

Spatial Scaling of Land Use/Land Cover And Ecosystem Services Across Urban Hierarchical Levels: Patterns And Relationships

Xiao Sun

Chinese Academy of Agricultural Sciences

Qun Ma (✉ maqun0127@shnu.edu.cn)

Shanghai Normal University

Research Article

Keywords: Ecosystem services, Land use/land cover, Scaling relations, Urban agglomeration, 44 Urban hierarchical levels

Posted Date: May 6th, 2021

DOI: <https://doi.org/10.21203/rs.3.rs-425853/v1>

License:  This work is licensed under a Creative Commons Attribution 4.0 International License.

[Read Full License](#)

Version of Record: A version of this preprint was published at Landscape Ecology on January 31st, 2022. See the published version at <https://doi.org/10.1007/s10980-021-01387-4>.

1 **Spatial scaling of land use/land cover and ecosystem services**
2 **across urban hierarchical levels: Patterns and relationships**

3 Xiao Sun, Qun Ma*

4

5 Xiao Sun

6 Key Laboratory of Agricultural Remote Sensing (AGRIRS), Ministry of Agriculture and Rural
7 Affairs/Institute of Agricultural Resources and Regional Planning, Chinese Academy of
8 Agricultural Sciences, Beijing 100081, China

9 Email: sunxiao@caas.cn

10

11 Xiao Sun

12 State Key Laboratory of Urban and Regional Ecology, Research Center for Eco-Environmental
13 Sciences, Chinese Academy of Sciences, Beijing 100085, China

14

15 Qun Ma*

16 School of Environmental and Geographical Sciences, Shanghai Normal University, Shanghai
17 200234, China

18 Email: maqun0127@shnu.edu.cn

19

20 **Abstract**

21 *Context* Land use/land cover (LULC) patterns seriously affect the ecosystem services (ESs),
22 especially in highly developed urban agglomerations. Exploring how LULC and ESs change
23 spatially across urban hierarchical levels and understanding the possible mechanisms can promote
24 the sustainable planning of urban landscapes.

25 *Objectives* By mapping the spatial patterns of LULC and ESs in three largest urban
26 agglomerations of China, this study aimed to (1) identify the scaling relations of LULC and ESs
27 across different urban hierarchical levels, (2) explore the possible mechanisms of these two types
28 of spatial scaling, and (3) explore how the scaling relations of ESs response to LULC and its
29 policy implications.

30 *Methods* Based on LULC, we used the Integrated Valuation of Ecosystem Services and
31 Tradeoffs (InVEST) model and other biophysical models to quantify ES indicators. Then,
32 scalograms were used to quantify the scaling relations of LULC and ESs with respect to changing
33 spatial extent.

34 *Results* Developed land and cropland exhibited the most predictable responses with changing
35 spatial extent. Compared to other ESs, provisioning services were the most predictable. The
36 predictable scaling relations of ESs at different urban hierarchical levels fell into two general types:
37 power laws at the city proper level and exponential relationships at the metropolitan region and
38 urban agglomeration levels.

39 *Conclusions* The scaling relations of both LULC and ESs varied across urban hierarchical levels.
40 The spatial scaling of ESs was closely related to LULC patterns. Integrating the scaling relations
41 of ESs into land use planning can help decision-makers formulate multi-scale landscape
42 conservation strategies.

43

44 **Keywords** Ecosystem services · Land use/land cover · Scaling relations · Urban agglomeration ·
45 Urban hierarchical levels

46

47 **Introduction**

48 For the past few decades, many parts of the world have been undergoing dramatic and rapid
49 urbanization (Stokes and Seto, 2019). As the world's most populous country, China has also
50 experienced dramatic urban expansion and socioeconomic development since the national reform
51 and opening up policies of 1978 (Chen et al. 2020; Liu et al. 2020). Various studies have used
52 population density (Normile 2016; Poku-Boansi 2021) or built-up land area (Seto et al. 2012; Xu
53 et al. 2018) to represent or measure urbanization. The urbanization population rate in China
54 increased from 17.9% in 1978 to 59.6% in 2018 (Wu et al. 2020), and the urban land area
55 expanded from 21,770 km² to 74,827 km² in the same period (Kuang 2020a). This rapid
56 urbanization has caused dramatic changes in land use/land cover (LULC) compositions and
57 structures with high spatial heterogeneity (Hasan et al. 2020; Lawler et al. 2014). As the basis of
58 assessing ESs, LULC changes have resulted in myriad impacts on natural resources and further
59 threatening ESs from a local to global scale, such as with water scarcity (Li et al. 2020), climate
60 change (Patra et al. 2018), soil erosion (Hu et al. 2019), and habitat loss (Swenson and Franklin
61 2000). Studying how LULC and ESs change spatially in rapidly urbanizing regions is important
62 for better understanding and predicting the patterns and processes of urbanization across multiple
63 scales, and the findings may further benefit urban sustainable planning and ecological protection
64 (Batty 2008; Wu et al. 2011; Zhao et al. 2018b).

65 Previous studies have explored the spatial characteristics of LULC patterns and ESs (Hasan et
66 al. 2020; Viglizzo et al. 2012). However, we still lack further information about the spatial scale
67 effects of LULC and ESs, and their underlying mechanisms. The hierarchical scaling strategy
68 provides an effective method of investigating spatial heterogeneity of complex urban systems over
69 a range of scales (O'Neill et al. 1986; Wu 1999; Wu and David 2002). First, urban systems exhibit
70 a hierarchically structured system wherein large urban regions consist of individual cities, which
71 in turn consist of a smaller city proper (Ma et al. 2019). Thus, the hierarchical perspective can
72 systematically characterize the impacts of landscape patterns on ecological processes in urban
73 regions (Bian et al. 2020; Ma et al. 2016b; Zhang et al. 2020b). Second, the scaling approach is a
74 powerful tool that can quantify multiscale characteristics explicitly and describe how various
75 urban attributes scale with city size (Bettencourt 2013; Lobo et al. 2020).

76 Until now, the scaling approach has predominantly been used to focused on two themes: The
77 first theme is how urban metrics (e.g., physical, demographic, economic, or environmental
78 attributes) change within city sizes (usually represented by the population size or urban area)
79 (Brock 1999; Fuller and Gaston 2009; Zhao et al. 2018a). Previous studies illustrated that most
80 urban attributes follow approximate power law scaling, with different scaling exponents (e.g., >1
81 for certain quantities reflecting wealth creation and innovation and <1 for certain material
82 infrastructural quantities) (Bettencourt et al. 2007; Bettencourt et al. 2010; Ribeiro et al. 2020).
83 The second theme is how landscape patterns (e.g., landscape shape) or ecological processes (e.g.,

84 the relations between impervious surface area and land surface temperature) change with spatial
85 scale (e.g., grain size or extent). For example, Wu (2004) systematically investigated the responses
86 of landscape metrics to changing scale and classified the responses into two categories: simple
87 scaling functions and unpredictable behavior. In general, most of the existing studies have focused
88 on landscape patterns (Kedron et al. 2018; Wu et al. 2004; Zhang et al. 2009) but neglected to
89 explore the scaling relations of ecological process and ES indicators. In the studies on landscape
90 patterns, researchers rarely investigate the specific scaling relations for different LULC types.
91 Especially, it is still unclear how the scaling relations of different LULC types and multiple ESs
92 change in rapidly urbanized regions with a hierarchical structure.

93 Urban agglomeration is a highly developed spatial form of integrated cities and exhibits a
94 hierarchically structured system (Ma et al. 2019; Wang et al. 2019b). As urban complex systems,
95 urban agglomerations are suffering the most serious changes in terms of LULC and ESs (Shen et
96 al. 2020; Yu et al. 2019; Zhou et al. 2018). Thus, we selected the top three urban agglomerations in
97 China, the Beijing-Tianjin-Hebei (BTH), the Yangtze River Delta (YRD), and the Pearl River
98 Delta (PRD) areas (Du et al. 2018), to explore the spatial features of LULC and ESs. The scaling
99 evolutions of various LULC types and ESs can help to predict or speculate the indicator values,
100 from one scale to another unknown scale (Xu et al. 2020). In addition, understanding the possible
101 processes and mechanisms behind the scaling functions of urban metrics is helpful for guiding
102 land use management and sustainable policy practice at different urban hierarchical levels
103 globally.

104 This study used the hierarchical scaling strategy to investigate how different LULC types and
105 ESs change within the spatial extent of the three largest urban agglomerations in China. We aimed
106 to achieve the following research objectives: (1) analyze the scaling relations of LULC and ESs
107 across different urban hierarchical levels; (2) understand the underlying mechanisms behind the
108 scaling relations of LULC and ESs; and (3) explore how the scaling relations of ESs respond to
109 LULC and its policy implications.

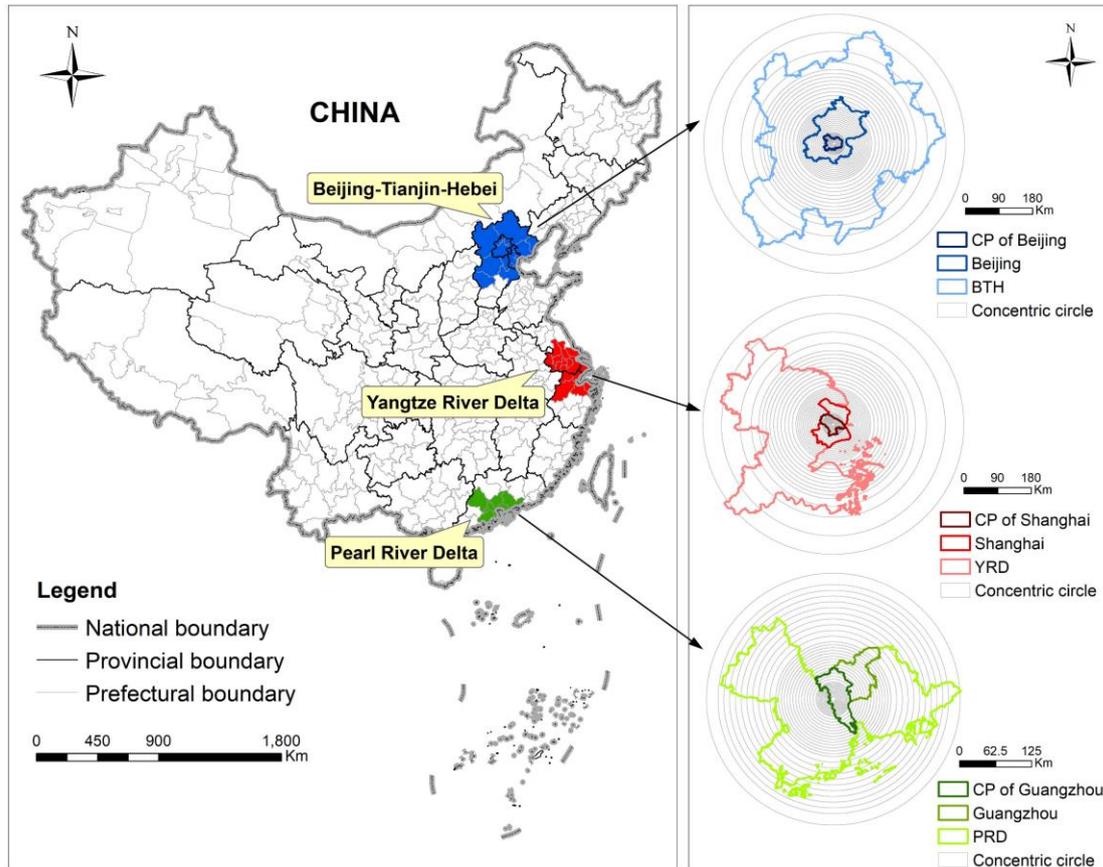
110 **Methods**

111 *Study area*

112 The top three urban agglomerations, BTH, YRD, and PRD (Fig. 1), cover only 5% of Chinese
113 territory (DUSNB, 2019) but generate 40% of the total national urban impervious surface (Ma et
114 al. 2019). They are the primary carriers of China's socioeconomic development and new
115 urbanization processes (Fang and Yu 2017). In 2018, BTH, YRD, and PRD accounted for 8.1%,
116 11.0%, and 5.1% of the total national urban population and 9.5%, 16.7%, and 9.2% of the total
117 gross domestic product (GDP), respectively (DUSNB 2019). In addition to their importance to
118 China's economic and social development, these three urban agglomerations constitute the typical
119 and core regions of prominent environmental issues in China (Zhang et al. 2017; Gao et al. 2019).

120 Administrative urban hierarchy is the most common and actual ecosystem management

121 boundary for regional planning and policy implementation (Forgione et al. 2016; Nikodinoska et
 122 al. 2018; Xing et al. 2021). BTH, YRD, and PRD consist of mature and complete nested urban
 123 hierarchical levels that extend from the city proper, metropolitan region to the broader urban
 124 agglomeration (Ma et al. 2019). The lowest city proper level (the city proper of Beijing, Shanghai,
 125 or Guangzhou) belongs to a metropolitan region (Beijing, Shanghai, or Guangzhou metropolitan
 126 region), and is also embedded in the highest urban agglomeration level (BTH, YRD, or PRD
 127 urban agglomeration) (Fig. 1).
 128



129
 130 **Fig. 1** Location and administrative hierarchy of study areas. The concentric circle illustrations on
 131 the right represent the spatial extension of three urban agglomerations at the administrative
 132 hierarchy levels. The CP of Beijing/Shanghai/Guangzhou refer to the city proper of
 133 Beijing/Shanghai/Guangzhou; Beijing/Shanghai/Guangzhou refer to the
 134 Beijing/Shanghai/Guangzhou metropolitan region; BTH/YRD/PRD refer to the
 135 Beijing-Tianjin-Hebei/Yangtze River Delta/Pearl River Delta urban agglomeration.

136 *Data acquisition*

137 For the three urban agglomerations, LULC data with a spatial resolution of 30×30 m in 2018
 138 were downloaded from the Resource and Environment Science and Data Center
 139 (<http://www.resdc.cn/>). Digital elevation model data with a spatial resolution of 30×30 m were
 140 derived from the Geospatial Data Cloud (<http://www.gscloud.cn/>). Climate data, including the

141 annual average precipitation and temperature in all meteorological stations, were collected from
142 the China Meteorological Data Service Center (<http://data.cma.cn/>). Potential evapotranspiration
143 data were collected from the Global Aridity and PET Database
144 (<https://cgiarcsi.community/data/global-aridity-and-pet-database/>). Soil properties data, including
145 bulk density, soil organic carbon, clay content, sand content, silt content, and soil depth, were
146 obtained from the World Soil Information database (<https://soilgrids.org/>).

147 *LULC maps in urban agglomerations*

148 LULC maps in 2018 for the BTH, YRD, and the PRD urban agglomerations were generated by
149 researchers at the Institute of Geographic Sciences and Natural Resources Research, Chinese
150 Academy of Sciences, through manual visual interpretation. The original data were Landsat 8
151 OLI_TIRS remote-sensing images that were downloaded from the United States Geological
152 Survey (<https://www.usgs.gov/>). The raster data comprised six LULC types: (1) cropland,
153 including irrigated cropland, dry cropland, and orchards; (2) forests, including natural forests,
154 artificial forests, coppice, shrub forests, and sparse forests; (3) grassland, including all types of
155 grasslands with coverage of more than 5%, pastures, and hay; (4) water bodies, including rivers,
156 lakes, reservoirs, ponds, swamps, tidal lands, and wetlands; (5) developed land, including urban
157 construction land, rural residential land, industrial land, traffic road, airport, and mining area; and
158 (6) barren land, including sandy land, desert, saline land, swampland, bare land, bare rock, and
159 other unutilized lands.

160 *Quantifying multiple ES indicators in urban agglomerations*

161 This study selected eight ES indicators: carbon storage, food production, water yield, air
162 pollution removal, nitrogen retention, soil retention, habitat quality, and recreational opportunity.
163 The selection was based on three criteria: (1) the indicators belong to the basic provisioning,
164 regulating, and cultural services (Millennium Ecosystem Assessment 2005); (2) they have great
165 significance to the sustainable development of three urban agglomerations (Liu et al. 2018; Luo et
166 al. 2020; Sun et al. 2018; Zhang et al. 2017); and (3) they can comprehensively reflect the
167 characteristics of complex urban ecosystems and represent the information of water resources,
168 food, climate, soil, habitat, and culture in urban agglomerations.

169 We used the Integrated Valuation of Ecosystem Services and Tradeoffs (InVEST) model to
170 quantify the carbon storage, water yield, nitrogen retention, soil retention, and habitat quality for
171 the three urban agglomerations. Food production was estimated by using the Global
172 Agro-Ecological Zones (GAEZ) model. The removal of PM_{2.5} (atmospheric aerosol particles with
173 a diameter of less than 2.5 μm) was estimated on the basis of the adsorption capacity of vegetation
174 to PM_{2.5} pollutants. Recreational opportunity was quantified by the Recreation opportunity
175 Spectrum (ROS) model. Table 1 listed the ES indicators and assessment descriptions. In addition,
176 the calculation process and detailed parameters were shown in Supplementary Information (SI).

177

178

Table 1

179

Descriptions of ecosystem service (ES) indicators and the assessment methods

ES types	Principles and descriptions of ES assessment models
Carbon storage	The InVEST Carbon Storage and Sequestration model maps the carbon densities of different land use/land cover (LULC) types and summarizes the results into aggregate storage values (Goldstein et al. 2012).
Food production	The food production potential was calculated by considering the limiting factors, such as water resource, soil properties, topographic conditions, cultivated land distribution, and management measures (Liu et al. 2015).
Water yield	The InVEST Water Yield model estimates the water of different landscape areas. It is based on the principle of water balance on a grid map (Sharp et al. 2020).
PM _{2.5} removal	The PM _{2.5} removal was calculated on the basis of the adsorption of PM _{2.5} per unit area for different LULC types.
Nitrogen retention	The InVEST Nutrient Delivery Ratio model assesses nitrogen retention according to the nutrient pollutants removed by vegetation and soil in surface runoff (Sharp et al. 2020).
Soil retention	The InVEST Sediment Delivery model computes sediment retention based on the amount of soil loss and its delivery ratio (Hamel et al. 2015).
Habitat quality	The InVEST Habitat Quality model combines the information of LULC types and threats to biodiversity to produce habitat quality maps (Sharp et al. 2020).
Recreational opportunity	The ROS model was associated with the recreational index and accessibility index for reaching the recreational locations (Lavorel et al. 2020).

180

181

Quantifying the scaling relations of LULC and ESs with respect to spatial extent

182

ES supply depends on ecosystem structures and processes. Thus, LULC change has become a major driving force that impacts regional ecosystem change (Hasan et al. 2020). Both LULC compositions and configuration are expected to significantly impact ESs (Lei et al. 2021). Therefore, exploring how LULC change responds to increased spatial extent is an important basis for understanding and explaining the spatial scaling effect of ESs. The form of scalograms (Frazier 2016; Ma et al. 2019; Wu 2004) was adopted in this study to reflect how different LULC types and ESs respond to changing spatial extents in three urban agglomerations. A series of concentric circles with gradually expanding radii were used to represent the expansion of spatial extent across urban hierarchical levels (Fig. 1, Table 2). The spatial extent was expressed as the radius of a concentric circle. Thus, scalograms were constructed by plotting the changes of LULC and ESs with respect to increasing spatial extents (Wu 2004).

193

194
195
196

Table 2

The radii list of concentric circles at different administrative hierarchical levels for BTH, YRD, and PRD urban agglomerations (Unit: km).

BTH urban agglomeration			YRD urban agglomeration			PRD urban agglomeration		
CP of Beijing	Beijing	BTH	CP of Shanghai	Shanghai	YRD	CP of Guangzhou	Guangzhou	PRD
4	4	4	4	4	4	4	4	4
6	6	6	6	6	6	6	6	6
8	8	8	8	8	8	8	8	8
10	10	10	10	10	10	10	10	10
12	12	12	12	12	12	12	12	12
14	14	14	14	14	14	14	14	14
16	16	16	16	16	16	16	16	16
18	18	18	18	18	18	18	18	18
20	20	20	20	20	20	20	20	20
22	22	22	22	22	22	22	22	22
24	24	24	24	24	24	24	24	24
26	26	26	26	26	26	26	26	26
28	28	28	28	28	28	28	28	28
30	30	30	30	30	30	30	30	30
	35	35	35	35	35	35	35	35
	40	40	40	40	40	40	40	40
	45	45		45	45	45	45	45
	50	50		50	50	50	50	50
	55	55		55	55	55	55	55
	60	60		60	60	60	60	60
	65	65		65	65	65	65	65
	70	70		70	70	70	70	70
	75	75			75		75	75
	80	80			80		80	80
	85	85			85		85	85
	90	90			90		90	90
	95	95			95		95	95
	100	100			100		100	100
	110	110			110		110	110
	120	120			120		120	120
	130	130			130			130
		140			140			140
		150			150			150
		160			160			160
		170			170			170
		180			180			180
		190			190			190
		200			200			200
		220			220			220
		250			250			
		300			300			
		350			350			

197 Note: The center of concentric circles is the administrative center of a city proper. For each urban hierarchical level,
198 the concentric circle with the maximum radius can cover the boundary of the corresponding level. BTH/YRD/PRD
199 represents the Beijing-Tianjin-Hebei/Yangtze River Delta/Pearl River Delta urban agglomeration, CP of
200 Beijing/Shanghai/Guangzhou refers to the city proper of Beijing/Shanghai/Guangzhou.

201

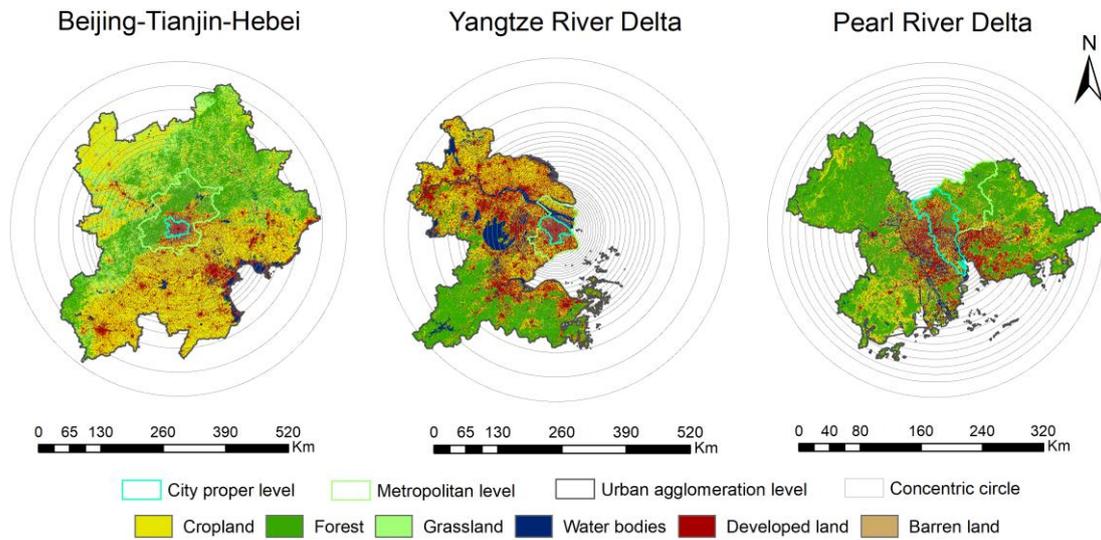
202 Using SPSS 25.0, we adopted the curve estimation regression models (Jomnonkwao et al. 2020)
203 to investigate the scaling relations. The coefficient of determination (R^2) was used to indicate the
204 goodness of fit of the curve (Massada and Radeloff 2010). The independent variable (X) was the
205 radius of the concentric circle (spatial extent). The dependent variable (Y) was the proportion of
206 LULC or the average ES value within the corresponding concentric circle area.

207 **Results**

208 *Scaling relations of LULC for three urban agglomerations*

209 For the three urban agglomerations, proportions of developed land decreased as the spatial
210 extent increased. In contrast, the proportions of forests and cropland increased with the expansion
211 of the spatial scale. The proportion of water bodies showed irregular fluctuations with the
212 expansion of scale. In addition, the proportion of grassland in BTH increased when the spatial
213 scale expanded. Further, the proportions of grassland and barren land were exceedingly small in
214 YRD and PRD (Fig. 2, Fig. 3).

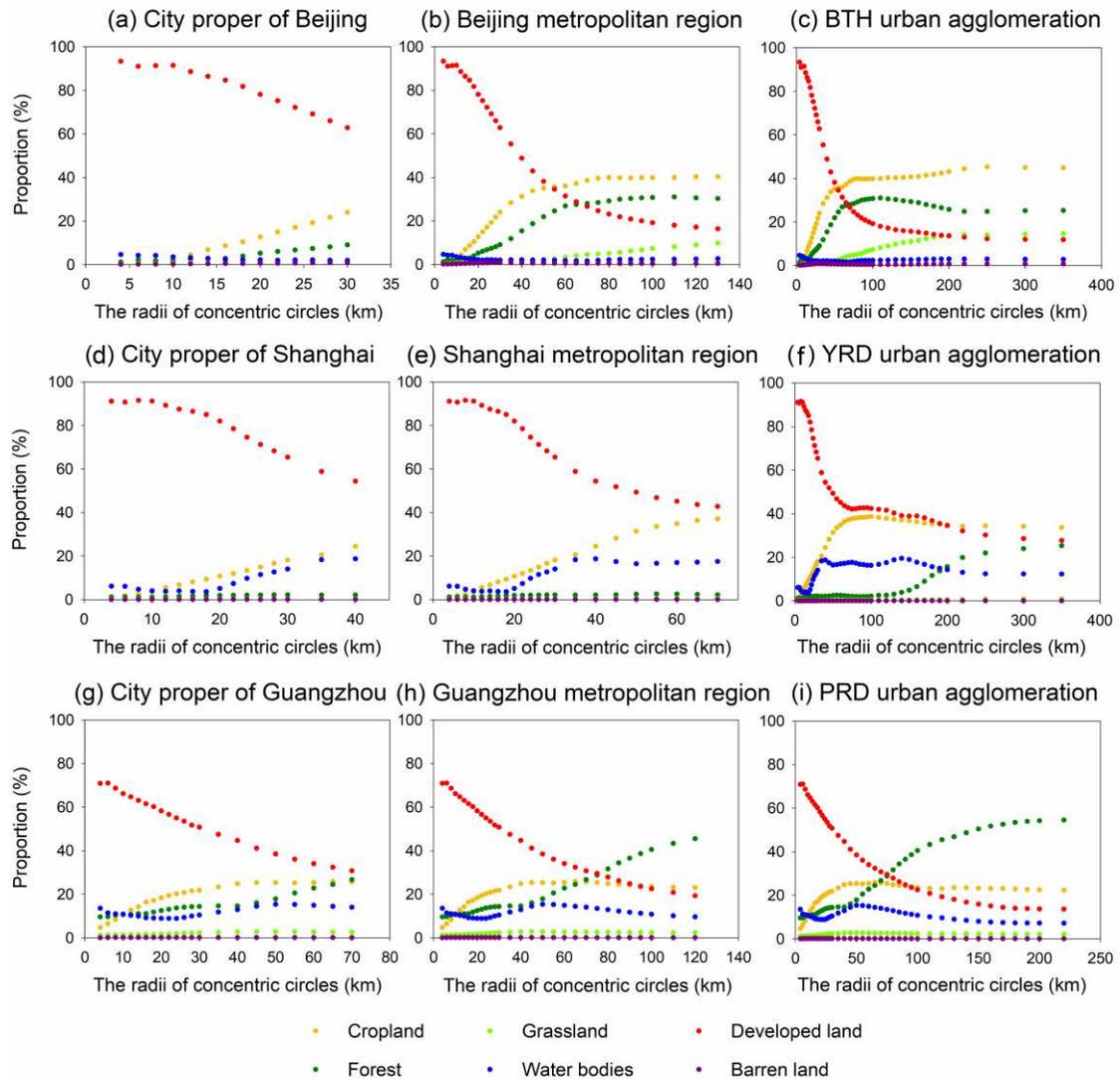
215



216

217 **Fig. 2** Spatial distributions of land use/land cover in 2018 for three urban agglomerations. The
218 radii list of concentric circles is shown in Table 2.

219



220

221 **Fig. 3** Scalograms of the proportion of different land use/land cover types with respect to
 222 increasing concentric circle radii in the three largest urban agglomerations of China
 223 (Beijing-Tianjin-Hebei (BTH), Yangtze River Delta (YRD), and Pearl River Delta (PRD)).

224

225 Overall, the proportions of cropland, forest, and developed land exhibited appropriate and
 226 predictable scaling laws at certain levels of urban hierarchy. The typical scaling curves for
 227 different urban hierarchical levels were classified into three rules: linear, sigmoidal, and
 228 exponential relationships. Among them, the exponential relationships were divided into
 229 exponential growth and exponential decay (Fig. A1). However, the scaling relations of water
 230 bodies and barren land were unpredictable at nearly all three urban hierarchical levels. The
 231 proportion of grassland showed scaling laws in BTH, but this was unpredictable in YRD and PRD
 232 (Table 3).

233 For the BTH area, the linear scaling relations were found at the city proper level for most of the
 234 land use types. Next, the general relationships were converted into exponential and sigmoidal
 235 functions at the broader metropolitan and urban agglomeration levels. For YRD, most of the

236 LULC types did not reveal the scaling relations and functions except for cropland and developed
 237 land. Both showed linear relationships at the city proper level and sigmoidal and exponential
 238 growth or exponential decay at higher urban hierarchical levels, respectively. For PRD, the
 239 cropland, forest, and developed land scaled in a predictable way. The proportions of cropland
 240 increased with exponential curves, while forest increased with sigmoidal curves. The proportions
 241 of developed land decreased in a linear relationship at the city proper level, while it converted to
 242 exponential decay at the higher levels (Table 3).

243

244 **Table 3**

245 Type of scaling relations for proportions of different land use/land cover with respect to increasing
 246 radii of concentric circles within the three largest urban agglomerations

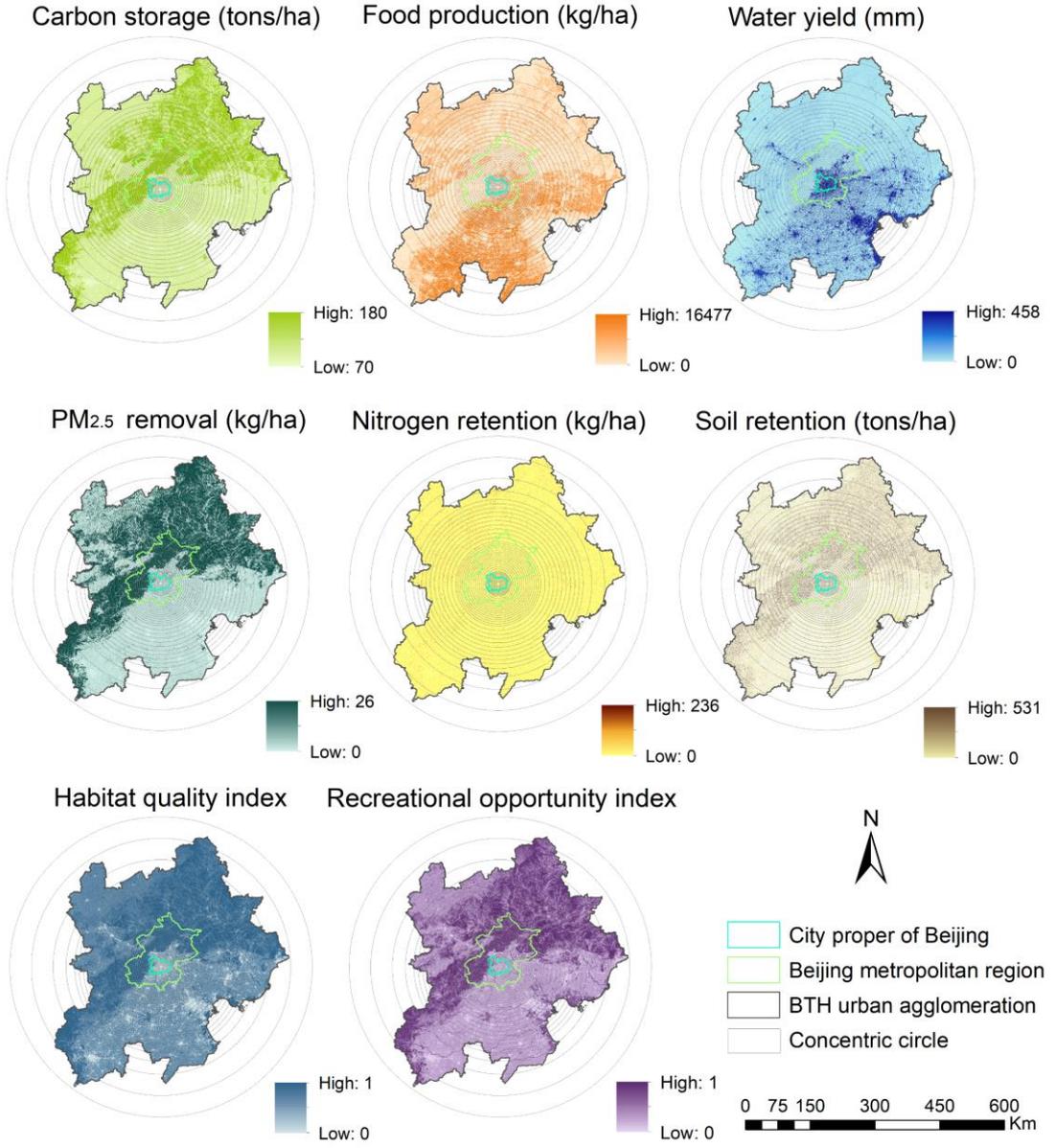
Urban agglomeration	Land use/land cover types	Urban hierarchical levels		
		City proper	Metropolitan region	Urban agglomeration
Beijing-Tianjin-Hebei	Cropland	Linear	Sigmoidal	Exponential growth
	Forest	Linear	Exponential growth	–
	Grassland	Linear	Exponential growth	Sigmoidal
	Water bodies	Exponential decay	–	–
	Developed land	Linear	Exponential decay	Exponential decay
	Barren land	Linear	–	–
Yangtze River Delta	Cropland	Linear	Sigmoidal	Exponential growth
	Forest	–	–	Sigmoidal
	Grassland	–	–	–
	Water bodies	–	–	–
	Developed land	Linear	Exponential decay	Exponential decay
	Barren land	–	–	–
Pearl River Delta	Cropland	Exponential growth	Exponential growth	Exponential growth
	Forest	Sigmoidal	Sigmoidal	Sigmoidal
	Grassland	Sigmoidal	–	–
	Water bodies	–	–	–
	Developed land	Linear	Exponential decay	Exponential decay
	Barren land	–	–	–

247 Note: The archetypes of scaling relations were shown in Fig. A1. For all scaling relation equations in Table 3, the
 248 relationships between the proportions of land use/land cover types and the increasing spatial extents were
 249 significant (**P<0.001, R²>0.990).

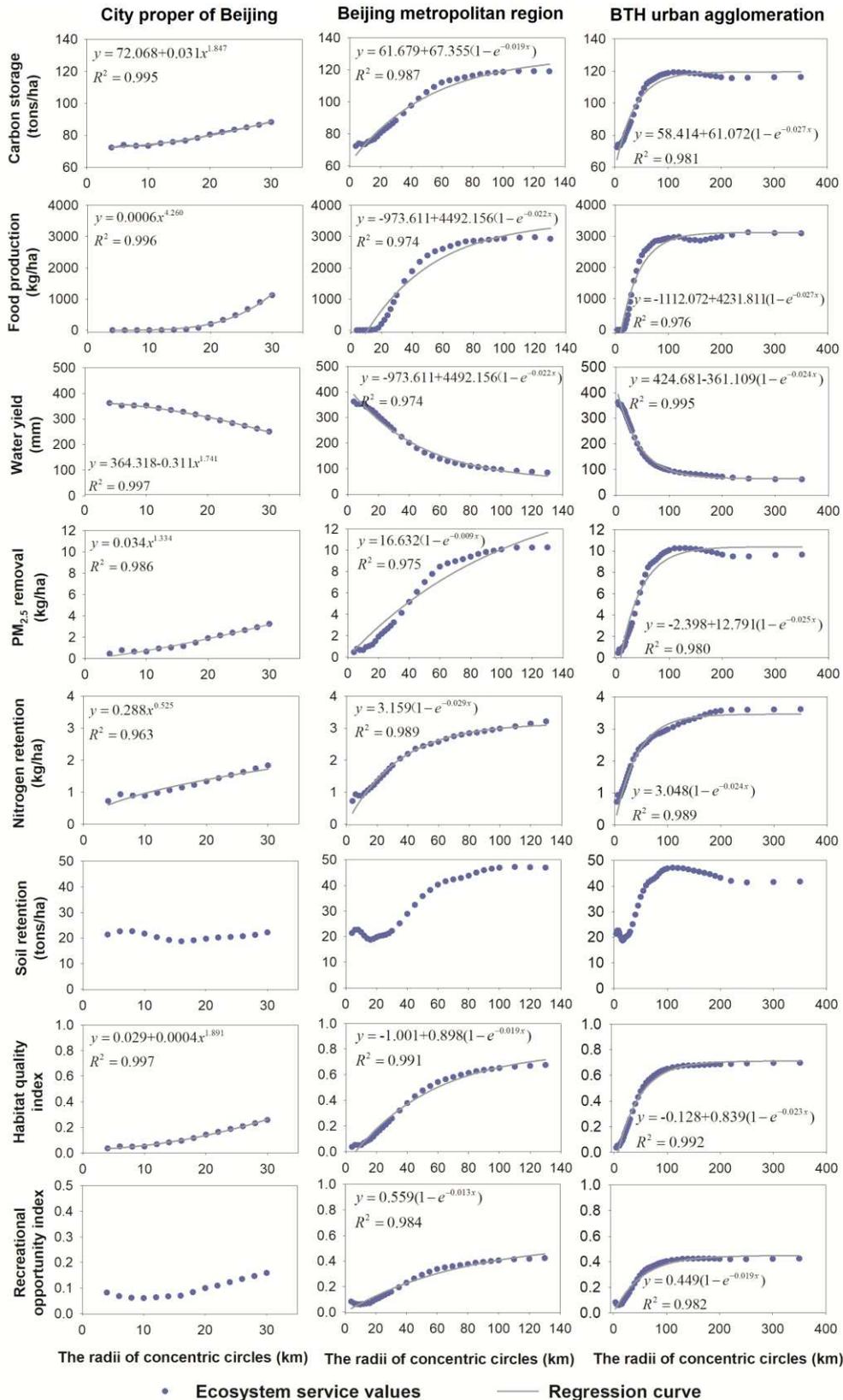
250 *Scaling relations of multiple ESs in BTH*

251 For the spatial patterns of multiple ESs in BTH, the values for most of ESs become higher with
 252 increasing spatial extent, except for water yield and soil retention (Fig. 4). According to the
 253 scalograms of multiple ESs, with respect to increasing spatial extent in BTH, the carbon storage,
 254 food production, water yield, PM_{2.5} removal, nitrogen retention, and the habitat quality revealed
 255 significant and predictable scaling relations at different urban hierarchical levels. However, the
 256 soil retention did not show any scaling relations. In general, the two form functions, which were
 257 power law and exponential functions, can fit the scaling relations of ESs better in the BTH

258 agglomeration. The ESs showed pervasive power law scaling relations at the city proper level
 259 while exhibiting exponential scaling relations at the higher metropolitan and urban agglomeration
 260 levels (Fig. 5).
 261



262
 263 **Fig. 4** Spatial distributions of multiple ecosystem services in 2018 for Beijing-Tianjin-Hebei
 264 (BTH) urban agglomeration. The radii list of concentric circles was shown in Table 2.
 265



266

267

Fig. 5 Scalograms of multiple ecosystem service indicators with respect to increasing concentric

268

circle radii at various urban hierarchical levels: the city proper of Beijing, Beijing metropolitan

269

region, and the Beijing-Tianjin-Hebei (BTH). For all regression curves and equations, the

270

significance level P-values < 0.001.

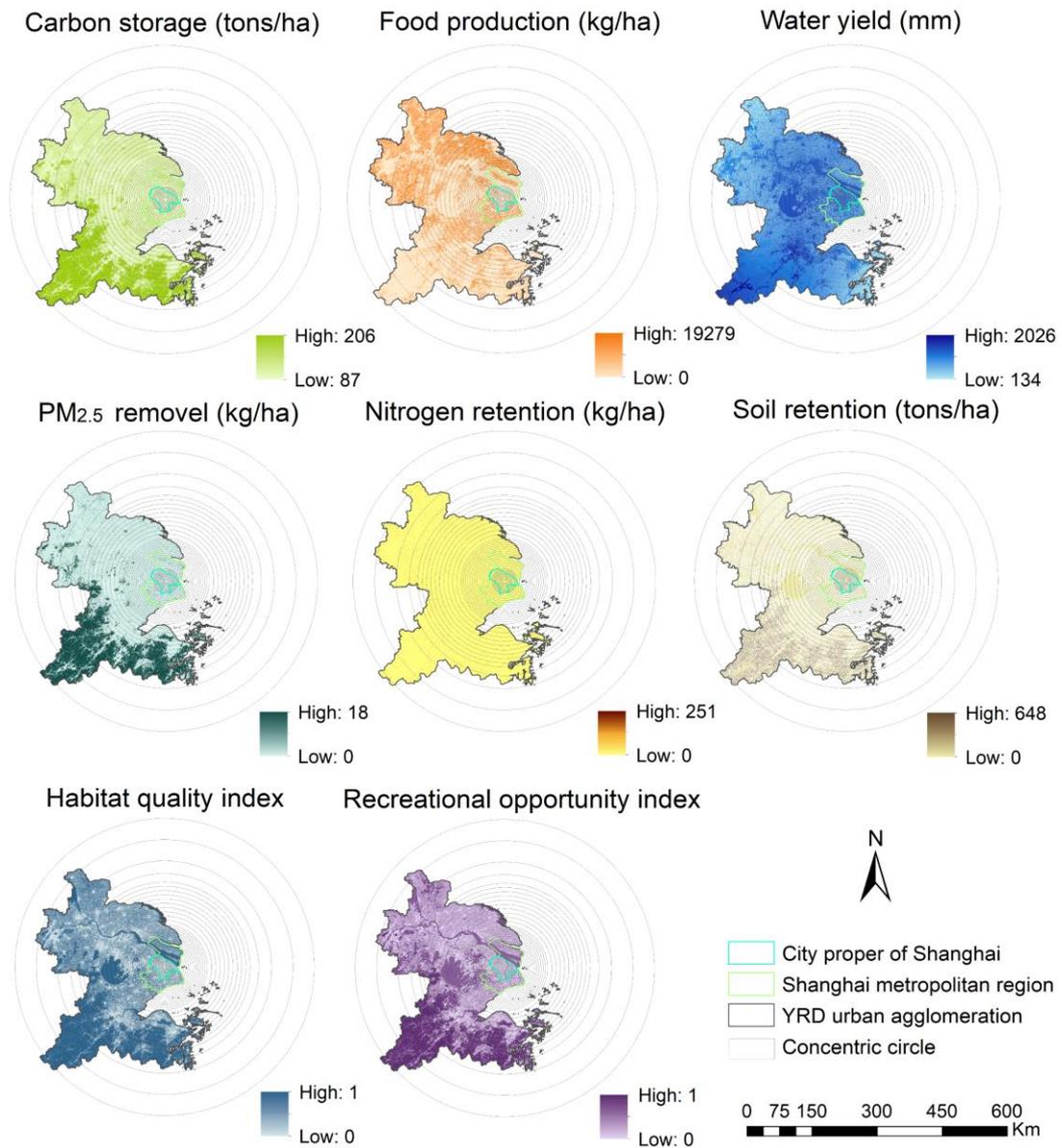
271

272 Specifically, the values of carbon storage, food production, PM_{2.5} removal, and habitat quality
273 increased consistently and followed a power law function with a scaling exponent of greater than
274 1 within the city proper of Beijing. As the spatial extent increased in the metropolitan region and
275 urban agglomeration, the indicator values showed exponential functions with negative exponents.
276 For nitrogen retention, the scaling exponent was smaller than 1 in the power law function within
277 the city proper of Beijing. As the spatial extent increased beyond the city proper, the nitrogen
278 retention values also showed an exponential function with a negative exponent and continued to
279 increase until it eventually remained stable. In addition to these increasing indicators, the water
280 yield showed a decreasing trend with the expansion of spatial extent. The index values followed a
281 power law decline within the city proper of Beijing. They continued to decrease with exponential
282 scaling relations at higher urban hierarchical levels. For recreational opportunities, the index
283 values first decreased and then increased with the increasing spatial extent within the city proper.
284 As the spatial extent increased, the index values followed exponential functions and continued to
285 grow until it stabilized (Fig. 5).

286 *Scaling relations of multiple ESs in YRD*

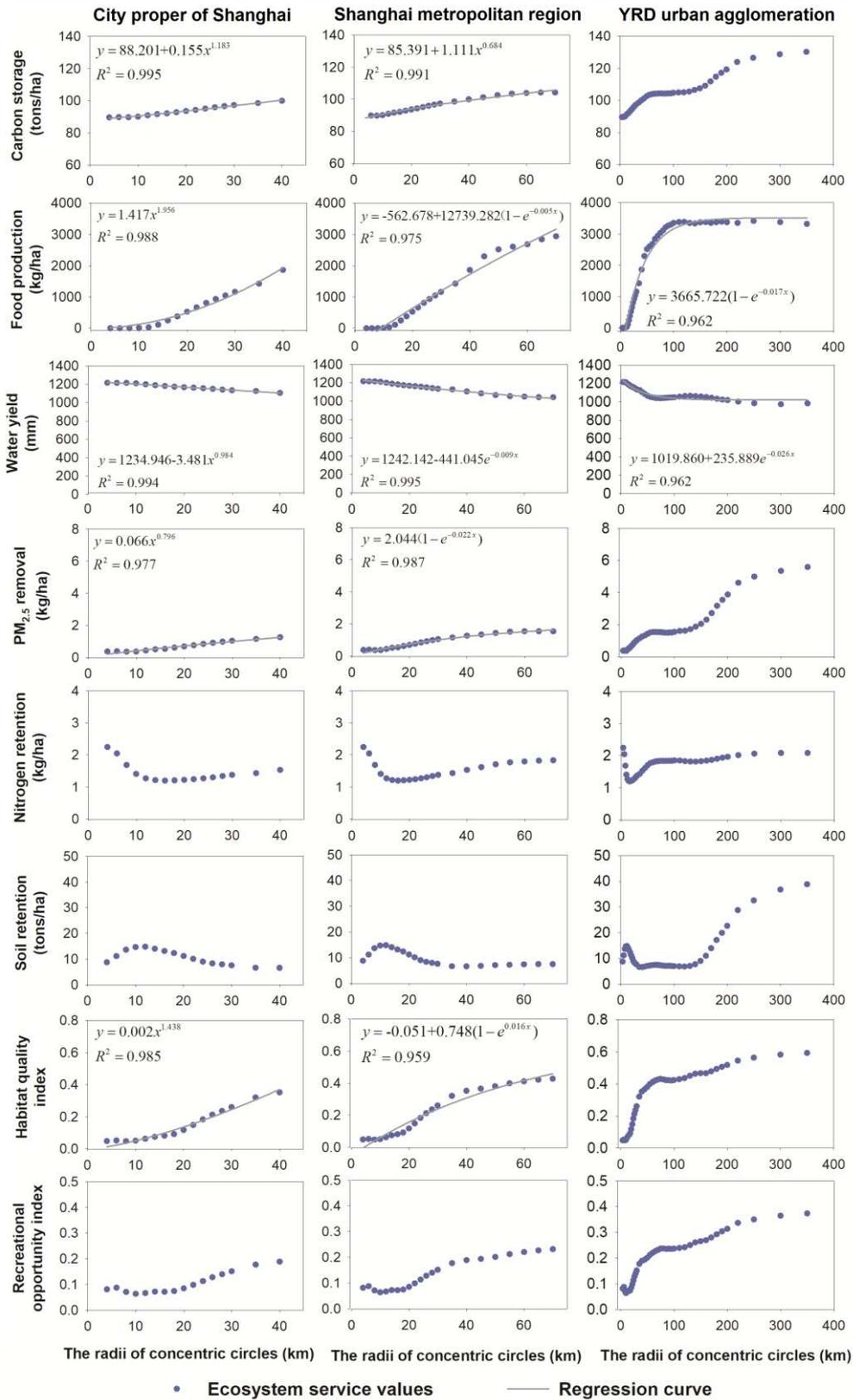
287 For the spatial patterns of multiple ESs in YRD, the values for most of ESs become higher with
288 increasing spatial extent, except for water yield, nitrogen retention, and soil retention (Fig. 6).
289 According to the scalograms of multiple ESs of the increasing spatial extent in YRD, only food
290 production and water yield displayed universal and predictable scaling relations at different urban
291 hierarchical levels. For carbon storage, PM_{2.5} removal, and habitat quality, the index values
292 followed predictable scaling relations within the Shanghai metropolitan region. As the spatial
293 extent further increased beyond the metropolitan region, the scaling relations became
294 unpredictable. In addition, for nitrogen retention, soil retention, and recreational opportunities,
295 there were no predictable scaling relations that could fit the curves (Fig. 7).

296



297
298
299
300

Fig. 6 Spatial distributions of multiple ecosystem services in 2018 for Yangtze River Delta (YRD) urban agglomeration. The radii list of concentric circles was shown in Table 2.



301

302

Fig. 7 Scalograms of multiple ecosystem service indicators with respect to increasing concentric

303

circle radii at various urban hierarchical levels: the city proper of Shanghai, Shanghai metropolitan

304

region, and the Yangtze River Delta (YRD). For all regression curves and equations, the

305

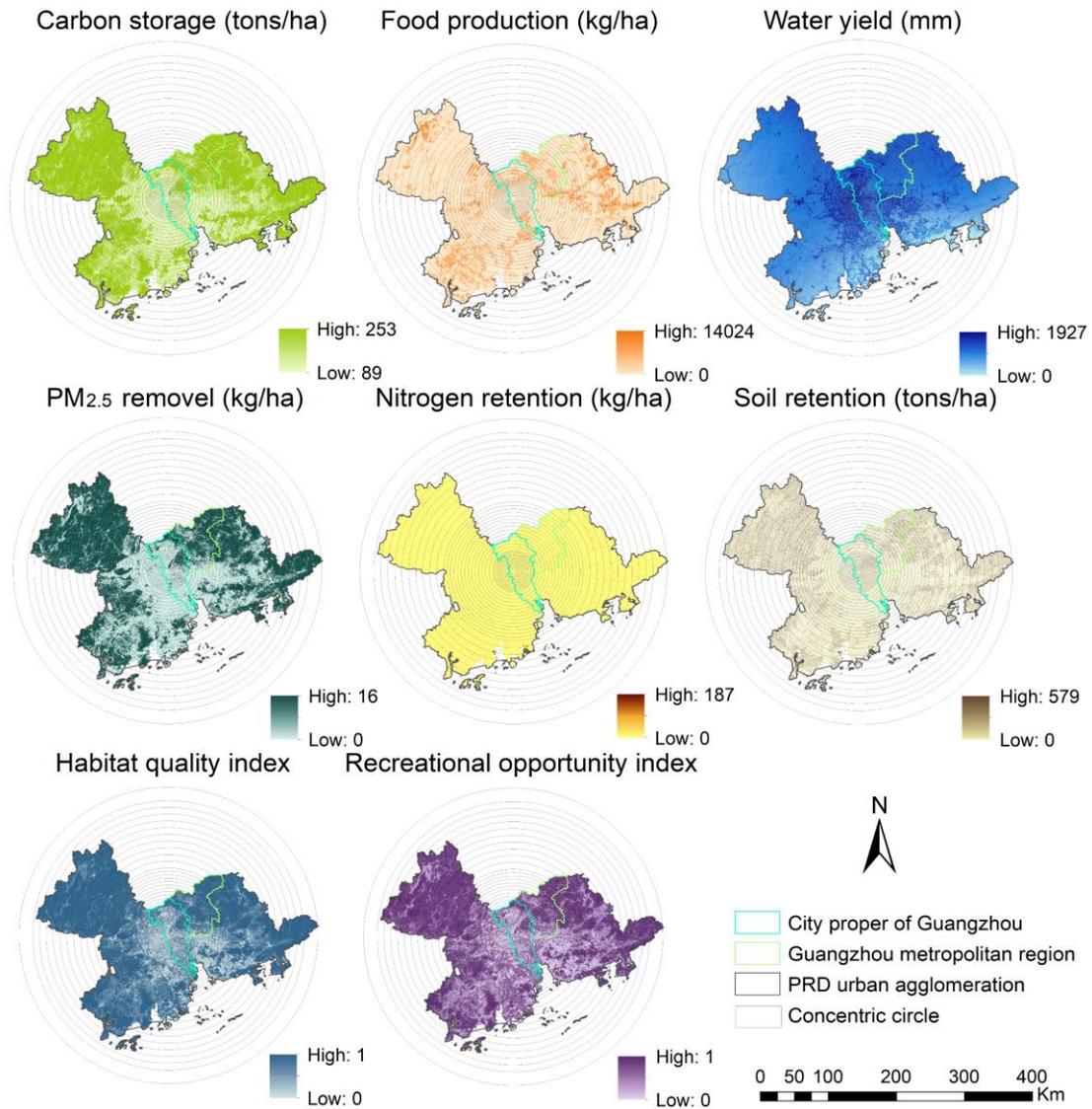
significance level P-values < 0.001.

306 Specifically, food production values increased with a power law scaling relationship at the city
307 proper level, while it showed exponential growth scaling relationships at higher urban hierarchical
308 levels. Water yield values decreased with a power law scaling relation at the city proper level,
309 followed by exponential decay scaling as the spatial extent covered the metropolitan region and
310 urban agglomeration. For carbon storage, PM_{2.5} removal, and habitat quality, the index values
311 exhibited power law scaling relations within the city proper of Shanghai. As the spatial extent
312 further increased to the Shanghai metropolitan region, carbon storage still followed a power law
313 function, while PM_{2.5} removal and habitat quality followed an exponential growth relationship.
314 The values of these three ESs showed upward staircase-like curves when the spatial extent
315 increased to urban agglomeration. In addition, the values of nitrogen retention did not change
316 significantly at various urban hierarchical levels. The soil retention values performed best when
317 spatial extent increased beyond the metropolitan region. The recreational opportunity index values
318 showed an upward staircase-like curve with increasing spatial extent (Fig. 7).

319 *Scaling relations of multiple ESs in PRD*

320 For the spatial patterns of multiple ESs in PRD, the values for most of ESs become higher with
321 increasing spatial extent, except for water yield and soil retention (Fig. 8). According to the
322 scalograms of multiple ESs, with respect to increasing spatial extent in PRD, the carbon storage,
323 food production, water yield, PM_{2.5} removal, nitrogen retention, and habitat quality all exhibited
324 predictable scaling relationships at different urban hierarchical levels. As with the BTH
325 agglomeration, the scaling relationships of soil retention in PRD were unpredictable. The
326 recreational opportunity showed a predictable scaling relation only at the urban agglomeration
327 level. In general, the scaling relationship also presented two types of functions, which were power
328 law and exponential relationships. Moreover, exponential relationships predominantly existed at
329 higher urban hierarchical levels (Fig. 9).

330

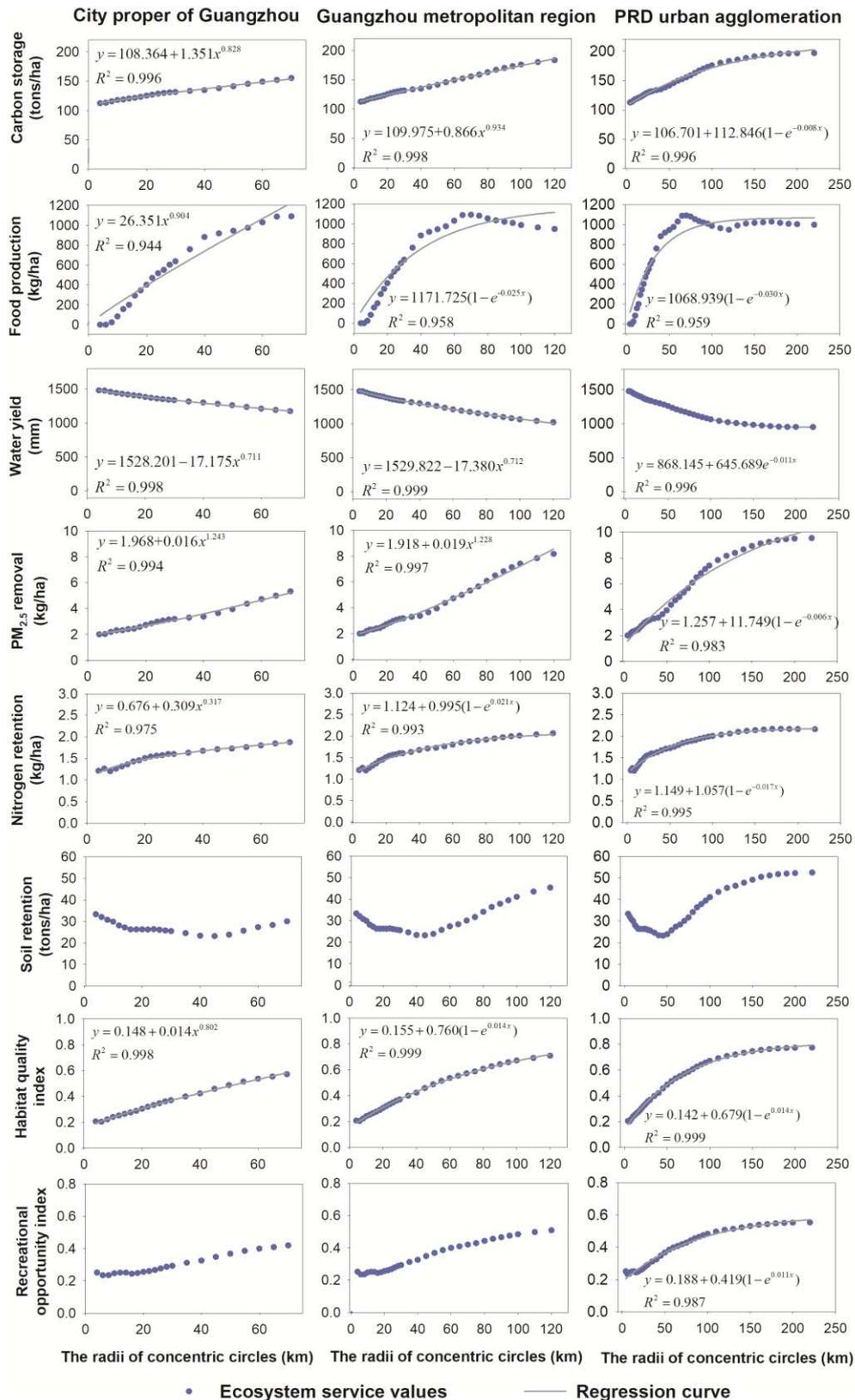


331

332 **Fig. 8** Spatial distributions of multiple ecosystem services in 2018 for Pearl River Delta (PRD)

333 urban agglomeration. The radii list of concentric circles was shown in Table 2.

334



335

336 **Fig. 9** Scalograms of multiple ecosystem service indicators with respect to increasing concentric
 337 circle radii at various urban hierarchical levels: the city proper of Guangzhou, Guangzhou
 338 metropolitan region, and the Pearl River Delta (PRD). For all regression curves and equations, the
 339 significance level P-values < 0.001.

340 Carbon storage and PM_{2.5} removal values followed power law functions within the Guangzhou
341 metropolitan region. Thereafter, they showed exponential functions as the spatial extent further
342 increased to the urban agglomeration. The values of food production, nitrogen retention, and
343 habitat quality followed power law functions within the city proper and then increased
344 exponentially as the spatial extent increased beyond the city proper. For water yield, the index
345 values showed a monotonously decreasing trend as the spatial extent increased. The index values
346 followed the power law functions within the Guangzhou metropolitan region and then showed the
347 exponential function as the spatial extent increased. For recreational opportunity, the index values
348 began to follow an exponential function only when the spatial extent increased to the urban
349 agglomeration level (Fig. 9).

350 **Discussion**

351 *How did the scaling relations of LULC change across different urban hierarchical levels and*
352 *agglomerations?*

353 Different LULC types showed diversified distribution trends when the spatial extent increased.
354 The developed lands exhibited decreasing trends from city proper outwards when croplands and
355 forests showed an increasing trend. This was mainly because croplands and forests were
356 substantially occupied by developed land during the rapid process of urbanization in China,
357 especially for large urban agglomerations (Zhou et al. 2021; Kuang et al. 2020b). In contrast,
358 water bodies and grasslands showed irregular and random characteristics with the increase of
359 spatial extent. This phenomenon was predominantly related to natural geographical conditions
360 (e.g., topography, climate, etc.), regional planning (e.g., park layouts, the demarcation of nature
361 reserves), and other socioeconomic factors (Luo et al. 2020; Sun et al. 2020; Yang et al. 2019). In
362 addition, barren lands had no obvious scaling relations, as its area within the three highly
363 urbanized agglomerations was exceedingly small.

364 Among all LULC types, developed land and cropland were the most predictable at three urban
365 hierarchical levels (Table 3). Developed land and cropland were most severely affected by human
366 activities, especially in urban centers (d'Amour et al. 2016). Socioeconomic factors such as
367 population agglomeration and gross domestic product growth predominantly affect the spatial
368 pattern of developed land and cropland (Arcaute et al. 2015; Gu 2019; Li et al. 2017). The scaling
369 relations of socioeconomic factors can be well approximated by common laws. For example, since
370 the population density and the Gross Domestic Product followed predictable functions in the three
371 urban agglomerations (Fig. A2), the two LULC types were also spatially predictable. The
372 predictable scaling relationships of this study can provide information for the extrapolation of
373 developed lands and croplands at different spatial amplitudes and urban hierarchical levels.

374 Different urban hierarchical levels presented various scaling relationships for LULC, of which
375 the city proper level was the most predictable. For example, water bodies and barren land showed
376 predictable scaling functions at the city proper level but were unpredictable at the broader

377 metropolitan and urban agglomeration levels (Table 3). This was predominantly because the
378 influencing factors of LULC distributions became more diverse with the gradual expansion of
379 spatial amplitude (i.e., the addition of more natural, geological, and policy factors, such as
380 topography, climate, and regional planning) (Du et al. 2014; Verburg et al. 2003). This made the
381 scaling relationships more complex and increased unpredictability. At the city proper level, linear
382 relationships (i.e., power law with scaling exponent = 1) existed for most of the LULC types. The
383 results were relatively consistent with those of previous studies, showing that developed lands
384 followed power law scaling relationships with increasing extent in highly urbanized areas (Ma et
385 al. 2019). In urban centers, human production and living activities severely disturbed the land
386 spatial composition. Previous studies have proved that high population densities mainly influenced
387 the spatial patterns of urban developed lands within the city proper (Ma et al. 2016a; Ma et al.
388 2019). As the spatial extent increased to broader metropolitan and urban hierarchical levels,
389 population densities became smaller and more stable and showed exponential decay functions at
390 the urban agglomeration level (Fig. A2). Thus, human disturbances gradually weakened outside
391 the suburbs (Lan et al. 2021; Liu et al. 2018). This led to the change rates of developed land and
392 cropland slow down and the corresponding regression curves became exponential rules at the
393 broader scales (Fig. 3; Table 3).

394 Different urban agglomerations also presented various scaling relationships for LULC, among
395 which the relations in BTH were the most predictable and those in YRD were the weakest. This
396 was related to the development modes of urbanization in the agglomerations. BTH was a simple
397 combined urban agglomeration, with Beijing as the core metropolitan area and Tianjin as the
398 sub-core city. PRD contains core cities concentrated in the central region while YRD was a more
399 complex urban agglomeration with several dispersed core cities (Fig. 2) (Fang and Yu 2017).
400 Especially for YRD, the spatial patterns of different LULC presented complex and unpredictable
401 scaling relations. The main reasons were as follows: (1) the city proper of Shanghai can only
402 expand westward due to topography conditional constraints (Fig. 1) and (2) YRD is a spatial
403 polycentric mega-city region (Chen et al. 2019), which led to the fragmentation of forest,
404 croplands, and other landscape patches (Lu et al. 2018). This kind of polycentric pattern was
405 induced by a series of regional integration development measures. For example, the General
406 Office of the State Council, PRC, proposed the “Regional Plan for Yangtze River Delta Region” in
407 2010 (www.gov.cn/) and implemented equal infrastructure construction, industrial planning, and
408 public service popularization in each core city, which aggravated polycentric development.

409 *How did the scaling relations of multiple ESs change and response to LULC across different*
410 *urban hierarchical levels and agglomerations?*

411 Most ES values increased with the increasing spatial extent, except for water yield, nitrogen
412 retention, and soil retention in three urban agglomerations. The average values of carbon storage,
413 PM_{2.5} removal, habitat quality, and recreational opportunity were lower in urban centers and

414 became higher in broader metropolitan and urban agglomeration regions, due to the increase of
415 ecological lands, such as forests, wetlands, and grasslands (Fig. 2). Previous studies have shown
416 that ecological land per unit can provide more similar regulation and cultural services, compared
417 to artificial land (Baumeister et al. 2020; Luo et al. 2018). In contrast, the values of water yield
418 were highest in urban centers. This was mainly because the reference evapotranspiration in urban
419 areas was less than in suburban vegetated areas (Benra et al. 2021; Yang et al. 2019). It is worth
420 noting that soil retention changed irregularly with increase in spatial extent. The phenomenon was
421 largely determined by comprehensive factors, such as topography (Sun et al. 2014), rainfall
422 intensity (Rodríguez-Caballero et al. 2013), and vegetation distribution (Korkanç and Dorum
423 2019).

424 For different types of ESs, provisioning services were most predictable. The food production
425 and water yield showed predictable scaling relations at all hierarchical levels and urban
426 agglomerations. For example, food production revealed power law and exponential growth
427 relationships at all three urban agglomerations. Landscape structure (e.g., composition and
428 configuration) was often considered linked to ecological processes and ESs (Botequilha Leitao
429 and Ahern 2002). To be specific, the scaling relations of food production in this study might be
430 associated with the predictable distribution of croplands. Both food production and croplands
431 increased rapidly on smaller scales (power law functions) and became stable at larger urban
432 agglomeration scales (exponential functions). In addition, some types of regulating ESs only
433 exhibited predictable functions at certain urban levels, such as for carbon storage and PM_{2.5}
434 removal within YRD (Fig. 7). When the spatial scale extended to the broader urban hierarchical
435 levels, the scaling relationships of these two indicators transformed into complex staircase-like
436 patterns. A previous study has shown that the scalograms of urban impervious surfaces exhibited
437 scale breaks (change points) that corresponded roughly to the urban administrative levels (Wu and
438 Li 2006). Similar to urban impervious surfaces, the breakpoint changes also existed in croplands,
439 forests, and developed lands (Fig. 3f), which might cause the staircase-like patterns of carbon
440 storage and PM_{2.5} removal within YRD.

441 Different urban hierarchical levels presented two types of scaling functions for ESs, among
442 which the power laws predominantly expressed ES patterns at smaller levels while exponential
443 relationships were more suitable for larger levels. Although the power law function was found to
444 be ubiquitous when representing the scaling relationship between ecological attributes and
445 measurable scales (Fisher et al. 2008; Newman 2005; Spence 2004), it cannot fully express the
446 scaling results of LULC and ESs for various types and urban hierarchical levels. Zhao and Liu
447 (2014) found that a similar power law relationship existed in the critical scale resolution of the
448 carbon cycle with spatial extent. They proposed that future studies should further investigate the
449 compatibility of power law with the scaling relations of ecological indicators. This study fills the
450 research gap in the scaling of ESs and demonstrates that power law scaling can be fitted well only
451 at certain ranges or urban hierarchical levels. At the city proper level, landscape structures were

452 often shaped by anthropogenic factors that were related to developed lands, such as demographic,
453 economic, and traffic (Li et al. 2013; Ma et al. 2019; Xie et al. 2017). Consequently, the ESs
454 revealed similar predictable power law functions that were consistent with developed lands. As the
455 spatial extent increased to broader metropolitan or urban agglomeration regions, cropland, forest,
456 and other ecological lands were incorporated into dominant LULC types (Fig. 3). Natural factors,
457 such as topography and climate conditions, which are macroscopic and stable, became the main
458 driving forces in shaping landscape patterns and processes (Peng et al. 2017; Smith et al. 2019).
459 Thus, the ES patterns became stable and exhibited significant exponential relationships on a
460 broader scale.

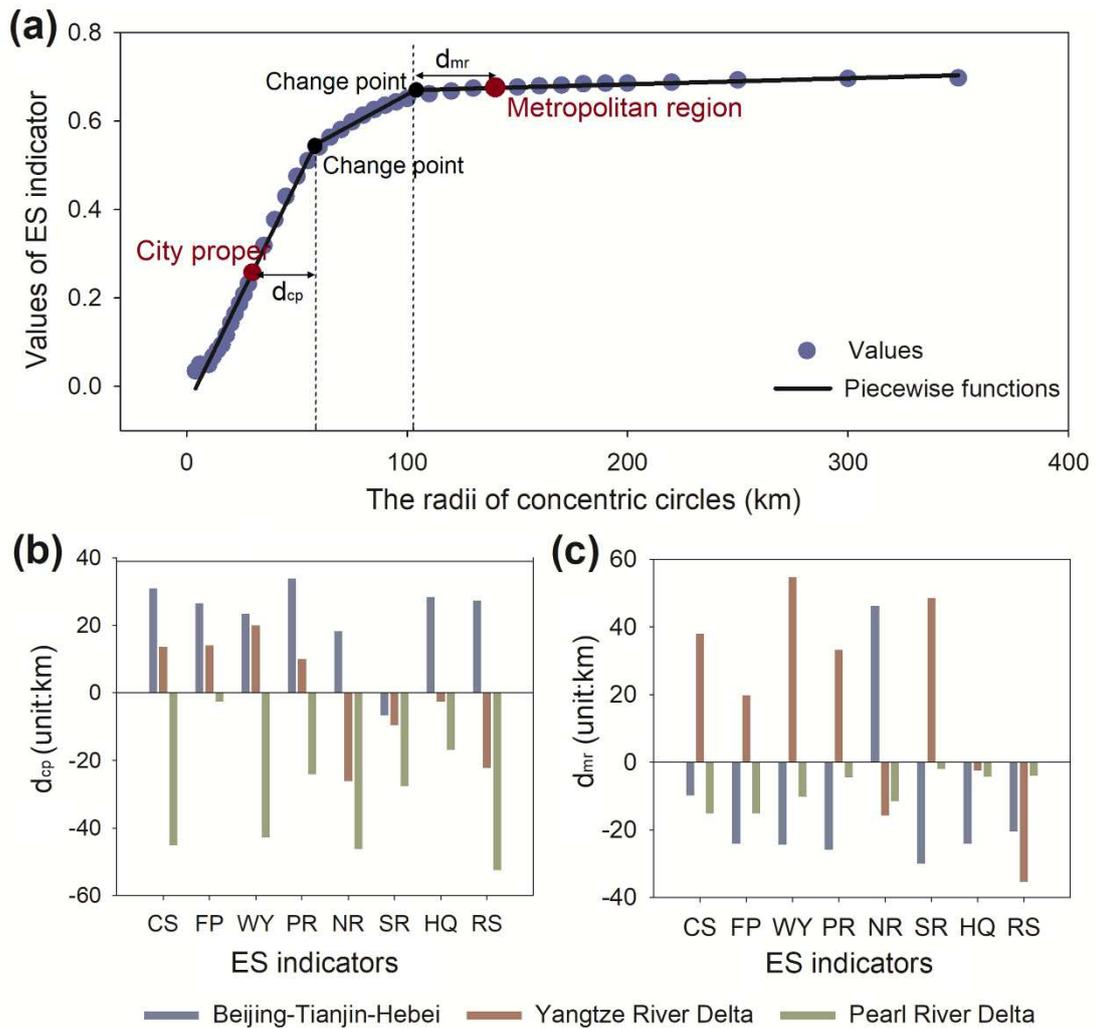
461 Different urban agglomerations also presented various scaling relationships for ESs, among
462 which the BTH was the most quantitatively predictable while the YRD was unpredictable. The
463 study showed that the more complicated the distribution of LULC, the greater the variability of ES
464 patterns (Yee et al. 2021) and the poorer the predictability of scaling relations. Therefore, the
465 predictability of spatial scaling of ESs may be closely related to the complexity of LULC patterns.
466 Compared with YRD and PRD, the distribution shape of BTH was more regular and the LULC
467 area was spread evenly in all directions of concentric circles (Fig. 1). The urbanization of BTH
468 was dominated by the “urban edge-expansion mode (the newly developed patches expanded from
469 the fringes of existing urban centers)” (Jia et al. 2020), which made ESs more scale-predictable.
470 However, the scaling relations of ESs in YRD were difficult to predict for the following two key
471 reasons: (1) the topographical constraints and polycentric urban development had complicated the
472 urban shape and increased landscape fragmentation (Li and Zhou 2019; Tao et al. 2020), and thus
473 intensified the spatial heterogeneity of ESs. (2) Large proportions of ESs were provided by forest
474 and water bodies in YRD; however, these two LULC types were unpredictable (Table 3), which
475 also affected the predictability of spatial scaling of ESs.

476 *How to integrate the scaling relations of ESs into land use planning and policy practice?*

477 The results of this study confirmed that the scaling of ESs were closely related to the land use
478 patterns in urban agglomerations. Thus, the supply of multiple ESs can be effectively improved by
479 managing the spatial patterns of LULC (Mitchell et al. 2015). The most vulnerable and the worst
480 performing area for multiple ESs was the city proper, the key human dominated region (Robinson
481 et al. 2009). It thus required land use optimization. At the city proper level, LULC patterns were
482 more predictable, and thus, the planning can promote ESs more predictably. However,
483 urbanization expansion and LULC spatial patterns emerged from a myriad of policy decisions and
484 planning processes (Batty 2008). To conserve ESs, diversified land management measures should
485 be implemented to avoid disorderly or leapfrog urban expansion. First, more compact urban
486 expansion forms should be encouraged, such as improving land use efficiency to maintain rational
487 densification (Shoemaker et al. 2019; Wang et al. 2019a). Second, conserving the natural land
488 cover, such as by setting the “ecological red line” to protect green and blue spaces in urban areas

489 (Lin and Li 2019), can effectively improve regulating and cultural services. Specifically, the
490 priorities are to reconcile the conflict between developed land and cropland in the city proper of
491 Beijing while protecting those ecological lands that are affected by urbanization most seriously,
492 such as water bodies in the city proper of Shanghai and forests in the city proper of Guangzhou
493 (Fig. 3) (Wang et al. 2020; Zhang et al. 2020a). Third, decision-makers can adjust landscape
494 structures by optimizing landscape configuration (Rieb and Bennett 2020), promoting landscape
495 connectivity, and using other measures to support the landscape multifunctionality (Bolliger et al.
496 2011) and increase the supply of ESs per unit area.

497 The scaling relations of ESs could reflect the actual spatial characteristics of the ecological
498 environment at each urban hierarchical level of the urban agglomerations. We used piecewise
499 function to simulate the scaling curve of ESs, which can reflect the critical thresholds (change
500 points) of various ES indicators (Momb Blanch et al. 2016) that were the turning points of the
501 change rates of scaling relations for ESs (Fig. 10a). The range of change point represents the
502 actual geographical boundary of the ES scaling changes, which was not consistent with the urban
503 administrative boundary. Fig. 10b and Fig. 10c illustrate the distance (variance) between the
504 change point of ES scaling relations and the administrative boundary for multiple ESs in the three
505 urban agglomerations. In the city proper, the change point values for most of the ESs were larger
506 than the boundary value of the city proper of Beijing, while the change point values for all ESs
507 were less than the boundary value of the city proper of Guangzhou. This reflected that the
508 eco-environmental effects of urbanization of Beijing city proper had gone beyond the scale of
509 administrative planning, and hence, the actual management extent of ESs should be broader than
510 the administrative boundary. In contrast, the eco-environmental effects of urbanization for
511 Guangzhou city proper had not extended to the administrative boundary, and the actual impact
512 extent of ESs were smaller than the administrative boundary. At the metropolitan level, the actual
513 impact extent of most of the ESs was broader than the administrative boundary in the Shanghai
514 metropolitan region, while the actual ES boundaries were smaller in the Beijing and Guangzhou
515 metropolitan regions. The urban sprawl and polycentric development in the Shanghai metropolitan
516 region (Tao et al. 2020) had exceeded the administrative boundary. Therefore, the Shanghai
517 metropolitan region should pay more attention to the impacts of urbanization on ESs. To
518 effectively protect the ecological environment and natural resources, policy makers should
519 incorporate the actual extent of ES management into sustainable land use planning, rather than
520 merely relying on administrative boundaries (Brunet et al. 2018; Hein et al. 2006). This means that
521 the proposed ES boundary changes can provide a theoretical basis for the boundary determination
522 of the actual natural resource management and landscape planning in urban agglomerations.



523

524 **Fig. 10 (a)** Example of a scalogram for ecosystem service (ES) change points versus the
 525 administrative boundary points. The solid black dots represent the turning points for the ES
 526 change rates; the solid red dots represent the administrative boundary points. The “ d_{cp} ” represents
 527 the distance between the first change point of ES scaling relations and the administrative boundary
 528 of the city proper; the “ d_{mr} ” represents the distance between the second change point of ES scaling
 529 relations and the administrative boundary of the metropolitan region. Distance was calculated by
 530 the change point value minus the administrative boundary value. **(b)** The “ d_{cp} ” for multiple ESs in
 531 Beijing-Tianjin-Hebei (BTH), Yangtze River Delta (YRD), and Pearl River Delta (PRD) urban
 532 agglomerations. CS = carbon storage; FP = food production; WY = water yield; PR = PM_{2.5}
 533 removal; NR = nitrogen retention; SR = soil retention; HQ = habitat quality; RS = recreational
 534 opportunity. **(c)** The “ d_{mr} ” for multiple ESs in BTH, YRD, and PRD urban agglomerations.

535 *Limitations and future perspectives*

536 Although the InVEST models contain uncertainties and do not fully reflect the actual indicator
 537 values (Daneshi et al. 2021; Pickard et al. 2017), previous studies have confirmed that they
 538 performed well in certain regions, which included urbanization areas similar to the ones in the

539 current study (Redhead et al. 2016; Sun et al. 2018). Thus, the results of this study in the three
540 urban agglomerations are relatively reliable. The models have been applied with multiple scales
541 and have proven effective (Bagstad et al. 2018; Grafius et al. 2016; Nelson et al. 2009). Therefore,
542 they could be used in the current study to explore the scaling relations and overall trends and
543 changes of ESs. However, uncertainties still existed in the identification of scaling relationships
544 for LULC and ESs. For example, the predictable scaling relationships for some ES indicators can
545 be characterized by either the power law or the exponential function. The fitting functions were
546 determined solely based on significance (P values) and determination coefficient (R^2) values.
547 Future research should increase the sampled database of urban agglomerations and validate the
548 scaling relations we proposed in this study. This will help in choosing more universal and
549 reasonable regression models and in analyzing the underlying biophysical mechanisms behind
550 them. Besides, previous research mainly focused on exploring how the natural and
551 socio-economic attributes changed with different city scale (where population was usually used as
552 the proxy indicator) (Bettencourt and Lobo 2016; Zhao et al. 2018a). These studies often adopted
553 power law functions to express the relationships between city attributes and scales (Bettencourt
554 2013; Bettencourt et al. 2010; Wu 2004). In the future, we can compare the fitting curves derived
555 from the scaling relations of the city scale indicator (e.g., population) or the spatial extent, to
556 further understand the scale dependence of urban landscape patterns and ecological processes. In
557 addition, we selected three highly developed national-level urban agglomerations in China to
558 reflect the scaling relations of multiple LULC types and ESs. Previous study proved that the
559 scaling relations of developed land in other six world metropolitan regions were closely resembled
560 those of Beijing, Shanghai, and Guangzhou (Ma et al. 2019). In the future study, it still needs to be
561 demonstrated whether the scaling relations of different LULC types and ESs would hold in urban
562 agglomerations with other development levels in China or other countries.

563 **Conclusions**

564 Exploring how LULC and ES patterns change spatially across multiple scales is helpful for
565 promoting urban landscape sustainability. Based on LULC maps and ES assessment results, this
566 study used the scalogram forms to explore how different LULC types and ES indicators respond
567 to changing spatial extents in the three largest urban agglomerations of China. The results revealed
568 that different LULC mainly exhibited three types of scaling relations, linear, exponential, and
569 sigmoidal relationships, among which developed land and cropland were the most predictable.
570 LULC types at the city proper level were more predictable than the metropolitan and urban
571 agglomeration levels. Most of the ES indicators increased when the spatial extent increased, but
572 their scaling relations varied; provisioning services were the most predictable, while soil retention
573 was unpredictable. The ESs predominantly presented two types of scaling functions, among which
574 power law expressed ES patterns mainly at the lower levels while the exponential relationships
575 were applicable to the higher levels. Among the three urban agglomerations, BTH performed the

576 best, while YRD performed the worst in the predictability of the ES spatial scaling. Since the
577 scaling relations of ESs were deeply affected by the LULC patterns, implementing land use
578 composition and configuring optimization strategies are conducive to ecological conservation
579 planning, especially in the city proper. The scaling relations of ESs can provide a scientific basis
580 for the boundary determination of urban land use management.

581

582 **Acknowledgements**

583 This work was supported by the National Natural Science Foundation of China [Grant No.
584 41901227 and U1901601], the Open Foundation of the State Key Laboratory of Urban and
585 Regional Ecology of China [Grant No. SKLURE2020-2-1], and the Fundamental Research Funds
586 for Central Non-profit Scientific Institution [Grant No. 1610132021001].

587

588 **Declarations**

589 **Conflict of interest** The authors declare that they have no conflict of interest.

590

591 **References**

- 592 Arcaute E, Hatna E, Ferguson P, Youn H, Johansson A, Batty M (2015) Constructing cities,
593 deconstructing scaling laws. *J R Soc Interface* 12(102): 20140745
- 594 Bagstad KJ, Cohen E, Ancona ZH, McNulty SG, Sun G (2018) The sensitivity of ecosystem
595 service models to choices of input data and spatial resolution. *Appl Geogr* 93: 25–36
- 596 Batty M (2008) The size, scale, and shape of cities. *Science* 319(5864): 769–771
- 597 Baumeister CF, Baumeister T, Plieninger T, Schraml U (2020) Exploring cultural ecosystem
598 service hotspots: Linking multiple urban forest features with public participation mapping
599 data. *Urban For Urban Gree* 48: 126561
- 600 Benra F, Frutos AD, Gaglio M, Álvarez-Garretón C, Felipe-Lucia M, Bonn A (2021) Mapping
601 water ecosystem services: Evaluating InVEST model predictions in data scarce regions.
602 *Environ Modell Softw* 3: 104982
- 603 Bettencourt LM (2013) The origins of scaling in cities. *Science* 340(6139): 1438–1441
- 604 Bettencourt LM, Lobo J (2016) Urban scaling in Europe. *J R Soc Interface* 13(116): 20160005
- 605 Bettencourt LM, Lobo J, Helbing D, Kühnert C, West GB (2007) Growth, innovation, scaling, and
606 the pace of life in cities. *Proc Natl Acad Sci USA* 104(17): 7301–7306
- 607 Bettencourt LM, Lobo J, Strumsky D, West GB (2010) Urban scaling and its deviations: revealing
608 the structure of wealth, innovation and crime across cities. *PLoS One* 5: e13541
- 609 Bian H, Gao J, Wu J, Sun X, Du Y (2020) Hierarchical analysis of landscape urbanization and its
610 impacts on regional sustainability: A case study of the Yangtze River Economic Belt of China.

611 J Clean Prod 279: 123267

612 Bolliger J, Bättig MB, Gallati J, Kläy A, Stauffacher M, Kienast F (2011) Landscape
613 multifunctionality: a powerful concept to identify effects of environmental change. *Reg*
614 *Environ Change* 11: 203–206

615 Leitão AB, Ahern J (2002) Applying landscape ecological concepts and metrics in sustainable
616 landscape planning. *Landsc Urban Plan* 59:65–93

617 Brock WA (1999) Scaling in economics: a reader's guide. *Ind Corp Change* 8(3): 409–446

618 Brunet L, Tuomisaari J, Lavorel S, Crouzat E, Bierry A, Peltola T, Arpin I (2018) Actionable
619 knowledge for land use planning: Making ecosystem services operational. *Land Use Policy*
620 72: 27–34

621 Chen G, Li X, Liu X, Chen Y, Liang X, Leng J, Xu X, Liao W, Qiu Y, Wu Q, Huang K (2020)
622 Global projections of future urban land expansion under shared socioeconomic pathways. *Nat*
623 *Commun* 11: 1–12

624 Chen W, Yenneti K, We YD, Yuan F, Wu JW, Gao JL (2019) Polycentricity in the Yangtze River
625 Delta Urban Agglomeration (YRDUA): More Cohesion or More Disparities? *Sustainability*
626 11: 3106

627 Daneshi A, Brouwer R, Najafinejad A, Panahi M, Zarandian A, Maghsood FF (2021) Modelling
628 the impacts of climate and land use change on water security in a semi-arid forested
629 watershed using InVEST. *J Hydrol* 593: 125621

630 D'Amour C, Reitsma F, Baiocchi G, Barthel S, Güneralp B, Erb K, Haberl H, Creutzig F, Seto KC
631 (2016) Future urban land expansion and implications for global croplands. *Proc Natl Acad*
632 *Sci USA* 114 (34): 8939–8944

633 Department of Urban Surveys of National Bureau of Statistics of China (DUSNB) (2018) China
634 city statistics yearbook 2019. China Statistics Press, Beijing (in Chinese)

635 Du S, Wang Q, Guo L (2014) Spatially varying relationships between land-cover change and
636 driving factors at multiple sampling scales. *J Environ Manage* 137: 101–110

637 Du Y, Sun T, Peng J, Fang K, Liu Y, Yang Y, Wang Y (2018) Direct and spillover effects of
638 urbanization on PM_{2.5} concentrations in China's top three urban agglomerations. *J Clean Prod*
639 190: 72–83

640 Fang C, Yu D (2017) Urban agglomeration: An evolving concept of an emerging phenomenon.
641 *Landscape Urban Plan* 162: 126–136

642 Fisher JL, Hurtt GC, Thomas RQ, Chambers JQ (2008) Clustered disturbances lead to bias in
643 large-scale estimates based on forest sample plots. *Ecol Lett* 11: 554–563

644 Forgione HM, Pregitzer CC, Charlop-Powers S, Gunther B (2016) Advancing urban ecosystem
645 governance in New York City: Shifting towards a unified perspective for conservation
646 management. *Environ Sci Pol* 62: 127–132

647 Frazier AE (2016) Surface metrics: scaling relationships and downscaling behavior. *Landscape*
648 *Ecol* 31: 351–363.

649 Fuller RA, Gaston KJ (2009) The scaling of green space coverage in European cities. *Biol Lett*
650 5(3): 352–355

651 Gao J, Yu ZW, Wang LC, Vejre H (2019) Suitability of regional development based on ecosystem
652 service benefits and losses: A case study of the Yangtze River Delta urban agglomeration,
653 China. *Ecol Indic* 107: 105579

654 Goldstein JH, Caldarone G, Duarte TK, Ennaanay D, Hannahs N, Mendoza G, Polasky S, Wolny
655 S, Daily GC (2012) Integrating ecosystem-service tradeoffs into land-use decisions. *P Natl*
656 *Acad Sci USA* 109(19): 7565–7570

657 Grafius DR, Corstanje R, Warren PH, Evans KL, Hancock S, Harris JA (2016) The impact of land
658 use/land cover scale on modelling urban ecosystem services. *Landscape Ecol* 31: 1509–1522

659 Gu C (2019) Urbanization: Processes and driving forces. *Sci China Earth Sci* 62(9): 1351–1360

660 Hamel P, Chaplin-Kramer R, Sim S, Mueller C (2015) A new approach to modeling the sediment
661 retention service (InVEST 3.0): Case study of the Cape Fear catchment, North Carolina, USA.
662 *Sci Total Environ* 524–525: 166–177

663 Hasan SS, Zhen L, Miah MG, Ahamed T, Samie A (2020) Impact of land use change on ecosystem
664 services: A review. *Environ Dev* 34: 100527

665 Hein L, van Koppen K, de Groot RS, van Ierland EC (2006) Spatial scales, stakeholders and the
666 valuation of ecosystem services. *Ecol Econ* 57: 209–228

667 Hu M, Li Z, Wang Y, Jiao M, Li M, Xia B (2019) Spatio-temporal changes in ecosystem service
668 value in response to land-use/cover changes in the Pearl River Delta. *Resour Conserv Recy*
669 149: 106–114

670 Jomnonkwao S, Uttra S, Ratanavaraha V (2020) Forecasting road traffic deaths in Thailand:
671 applications of Time-Series, curve estimation, multiple linear regression, and path analysis
672 models. *Sustainability* 12(1): 395

673 Kedron PJ, Frazier AE, Ovando-Montejo GA, Wang J (2018) Surface metrics for landscape
674 ecology: a comparison of landscape models across ecoregions and scales. *Landscape Ecol* 33:
675 1489–1504.

676 Korkanç SY, Dorum G (2019) The nutrient and carbon losses of soils from different land cover
677 systems under simulated rainfall conditions. *CATENA* 172: 203–211

678 Kuang B, Lu X, Han J, Fan X, Zuo J (2020) How urbanization influence urban land consumption
679 intensity: Evidence from China. *Habitat Int* 100: 102103

680 Kuang W (2020a) 70 years of urban expansion across China: trajectory, pattern, and national
681 policies. *Sci Bull* 65: 1970–1974

682 Kuang W (2020b) National urban land-use/cover change since the beginning of the 21st century
683 and its policy implications in China. *Land Use Policy* 97: 104747

684 Lan T, Shao G, Xu Z, Tang L, Sun L (2021) Measuring urban compactness based on functional
685 characterization and human activity intensity by integrating multiple geospatial data sources.
686 *Ecol Indic* 121: 107177

- 687 Lan X, Tang H, Liang H (2017) A theoretical framework for researching cultural ecosystem
688 service flows in urban agglomerations. *Ecosyst Serv* 28: 95–104
- 689 Lavorel S, Rey P, Grigulis K, Zawada M, Byczek C (2020) Interactions between outdoor
690 recreation and iconic terrestrial vertebrates in two French alpine national parks. *Ecosyst Serv*
691 45: 101155
- 692 Lawler J, Lewis D, Nelson E, Plantinga A, Polasky S, Withey J, Helmers D, Martinuzzi S,
693 Pennington D, Radeloff V (2014) Projected land-use change impacts on ecosystem services
694 in the United States. *Proc Natl Acad Sci USA* 111(20): 7492–7497
- 695 Lei J, Wang S, Wu J, Wang J, Xiong X (2021) Land-use configuration has significant impacts on
696 water-related ecosystem services. *Ecol Eng* 160: 106133
- 697 Li C, Li J, Wu J (2013) Quantifying the speed, growth modes, and landscape pattern changes of
698 urbanization: a hierarchical patch dynamics approach. *Landscape Ecol* 28(10): 1875–1888
- 699 Li H, Peng J, Liu Y, Hu Y (2017) Urbanization impact on landscape patterns in Beijing City, China:
700 A spatial heterogeneity perspective. *Ecol Indic* 82: 50–60
- 701 Li W, Hai X, Han L, Mao J, Tian M (2020) Does urbanization intensify regional water scarcity?
702 Evidence and implications from a megaregion of China. *J Clean Prod* 244: 118592
- 703 Li F, Zhou T (2019) Effects of urban form on air quality in China: An analysis based on the spatial
704 autoregressive model. *Cities* 89: 130–140
- 705 Lin JY, Li X (2019) Large-scale ecological red line planning in urban agglomerations using a
706 semi-automatic intelligent zoning method. *Sustain Cities Soc* 46: 101410
- 707 Liu L, Xu X, Chen X (2015) Assessing the impact of urban expansion on potential crop yield in
708 China during 1990–2010. *Food Sec* 7: 33–43
- 709 Liu W, Zhan J, Zhao F, Yan H, Zhang F, Wei X (2019) Impacts of urbanization-induced land-use
710 changes on ecosystem services: A case study of the Pearl River Delta Metropolitan Region,
711 China. *Ecol Indic* 98: 228–238
- 712 Liu Y, Zhang X, Kong X, Wang R, Chen L (2018) Identifying the relationship between urban land
713 expansion and human activities in the Yangtze River Economic Belt, China. *Appl Geogr* 94:
714 163–177
- 715 Liu Y, Zhang X, Pan X, Ma X, Tang M (2020) The spatial integration and coordinated industrial
716 development of urban agglomerations in the Yangtze River Economic Belt, China. *Cities* 104:
717 102801
- 718 Lobo J, Bettencourt LM, Smith ME, Ortman S (2020) Settlement scaling theory: Bridging the
719 study of ancient and contemporary urban systems. *Urban Stud* 57(4): 731–747
- 720 Lu D, Mao W, Yang D, Zhao J, Xu J (2018) Effects of land use and landscape pattern on PM_{2.5} in
721 Yangtze River Delta, China. *Atmos Pollut Res* 9(4): 705–713
- 722 Luo Q, Zhang X, Li Z, Yang M, Lin Y (2018) The effects of China's Ecological Control Line
723 policy on ecosystem services: The case of Wuhan City. *Ecol Indic* 93: 292–301
- 724 Luo Q, Zhou J, Li Z, Yu B (2020) Spatial differences of ecosystem services and their driving

725 factors: A comparison analysis among three urban agglomerations in China's Yangtze River
726 Economic Belt. *Sci Total Environ* 725: 138452

727 Ma Q, He C, Wu J (2016) Behind the rapid expansion of urban impervious surfaces in China:
728 Major influencing factors revealed by a hierarchical multiscale analysis. *Land Use Policy*, 59,
729 434–445

730 Ma Q, Wu J, He C (2016b). A hierarchical analysis of the relationship between urban impervious
731 surfaces and land surface temperatures: spatial scale dependence, temporal variations, and
732 bioclimatic modulation. *Landscape Ecol* 31(5): 1139–1153

733 Ma Q, Wu J, He C, Hu G (2019) Reprint of “Spatial scaling of urban impervious surfaces across
734 evolving landscapes: From cities to urban regions”. *Landscape Urban Plan* 187: 132–144

735 Massada AB, Radeloff VC (2010) Two multi-scale contextual approaches for mapping spatial
736 pattern. *Landscape Ecol* 25: 711–725

737 Millennium Ecosystem Assessment (2005) *Ecosystems and Human Well-Being: Synthesis*. Island
738 Press, Washington, DC

739 Mitchell MGE, Bennett EM, Gonzalez A (2015) Strong and nonlinear effects of fragmentation on
740 ecosystem service provision at multiple scales. *Environ Res Lett* 10: 094014

741 Momblanch A, Connor JD, Crossman ND, Paredes-Arquiola J, Andreu J (2016) Using ecosystem
742 services to represent the environment in hydro-economic models. *J Hydrol* 522: 95–109

743 Nelson E, Mendoza G, Regetz J, Polasky S, Tallis H, Cameron D, Chan KMA, Daily GC,
744 Goldstein J, Kareiva PM, Lonsdorf E, Naidoo R, Ricketts TH, Shaw M (2009) Modeling
745 multiple ecosystem services, biodiversity conservation, commodity production, and tradeoffs
746 at landscape scales. *Front Ecol Environ* 7(1): 4–11

747 Newman MEJ (2005) Power laws, Pareto distributions and Zipf’s law. *Contemp Phys* 46: 323–351

748 Nikodinoska N, Paletto A, Pastorella F, Granvik M, Franzese PP (2018) Assessing, valuing and
749 mapping ecosystem services at city level: The case of Uppsala (Sweden). *Ecol Model* 121:
750 107028

751 Normile D (2016) China rethinks cities. *Science* 352(6288): 916–918

752 O'Neill RV, Deangelis DL, Waide JB, Allen GE (1986) *A hierarchical concept of ecosystems*.
753 Princeton University Press, New Jersey

754 Patra S, Sahoo S, Mishra P, Mahapatra SC (2018) Impacts of urbanization on land use/cover
755 changes and its probable implications on local climate and groundwater level. *J Urban Manag*
756 7(2): 70–84

757 Peng J, Liu Y, Liu Z, Yang Y (2017) Mapping spatial non-stationarity of human-natural factors
758 associated with agricultural landscape multifunctionality in Beijing–Tianjin–Hebei region,
759 China. *Agric Ecosyst Environ* 246: 221–233

760 Pickard BR, Berkel DV, Petrasova A, Meentemeyer RK (2017) Forecasts of urbanization
761 scenarios reveal trade-offs between landscape change and ecosystem services. *Landscape*
762 *Ecol* 32: 617–634

763 Poku-Boansi M (2012) Contextualizing urban growth, urbanisation and travel behaviour in
764 Ghanaian cities. *Cities* 110: 103083

765 Redhead JW, Stratford C, Sharps K, Jones L, Ziv G, Clarke D, Oliver TH, Bullock JM (2016)
766 Empirical validation of the InVEST water yield ecosystem service model at a national scale.
767 *Sci Total Environ* 569–570: 1418–1426

768 Ribeiro FL, Meirelles J, Netto VM, Neto CR, Baronchelli A (2020) On the relation between
769 transversal and longitudinal scaling in cities. *PLoS One* 15(5): 1–20

770 Rieb JT, Bennett EM (2020) Landscape structure as a mediator of ecosystem service interactions.
771 *Landscape Ecol* 35: 2863–2880

772 Robinson DT, Brown DG, Currie WS (2009) Modelling carbon storage in highly fragmented and
773 human-dominated landscapes: Linking land-cover patterns and ecosystem models. *Ecol*
774 *Modell* 220(9): 1325–1338

775 Rodríguez-Caballero E, Cantón Y, Chamizo S, Lázaro R, Escudero A (2013) Soil loss and runoff
776 in semiarid ecosystems: a complex interaction between biological soil crusts,
777 micro-topography, and hydrological drivers. *Ecosystems* 16: 529–546

778 Seto KC, Reenberg A, Boone CG, Fragkias M, Haase D, Langanke T, Marcotullio P, Munroe DK,
779 Olah B, Simon D (2012) Urban land teleconnections and sustainability. *Proc Natl Acad Sci*
780 *USA* 109(20): 7687–7692

781 Smith P, Adams J, Beerling DJ, Beringer T, Calvin KV, Fuss S, Griscom B, Hagemann N,
782 Kammann C, Kraxner F, Minx JC, Popp A, Renforth P, Luis J, Vicente V, Keesstra S (2019)
783 Land-Management options for greenhouse gas removal and their impacts on ecosystem
784 services and the sustainable development goals. *Annu Rev Environ Resour* 44: 255–286

785 Sharp R, Douglass J, Wolny S, Arkema K, Bernhardt J, Bierbower W, Chaumont N, Denu D,
786 Fisher D, Glowinski K, Griffin R, Guannel G, Guerry A, Johnson J, Hamel P, Kennedy C,
787 Kim CK, Lacayo M, Lonsdorf E, Mandle L, Rogers L, Silver J, Toft J, Verutes G, Vogl AL,
788 Wood S, Wyatt K (2020) InVEST 3.8.9.post9+ug.ga009fc0 User’s Guide. The Natural
789 Capital Project, Stanford University, University of Minnesota, The Nature Conservancy, and
790 World Wildlife Fund

791 Shen J, Li S, Liang Z, Liu L, Li D, Wu S (2020) Exploring the heterogeneity and nonlinearity of
792 trade-offs and synergies among ecosystem services bundles in the Beijing-Tianjin-Hebei
793 urban agglomeration. *Ecosyst Serv* 43: 101103

794 Shoemaker DA, BenDor TK, Meentemeyer RK (2019) Anticipating trade-offs between urban
795 patterns and ecosystem service production: Scenario analyses of sprawl alternatives for a
796 rapidly urbanizing region. *Comput Environ Urban* 74: 114–125

797 Spence AJ (2009) Scaling in biology. *Curr Biol* CB 19: R57–R61

798 Stokes EC, Seto KC (2019) Characterizing and measuring urban landscapes for sustainability.
799 *Environ Res Lett* 14: 045002

800 Sun W, Shao Q, Liu J, Zhai J (2014) Assessing the effects of land use and topography on soil

801 erosion on the Loess Plateau in China. *CATENA* 121: 151–163

802 Sun X, Lu ZM, Li F, Crittenden JC (2018) Analyzing spatio-temporal changes and trade-offs to
803 support the supply of multiple ecosystem services in Beijing, China. *Ecol Indic* 94: 117–129

804 Sun X, Tan X, Chen K, Song S, Zhu X, Hou D (2020) Quantifying landscape-metrics impacts on
805 urban green-spaces and water-bodies cooling effect: The study of Nanjing, China. *Urban For*
806 *Urban Gree* 55: 126838

807 Swenson JJ, Franklin J (2000) The effects of future urban development on habitat fragmentation in
808 the Santa Monica Mountains. *Landscape Ecol* 15: 713–730

809 Tao Y, Zhang Z, Ou WX, Guo J, Pueppke SG (2020) How does urban form influence PM_{2.5}
810 concentrations: Insights from 350 different-sized cities in the rapidly urbanizing Yangtze
811 River Delta region of China, 1998–2015. *Cities* 98: 102581

812 Verburg PH, de Groot WT, Veldkamp AJ (2003) Methodology for Multi-Scale Land-Use Change
813 Modelling: Concepts and Challenges. *Global Environmental Change and Land Use*. Springer,
814 Dordrecht. https://doi.org/10.1007/978-94-017-0335-2_2

815 Viglizzo EF, Paruelo JM, Lateralra P, Jobbágy EG (2012) Ecosystem service evaluation to support
816 land-use policy. *Agr Ecosyst Environ* 154: 78–84

817 Wang LZ, Omrani H, Zhao Z, Francomano D, Li K, Pijanowski B (2019) Analysis on urban
818 densification dynamics and future modes in southeastern Wisconsin, USA. *PLoS One* 14(3):
819 0211964

820 Wang Z, Liang L, Sun Z, Wang X (2019b) Spatiotemporal differentiation and the factors
821 influencing urbanization and ecological environment synergistic effects within the
822 Beijing-Tianjin-Hebei urban agglomeration. *J Environ Manage* 243: 227–239

823 Wang Z, Zhang S, Peng Y, Wu C, Lv Y, Xiao K, Zhao J, Qian G (2020) Impact of rapid
824 urbanization on the threshold effect in the relationship between impervious surfaces and
825 water quality in Shanghai, China. *Environ Pollut* 267: 115569

826 Wu J (1999) Hierarchy and scaling: extrapolating information along a scaling ladder. *Can J*
827 *Remote Sens* 25(4): 367–380

828 Wu J (2004) Effects of changing scale on landscape pattern analysis: scaling relations. *Landscape*
829 *Ecol* 19: 125–138

830 Wu J, David JL (2002) A spatially explicit hierarchical approach to modeling complex ecological
831 systems: theory and applications. *Ecol Model* 153(1): 7–26

832 Wu J, Jenerette GD, Buyantuyev A, Redman CL (2011) Quantifying spatiotemporal patterns of
833 urbanization: The case of the two fastest growing metropolitan regions in the United States.
834 *Ecol Complex* 8(1): 1–8

835 Wu J, Li H (2006) Perspectives and methods of scaling. In Wu J, Jones B, Li H, Loucks OL (Eds.).
836 *Scaling and uncertainty analysis in ecology*. Springer, Dordrecht, pp 17–44

837 Wu J, Shen WJ, Sun WZ, Tueller PT (2002) Empirical patterns of the effects of changing scale on
838 landscape metrics. *Landsc Ecol* 17: 761–782

839 Wu R, Li Z, Wang S (2020) The varying driving forces of urban land expansion in China: Insights
840 from a spatial-temporal analysis. *Sci Total Environ* 8: 142591

841 Xie H, He Y, Xie X (2017) Exploring the factors influencing ecological land change for China's
842 Beijing–Tianjin–Hebei Region using big data. *J Clean Prod* 142: 677–687

843 Xing L, Zhu Y, Wang J (2021) Spatial spillover effects of urbanization on ecosystem services
844 value in Chinese cities. *Ecol Indic* 121: 107028

845 Xu C, Zhao S, Liu S (2020) Spatial scaling of multiple landscape features in the conterminous
846 United States. *Landsc Ecol* 35: 223–247

847 Xu Q, Dong Y, Yang R (2018) Influence of land urbanization on carbon sequestration of urban
848 vegetation: A temporal cooperativity analysis in Guangzhou as an example. *Sci Total Environ*
849 635: 26–34

850 Yang D, Liu W, Tang L, Chen L, Li X, Xu X (2019) Estimation of water provision service for
851 monsoon catchments of South China: Applicability of the InVEST model. *Landscape Urban*
852 *Plan* 182: 133–143

853 Yang J, Yang J, Luo X, Huang C (2019) Impacts by expansion of human settlements on nature
854 reserves in China. *J Environ Manag* 248: 109233

855 Yee SH, Paulukonis E, Simmons C, Russell M, Fulford R, Harwell L, Smith LM (2021) Projecting
856 effects of land use change on human well-being through changes in ecosystem services. *Ecol*
857 *Modell* 440: 109358

858 Yu J, Zhou K, Yang S (2019) Land use efficiency and influencing factors of urban agglomerations
859 in China. *Land Use Policy* 88: 104143

860 Zhang D, Huang Q, He C, Wu J (2017) Impacts of urban expansion on ecosystem services in the
861 Beijing–Tianjin–Hebei urban agglomeration, China: A scenario analysis based on the Shared
862 Socioeconomic Pathways. *Resour Conserv Recycl* 125: 115–130

863 Zhang Y, Hu Y, Zhuang D (2020a) A highly integrated, expansible, and comprehensive analytical
864 framework for urban ecological land: A case study in Guangzhou, China. *J Clean Prod* 268:
865 122360

866 Zhang P, Kohli D, Sun Q, Zhang Y, Liu S, Sun D (2020b) Remote sensing modeling of urban
867 density dynamics across 36 major cities in China: Fresh insights from hierarchical urbanized
868 space. *Landscape Urban Plan* 203: 103896

869 Zhang X, Zhong T, Wang K, Cheng Z (2009) Scaling of impervious surface area and vegetation as
870 indicators to urban land surface temperature using satellite data. *Int J Remote Sens* 30(4):
871 841–859

872 Zhao S, Liu S (2014) Scale criticality in estimating ecosystem carbon dynamics. *Glob Change*
873 *Biol* 20: 2240–2251

874 Zhao S, Liu S, Xu C, Yuan W, Sun Y, Yan W, Zhao M, Henebry GM, Fang J (2018a)
875 Contemporary Evolution and Scaling of 32 Major Cities in China. *Ecol Appl* 28(6): 1655–
876 1668

- 877 Zhou D, Tian Y, Jiang G (2018b) Spatio-temporal investigation of the interactive relationship
878 between urbanization and ecosystem services: Case study of the Jingjinji urban
879 agglomeration, China. *Ecol Indic* 95: 152–164
- 880 Zhou Y, Chen M, Tang Z, Mei Z (2021) Urbanization, land use change, and carbon emissions:
881 Quantitative assessments for city-level carbon emissions in Beijing-Tianjin-Hebei region.
882 *Sustain Cities Soc* 66: 102701

Figures

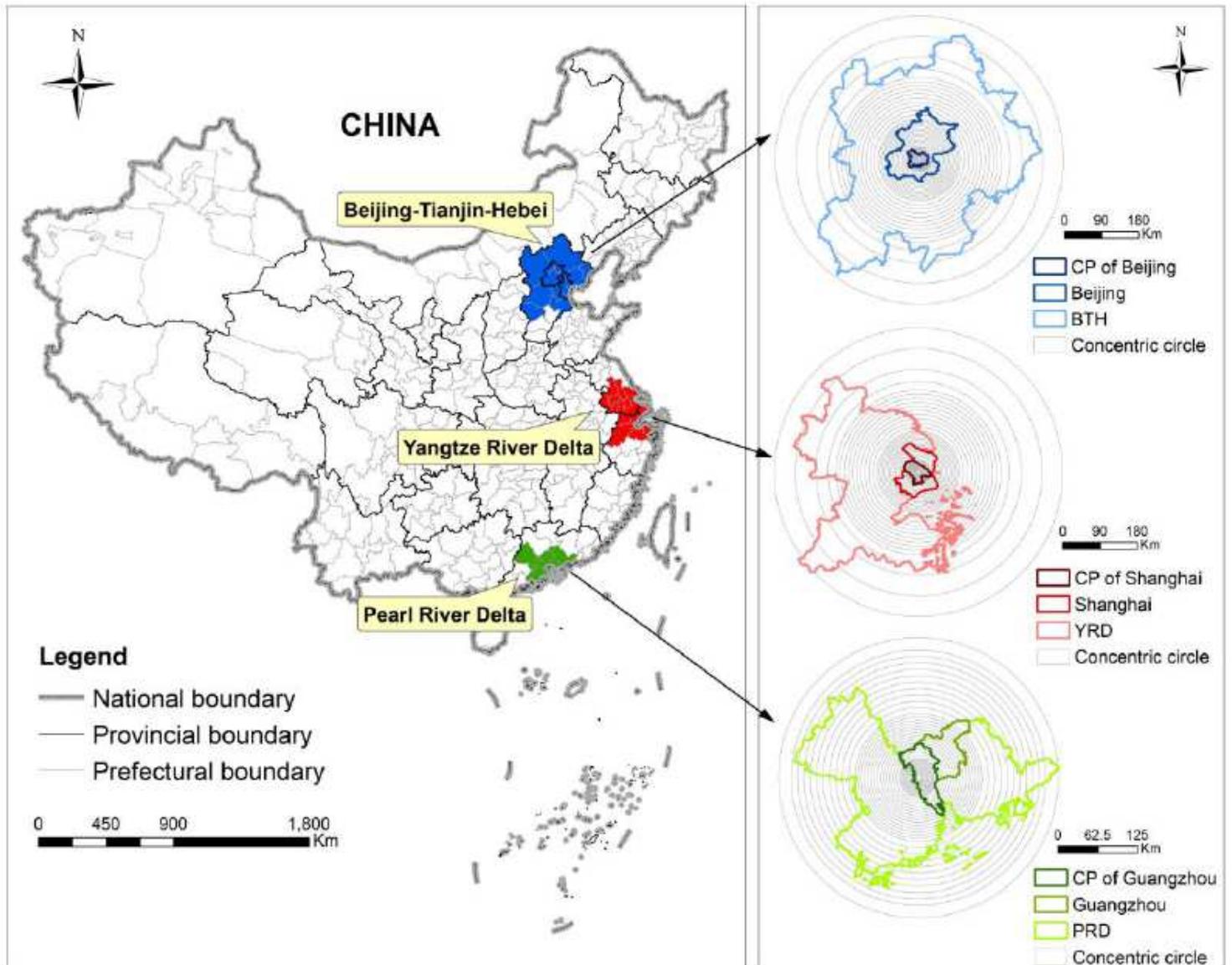


Figure 1

Location and administrative hierarchy of study areas. The concentric circle illustrations on the right represent the spatial extension of three urban agglomerations at the administrative hierarchy levels. The CP of Beijing/Shanghai/Guangzhou refer to the city proper of Beijing/Shanghai/Guangzhou; Beijing/Shanghai/Guangzhou refer to the Beijing/Shanghai/Guangzhou metropolitan region; BTH/YRD/PRD refer to the Beijing Tianjin Hebei/Yangtze River Delta/Pearl River Delta urban agglomeration. Note: The designations employed and the presentation of the material on this map do not imply the expression of any opinion whatsoever on the part of Research Square concerning the legal status of any country, territory, city or area or of its authorities, or concerning the delimitation of its frontiers or boundaries. This map has been provided by the authors.

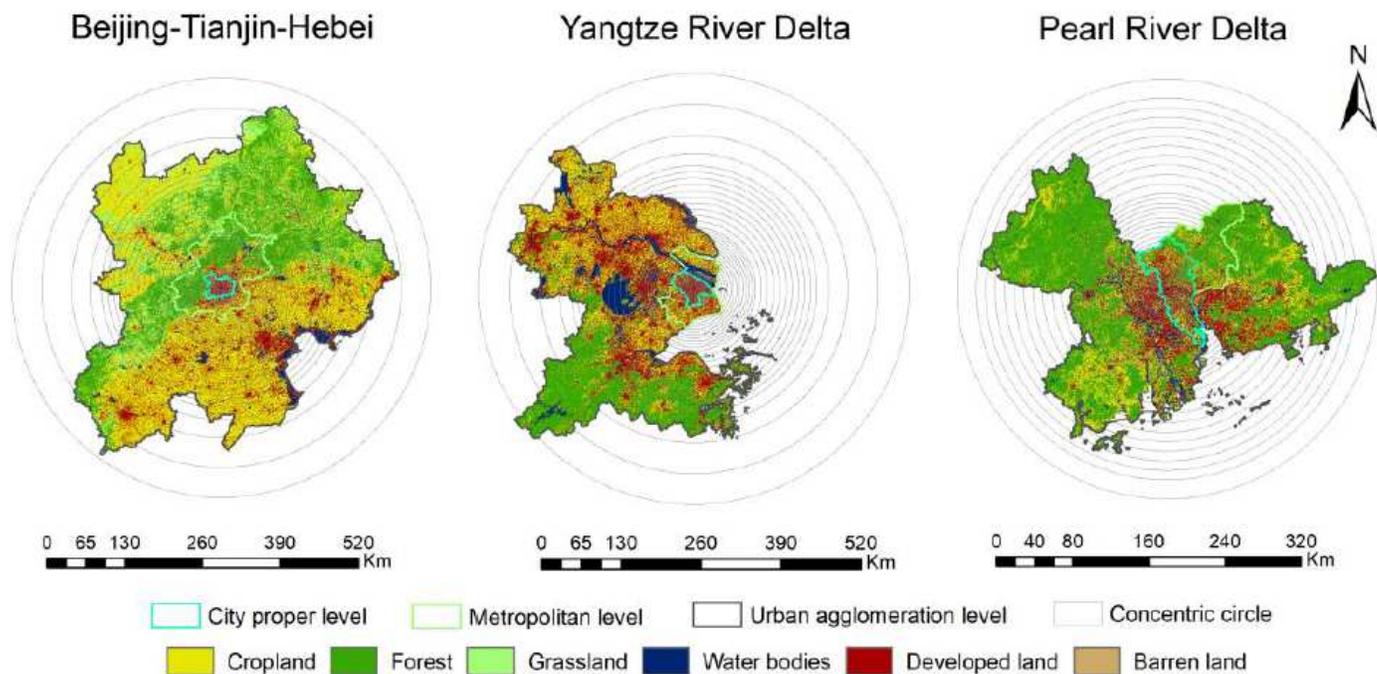


Figure 2

Spatial distributions of land use/land cover in 2018 for three urban agglomerations. The radii list of concentric circles is shown in Table 2. Note: The designations employed and the presentation of the material on this map do not imply the expression of any opinion whatsoever on the part of Research Square concerning the legal status of any country, territory, city or area or of its authorities, or concerning the delimitation of its frontiers or boundaries. This map has been provided by the authors.

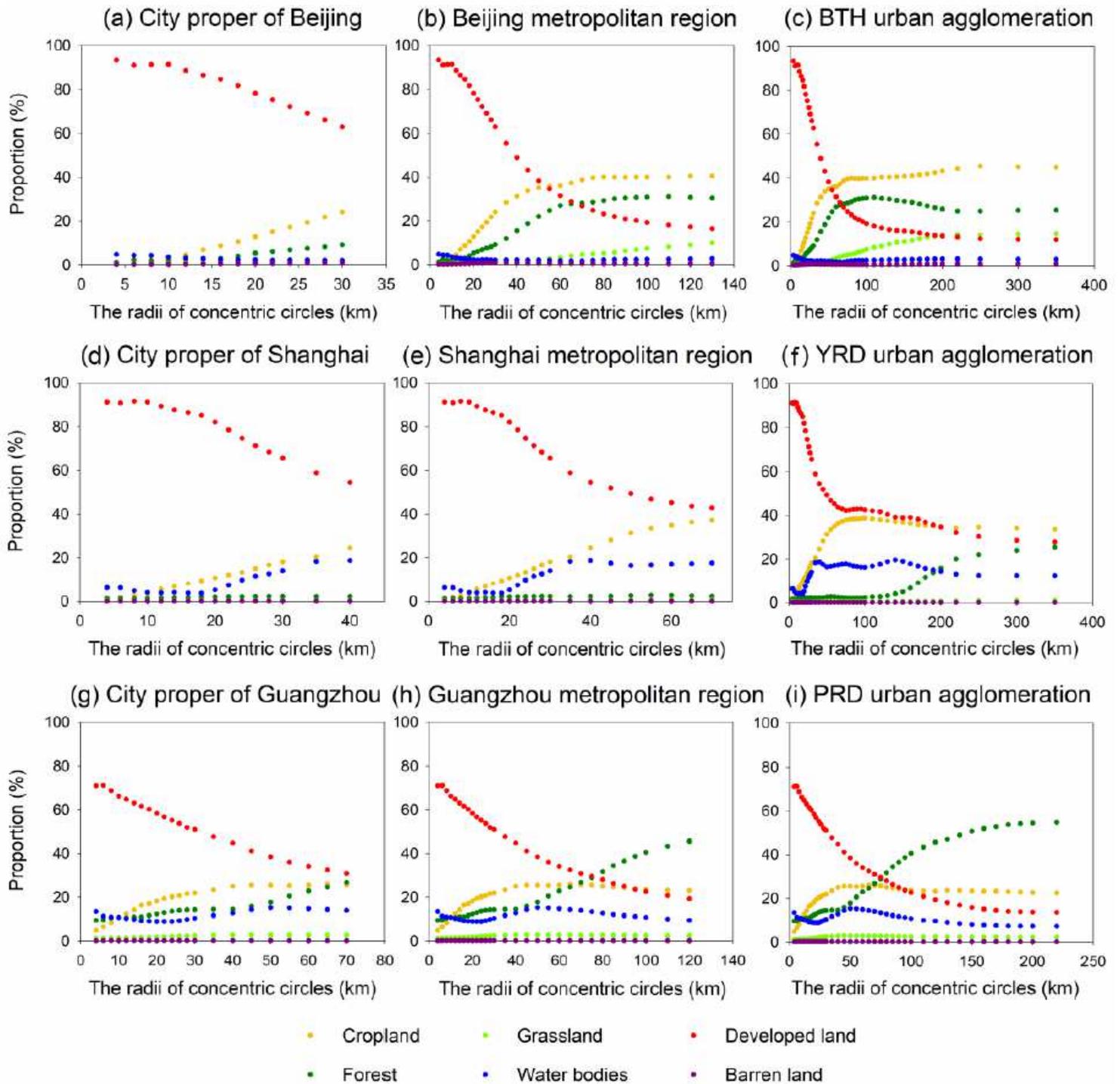


Figure 3

Scalograms of the proportion of different land use /land cover types with respect to increasing concentric circle radii in the three largest urban agglomerations of China (Beijing Tianjin Hebei (BTH), Yangtze River Delta (YRD), and Pearl River Delta (PRD)).

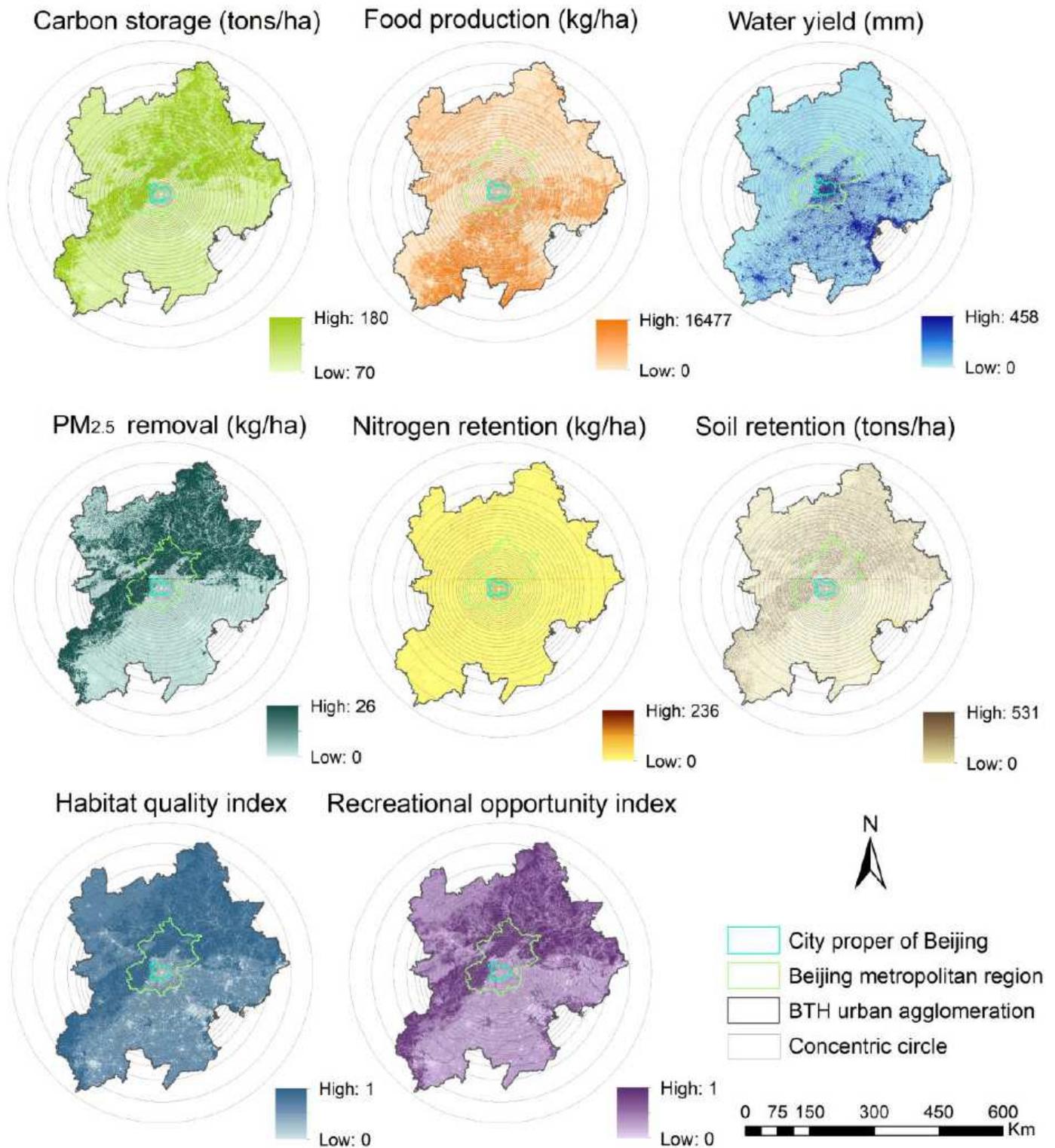


Figure 4

Spatial distributions of multiple ecosystem services in 2018 for Beijing Tianjin-Hebei (BTH) urban agglomeration. The radii list of concentric circles was shown in Table 2. Note: The designations employed and the presentation of the material on this map do not imply the expression of any opinion whatsoever on the part of Research Square concerning the legal status of any country, territory, city or

area or of its authorities, or concerning the delimitation of its frontiers or boundaries. This map has been provided by the authors.

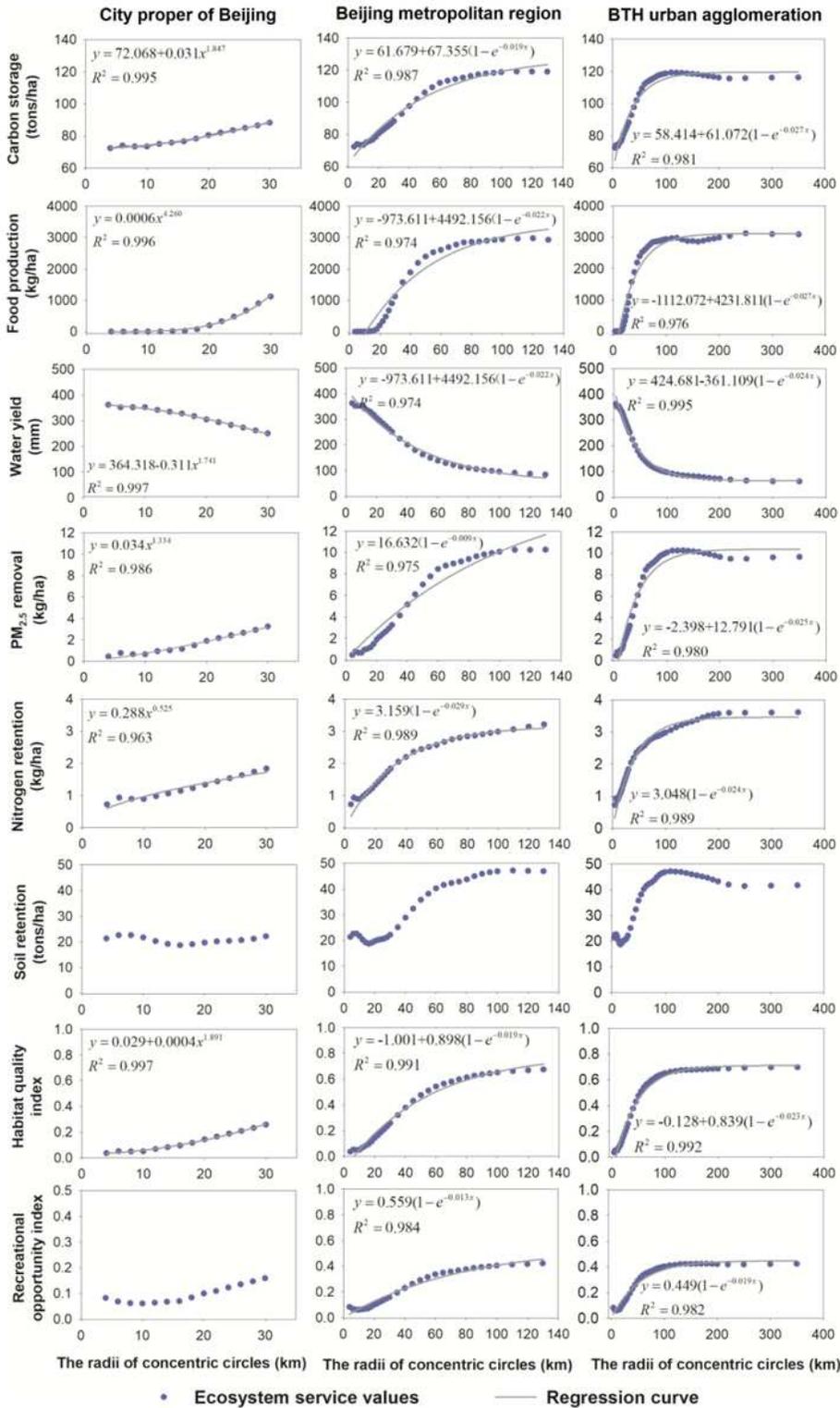


Figure 5

Scalograms of multiple ecosystem service indicators with respect to increasing concentric circle radii at various urban hierarchical levels: the city proper of Beijing, Beijing metropolitan region, and the Beijing Tianjin Hebei (BTH). For all regression curves and equations, the 269 significance level P values < 0.001 .

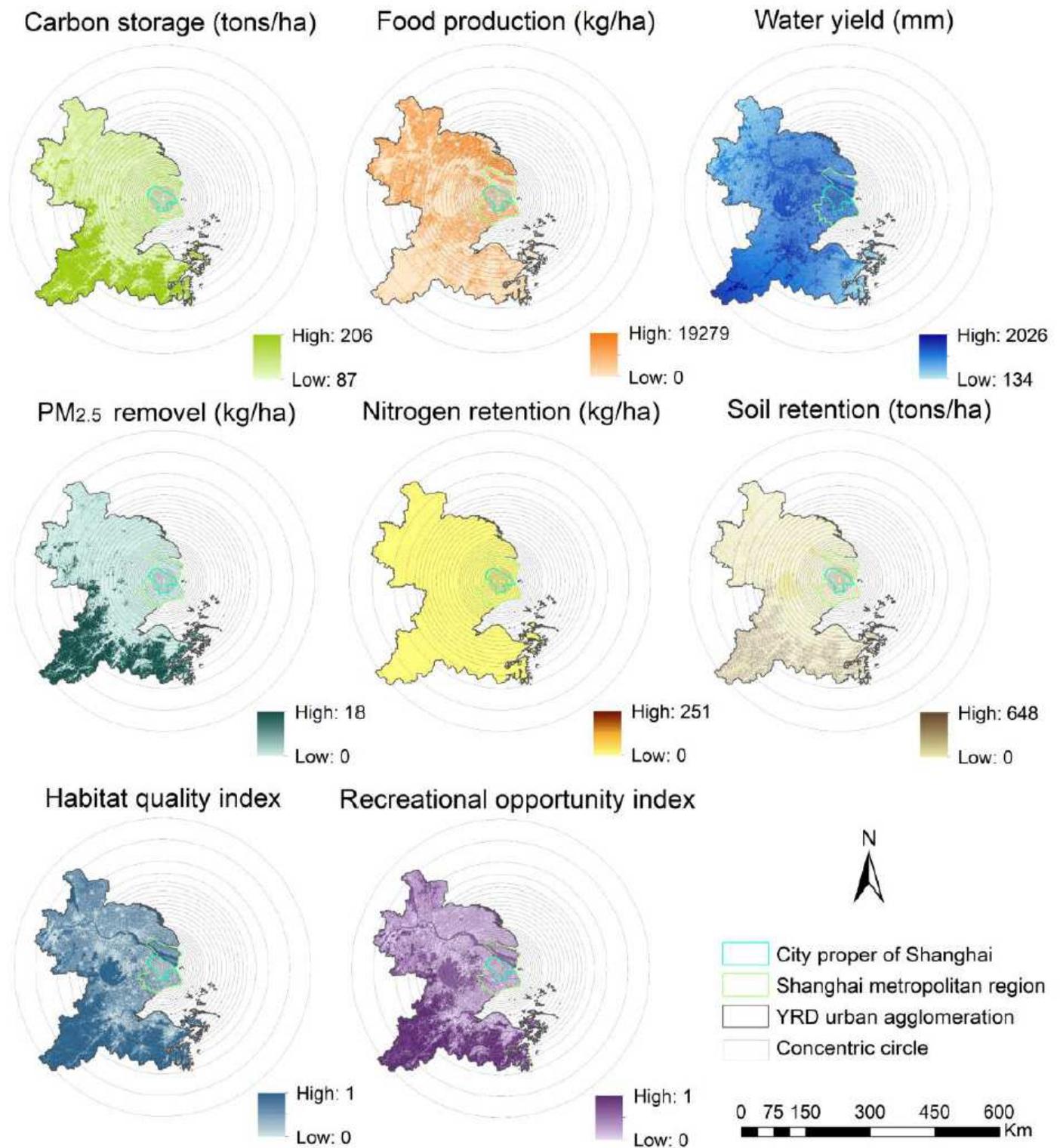


Figure 6

Spatial distributions of multiple ecosystem services in 2018 for Yangtze River Delta (YRD) urban agglomeration. The radii list of concentric circles was shown in Table 2. Note: The designations employed and the presentation of the material on this map do not imply the expression of any opinion whatsoever on the part of Research Square concerning the legal status of any country, territory, city or

area or of its authorities, or concerning the delimitation of its frontiers or boundaries. This map has been provided by the authors.

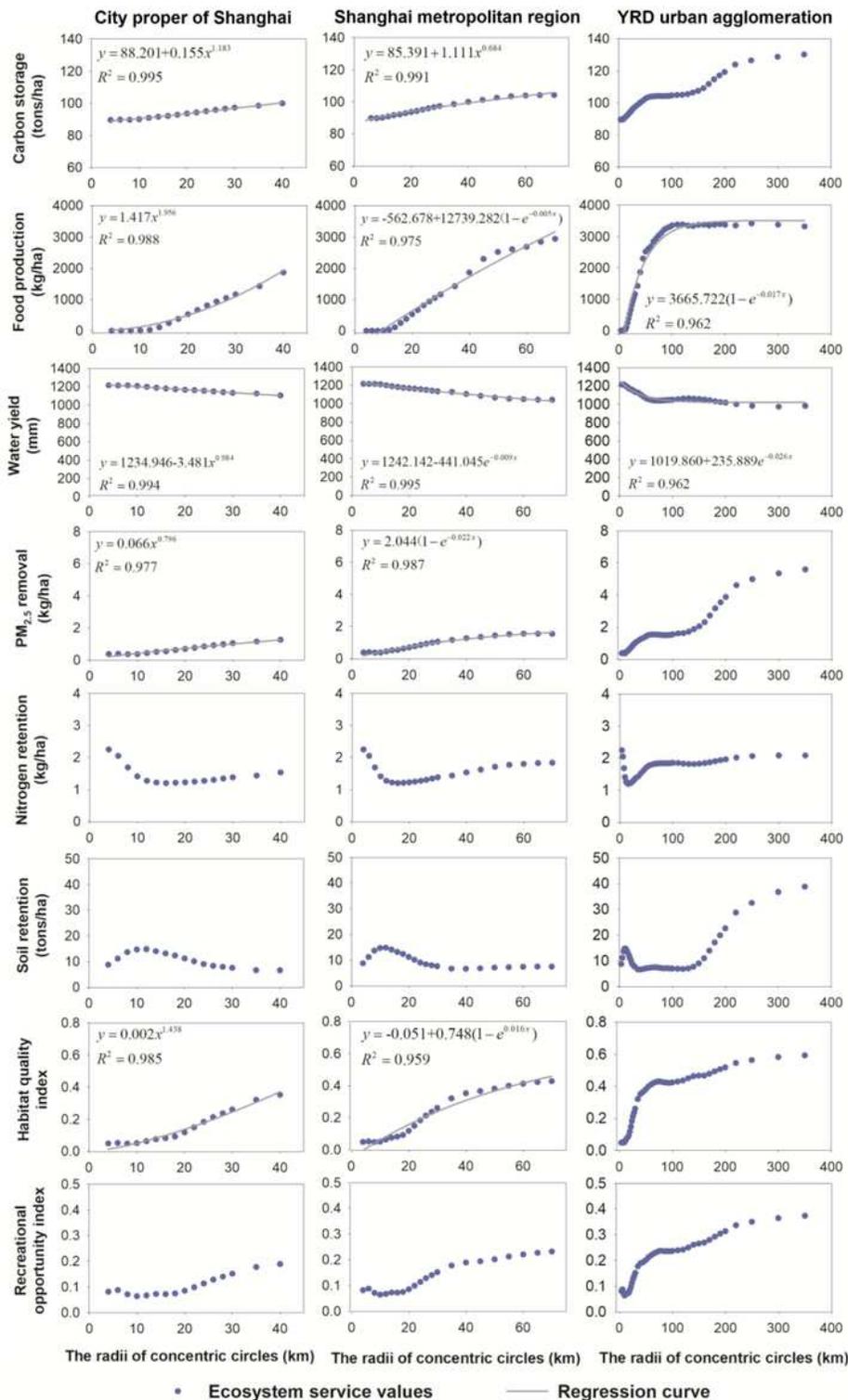


Figure 7

Scalograms of multiple ecosystem service indicators with respect to increasing concentric circle radii at various urban hierarchical levels: the city proper of Shanghai, Shanghai metropolitan region, and the

Yangtze River Delta (YRD). For all regression curves and equations, the significance level P values < 0.001.

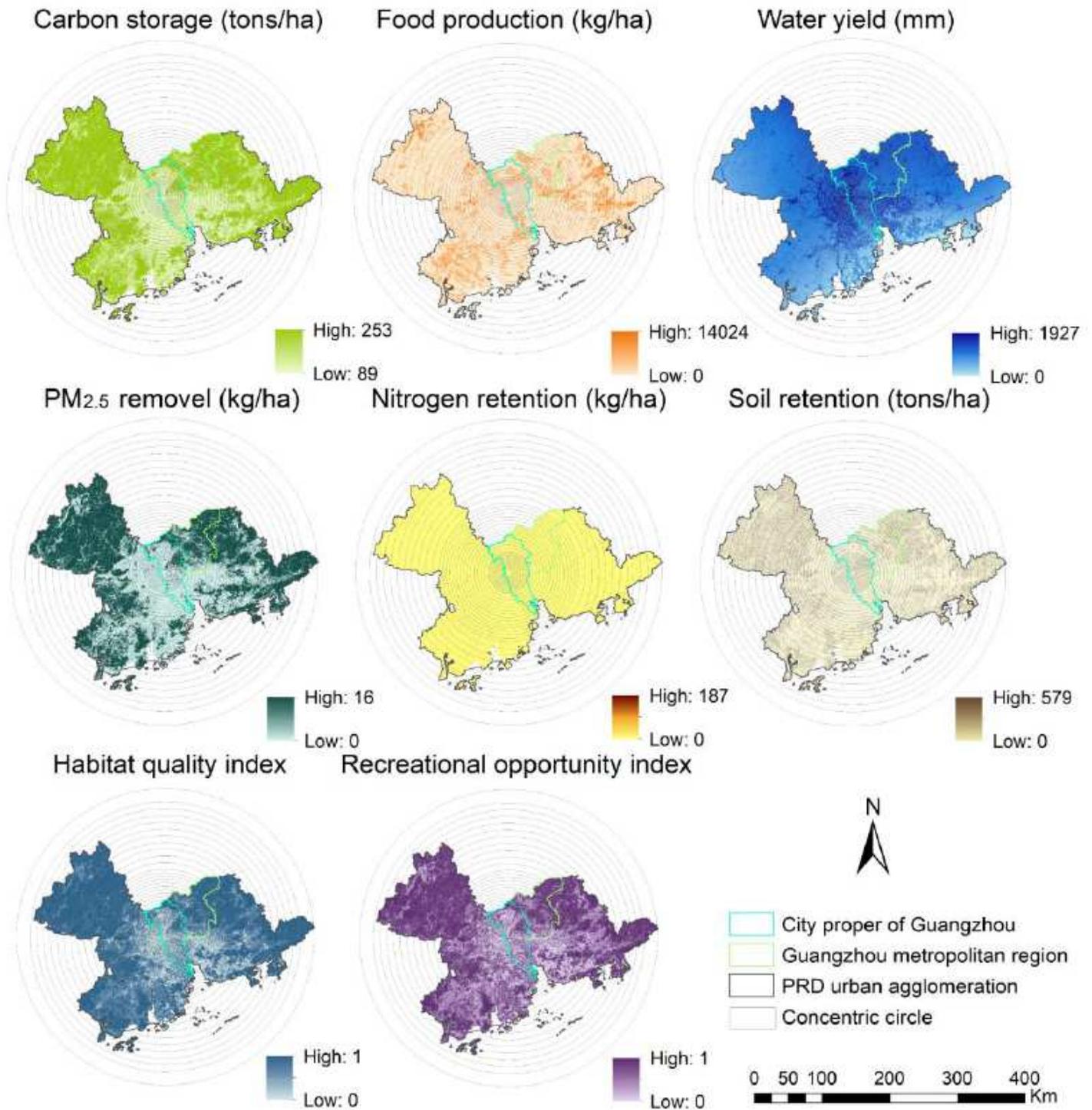


Figure 8

Spatial distributions of multiple ecosystem services in 2018 for Pearl River Delta (PRD) urban agglomeration. The radii list of concentric circles was shown in Table 2. Note: The designations employed and the presentation of the material on this map do not imply the expression of any opinion

whatsoever on the part of Research Square concerning the legal status of any country, territory, city or area or of its authorities, or concerning the delimitation of its frontiers or boundaries. This map has been provided by the authors.

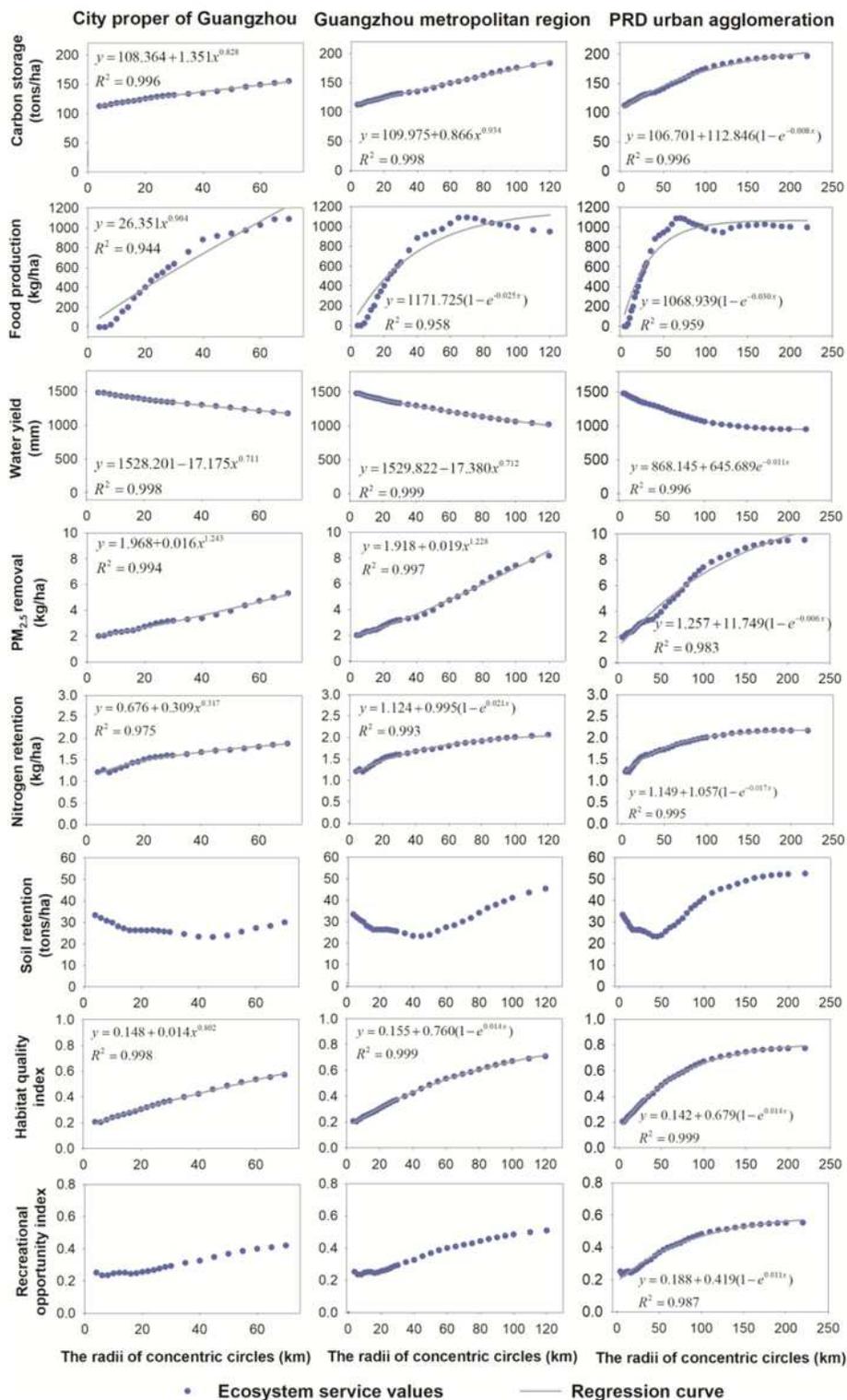


Figure 9

Scalograms of multiple ecosystem service indicators with respect to increasing concentric circle radii at various urban hierarchical levels: the city proper of Guangzhou, Guangzhou metropolitan region, and the

Pearl River Delta (PRD). For all regression curves and equations, the significance level P values < 0.001.

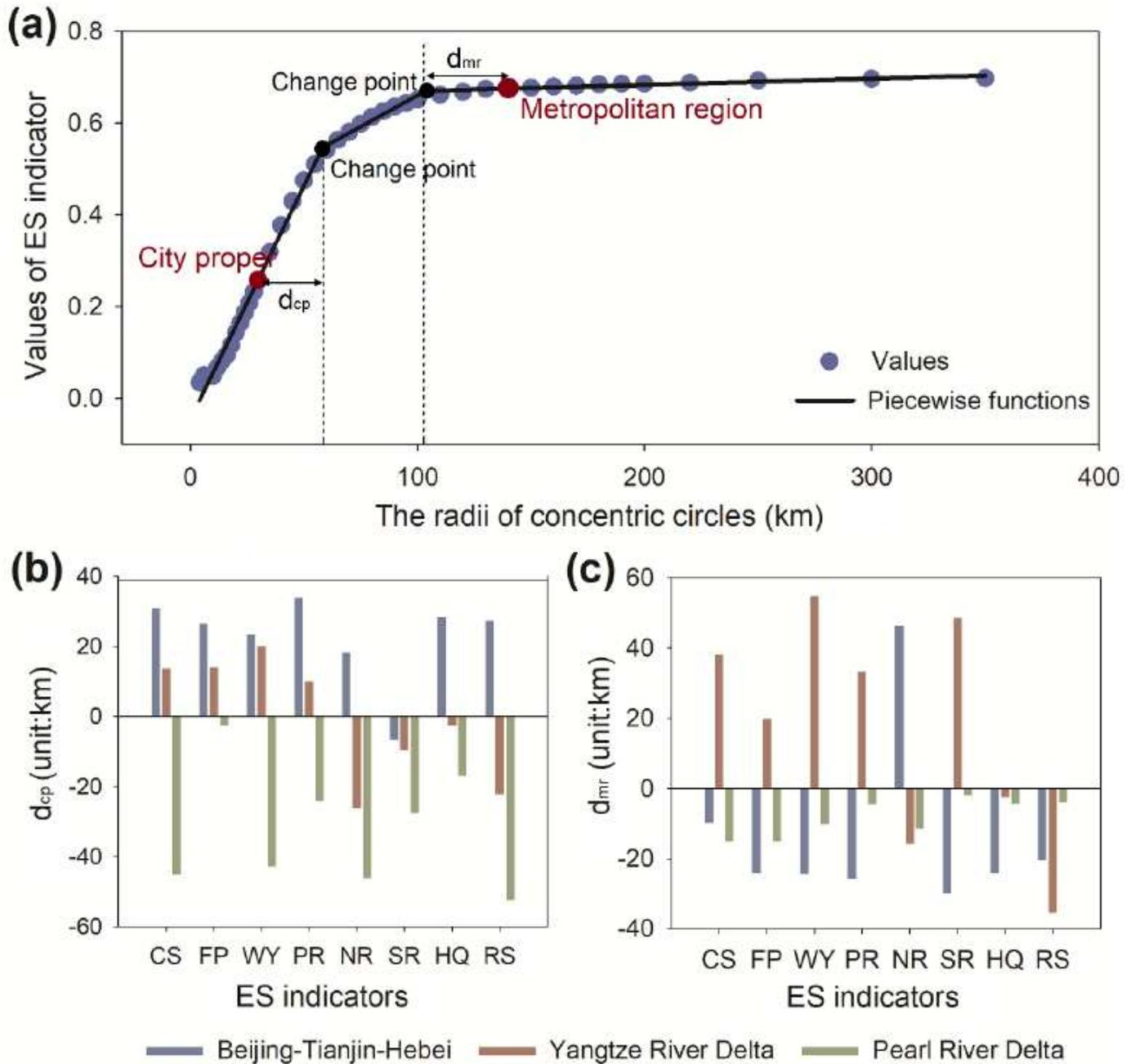


Figure 10

(a) Example of a scalogram for ecosystem service (ES) change points versus the administrative boundary points. The solid black dots represent the turning points for the ES change rates; the solid red dots represent the administrative boundary points. The “ d_{cp} ” represents the distance between the first change point of ES scaling relations and the administrative boundary of the city proper; the “ d_{mr} ” represents the distance between the second change point of ES scaling relations and the administrative boundary of the metropolitan region. Distance was calculated by the change point value minus the administrative boundary value. (b) The “ d_{cp} ” for multiple ESs in Beijing Tianjin Hebei (BTH), Yangtze River Delta (YRD), and Pearl River Delta (PRD) urban agglomerations. CS = carbon storage; FP = food

production; WY = water yield; PR = PM2.5 removal; NR = nitrogen retention; SR = soil retention; HQ = habitat quality; RS = recreational opportunity. (c) The “dmr” for multiple ESs in BTH, YRD, and PRD urban agglomerations.

Supplementary Files

This is a list of supplementary files associated with this preprint. Click to download.

- [SupplementaryInformationSI.docx](#)
- [Supplementaryfigures.docx](#)