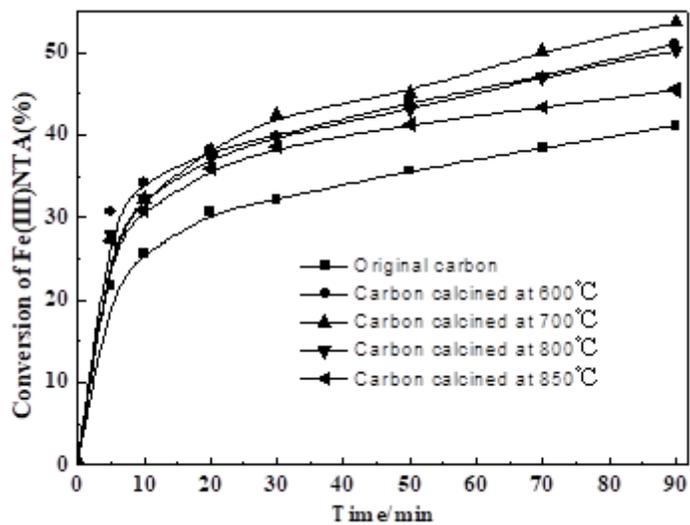


Figure 7

Effect of calcination temperature on the catalytic performance of activated carbon

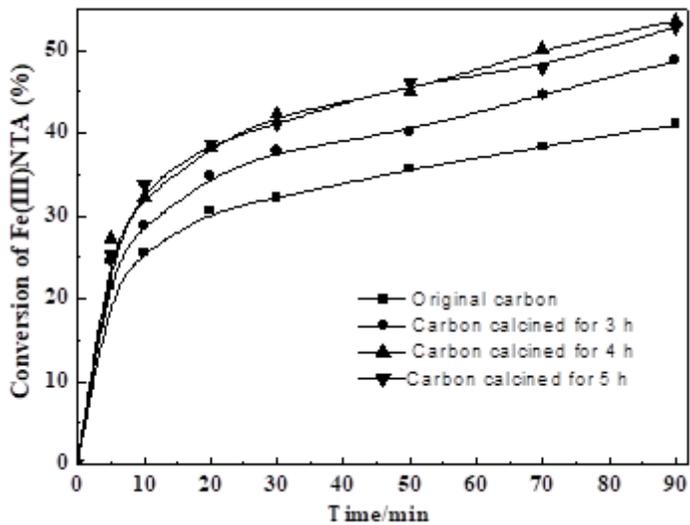


Figure 8

Effect of calcination time on the catalytic performance of activated carbon

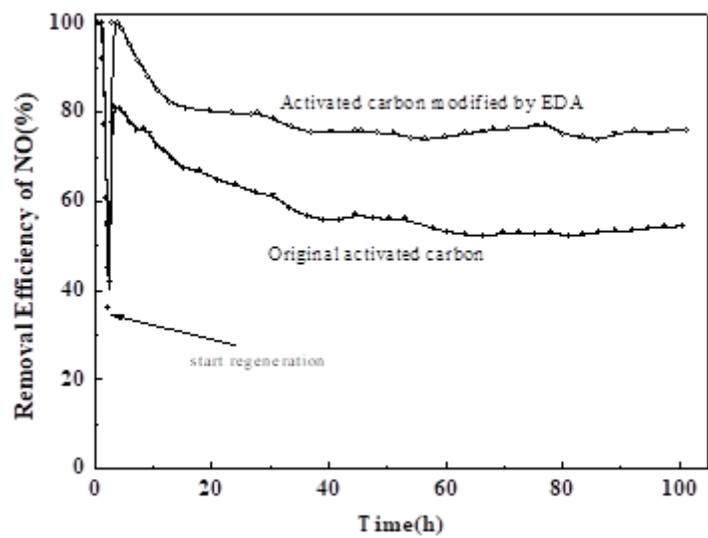


Figure 9

NO removal coupled with Fe(II)NTA- regeneration catalyzed by activated carbon