

Pronounced loss of Amazon rainforest resilience since the early 2000s

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Article

Keywords: global carbon cycle, climate change, biodiversity, carbon storage

Posted Date: April 14th, 2021

DOI: <https://doi.org/10.21203/rs.3.rs-379902/v1>

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Version of Record: A version of this preprint was published at Nature Climate Change on March 7th, 2022.
See the published version at <https://doi.org/10.1038/s41558-022-01287-8>.

1 Pronounced loss of Amazon rainforest resilience since the 2 early 2000s

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13 **The resilience of the Amazon rainforest to climate and land-use change is of critical**
14 **importance for biodiversity, regional climate, and the global carbon cycle. Some**
15 **models project future climate-driven Amazon rainforest dieback and transition to**
16 **savanna¹. Deforestation and climate change, via increasing dry-season length^{2,3} and**
17 **drought frequency – with three 1-in-100-year droughts since 2005⁴⁻⁶ – may already**
18 **have pushed the Amazon close to a critical threshold of rainforest dieback^{7,8}.**
19 **However, others argue that CO₂ fertilization should make the forest more resilient^{9,10}.**
20 **Here we quantify Amazon resilience by applying established indicators¹¹ to remotely-**
21 **sensed vegetation data with focus on vegetation optical depth (1991-2016), which**
22 **correlates well with broadleaf tree coverage. We find that the Amazon rainforest has**
23 **been losing resilience since 2003, consistent with the approach to a critical transition.**
24 **Resilience is being lost faster in regions with less rainfall, and in parts of the**
25 **rainforest that are closer to human activity. Given observed increases in dry-season**
26 **length^{2,3} and drought frequency⁴⁻⁶, and expanding areas of land use change, loss of**
27 **resilience is likely to continue. We provide direct empirical evidence that the Amazon**

28 **rainforest is losing stability, risking dieback with profound implications for**
29 **biodiversity, carbon storage and climate change at a global scale.**

30

31 There is widespread concern about the resilience of the Amazon rainforest to land-use
32 change and climate change. The Amazon is recognised as a potential tipping element in the
33 Earth's climate system¹², is a crucible of biodiversity¹³, and usually acts as a large terrestrial
34 carbon sink¹⁴. The net ecosystem productivity (carbon uptake flux) of the Amazon has,
35 however, been declining over the last four decades and during two major droughts in 2005
36 and 2010, the Amazon temporarily turned into a carbon source, due to increased tree
37 mortality¹⁵⁻¹⁷. Several studies have suggested that deforestation¹⁸ and anthropogenic global
38 warming^{1,5}, especially in combination, could push the Amazon rainforest past critical
39 thresholds^{7,8} where positive feedbacks propel abrupt and substantial further forest loss. Two
40 types of positive feedback are particularly important. First, localised fire feedbacks amplify
41 drought and associated forest loss by destroying trees¹⁹, and the fire regime itself may 'tip'
42 from localised to 'mega-fires'²⁰. Second, deforestation and forest degradation, whether due
43 to direct human intervention or droughts, reduce evapotranspiration and hence the moisture
44 transported further westward, reducing rainfall and forest viability there²¹ and establishing a
45 large-scale moisture recycling feedback. Net rainfall reduction may in turn reduce latent
46 heating over the Amazon to the extent that it weakens the low-level circulation of the South
47 American monsoon¹⁸. Model projections of future changes in the Amazon rainforest differ
48 widely^{1,9,10,22}. Early studies showed that the Amazon rainforest may exhibit strong dieback by
49 the end of the 21st century^{1,23}. Both pronounced drying in tropical South America and a weak
50 CO₂ fertilisation effect⁹ contributed to this result, with dieback also more common under
51 stronger greenhouse gas emission scenarios¹⁰. Other studies based on varying general
52 circulation and vegetation model components show a wider range of results^{24,25}.
53 Nevertheless, the forest may be 'committed' to dieback despite appearing stable at the end
54 of model runs²⁶. This highlights the importance of measuring the changing dynamical

55 stability of the forest alongside its mean state. Given the uncertainty in model projections, we
56 directly analyse observational data for signs of resilience loss in the Amazon.

57 The mean state of a system is not usually informative of changes in resilience; either can
58 change whilst the other remains constant. Thus, higher-order statistical characteristics that
59 respond more sensitively to destabilisation than the mean need to be considered to quantify
60 resilience. To measure the changing resilience of the Amazon rainforest, we use a stability
61 indicator used to predict the approach of a dynamical system towards a bifurcation-induced
62 critical transition. The predictability arises from the phenomenon of critical slowing down^{27,28}
63 (CSD): as the currently occupied equilibrium state of a system becomes less stable, it
64 responds more sluggishly to short-term perturbations (e.g. weather variability for the
65 Amazon). This loss of resilience (defined²⁹ as return rate from perturbation) reflects a
66 weakening of negative feedbacks that maintain stability. The behaviour can be detected by
67 an increase in lag-1 autocorrelation (AR(1)) in time series capturing the system
68 dynamics^{30,31}. It may also manifest as an increase in variance over time, but variance can
69 also be easily influenced by changing variability of the perturbations driving the system³².
70 Increasing AR(1) has been used to detect critical slowing down prior to bifurcation-induced
71 state transitions in a number of systems, including but not limited to climate^{30,33} and
72 ecology³⁴. A caveat, highlighted by analysis of model projections prior to Amazon dieback³²,
73 is that a system should be forced slower than its intrinsic response timescale for CSD to
74 occur (see Methods). Hence, the absence of CSD may not rule out the possibility of a
75 forthcoming critical transition. Conversely, increasing AR(1) can sometimes occur for other
76 physical reasons. A space-for-time substitution has previously revealed that tropical forest
77 resilience as measured by mean AR(1) (on a grid point basis) is lower for less annual rainfall
78 sums¹¹, but changes of Amazon resilience over time have not been investigated so far.

79 We investigate controls on the resilience of the Amazon vegetation system and how its
80 resilience has changed over the last three decades, in terms of a changing AR(1) coefficient
81 as estimated from satellite-derived vegetation data. The main dataset we use is from the

82 Vegetation Optical Depth Climate Archive (VODCA)³⁵, but we also analyse the NOAA
83 Advanced Very-High-Resolution Radiometer's (AVHRR) normalized difference vegetation
84 index NDVI³⁶ for comparison. Vegetation Optical Depth (VOD) has been previously used to
85 estimate changes in vegetation biomass³⁷, whereas NDVI is more commonly used to
86 measure the greenness of vegetation, i.e. photosynthetic activity³⁸. We use the Ku-band
87 product from VODCA, which has a resolution of 0.25°x0.25°, and for direct comparison we
88 rescale the NDVI data to the same resolution. We focus on two stressors of the Amazon
89 that may cause resilience changes – precipitation and human influence.

90 We use the Amazon basin as our study region and focus on those grid boxes which have a
91 broadleaf fraction greater than or equal to 80% evergreen broadleaf (BL) fraction according
92 to the MODIS Land Cover Type product in 2001³⁹ (See Methods). Figure 1 shows that when
93 comparing BL fraction in 2001 to mean annual precipitation (MAP) from 2001-2016 (from
94 CHIRPS⁴⁰; see Methods), there is a clear bimodal region visible between approximately
95 1500-2250mm, which has been reported previously⁴¹⁻⁴³ (Fig.1a). Bi-stability, where a
96 forested or non-forested area can exist under the same MAP, suggests the potential for
97 bifurcation- and noise-induced transitions, the latter potentially triggered by single
98 perturbations such as droughts or fires. Over most of the region, BL fraction has not
99 changed significantly between 2001 and 2016 (Fig. 1b). However, deforestation has
100 occurred along parts of the southern and eastern edges of the rainforest, along parts of the
101 Amazon river, and in some northern parts of the basin.

102 With a very similar spatial pattern as the change in BL fraction (Fig. 1b), we find decreases
103 in VOD around the south-eastern edges of the forest (Fig. 1c). Averaged across the Amazon
104 study region we find overall decreasing VOD, which matches with the observed decrease in
105 the number of grid boxes that have BL \geq 80% each year (Fig. 1d). NDVI, in contrast, does
106 not agree spatially with the changes in BL fraction – rather, NDVI increases in the south-
107 eastern parts of the Amazon where deforestation rates are known to be high (Supplementary
108 Fig. 1). Changes in BL fraction from 2001-2016 are strongly correlated with changes in VOD

109 over the same period (Fig. 1e), whereas changes in NDVI are not (Fig. 1f), echoing previous
110 in-situ comparisons between VOD and NDVI⁴⁴. Hence, we focus our analysis on VOD in the
111 following, with results for NDVI in the Supplementary Figures.

112 We begin our resilience analysis by focusing on the temporal changes of AR(1), computed in
113 sliding windows from the nonlinearly detrended and de-seasonalised VOD time series (see
114 Fig. 2, and Methods). The time series calculated from the mean AR(1) value across our
115 study area each month shows a substantial increase over time, particularly from ~2003 (Fig.
116 2a). The spatial distribution of the AR(1) tendency, measured by the Kendall rank correlation
117 coefficient τ (see Methods) at each grid box, shows that decreases in AR(1) (increases in
118 resilience) are mostly restricted to parts of the region with higher mean annual precipitation
119 (MAP) (Fig. 2b). We also observe stable or decreasing AR(1) values around the tributaries of
120 the Amazon river, where vegetation growth will be less dependent on precipitation for water
121 availability. Overall, the majority (74.6%) of grid boxes show increasing AR(1) values and
122 hence, loss of resilience (Fig. 2c). Using alternative methods of detrending the VOD time
123 series (see Methods) yields similar results (Supplementary Fig. 2). A predominance of
124 increasing AR(1) trends is also found for the NDVI time series since 2003 (Supplementary
125 Fig. 3).

126 To further explore the relationship between MAP and AR(1) trend, we create mean AR(1)
127 time series on a moving MAP-band of 500mm (see Methods). These bands show broadly
128 the same behaviour as the region overall (Fig. 3a), with all bands showing a significant
129 decrease in resilience post-2003 ($p < 0.001$). The increase in AR(1) post-2003 appears least
130 pronounced for the highest rainfall band (3500-4000mm). Sure enough, the intensity of
131 resilience loss increases as the MAP-band decreases below 3500-4000mm (Fig. 3b). For
132 NDVI, the same relationship is also observed (Supplementary Fig. 4a,b). However, due to a
133 large decrease in NDVI AR(1) pre-2003 across the region, analysing the full AR(1) time
134 series yield decreasing AR(1) Kendall τ coefficients for the higher MAP-bands.

135 It has previously been suggested that the forest near human land-use areas is less
136 resilient⁴². To determine if this is shown by VOD, we measure the distance of each grid box
137 from human land use in 2016 (see Methods, Supplementary Fig. 5). Calculating mean AR(1)
138 time series on 50km distance bands, shows increases in AR(1) post-2003 are stronger for
139 grid boxes closer to human land use (Fig. 4a). Grid boxes that are in more remote locations
140 still show a loss of resilience but the AR(1) time series for these are more variable – likely
141 because the area they are averaged over shrinks and becomes more disconnected (Fig. 4a).
142 Above 200-250km away from human land use the signal of loss of resilience becomes less
143 pronounced (Fig. 4b). NDVI time series also show there is a loss of resilience from 2003, in
144 grid boxes that are closer than 200km from human land use (Supplementary Fig. 4c,d).

145 Our results suggest that the loss of resilience of the Amazon rainforest that is especially
146 pronounced since 2003 (Fig. 2), could be due to a combination of changing precipitation
147 patterns (Fig. 3) and changing human interference in the region (Fig. 4). Here we reason that
148 as lower baseline MAP and greater proximity to human interference are both associated with
149 greater loss of resilience, declining MAP and/or increasing human interference may be
150 expected to cause increased resilience loss. We find increases in human land use areas
151 using the MODIS Land Cover data over the time period, both in reach and intensity
152 (Supplementary Fig. 6). However, although there are large parts of the study region with
153 decreasing MAP, by comparing the spatial pattern of MAP decreases to the AR(1) increases
154 (Supplementary Fig. 7), it is unlikely that the changes in MAP are the dominant driver of
155 Amazon rainforest resilience loss. Rather, increases in dry-season length as reported in
156 several recent studies^{2,3,45,46} may explain the loss in vegetation resilience since the early
157 2000s detected here. With a longer study period to measure trends in MAP, it is possible
158 that a stronger correlation between MAP changes and changes in resilience over time may
159 be found.

160

161 The changes in forest resilience observed as increasing AR(1) in both vegetation indices are
162 supported by another indicator of critical slowing down, namely increasing variance³³ – of
163 both VOD (Supplementary Figure 8) and NDVI (Supplementary Figure 9). We note that
164 variance is more strongly affected by changes in the frequency and amplitude of the forcing
165 of a system, and as such results could be biased towards individual events. This, along with
166 other issues, has led AR(1) to be considered a more robust indicator⁴⁷.

167 As emphasized above, changes in BL fraction do not directly relate to changes in resilience.
168 Indeed, we infer a marked loss of resilience in terms of increasing AR(1) in vast areas where
169 the BL fraction does not strongly decrease (compare Figs. 1b and 2b). One possible
170 interpretation of this from model behaviour is that part of the Amazon rainforest might
171 already be committed to dieback²⁶ despite not yet showing a strong change in mean state.
172 Our results suggest that the overall loss of Amazon resilience we find since the early 2000s
173 is attenuated in regions with higher rainfall and amplified in areas closer to human land use
174 change. This suggests that reducing deforestation will not just protect the parts of the forest
175 that are directly threatened but also benefit Amazon rainforest resilience over a much larger
176 area.

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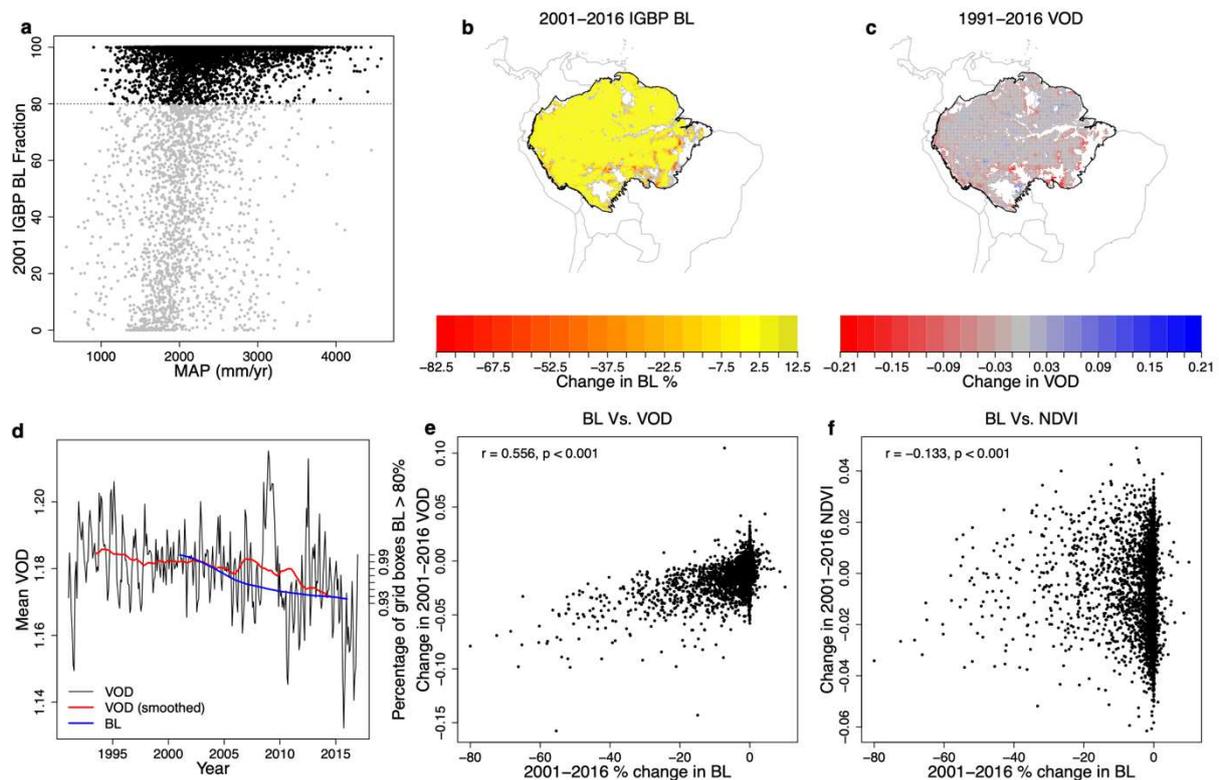
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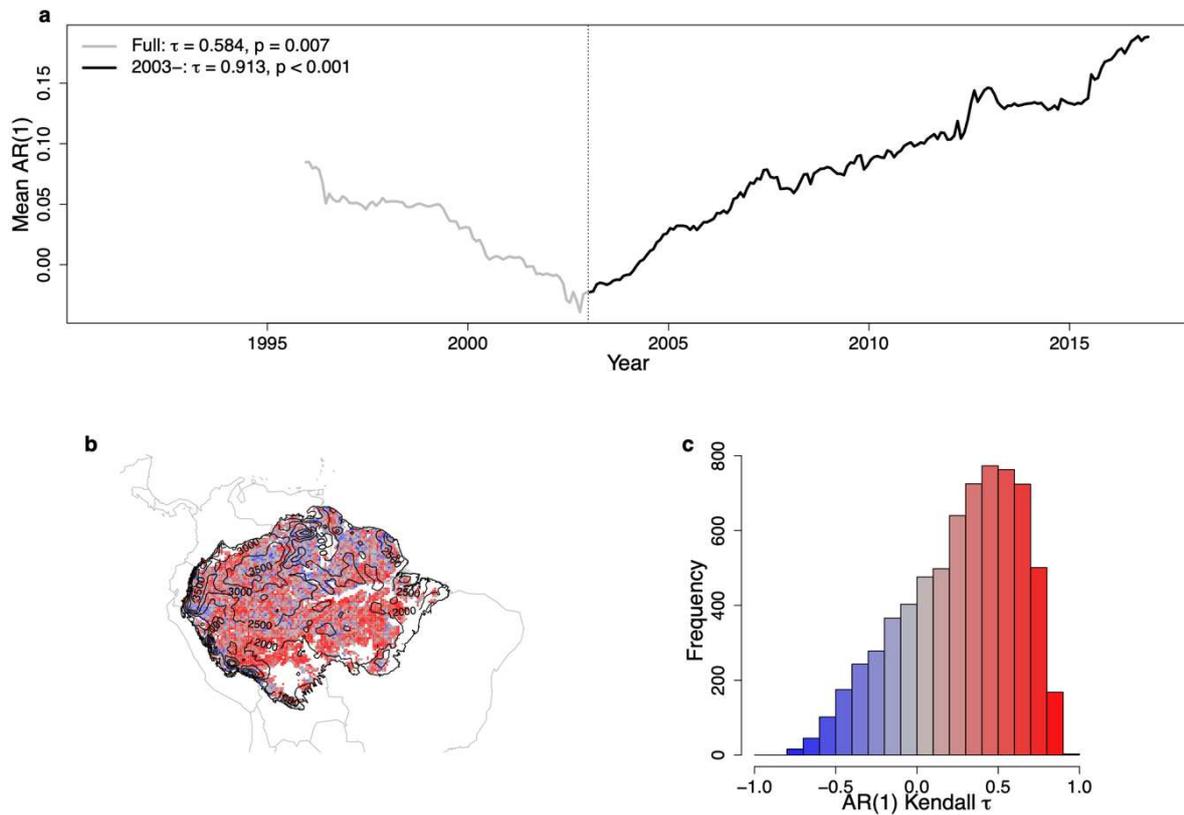
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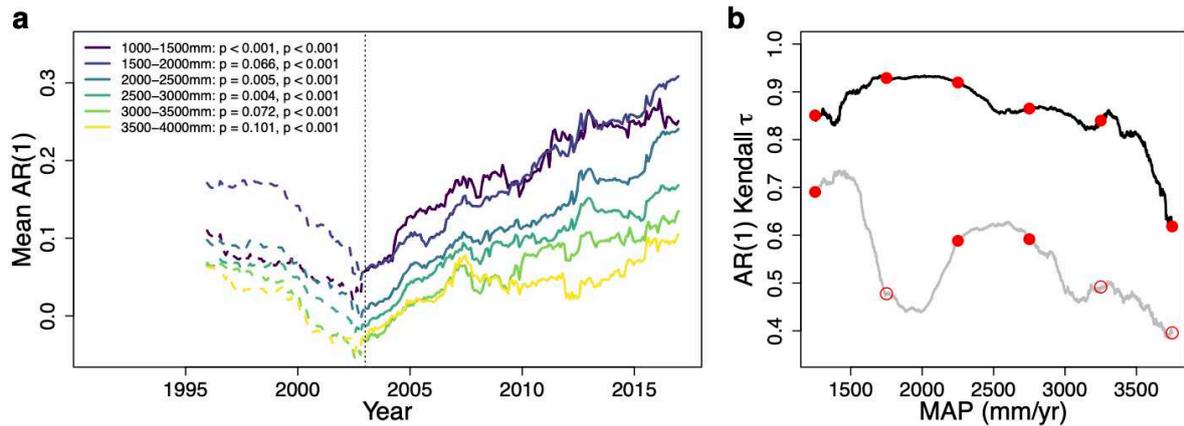
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315 **Figure 1: Relationships between different vegetation and rainfall data for the Amazon**
 316 **basin.** (a) The relationship between 2001-2016 mean annual precipitation (MAP) from
 317 CHIRPS⁴⁰ and 2001 MODIS Land Cover Evergreen Broadleaf (BL) fraction³⁹. Points
 318 coloured black are where 2001 BL \geq 80%. (b) Change in the BL fraction from 2001 to 2016
 319 for grid points where BL > 80% in 2001. Points that are predominantly BL in 2001 according
 320 to the MODIS 2001 dataset, but <80% are shown in grey. (c,d) Change in Vegetation Optical
 321 Depth (VOD) Climate Archive Ku-Band product³⁵ from 1991-2016 (difference between the
 322 2012-2016 and 1991-1995 means) for the grid points where BL > 80% in 2001, along with
 323 the median time series in these points. Also shown in blue on (d) is the annual percentage of
 324 grid boxes that have BL > 80%, from those that have BL > 80% in 2001. Sharp decreases in
 325 BL fraction and VOD could be directly attributed to deforestation. In these cases of externally
 326 forced forest loss, we may not see changes in AR(1) unless there was an underlying loss of
 327 resilience beforehand.



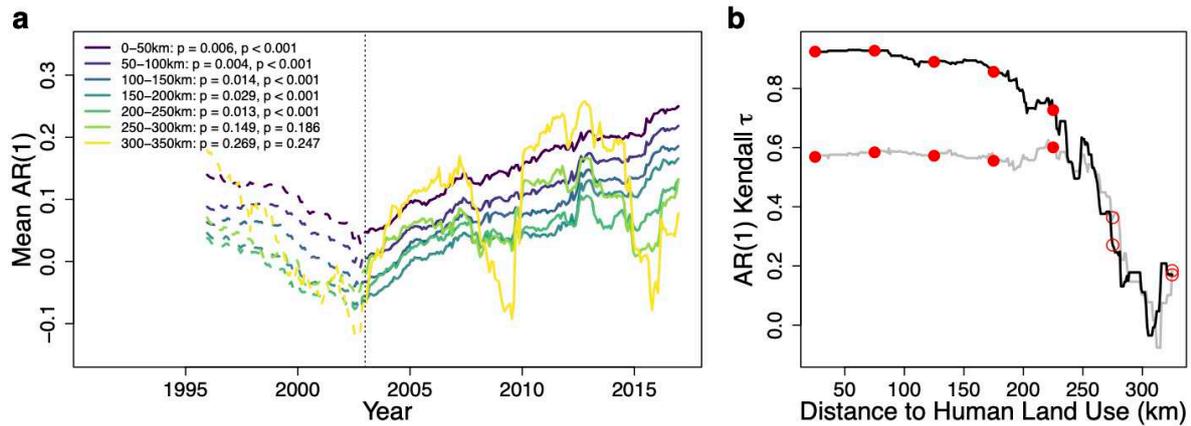
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329 **Figure 2: Changes in Amazon vegetation resilience since the early 1990s and from**
 330 **2003.** (a) Mean VOD AR(1) time series created from grid points that have $\geq 80\%$ BL
 331 fraction in the Amazon basin. The full AR(1) time series from 1991 (grey) has a Kendall τ
 332 value of 0.584 ($p = 0.007$) and from 2003 (black), a value of 0.913 ($p < 0.001$). (b) A map of
 333 the Kendall τ values of individual grid boxes from 2003, shown alongside contours of MAP
 334 (mm/year) over the same time period. (c) A histogram of the Kendall τ values from the map.



335

336 **Figure 3: The relationship between annual rainfall sums and vegetation resilience.** (a)
 337 Example VOD AR(1) time series for 500mm MAP-bands from 1991 (dotted lines) and from
 338 2003 (solid lines). (b) Full VOD AR(1) Kendall τ series for a sliding MAP-band, from 1991
 339 (grey) and from 2003 (black). Red circles show the results from panel (a) and are closed if
 340 the Kendall τ value is significantly positive ($p < 0.05$) and open otherwise. The tendency of
 341 the relationships in (b) are $\tau = -0.423$ (grey) and $\tau = -0.553$ (black), confirming there is a
 342 more severe decrease in resilience with lower rainfall values.



343

344 **Figure 4: The relationship between human activity and vegetation resilience.** (a)
 345 Example VOD AR(1) time series for 25km bands measuring the distance a forested grid box
 346 is from a human land use grid box (defined in the Methods from the MODIS Land Cover
 347 product³⁹ and shown in Supplementary Fig. 6), from 1991 (dotted lines) and from 2003 (solid
 348 lines). (b) Full VOD AR(1) Kendall τ series for a sliding distance-band, from 1991 (grey) and
 349 from 2003 (black). Red circles show the results from panel (a) and are closed if the Kendall τ
 350 value is significantly positive ($p < 0.05$) and open otherwise. The tendency of these
 351 relationships are $\tau = -0.553$ (grey) and $\tau = -0.857$ (black), showing there is a more severe
 352 decrease in resilience with increasing proximity to human land use.

353 **Methods**

354 **Datasets.** We use the Amazon basin

355 (http://worldmap.harvard.edu/data/geonode:amapoly_ivb) as our region of study. The main
356 dataset used to determine forest health is from the Vegetation Optical Depth Climate Archive
357 (VODCA)³⁵, of which we use the Ku-band product. This data is available at 0.25°x0.25° at a
358 monthly resolution from January 1988 to December 2016. We also use NOAA AVHRR
359 NDVI³⁶. For precipitation data, we use the CHIRPS dataset⁴⁰ downloaded from Google Earth
360 Engine (GEE) at a monthly resolution. Finally, to determine land cover types, we used the
361 IGBP MODIS land cover dataset MCD12C1³⁹. All of these datasets are at a higher spatial
362 resolution than the VODCA dataset and thus we linearly interpolate them to match the lower
363 resolution.

364 For the vegetation datasets that we measure the resilience indicators on (see below), we
365 use STL decomposition⁴⁸ using the `stl()` function in R. This splits time series in each grid box
366 into an overall trend, a repeating annual cycle (by using the 'periodic' option for the seasonal
367 window), and a residual component. We use the residual component in our resilience
368 analysis. Finding the first 3 years had large jumps in VOD which were seen when testing
369 other regions of the world as well as in the Amazon region, we restrict our analysis to
370 January 1991 to December 2016.

371 To test the robustness of the detrending, we also vary the size of the trend window in the
372 `stl()` function. The results from these alternatively detrended time series are shown
373 Supplementary Figure 2.

374 **Grid box selection.** We use the IGBP MODIS land cover dataset at the resolution described
375 above to determine which grid boxes to use in our analysis. The dataset is at an annual
376 resolution from 2001 to 2018 (but we only use the time series up to 2016 to match the time
377 span of our VOD and NDVI datasets). To focus on changes in forest resilience, we use grid
378 boxes where the evergreen broadleaf fraction is greater than or equal to 80% in 2001. Grid

379 boxes are treated as human land use area if the built-up, croplands, or vegetation mosaics
380 fraction is greater than 0% in 2016. We believe using these years to determine these
381 factors is the most cautious and least biased way to choose which grid boxes to use.

382 We measure the minimum distance between forested Amazon basin grid boxes and human
383 land use grid boxes using the latitude and longitude of each grid point. We do not restrict
384 human land use grid boxes to the Amazon basin region when determining the forested grid
385 boxes distance from them. This ensures that human land use grid boxes just outside the
386 region which could be the closest, are not ignored.

387 To ensure that the pattern of changes in resilience is not a consequence of more settlements
388 being in the south east of the region combined with the gradient of rainfall from northwest to
389 southeast typical of the rainforest, we measure the correlation between MAP and the
390 distances from the urban grid boxes. Although this is statistically significant, it is relatively
391 weak (Spearman's $\rho=0.109$, $p<0.001$) and as such we are confident that there are separate
392 processes that causes these relationships.

393 **Resilience indicator AR(1).** We measure our resilience indicator on the residual component
394 of the decomposed vegetation time series. We focus on lag-1 autocorrelation (AR(1)), which
395 provides the most robust indicator for critical slowing down prior to bifurcation-induced
396 transitions and has been widely used for this purpose^{11,28,30}. We measure it on a sliding
397 window length equal to 5 years (60 months). The sliding window creates a time series of the
398 AR(1) coefficient in each location.

399 From linearization and the analogy to the Ornstein-Uhlenbeck process, it holds
400 approximately that for discrete time steps of width Δt (one month in the case at hand):

$$401 \quad AR(1) = e^{-\kappa\Delta t},$$

402 where κ is the linear recovery rate. A decreasing recovery rate κ implies that the system's
403 capability to recover from perturbations is progressively lost, corresponding to diminishing

404 stability or resilience of the attained equilibrium state. From the above equation it is clear that
405 the AR(1) increases with decreasing κ . The point at which stability is lost and the system will
406 undergo a critical transition to shift to a new equilibrium state, corresponds to $\kappa = 0$ and
407 AR(1) = 1, respectively.

408 Measuring AR(1) across the whole time series provides information about the characteristic
409 timescales of the two vegetation datasets we use³¹. Inverting κ gives the characteristic time
410 scale of the system; for the VOD, we find $1/\kappa = 1.240$ months, whereas for the NDVI, we find
411 $1/\kappa = 0.838$ months when using the mean AR(1) value across the region. This suggests that,
412 in accordance with our interpretation of the two satellite-derived variables, the NDVI is more
413 sensitive to shorter-term vegetation changes such as leaf greenness, while the VOD's Ku
414 band is sensitive to longer-term changes such as variability in the thickness of forest stems.

415 **Creation and tendency of AR(1) and variance time series.** For analysis where either
416 MAP- or distance-bands are used to create an AR(1) or variance series, we calculate the
417 mean AR(1) or variance value in each month for forested Amazon basin grid boxes, from
418 which the tendency of this mean series can be calculated. Alternatively, the Kendall τ for
419 each band can be calculated by taking the mean Kendall τ for each individual grid box that is
420 within the band. Results from this method are shown in Supplementary Figure 10 for AR(1).

421 The tendency of the indicator is determined in terms of Kendall's τ . This is a rank correlation
422 coefficient with one variable taken to be time. Kendall's tau values of 1 imply that the time
423 series is always increasing, -1 always decreasing, and 0 no overall trend. Following previous
424 work^{30,49,50}, we test the statistical significance of positive tendencies using a test based on
425 phase surrogates that preserve both the variance and the serial correlations of the time
426 series from which the surrogates are constructed. Specifically, we compute the Fourier
427 transform of each time series for which we want to test the significance of Kendall's τ , then
428 randomly permute the phases and finally apply the inverse Fourier transform. Since this
429 preserves the power spectral density, it also preserves the autocorrelation function due to

430 the Wiener-Khinchin theorem. For each time series this procedure is repeated 100,000 times
431 to obtain the surrogates. Kendall's τ is computed for each surrogate to obtain the null model
432 distribution (corresponding to the assumption of the same variance and autocorrelation but
433 no underlying trend), from which the significance thresholds are computed as the 95th
434 percentiles.

435

436 **Data and code availability**

437 Data is available from the sources listed. R code is available on request.

438 **Methods references**

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447 **Acknowledgements.** NB acknowledges funding by the Volkswagen foundation. This is
448 TiPES contribution #X. The TiPES project ('Tipping Points in the Earth System') has
449 received funding from the European Union's Horizon 2020 research and innovation
450 programme under grant agreement No. 820970. CAB and TML were supported by the
451 Leverhulme Trust (RPG-2018-046). TML was also supported by a Turing Fellowship.

452 **Author contributions.** NB and CAB conceived and designed the study with input from TML.

453 CAB performed the numerical analysis with contributions from NB. All authors discussed

454 results and drew conclusions. CAB wrote the paper with contributions from NB and TML.

455 **Competing interests.** The authors declare no competing interests.

456 **Additional Information**

457 **Supplementary Information** accompanies the paper.

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Figures

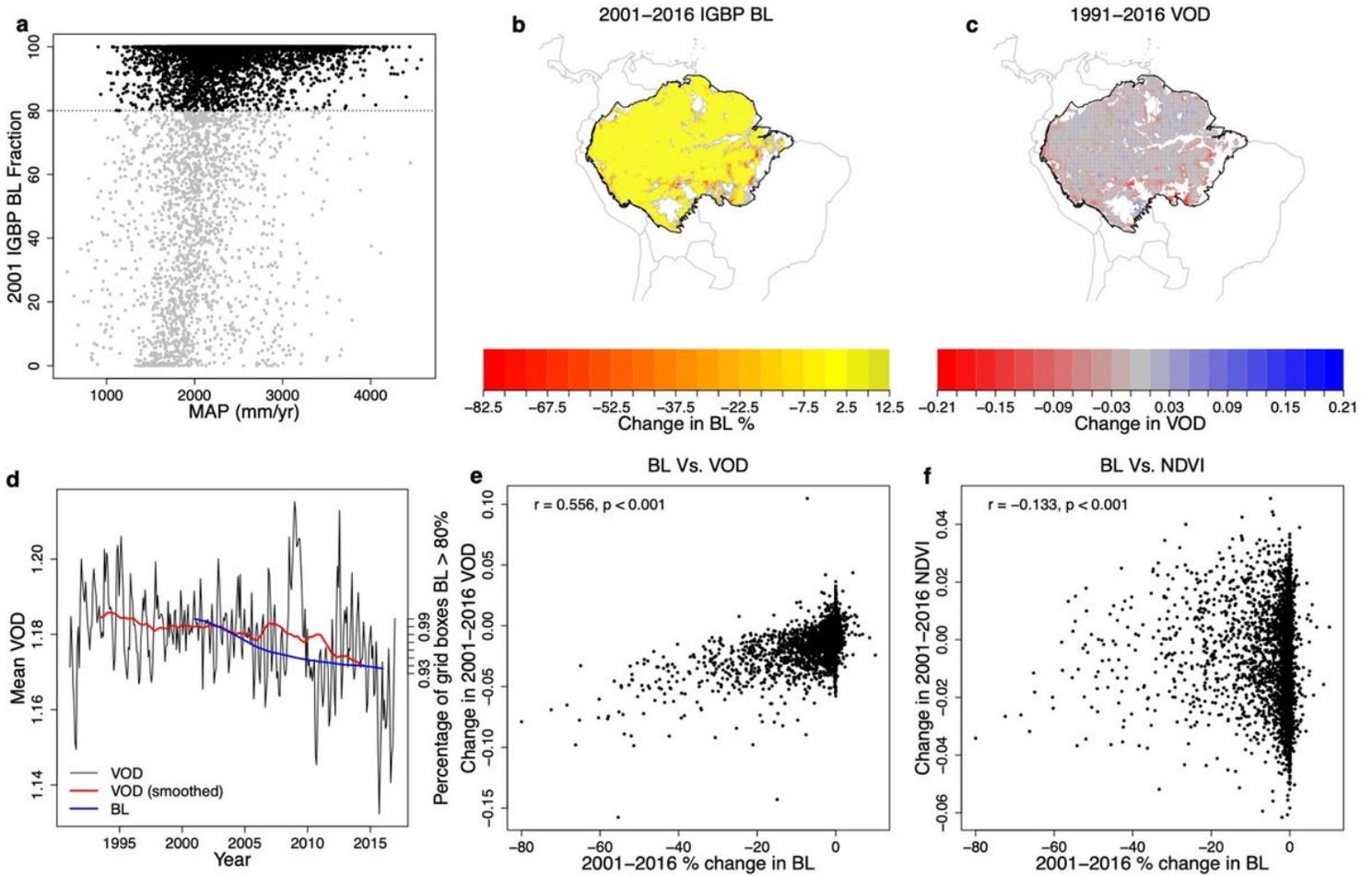


Figure 1

Relationships between different vegetation and rainfall data for the Amazon basin. (a) The relationship between 2001-2016 mean annual precipitation (MAP) from CHIRPS40 and 2001 MODIS Land Cover Evergreen Broadleaf (BL) fraction. Points colored black are where 2001 BL $\geq 80\%$. (b) Change in the BL fraction from 2001 to 2016 for grid points where BL $> 80\%$ in 2001. Points that are predominantly BL in 2001 according to the MODIS 2001 dataset, but $< 80\%$ are shown in grey. (c,d) Change in Vegetation Optical Depth (VOD) Climate Archive Ku-Band product from 1991-2016 (difference between the 2012-2016 and 1991-1995 means) for the grid points where BL $> 80\%$ in 2001, along with the median time series in these points. Also shown in blue on (d) is the annual percentage of grid boxes that have BL $> 80\%$, from those that have BL $> 80\%$ in 2001. Sharp decreases in BL fraction and VOD could be directly attributed to deforestation. In these cases of externally forced forest loss, we may not see changes in AR(1) unless there was an underlying loss of resilience beforehand.

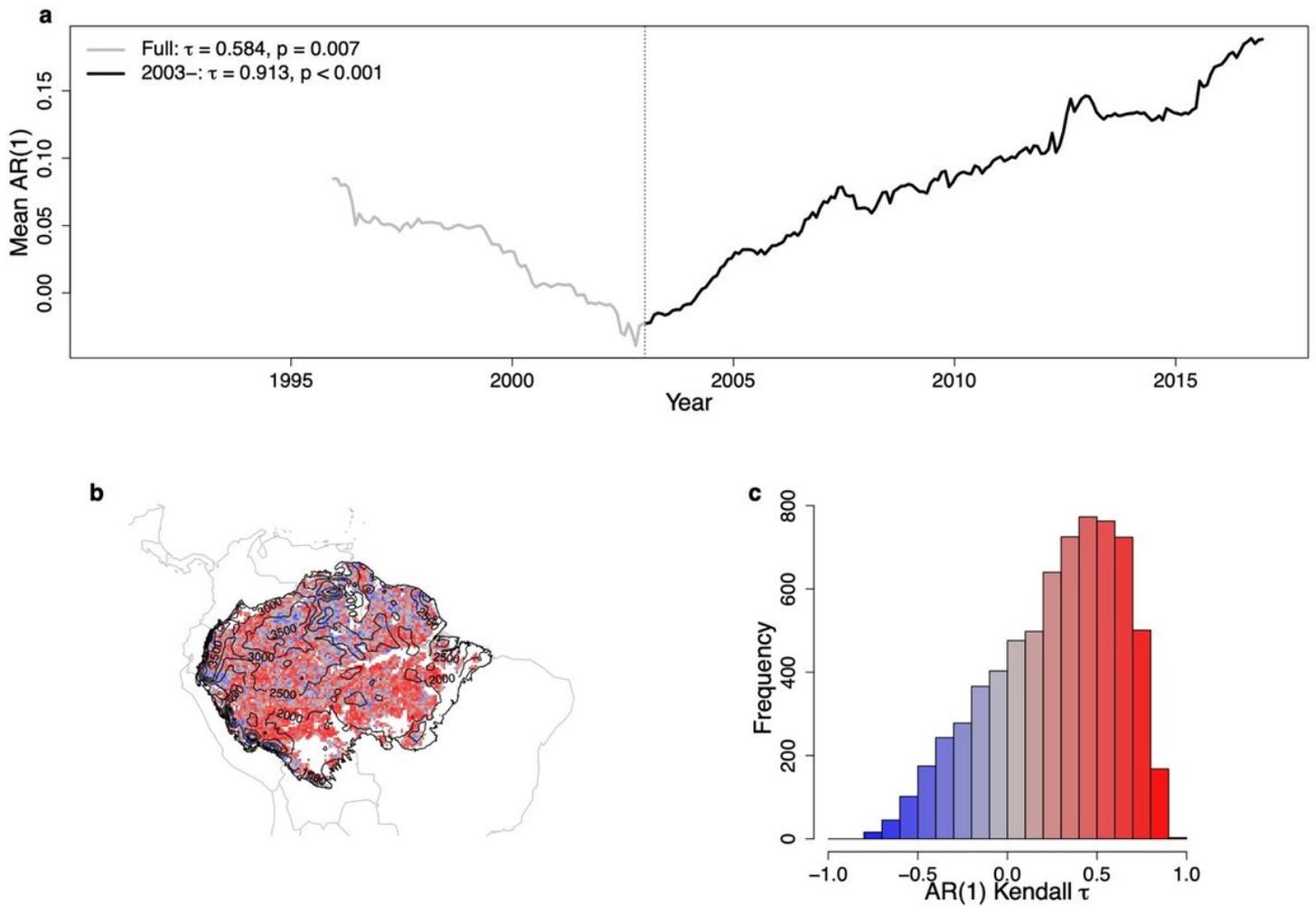


Figure 2

Changes in Amazon vegetation resilience since the early 1990s and from 2003. (a) Mean VOD AR(1) time series created from grid points that have $\geq 80\%$ BL fraction in the Amazon basin. The full AR(1) time series from 1991 (grey) has a Kendall t value of 0.584 ($p = 0.007$) and from 2003 (black), a value of 0.913 ($p < 0.001$). (b) A map of the Kendall t values of individual grid boxes from 2003, shown alongside contours of MAP (mm/year) over the same time period. (c) A histogram of the Kendall t values from the map.

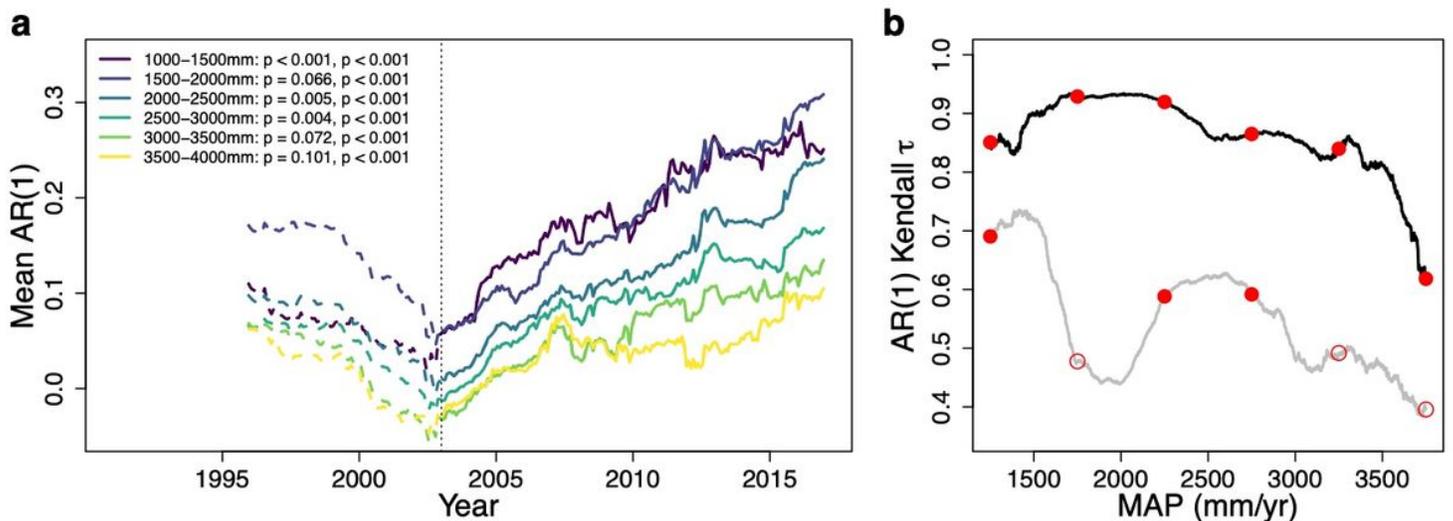


Figure 3

The relationship between annual rainfall sums and vegetation resilience. (a) Example VOD AR(1) time series for 500mm MAP-bands from 1991 (dotted lines) and from 2003 (solid lines). (b) Full VOD AR(1) Kendall t series for a sliding MAP-band, from 1991 (grey) and from 2003 (black). Red circles show the results from panel (a) and are closed if the Kendall t value is significantly positive ($p < 0.05$) and open otherwise. The tendency of the relationships in (b) are $t = -0.423$ (grey) and $t = -0.553$ (black), confirming there is a more severe decrease in resilience with lower rainfall values.

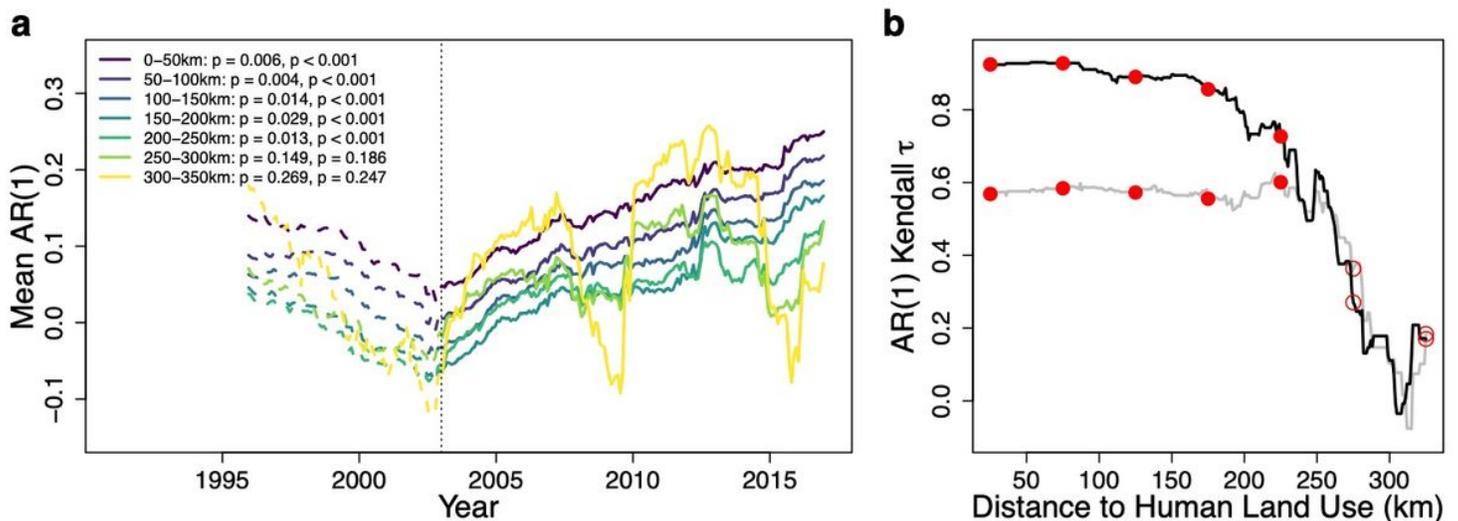


Figure 4

The relationship between human activity and vegetation resilience. (a) Example VOD AR(1) time series for 25km bands measuring the distance a forested grid box is from a human land use grid box (defined in the Methods from the MODIS Land Cover product39 and shown in Supplementary Fig. 6), from 1991 (dotted lines) and from 2003 (solid lines). (b) Full VOD AR(1) Kendall t series for a sliding distance-band,

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Supplementary Files

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