

Fixed-bed Column Dynamics of Ultrasound and Na-functionalized Diatomite to Remove Phosphate from Water

Junxiu Ye (✉ 2427182426@qq.com)

Yunnan Minzu University

Min Yang

Yunnan Minzu University

Xuemei Ding

Yunnan Minzu University

Wei Tan

Yunnan Minzu University

Guizhen Li

Yunnan Minzu University

Shuju Fang

Yunnan Minzu University

Hongbin Wang

Yunnan Minzu University

Research Article

Keywords: Diatomite, modify, ultrasonic, sodium chloride, dynamic adsorption, Phosphate, Sewage treatment, Thomas model

Posted Date: April 13th, 2021

DOI: <https://doi.org/10.21203/rs.3.rs-406797/v1>

License:   This work is licensed under a Creative Commons Attribution 4.0 International License.

[Read Full License](#)

Version of Record: A version of this preprint was published at Environmental Science and Pollution Research on July 6th, 2021. See the published version at <https://doi.org/10.1007/s11356-021-15126-4>.

1 **Fixed-bed column dynamics of ultrasound and Na-functionalized diatomite to remove phosphate**
2 **from water**

3 Junxiu Ye, Min Yang, Xuemei Ding, Wei Tan, Guizhen Li , Shuju Fang, Hongbin Wang*

4 School of Chemistry and Environment, Yunnan Minzu University, Kunming 650500, Yunnan, P.R.

5 China

6 **Corresponding author**

7 Hongbin Wang, School of Chemistry and Environment, Yunnan Minzu University, Kunming 650500,

8 Yunnan, China (E-mail: wanghb2152@126.com; Tel. +13708440749)

9 **Abstract:** A continuous fixed-bed column study has been used to evaluate phosphate adsorption
10 performance of U-D-Na which was functionalized by the cheap NaCl reagent after simple ultrasonic
11 purification of diatomite. Experimentally, various effect factors, the flow rate, the initial phosphate
12 concentration, and the bed height on breakthrough time of fixed column were studied. Experimental
13 results showed that the breakthrough time declined with the increase of inlet phosphorous
14 concentration and feed rate, whereas the increase of bed height turned out to significantly prolong the
15 breakthrough time. The dynamic adsorption data could better be fitted by the Thomas model, with the
16 correlation coefficients obtained, $R^2 > 0.9000$ at the majority of operating conditions (5/7). At least
17 thrice loop of adsorption and desorption was achieved with 0.1 M hydrochloric acid eluent and
18 deionized water. The results proved that U-D-Na could be used as a better alternative phosphate
19 adsorbent from wastewater in a continuous column process.

20 **Key words:** Diatomite; modify; ultrasonic; sodium chloride; dynamic adsorption; Phosphate;

21 Sewage treatment; Thomas model

22 **1. Introduction**

23 Eutrophication caused by the high content of nutrients such as phosphate in water has always been
24 concerning in the field of water treatment (Si et al. 2000). It was reported that the inland river basins in
25 many developed countries (e.g. the United States) and some developing countries (e.g. China) were
26 suffering from water eutrophication. And the trend of eutrophication is largely related to human
27 activities; the level of pollution in the river nearby densely populated areas was more serious
28 (Bhagowati & Ahamad 2018; Bu & Xu 2013; Yi et al. 2011). According to The U.S. Environmental
29 Protection Agency (USEPA), the phosphate concentration in water exceeding 0.02 mg/L might lead to
30 eutrophication. And China's "Pollutant Discharge Standard for Municipal Wastewater Treatment
31 Plants" stipulated that the phosphate content in urban sewage discharge should be less than or equal to

32 0.50 mg/L. In recent years, in order to remove phosphate in sewage, plenty of operation methods, such
33 as membrane filtration, adsorption, flocculation, precipitation, and crystallization had been carried out
34 by many scholars (Duan et al. 2013; Fan et al. 2017; Greenlee et al. 2009; Wen & Jian 2008; Wu et al.
35 2018). Among these, the adsorption technology has been considered to be a potential option due to its
36 simple operation, rapid adsorption performance, low cost, and high utilization rate of resources. This is
37 of great significance in sewage treatment and environmental protection. (Cong et al. 2014; Zhao et al.
38 2018)

39 The adsorption technology has been widely applied to many water treatment plants for phosphate
40 removal. This is an increasing tendency in fully utilizing effective adsorbent to remove target
41 substances in sewage. Generally, the adsorbate is attached to the surface of the adsorbent due to a
42 certain unbalanced attractive force, such as electrostatic attraction, chemical bonding, and van der
43 Waals force. Meanwhile, most of the adsorption processes were accompanied by physical
44 sedimentation. The strength of the acting force that existed in the adsorbent would largely affect the
45 sorption performance. Therefore, the selection and optimization of the adsorbent have become a key
46 factor in whether the technology can be applied to actual production. In the past few years, various
47 adsorbents for phosphate removal had been reported by many scholars. such as the hydrated metal
48 adsorbent reported (Chouyyok et al. 2010), in which HZrO and HFeO were dispersed into microporous
49 of anion ion exchange resin (IRA-400); it was found that the removal rate for phosphate reached 83%.
50 This might be because strong ligand sorption existed in poly-valent metals and phosphate. Furthermore,
51 Zhou reported a novel adsorbent (LDHs) that was synthesized by a simple hydrothermal method, with
52 the adsorption capacity reached 54.1 to 232mg/g (Zhou et al. 2011). At present, there are also several
53 sporadic reports on the adsorption performance of diatomite at home and abroad.

54 The natural diatomite is an important silicate material evolved from seafloor diatoms. With a large
55 number of micropores from surface to the interior, specific surface area, strong acid and alkali
56 resistance, a lot of striking chemical and physical properties of diatomite have been observed. And the
57 surface of diatomite was rich in silyl hydroxyl groups and hydrogen bonds (Xiao et al. 2010), which
58 could cause a series of chemical reactions with particles. Many reaction mechanisms of diatomite were
59 also based on those functional groups. According to the structural character of raw diatomite, a certain
60 adsorption effect could be observed; moreover, modification of natural diatomite is a vital approach to
61 realize its efficient adsorption. The majority of previous studies have been found that the functionalized

62 diatomite showed a promising sorption performance on phosphate ions. For instance, Xu and Pang
63 pointed out that the phosphate removal efficiency of the diatomite could be largely improved by
64 increasing the specific surface area, void volume of diatomite, and removing surface moisture and
65 organic matter (Xu &Pang 2009). And also, a kind of diatomite modified by hydrated lanthanum
66 chloride (La-diatomite) was reported (Wu et al. 2018) in which he demonstrated efficient phosphate
67 adsorption, with the maximum adsorption capacity reaching 58.7 mg/g. Besides, it was also reported
68 that a sizable high dephosphorization efficiency (close to 100%) could be achieved by diatomite
69 functionalized with zeolite and aluminum metal modifier (Duan et al. 2013; Wu &Chen 2011)
70 Numerous previous researchers had validated that the modified diatomite had an excellent sorption
71 capacity; and some information of thermodynamics, kinetics, and adsorption mechanism had been
72 obtained at the stage of static experiments. However, it is also extremely necessary for the researchers
73 to carry out fixed-bed experiments to evaluate the practical performance of the novel adsorbent.

74 The column experiment was a continuous process of interaction between the target pollutant and
75 the adsorbent. Many key factors such as initial concentration, feed rate, and bed depth that affect
76 column efficiency need to be considered during the adsorption process. The large-scale sewage in
77 industries was generally treated by fixed-bed adsorption. This process filled the gap that the static
78 adsorption could only clean up small batches of solution (Singh et al. 2012). In the dynamic sorption
79 test, with large amounts of target ions in the solution continuously enter and leave the column, the height
80 of the saturation layer continuously increased while the height of the adsorption layer declined until the
81 entire column almost lost its adsorption capacity that it reached the saturation state. The fixed-bed
82 reached the above-mentioned saturation phase, which called the dynamic equilibrium. Indicators like
83 breakthrough time, equilibrium adsorption capacity, and regeneration in the dynamic system are vital to
84 evaluate the performance of columns. The dynamic sorption performance of an adsorbent could be
85 measured by the breakthrough curves at different conditions. Generally, changes in flow rate and initial
86 concentration would affect the mass transfer process of phosphate in the column, and further affecting
87 the transfer rate and breakthrough time. According to previous researches, the feed rates, inlet
88 concentrations, and bed heights were the main factors for breakthrough performance in fixed-bed
89 experiments.

90 For those reasons, based on the static adsorption experiments of U-D-Na recently conducted by
91 our research team, various test conditions were carried out to further explore the phosphate removal

92 ability in the fixed-bed column. We intend to draw conclusions from this research on the various
93 influence factors of dynamic adsorption and regeneration experiment to provide some useful
94 information for the design and operation of the U-D-Na on phosphate removal process from sewage in
95 practical applications.

96 **2 Materials and methods**

97 **2.1 Reagents and instruments**

98 Natural diatomite used as the adsorbent precursor in this study was collected from the local
99 province, Xundian, Yunnan, which could be developed for phosphate removal in water after simple
100 modification treatment. In addition, the main chemicals such as KH_2PO_4 (GR), NaCl (GR) and HCl
101 (AR) were purchased from Tianjin Guangfu Fine Chemical Research Institute, Xilong Chemical Co.
102 LTD and Tianjin Fengchuan Chemical Reagent Technology Co. LTD, respectively. The above reagents
103 are prepared with secondary deionized water.

104 Electric heating blast drying oven DL402 and Ultrasonic cleaner AS10200AD produced by Tianjin
105 Experimental Instrument Factory. The adsorbent was weighed by using the American Ohaus AR2140
106 electronic analytical balance. The anti-molybdenum antimony spectrophotometer 7200 was introduced
107 in Uniko Shanghai Instrument Co., LTD. Dynamic sorption experiments were operated in a fixed-bed
108 column with an inside diameter of 0.8 cm and a bed height of 35 cm. A constant flow rate was
109 guaranteed by a commercial peristaltic pump.

110 **2.2 Experimental methods**

111 **2.2.1 Preparation of U-D-Na**

112 The preparation methods of U-D-Na in this work can consult our research group's static
113 adsorption experiment. Briefly, a proper amount of diatomite was crushed, screened and ultrasonic
114 cleaned for 1 h, and keep them into the dry container after drying at 110 °C. Mix the raw reagents
115 treated above with 5% NaCl solution sufficiently for 2 hours, which was dried under 95 °C and
116 preserved at 110 °C. Finally, applying an adequate amount of diatomite in muffle furnace calcining at
117 400 °C for 2 hours, and then the adsorbent U-D-Na required was prepared successfully. (Specific
118 details and data on the preparation of U-D-Na seeing the research group's unpublished paper attached
119 in S.1.)

120 To acquire the microporous structure of adsorbent and its elementary composition, the
121 U-D-Na was characterized by Scanning Electron Microscopy (SEM), Energy Disperse Spectroscopy

122 (EDS) . Microgram obtained from SEM could be seen that the surface micropores of U-D-Na were
123 loosely arranged and ordered smoothly. Obviously, the sodium chloride flat attached diatomaceous
124 earth surface and matched with the results on EDS. The result showed that sodium content in diatomite
125 was increased after being functionalized. Relevant characterization data can be referred to support
126 material S.1.

127 **2.2.2 Dynamic adsorption experiment**

128 The simple diagram of the fixed-bed column in the experiment was shown in Fig.1, and the
129 column height was 35cm and inner diameter was 0.8cm. In the experiment, the peristaltic pump was
130 used to transport the liquid and controlled the injection flow rate. A tank was placed at the outlet of the
131 separation column to receive the tail liquid. In order to facilitate the connection between the column
132 and peristaltic pump, and prevent the U-D-Na from floating; using suitable absorbent cotton to attach
133 the steel sieve to the bottom of the column, and seal the top of the column with a ground-glass piston.

134 Firstly, a known quantity of U-D-Na was packed in the column to the desired bed height of 1, 3,
135 and 5 cm (equivalent to 1.00, 3.00, and 5.00 g of U-D-Na); various concentrations of phosphate liquid
136 (15.00, 20.00, and 25.00mg/L) were pumped into the column at the flow rate of 1.00, 2.00, and 3.00
137 mL/min, respectively. The effluent sample concentration was measured at the exit of the column by
138 molybdenum antimony spectrophotometry (SEPA 1989) at a regular interval of 10 minutes.

139 **2.2.3 Calculation of breakthrough curve parameters**

140 The adsorption performance of a column generally could be evaluated by the breakthrough curve
141 and the sorption capacity(Rout et al. 2017). The breakthrough curve was a function defined by the ratio
142 of the outflow concentration C_t and the inlet concentration C_0 against time t . The breakthrough time (t_b)
143 was determined as the time when the outlet phosphate concentration (C_t) reached 10% of the inlet
144 phosphorus concentration ($C_t / C_0 = 0.1$). Similarly, the exhaustion time (t_s) was defined as the time
145 when the outlet phosphorus concentration (C_t) reached 90% of the inlet phosphorus concentration ($C_t /$
146 $C_0 = 0.9$).

147 Calculation of total adsorption capacity :

$$148 \quad q_{total} = \frac{Q}{1000} * \int_{t=0}^{t=total} C_{ad} dt \quad (1)$$

149 Where, q_{total} is the total amount of phosphorous adsorbed onto U-D-Na column, mg/g; Q is
150 volumetric flow rate, mL/min; C_{ad} is the difference in the phosphorus concentration at the initial time

151 and the t time caused by adsorption, mg/L; t_{total} is the total time for the column to reach saturation,
152 min.

153 Calculation formula of equilibrium adsorption capacity :

$$154 \quad q_e = \frac{q_{total}}{M} \quad (2)$$

155 Where, q_e is the equilibrium uptake capacity, mg/g; Which is derived as the quantity adsorbed
156 (q_{total}) per weight of adsorbent (M is the amount of U-D-Na packed in the column, g)

157 In addition to the total adsorption capacity and equilibrium adsorption capacity, the column
158 adsorption capacity could also be evaluated by removal rate, which was calculated by dividing the
159 difference between the initial concentration and equilibrium concentration of the solution by the initial
160 concentration with formula as follows.

$$161 \quad R(\%) = \frac{C_0 - C_t}{C_0} * \% \quad (3)$$

162 **2.2.4 Theoretical model**

163 It was essential to deeply understand the breakthrough performance of fixed-bed systems by
164 appropriate mathematical models for the successful design and optimization of the adsorption process
165 of packed columns. In actual operation, the variation of concentration of the column outlet with time
166 and bed height could be evaluated and predicted by an appropriate mathematical model, which was
167 convenient and necessary for expanding the experimental scale. Meanwhile, a suitable mathematical
168 model could be also helpful for practitioners to know and explain the affinity, surface properties, and
169 adsorption process between adsorbents and adsorbate (Foo et al. 2013). In the present research, to
170 achieve the above objectives, the continuous adsorption process of U-D-Na was evaluated by using
171 Thomas and Yoon-Nelson model.

172 **2.2.5 Successive cycles of sorption and desorption by column**

173 The experimental procedure of adsorption/desorption was conducted using HCl (0.10 mol/L)
174 solution as eluent to assess the reusability of U-D-Na. In the process of operation, firstly, the phosphate
175 solutions (25.00 mg/L) were passed through the fixed bed of 3g of U-D-Na at a constant flow rate (2
176 mL/min), at room temperature. Samples were collected after passing through the column to measure
177 their concentrations at different time intervals. The adsorption experiment was terminated when the
178 column reached saturation. At the first cycle, the column had reached saturation when the outflow
179 phosphate ions were close to the initial concentration and then followed by 7.00% HCl and a moderate

180 volume of deionized water to pass through the column until the exit phosphate concentration was nearly
181 equal to zero. The column then finished its regeneration and could be used for subsequent
182 adsorption/desorption cycles experiments. Performing operations under the same condition three times
183 repeatedly and measuring each phosphate removal rate of the column, and the regeneration capacity
184 was obtained visually through the breakthrough curves.

185 **3 Results and discussions**

186 **3.1 Study on dynamic adsorption using fixed-bed column**

187 **3.1.1 Effect of flow rate**

188 Fig. 2 illustrated the breakthrough curves of the column packed with U-D-Na at various flow rates
189 for an initial phosphate concentration of 25.00mg/L and bed height of 3 cm. As can be seen from Fig. 2,
190 the breakthrough time (50min, 30min, and 20min) and the exhaustion time (320, 110, and 90min)
191 decreased with the increasing flow rates (1, 2, and 3 ml/min). Higher feed rates could result in a weaker
192 interaction between the adsorbate molecules and adsorbent for the adsorption to be taken place
193 (Dwivedi et al. 2008). A similar tendency was reported in Zhang's research (Zhang et al. 2014) that the
194 activated laterite was used to remove phosphate. Also, the trends in the slopes of the breakthrough
195 curves could be used to measure the sorption performance of the column. From Fig. 2, it was clear that
196 the curves become steeper gradually with the flow rates increased, which indicated an excellent
197 sorption capacity of the column at a higher feed rate. This is possibly due to a majority of active sites of
198 U-D-Na were quickly occupied by massive phosphate ions available at a higher flow rate, urging them
199 to get saturated more quickly. While the flow rate increased from 2 to 3 mL/min, the slope of curves
200 was close to each other under those two conditions, the breakthrough curve did not change much.
201 Therefore, a faster flow rate did not necessarily result in a sorption effect. A shorter breakthrough time
202 and an exhaustion time at a higher flow rate were also reported elsewhere for sorption onto different
203 adsorbent and adsorbate (Singh et al. 2012). The sorption capacity at different flow rates summarized
204 in Table 1 showed a dropping trend with the uprisng flow rates. It can be explained by a lower flow
205 rate leading to more residence time of the phosphate ions in the column, longer contact occurred
206 between phosphate ions and U-D-Na; the effective sorption interaction could be reached before
207 phosphate ions flowed out of the column. Thus the higher adsorption capacity (1.43 mg/g) was
208 observed at the lowest flow rate of 1 mL/min. Another possible explanation for the higher sorption
209 capacity of U-D-Na was that those active sites with weaker affinity could be effectively captured by

210 phosphate ions in a lower flow, which agreed with the study (Bulgariu & Bulgariu 2013). Moreover,
211 such performance at a higher flow rate was also one of the reasons for the earlier breakthrough and
212 exhaustion time. Comprehensive analysis of breakthrough curves depicted in Fig. 2 and the
213 experimental parameters from Table 1, the best sorption performance was obtained at the lowest flow
214 rate (2 mL/min)

215 **3.1.2 Effect of bed height**

216 The breakthrough curves under various bed depths of 1, 3, and 5 cm were conducted under the
217 initial concentration of 25.00 mg/L and a constant flow rate of 2 mL/min, the variation tendency was
218 shown in Fig. 3. The amount of adsorbent in the column was approximately 1, 3, and 5 g, respectively.
219 As can be seen from this graph, the breakthrough time (30 to 90 min) increased with the increasing bed
220 height (1 to 5 cm). Likewise, the time requirement of phosphate solution for reaching saturation point
221 was concomitantly increased with the growing bed height as expected. The exhaustion time listed in
222 table 1 was 95, 110, and 280 min for the bed heights of 1, 3, and 5 g, respectively. This might be
223 explained by the possibility of more binding sites available at taller bed heights for adsorption and
224 leading to a higher MTZ (Jain et al. 2013). However, at a lower bed height, the diffusion of U-D-Na in
225 the column was reduced due to the axial dispersion phenomenon (Vijayaraghavan et al. 2004). In
226 addition, at greater depth, more surface area of adsorbent is available, which resulted in a large surface
227 of active sites for sorption (Saha et al. 2012). Likewise, as depicted in table 1, the total phosphate
228 uptake capacity (q_{total}) of U-D-Na was 2.15, 2.78 and 6.63 g for the bed heights of 1, 3 and 5 cm,
229 respectively, but the equilibrium adsorption capacity (q_e) was 2.15, 0.93 and 1.33 mg/g, respectively.
230 An obvious conclusion we could draw that q_e did not increase with the uprising q_{total} . Inversely,
231 increasing the amount of U-D-Na from 1 to 5 g, q_e declined from 2.15 to 1.33 mg/g, which meant that
232 the sorption capacity per gram weight adsorbent did not increase with the increase in bed height, and
233 such sorption behavior was apparently attributed to the infinite phosphate ions in the column (Singh et
234 al. 2012). However, it is found a positive correlation between the adsorption capacity and bed height
235 (Kumar et al. 2011), which could ensure the handling of large amounts of phosphate solutions.

236 **3.1.3 Effect of influent phosphorus concentration**

237 The influence of different inlet phosphate ions concentrations on the breakthrough curves at a
238 given flow rate of 2 mL/min and bed depth of 3 cm was depicted in Fig. 4. The result indicated that the
239 breakthrough time was reversely related to the influent phosphate concentration. The breakthrough

240 time was found to be 70, 60, and 30 min for the concentration of 15.00, 20.00, and 25.00 mg/L,
241 respectively. As summarized in Table 1, a similar tendency was observed in exhaustion time (280, 190,
242 and 110 min), which clearly decreased with a rise in initial phosphate concentration. Besides, the
243 curves grew much steeper with a longer MTZ at a higher initial phosphate concentration. This might be
244 ascribed to much more phosphate ions available at a higher concentration occupied those binding sites
245 to be adsorbed. A lower concentration gradient could reduce the mass transfer coefficient or diffusion
246 (Zhang et al. 2011). The steepness in the slope of the curves depicted in Fig. 4 showed an increasing
247 uprise with the growing phosphate concentration. The steeper the curves, the better the sorption
248 capacity of the column (Cruz-Olivares et al. 2013). A sharpness in slope indicated a shorter time for the
249 column to reach saturation with a stronger adsorption rate. Nevertheless, the result in Table 1 illustrated
250 that the equilibrium uptake (q_e) of U-D-Na for phosphate lightly fallen from 1.51 to 0.93 mg/g with the
251 uprising phosphate inlet concentration from 15.00 to 25.00 mg/L. An uprising concentration did not
252 significantly result in a considerable sorption capacity this study, reason behind this still needs to be
253 explored. It can be merely inferred from the existing results that the increase in initial concentration
254 mainly affects the adsorption rate and breakthrough time.

255 **3.1.4 Analysis of experimental parameters**

256 The effects of various parameters such as bed depth, initial phosphate concentration, and feed
257 rate for breakthrough and exhaustion time of U-D-Na packaged fixed bed were discussed above.
258 Similar to the findings of most researchers that the breakthrough and exhaustion time prolonged with
259 the increase in bed height and feed rate, while decreased with the increasing initial concentration. This
260 could be explained by the two aspects that (1) high concentration phosphate ions resulted in a lower
261 mass transfer resistance ; (2) sufficient contact with abundant active sites was achieved per unit time
262 during the adsorption process. Through a series of tests on the breakthrough time and adsorption
263 capacity of the column packed with U-D-Na, the optimum sorption condition could be obtained at the
264 feed rate of 2 ml/min, initial concentration of 25.00mg/L, and the bed depth of 3 cm. Follow-up
265 experiments for column regeneration could be conducted in this condition.

266 **3.2 Breakthrough curve model**

267 **3.2.1 Thomas model**

268 The Thomas model was one of the most widely used models to evaluate column performance
269 and predict the phosphate adsorption breakthrough curve(Foo et al. 2013). It was based on two

270 assumptions (1) the adsorption interface is not affected by chemical reaction with an absence of axial
 271 dispersion, and (2) data fits Langmuir isotherm and the second-order reversible reaction kinetics.
 272 Compared with the Adams-Bohart model, it was more suitable for describing the entire process of
 273 dynamic adsorption ($C_t/C_0 = 0-1$), while the Adams-Bohart model was only suitable for describing
 274 the initial stage of the curve, namely the $C_t/C_0 = 0-0.5$, resulting in a limit of its validity in many ranges
 275 of conditions. The linear equation of the Thomas model and related symbols are shown below.
 276 Equilibrium adsorption capacity, q_0 and Thomas rate constant, K_{TH} were obtained by intercept and
 277 slope from the linear regression analysis of $\ln(C_0/C_t-1)$ against t , respectively.

$$278 \quad \ln\left(\frac{C_0}{C_t} - 1\right) = \frac{K_{TH}q_0m}{Q_0} - K_{TH}C_0t$$

279 (4)

280 Where, C_0 is the inlet phosphate concentrations, mg/L; C_t is the effluent phosphate
 281 concentrations, mg/L; q_0 is the adsorption capacity, mg/g; m is the mass of adsorbent, g; K_{TH} is the
 282 kinetic constant, mL/min.mg; Q_0 is the feed flow rate, mL/min.

283 The values of characteristic parameters of the Thomas model obtained from different
 284 experimental conditions were listed in Table 2. Thomas rate constant K_{TH} (0.70, 1.22, and 1.90
 285 mL/min.mg) increased with the increase in feed rate from 1 to 3 mL/min, as the increase in initial
 286 phosphate concentration from 15.00 to 25.00mg/L, the value of K_{TH} increased from 0.80 to 1.22
 287 mL/min.mg, whereas the sorption capacity, q_0 reduced from 1.53 to 1.34 mg/g. A larger value of K_{TH}
 288 was observed at a faster feed rate and a higher inlet concentration. This might be due to the larger
 289 concentration gradient and larger feed rate which could result in a higher driving force in this case.
 290 However, the K_{TH} was found reversely related to the bed height. This could be explained by the
 291 increase in bed height, namely an increase in the amount of adsorbent, leading to a large steric
 292 hindrance. A similar explanation was also reported (Paudyal et al. 2013). The equilibrium adsorption
 293 capacity, q_0 did not present a continuous variation trend at constant flow rates and bed heights in this
 294 experiment. In general, a rise both in flow rate and bed depth led to a q_0 value decrease. Analysis of the
 295 majority of data summarized in Table 2, Thomas model was found fitting well the dynamic sorption
 296 process of U-D-Na with the regression coefficient obtained $R^2 > 0.9000$ under most of the parameters
 297 (5/7). It could also be inferred that neither the internal nor the external diffusions of the adsorbent
 298 surface was the rate control step of the process (Chen et al. 2012). The predicted values (1.53, 1.19, and
 299 1.34 mg/g) calculated by the Thomas model were shown in Table 2, which indicated a good agreement

300 with the experimental data (1.43, 0.93, and 1.13) at an uprising flow rate (1, 2, and 3 mL/min). Those
301 results validated that the Thomas model is suitable for representing the dynamic sorption process.

302 3.2.2 Yoon–Nelson model

303 Yoon-Nelson model, developed by Yoon and Nelson in 1981, is the simplest mathematical and
304 semiempirical expression with fewer parameters. This model need not consider the property of
305 adsorbent and physical characters of the column. Similar function with the Thomas model is suitable to
306 describe the whole process of dynamic adsorption. The Yoon–Nelson model was based on the
307 assumption: the rate of decrease in the probability of adsorption for each adsorbate molecule was
308 proportional to the probability of adsorbate and the probability of adsorbate breakthrough on the
309 adsorbent (Yoon & Nelson 1984). The Yoon–Nelson rate constant K_{YN} and the time reaching 50%
310 phosphorous breakthrough τ could be calculated from the linear plot of $\ln\left(\frac{c_t}{c_0 - c_t}\right)$ against t . The linear
311 expression of the Yoon–Nelson model was given by the following equation.

$$312 \quad \ln\left(\frac{c_0}{c_0 - c_t}\right) = K_{YN}t - \tau K_{YN} \quad (5)$$

313 Where, K_{YN} is the Yoon–Nelson rate constant (min^{-1}), and τ is the time required for 50%
314 phosphorous breakthrough (min).

315 The data obtained from Yoon-Nelson were listed in Table 2. It was found that the K_{YN} (0.0175,
316 0.0298, and 0.0426 min^{-1}) increased while τ (184.58, 71.07, and 52.32 min) decreased with the flow
317 rate (1 to 3 mL/min) and initial concentration (15.00 to 25.00 mg/L) increased. A reasonable
318 explanation is that a faster time of τ could be reached at a higher flow rate and initial concentration.
319 A Similar tendency had been reported previously (Yuan et al. 2020). Nevertheless, a growth in bed
320 depth (1 to 5 cm) resulted in a fall trend in τ (158.67 to 71.07 min) and an uprising trend in K_{YN}
321 (0.0199 to 0.0298 min^{-1}). The results listed in Table 2 indicated that the flow rate had a huge effect on
322 τ , the value of τ decreased roughly three times, with the flow rate increased from 1 to 3 mL/min
323 when the sorption process was reaching 50% breakthrough point. According to the assumption of the
324 Yoon-Nelson model, it could be inferred that the probability of phosphate ions passing through the
325 column was enhanced at a higher flow rate. The correlation coefficient R^2 obtained from Yoon-Nelson
326 model was ranged from 0.8428 to 0.9519, under the majority of experimental conditions (5/7), the
327 values of R^2 were higher than 0.9000. However, the τ predicted values calculated by Yoon-Nelson did
328 not satisfactorily agree with the experimental values. At the flow rate of 1, 2, and 3 mL/min, the τ

329 predicted values were found to be 140.00, 52.00, and 50.00 min for the experimental values of 184.58,
330 71.07, and 52.32, respectively. The data gap between predicted and experimental values was listed in
331 table 2, a conforming lower value in predicted than the experimental was observed under all of the
332 conditions. Those results based on table 2 could be speculated that the Yoon-Nelson model neglect of
333 axial dispersion is not suitable to describe the phosphate-U-D-Na sorption system.

334 **3.2.3 Mathematical model analysis**

335 In this experiment, the Thomas model and Yoon-Nelson model were used to fit the adsorption
336 process of the phosphate-U-D-Na system. In comparison to Thomas and Yoon-Nelson model, the
337 results given above validated that the Thomas model is more suitable to represent the sorption process
338 of U-D-Na than the Yoon-Nelson model. As listed in Table 1, the values of R^2 obtained from the
339 Thomas model ranged from 0.8482 to 0.9480, and under the majority of conditions (5/7) were higher
340 than 0.9000. The q_e predicted values were found to be agreed well with those acquired from the
341 experiment in the changes of flow rates and bed depths. On the other hand, the time for phosphate
342 reaching 50% breakthrough, τ values predicted by Yoon-Nelson existed a large difference with the
343 experimental values. At all the conditions, the values of τ were evidently lower than those obtained
344 from experiments. The maximum absolute error among them was even reaching 57.67 min. These
345 results demonstrated that the Yoon-Nelson model which ignored the properties of U-D-Na to describe
346 the dynamic adsorption process is flawed.

347 **3.3 Successive adsorption-desorption cycles**

348 **3.3.1 Regeneration of U-D-Na adsorbed phosphate.**

349 In order to investigate the practical application ability of U-D-Na, the elution from a saturated bed
350 was carried out with 7.00% HCl eluent at the same flow rate of 2 mL/min. The trend plot between the
351 outlet concentration and the elution time was depicted in Fig. 5. It could be clearly seen from Fig.5 that
352 with continuous addition of the eluent, the outlet phosphate concentration in the eluent represented a
353 continuous declination. A sharp drop trend occurred in 60 min which indicated a large elution rate at
354 this period. The outlet phosphate concentration was found to be 94.06 mg/L after a flow of dilute acid
355 7.00% HCl solution in 10 min, near four times higher than the initial concentration (25.00 mg/L), which
356 indicated that the HCl solution could be considered as a highly effective eluent for phosphate ions in
357 the batch system. Roughly 100 min the concentration of eluent could be reduced below 10% of the

358 initial elution concentration. A short regeneration time for the U-D-Na could be achieved. Besides,
359 almost all phosphate ions attached to the column were cleared off within 180 min.

360 **3.3.2 Breakthrough curves of the column after three times cycles**

361 Fig. 6 depicted the breakthrough curves that the column of U-D-Na regenerated with HCl solution
362 undergone three consecutive phosphate adsorption cycles to evaluate its reusability. It is clear that the
363 breakthrough time did not change too much after being regenerated three times in the experiments. As
364 depicted in Fig.6, along with the increase in cycle times (1, 2, and 3 cycles) of regeneration, the
365 breakthrough time (30, 20, and 20 minutes) decreased slightly. However, the removal efficiency was
366 almost unchanged, with similar high adsorption efficiencies for phosphate to the first two times. A
367 declination in breakthrough time might be ascribed to the rupturing of certain binding sites, because
368 part of the column damage was generated by those physical or chemical behavior in the
369 adsorption/desorption process. According to the result of the breakthrough performance shown in Fig.
370 6, the U-D-Na was expected to be considered as a potential adsorbent because of its low-cost operation
371 with an effective phosphate removal rate and resource-saving.

372 **4. Conclusion**

373 U-D-Na was found to be a promising candidate adsorbent for phosphate removal from the solution
374 using a fixed-bed column in this study. The results showed that the breakthrough time was positively
375 correlated with the bed height, but negatively correlated with the initial concentration and flow rate.
376 Under the optimum experimental condition, the equilibrium sorption capacity (q_e) and the removal rate
377 of U-D-Na for phosphate could reach 1.19 mg/g and 98.14%, respectively. The phosphate-U-D-Na
378 sorption system was found well to be fitted by the Thomas model, with correlation coefficients,
379 $R^2 > 0.9000$ under the majority of conditions (5/7), the surface sorption is the main rate-controlling step.
380 Considerable removal efficiency after three desorption-adsorption cycles could near reach 98%. The
381 effectiveness of continuous operations was evidenced successfully by the effective removal of
382 phosphate from water onto the column. These results revealed that U-D-Na is an economical and
383 environmentally friendly adsorbent for water treatment.

384 **Declarations**

385 **Ethics approval and consent to participate**

386 Not applicable.

387 **Consent for publication**

388 Not applicable.

389 **Availability of data and materials**

390 The datasets used and/or analyzed during the current study are available from the corresponding
391 author on reasonable request.

392 **Competing interests**

393 The authors declare that they have no competing interests.

394 **Funding information**

395 This work was supported by the Science Research Fund Projects of Department of Education of
396 Yunnan Province (Grant No. 2020J0329) and the Science Research Fund Projects of Department of
397 Education of Yunnan Province (Grant No. 2020J0327)

398 **Authors' Contributions**

399 Funding acquisition, Wei Tan and Guizhen Li; methodology, Min Yang; experiment operation,
400 Xuemei Ding and Junxiu Ye ; data curation, Xuemei Ding and Shuju Fang; writing—original draft,
401 Junxiu Ye; writing—review and editing, Junxiu Ye, Guizhen Li and Hongbin Wang; All authors have
402 read and agreed to the published version of the manuscript.

403 **Acknowledgments**

404 The main author sincerely acknowledges the Yunnan Minzu University for providing a good
405 research platform. We all greatly thank professor Wang for providing a lot of valuable helps to guide
406 our experiments. We would like to thank the anonymous reviewers and the editor.

407 **Reference**

- 408 Bhagowati B, Ahamad KU (2018) A review on lake eutrophication dynamics and recent developments in
409 lake modeling. *Ecohydrology and Hydrobiology*, S164235931730143X.
410 <https://doi.org/10.1016/j.ecohyd.2018.03.002>
- 411 Bu FP, Xu XY (2013) Planted floating bed performance in treatment of eutrophic river water.
412 *Environmental Monitoring & Assessment* 185, 9651-9662.
413 <https://doi.org/10.1007/s10661-013-3280-6>
- 414 Bulgariu D, Bulgariu L (2013) Sorption of Pb(II) onto a mixture of algae waste biomass and anion
415 exchanger resin in a packed-bed column. *Bioresource Technology* 129, 374-380.
416 <https://doi.org/10.1016/j.biortech.2012.10.142>

417 Chen SH, Yue QY, Gao BY et al (2012) Adsorption of hexavalent chromium from aqueous solution by
418 modified corn stalk: A fixed-bed column study. *Bioresource Technology* 113, 114-120.
419 <https://doi.org/10.1016/j.biortech.2011.11.110>

420 Chouyyok WI, Wiacek RJ, Pattamakomsan K et al (2010) Phosphate Removal by Anion Binding on
421 Functionalized Nanoporous Sorbents. *Environmental Science & Technology* 44, 3073-3078.
422 <https://doi.org/10.1021/es100787m>

423 Cong HB, Mao ZL, Huang YL (2014) Study on the Phosphorus Removal Ability of Modified Calcite.
424 *Environmental Pollution & Control* 36, 29-33+38.
425 <https://doi.org/10.15985/j.cnki.1001-3865.2014.10.034>

426 Cruz-Olivares J, Pérez-Alonso C, Barrera-Díaz C et al (2013) Modeling of lead (II) biosorption by
427 residue of allspice in a fixed-bed column. *Chemical Engineering Journal* 228, 21-27.
428 <https://doi.org/10.1016/j.cej.2013.04.101>

429 Duan N, Zhang YF, Wu KM et al (2013) Preparation of diatomite composite adsorbent and its
430 performance in removing nitrogen and phosphorus from wastewater. *Bulletin of the Chinese
431 Ceramic Society* 32, 1528-1533. <https://doi.org/10.16552/j.cnki.issn1001-1625.2013.08.041>

432 Dwivedi CP, Sahu JN, Mohanty CR et al (2008) Column performance of granular activated carbon
433 packed bed for Pb(II) removal. *Journal of Hazardous Materials* 156, 596-603.
434 <https://doi.org/10.1016/j.jhazmat.2007.12.097>

435 Fan Y, Wang Z, Zhao LQ et al (2017) Study on the adsorption of phosphorus by zirconium modified
436 diatomite. *Environmental Science* 38, 1490-1496. <https://doi.org/10.13227/j.hjhx.201609132>

437 Foo KY, Lee LK, Hameed BH (2013) Preparation of tamarind fruit seed activated carbon by microwave
438 heating for the adsorptive treatment of landfill leachate: A laboratory column evaluation. *Bioresour
439 Technol* 133, 599-605. <https://doi.org/10.1016/j.biortech.2013.01.097>

440 Greenlee LF, Lawler DF, Freeman BD et al (2009) Reverse osmosis desalination: Water sources,
441 technology, and today's challenges. *Water Research* 43.
442 <https://doi.org/10.1016/j.watres.2009.03.010>

443 Jain M, Garg VK, Kadirvelu K (2013) Cadmium(II) sorption and desorption in a fixed bed column using
444 sunflower waste carbon calcium-alginate beads. *Bioresource Technology* 129.
445 <https://doi.org/10.1016/j.biortech.2012.11.036>

446 Kumar R, Bhatia D, Singh R et al (2011) Sorption of heavy metals from electroplating effluent using
447 immobilized biomass *Trichoderma viride* in a continuous packed-bed column. *International*
448 *Biodeterioration & Biodegradation* 65, 1133-1139. <https://doi.org/10.1016/j.ibiod.2011.09.003>

449 Paudyal H, Pangeni B, Inoue K et al (2013) Adsorptive removal of fluoride from aqueous medium using
450 a fixed bed column packed with Zr(IV) loaded dried orange juice residue. *Bioresour Technol* 146,
451 713-720. <https://doi.org/10.1016/j.biortech.2013.07.014>

452 Rout PR, Bhunia P, Dash RR (2017) Evaluation of kinetic and statistical models for predicting
453 breakthrough curves of phosphate removal using dolochar-packed columns. *Journal of Water*
454 *Process Engineering* 17. <https://doi.org/10.1016/j.jwpe.2017.04.003>

455 Saha PD, Chowdhury S, Mondal M et al (2012) Biosorption of Direct Red 28 (Congo Red) from
456 Aqueous Solutions by Eggshells: Batch and Column Studies. *Separation Science & Technology* 47,
457 112-123. <https://doi.org/10.1080/01496395.2011.610397>

458 SEPA (1989) *Water and wastewater monitoring and analysis methods*. China Environmental Science
459 Press

460 Si YB, Wang SQ, Chen HM (2000) Loss of nitrogen and phosphorus in farmland and eutrophication of
461 water body. *Soils* 32, 188-188. <https://doi.org/10.13758/j.cnki.tr.2000.04.005>

462 Singh A, Kumar D, Gaur JP (2012) Continuous metal removal from solution and industrial effluents
463 using *Spirogyra* biomass-packed column reactor. *Water Research* 46, 779-788.
464 <https://doi.org/10.1016/j.watres.2011.11.050>

465 Vijayaraghavan K, Jegan J, Palanivelu K et al (2004): Removal of nickel(II) ions from aqueous solution
466 using crab shell particles in a packed bed up-flow column. *Journal of Hazardous Materials* 113,
467 223-230. <https://doi.org/10.1016/j.jhazmat.2004.06.014>

468 Wen HX, Jian P (2008) Development and characterization of ferrihydrite-modified diatomite as a
469 phosphorus adsorbent. *Water Research* 42. <https://doi.org/10.1016/j.watres.2008.09.030>

470 Wu L, Chen YF (2011) Study on Modified Diatomite for Dephosphorization of Chaohu Lake Water.
471 *Chinese Journal of Environmental Engineering* 5, 777-782.
472 <https://doi.org/10.3724/SP.J.1011.2011.00468>

473 Wu Y, Li XM, Yang Q (2018) Hydrated lanthanum oxide-modified diatomite as highly efficient
474 adsorbent for low-concentration phosphate removal from secondary effluents. *Journal of*
475 *environmental management*. <https://doi.org/10.1016/j.jenvman.2018.10.059>

476 Xiao LG, Zhao Z, Yu WZ (2010) Development status and prospect of diatomite at home and abroad.
477 Journal of Jilin Jianzhu University 27, 26-30. <https://doi.org/CNKI:SUN:JLJZ.0.2010-02-008>

478 Xu RR, Pang WQ (2009) Inorganic Synthesis and Preparative Chemistry. Higher Education Press.
479 <https://doi.org/10.3321/j.issn:1005-281X.2000.04.011>

480 Yi LL, Jiao WT, Chen XN et al (2011) An overview of reclaimed water reuse in China. Journal of
481 Environmental Sciences 23, 1585-1593. [https://doi.org/10.1016/S1001-0742\(10\)60627-4](https://doi.org/10.1016/S1001-0742(10)60627-4)

482 Yoon YH, Nelson JH (1984) Application of Gas Adsorption Kinetics I. A Theoretical Model for
483 Respirator Cartridge Service Life. Am Ind Hyg Assoc J 45, 517-524.
484 <https://doi.org/10.1080/15298668491400205>

485 Yuan L, Chen Y, Liu M et al (2020) Dynamic adsorption and regeneration of phosphorus by modified
486 nano-cellulose. Chemical Industry and Engineering Progress 39, 2907-2914.
487 <https://doi.org/10.16085/j.issn.1000-6613.2019-1757>

488 Zhang L, Wu WT, Liu JY et al (2014) Removal of phosphate from water using raw and activated laterite:
489 batch and column studies. Desalination & Water Treatment 52, 778-783.
490 <https://doi.org/10.1080/19443994.2013.826786>

491 Zhang WX, Dong L, Yan H et al (2011) Removal of methylene blue from aqueous solutions by straw
492 based adsorbent in a fixed-bed column. Chemical Engineering Journal 173, 429-436.
493 <https://doi.org/10.1016/j.cej.2011.08.001>

494 Zhao YF, Su X, Zhang WH et al (2018) Study on the adsorption of phosphorous wastewater by modified
495 diatomite. Applied Chemical Industry 47, 883-886.
496 <https://doi.org/10.16581/j.cnki.issn1671-3206.20180330.083>

497 Zhou JB, Yang SL, Yu JG et al (2011) Novel hollow microspheres of hierarchical zinc-aluminum
498 layered double hydroxides and their enhanced adsorption capacity for phosphate in water. Journal
499 of Hazardous Materials 192, 1114-1121. <https://doi.org/10.1016/j.jhazmat.2011.06.013>

500

Figures

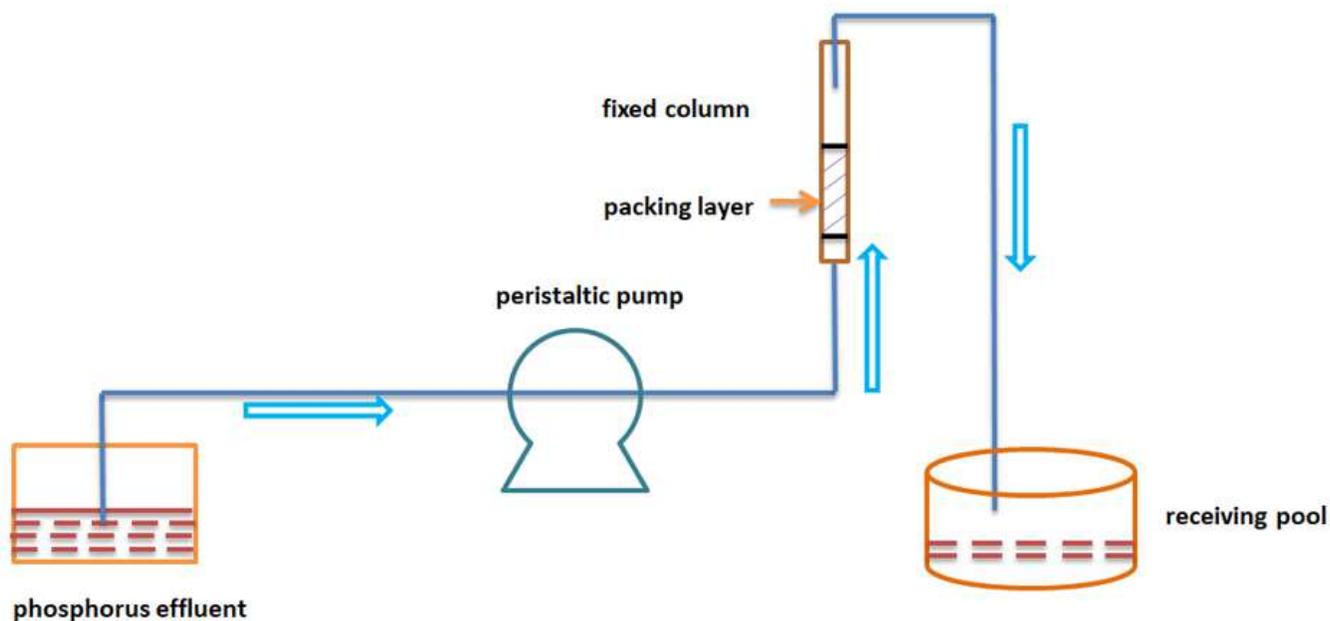


Figure 1

Simple diagram of phosphorus removal experimental apparatus

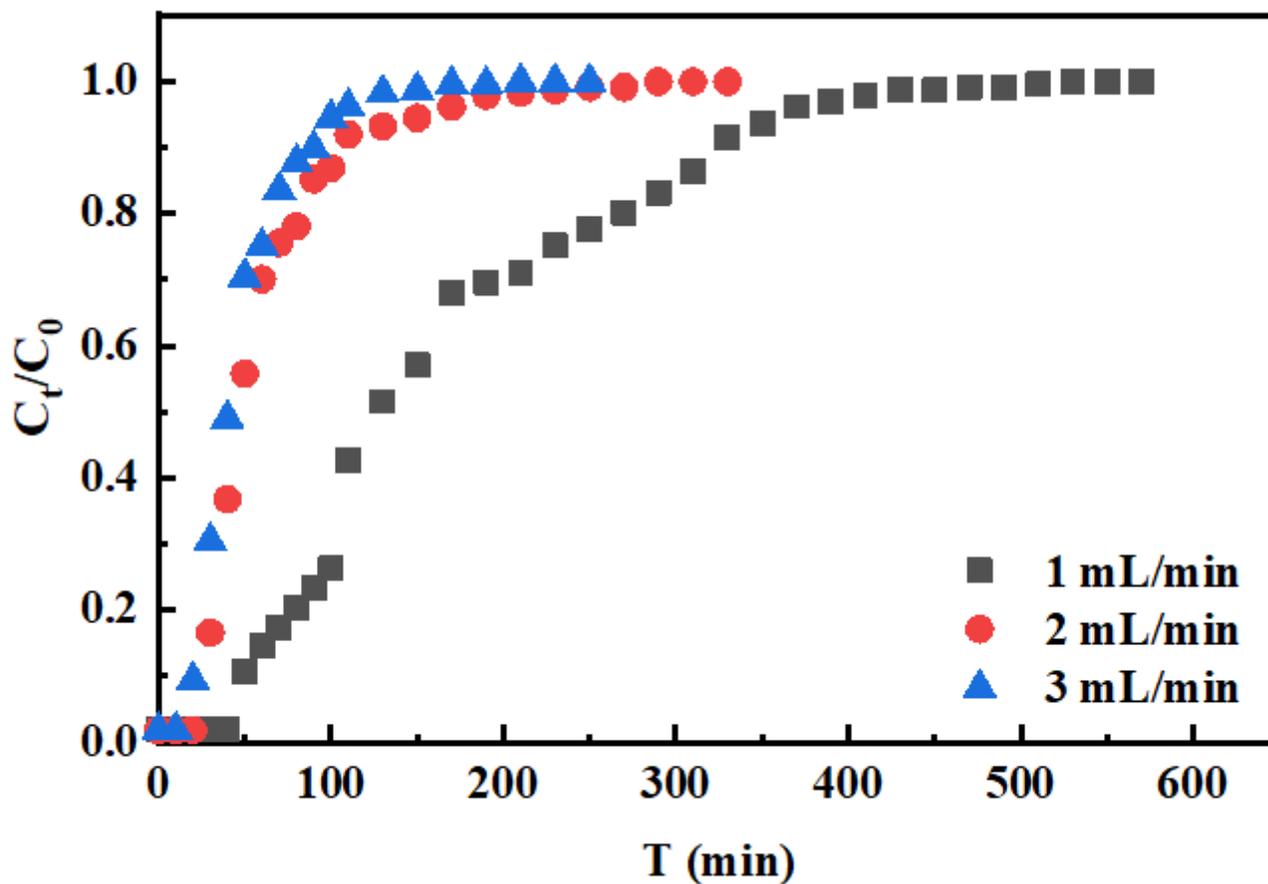


Figure 2

Effect of flow rate on the breakthrough curve of phosphate adsorption onto U-D-Na (natural pH, initial phosphate concentration of 25 mg/L, bed height of 3 cm)

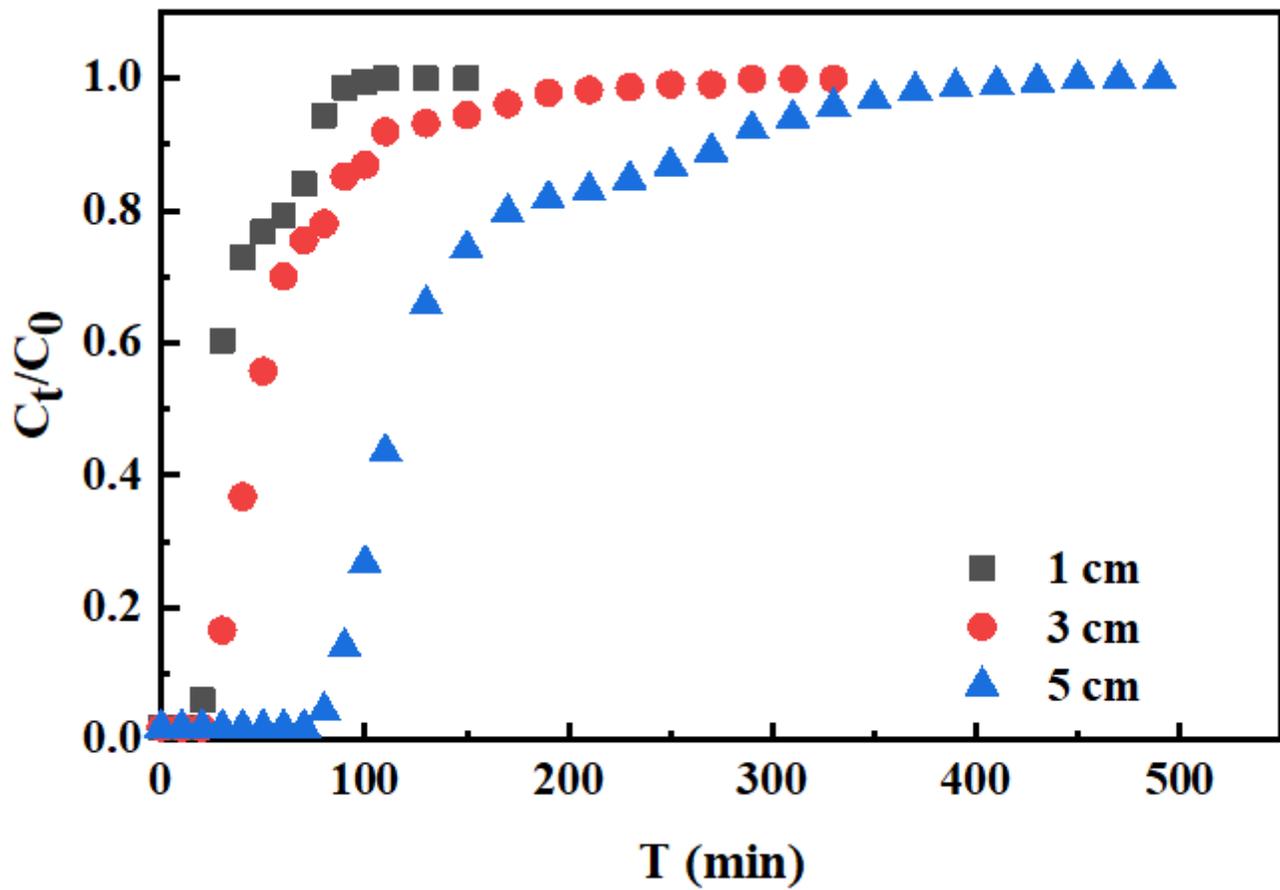


Figure 3

Effect of bed height on the breakthrough curve of phosphate adsorption onto U-D-Na (natural pH, flow rate of 2 mL/min, initial phosphate concentration of 25 mg/L)

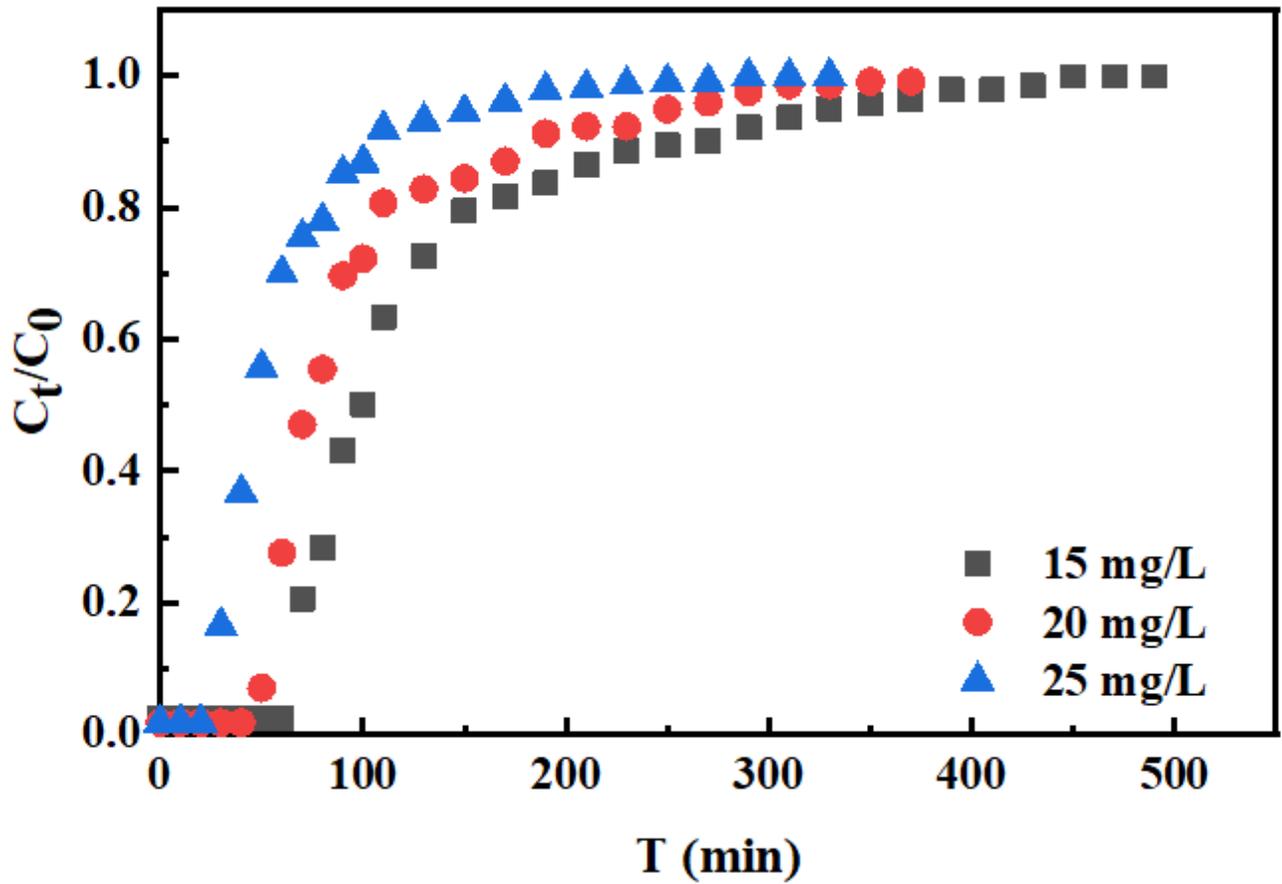


Figure 4

Effect of influent phosphorus concentration on the breakthrough curve of phosphate adsorption onto U-D-Na (natural pH, flow rate of 2 mL/min, bed height of 3 cm)

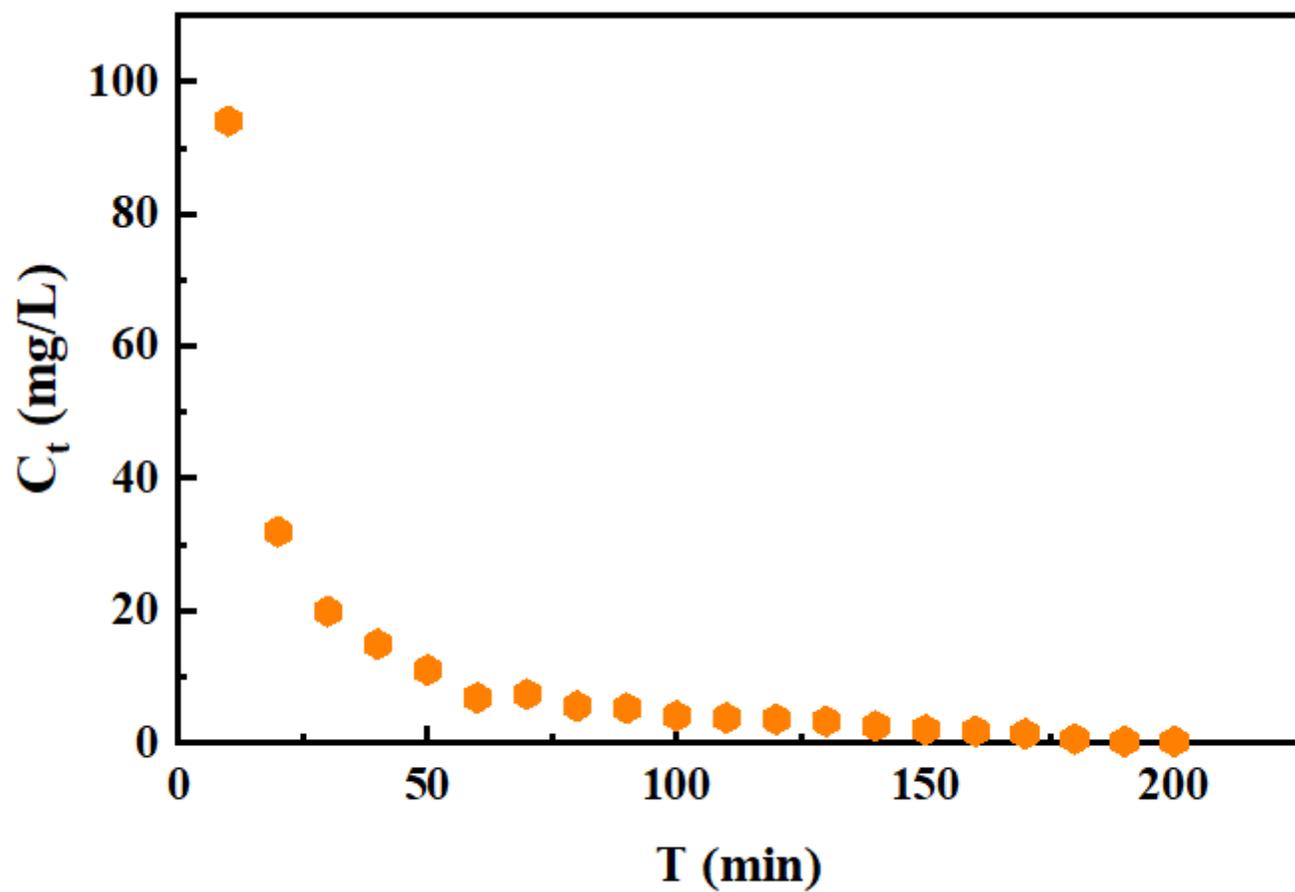


Figure 5

Change of solution concentration with elution time (Eluent of 7.00% HCl, flow rate of 2.00 ml/min)

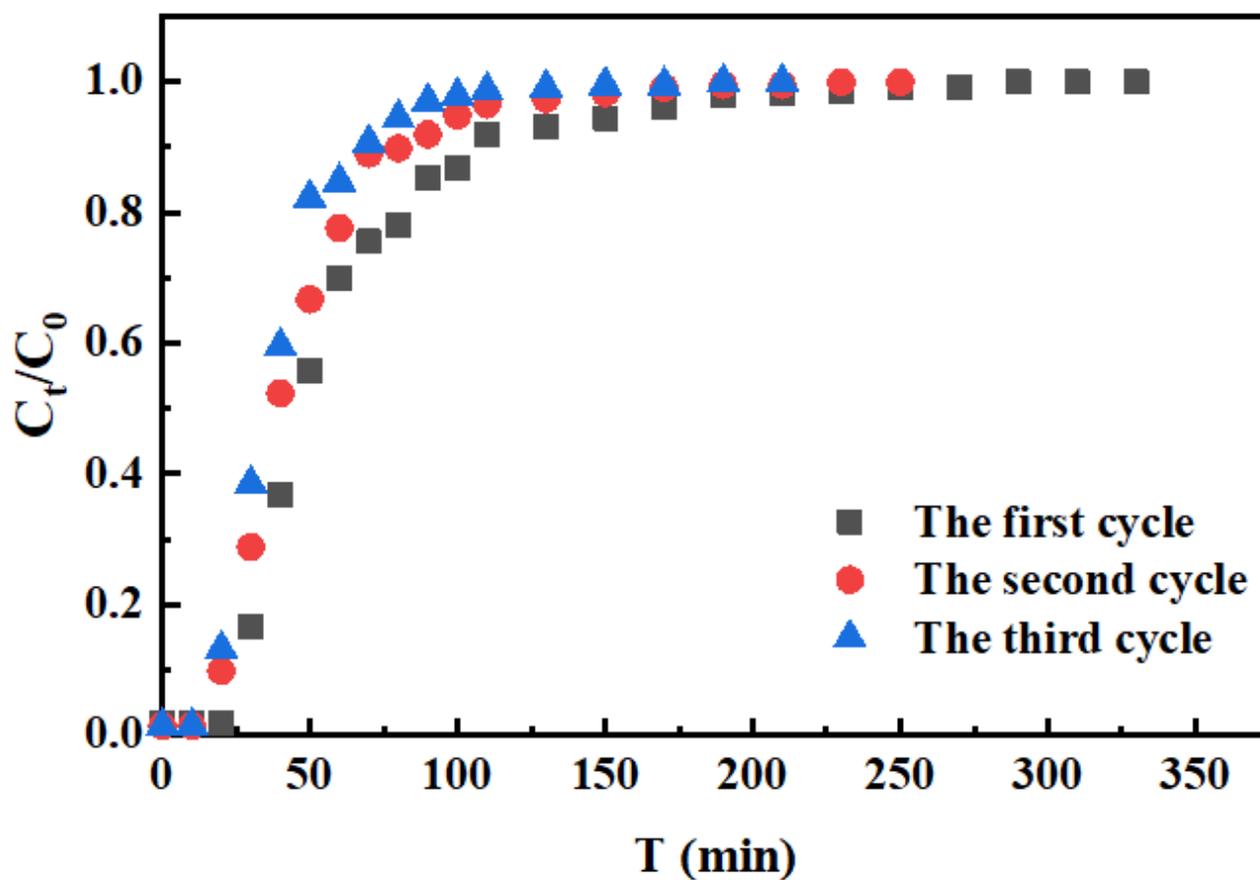


Figure 6

Influence of the number of regenerations in the column on the breakthrough time

Supplementary Files

This is a list of supplementary files associated with this preprint. Click to download.

- [SupplementaryMaterials.docx](#)