

Scenarios of environmental change in a post-conflict Colombia

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1 **Scenarios of environmental change in a post-conflict Colombia**

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21

22 **Abstract**

23 *Colombia is a notorious biodiversity hotspot that came to the spotlight of conservation upon*
24 *the signing of the peace agreement in 2016. Here we used a counterfactual approach to*
25 *forecast by 2036 the impact of deforestation on Colombia's biodiversity and carbon stocks*
26 *under three scenarios: (1) pre-signing of the peace agreement, (2) post-signing and (3)*
27 *business-as-usual. We found that if deforestation rates continued at the same pace of post-*
28 *signing, up to 41,000 km² of forest area may be lost by 2036, whereas pre-signing rates*
29 *would save nearly 25,000 km² (equivalent to the total forest loss observed between 2000 and*
30 *2018). Under the pre-signing scenario, between 2018-2036 Colombia would reduce the*
31 *average impact on the range of forest-specific species by nearly 50% of habitat area relative*
32 *to 2000-2018, whereas under the post-signing scenario, it would increase by 33%. Moreover,*
33 *losses of 312-807 Mm³ of growing stock volume and 267-688 Mt of aboveground biomass*
34 *were projected by 2036, jeopardizing the country's commitments towards international*
35 *conservation as well as climate targets. Importantly, we found a mismatch at the department*
36 *level on biodiversity and biomass losses, which highlight an urgent need to generate coherent*
37 *policies at a national level aiming to tackle both issues.*

38

39 Colombia is recognized as one of the most important hotspots of biodiversity
40 worldwide¹ and unlike many other countries is still largely covered by its natural ecosystems,
41 with more than 50% of its area covered by forests². The civil war that ravaged the country
42 from the 1960s until 2016 has had social, political, economic and environmental
43 consequences³⁻⁷. The Revolutionary Armed Forces of Colombia—People's Army (FARC,
44 Spanish acronym) was the first armed group with political goals to profit from illegal crops,
45 often cultivated under tree canopies⁸. The paramilitaries, on the other hand, had the support
46 of the main drug traffickers⁴ as well as wealthy cattle ranchers forcing population

47 displacement⁴. Such factors are widely recognised drivers of land cover change across
48 tropical forests⁹⁻¹¹. In a context of violence and economic interests, institutions are less able
49 to take actions to safeguard the environment¹². Furthermore, as the population is forced to
50 leave, the land can be used for further legal and illegal activities that contribute to
51 deforestation, such as mineral exploitation¹². The so-called “gunpoint conservation”, i.e.
52 hostile regions where FARC and paramilitaries enforce and maintain forest preservation
53 through the threat of violence¹³, was used by armed groups as refuge from air surveillance by
54 government forces⁸, to house their headquarters, and for their role in the local hydrology⁴.

55 Such social dynamics have led to important changes in the country’s forest cover. In
56 fact, the Institute of Hydrology, Meteorology and Environmental Studies (IDEAM, Spanish
57 acronym) reported that from 1990 through 2015, the total amount of forest lost was 52,342
58 km², representing 8% of the forest area in 1990^{14,15}. Particularly critical was the twenty-year
59 period between 1990 and 2010 in which the annual deforestation rate increased from 1,300
60 km² between 1990 and 2000 to 3,630 km² between 2000 and 2010^{16,17}.

61 After three years of negotiations, the Government of Colombia and the FARC signed
62 the peace agreement in La Havana, Cuba¹⁸ in November 2016. In parallel to this process, the
63 Government of Colombia developed the National Reducing Emissions from Deforestation
64 and forest Degradation (REDD+) Strategy that intends to mitigate climate change through
65 reforestation and sustainable forest management¹⁸. To achieve this goal, the peace accords
66 promise land titles to former combatants in areas historically controlled by the FARC and
67 allocates funding to build roads throughout former FARC territories. As a consequence, a
68 vast area of highly biodiverse forest historically used as coverage from air-assisted detection
69 by authorities is now accessible for extractive industries and agricultural expansion¹⁹, and
70 early signs of increased deforestation are already noticeable with important consequences in
71 terms of carbon emissions and biodiversity loss²⁰. Therefore, it is crucial to try and anticipate

72 the environmental impacts of these increased deforestation rates if Colombia is to ever
73 achieve its international commitments under the Paris Agreement (20-30% reduction in
74 carbon emissions by 2030) and consolidate biodiversity as the differential factor on global
75 competitiveness based on its natural capital.

76 Using a counterfactual approach based on a spatially-explicit model²¹, we projected
77 deforestation in Colombia up to 2036 under three scenarios, reflecting three different model
78 calibration periods (see Methods). The business-as-usual (BAU) scenario reflects 18 years of
79 deforestation in Colombia (2000-2018), whereas the Low scenario reflects the three years
80 leading to the signing of the peace agreement (2013-2015) and the High scenario reflects the
81 three years post-signing of such agreement (2016-2018). The model was used with a
82 combination of drivers that include the most used predictors of tropical deforestation^{20,22,23}.

83 The historical (2000-2018) and future (2018-2036) deforestation maps at 1 km² were
84 analysed for the whole country as well as per department to assess which areas are under
85 greater risk of future forest conversion. Additionally, by overlapping these forest maps
86 (historical and future for the three scenarios) with expert-validated freely available species
87 distribution maps for Colombia²⁴ (Figure S1) as well as maps of aboveground biomass and
88 growing stock volume²⁵ (Figure S2), we determined the environmental impacts of expected
89 conversions, and compared it with the historical impacts.

90 Our analyses showed that the main drivers of deforestation were mostly consistently
91 across the three time periods analysed, however, there were important differences to notice.
92 For instance, the location of previous deforestation was an important variable in the BAU and
93 Low scenarios but not in the High scenario, which suggests that deforestation moved to new
94 areas that were previously relatively intact (Table 1). Further, we observed less loss in areas
95 of higher armed conflict, but that effect disappeared in the most recent period, although
96 deforestation seem to take place closer to areas of coca plantations in all periods. As it is

97 typical, roads consistently played an important role in driving deforestation, while elevation,
98 slope and areas of protection as well as indigenous and black communities territories prevent
99 further losses across all time periods in Colombia (Table 1). Importantly, the fact that the
100 strength of the parameters' values for conservation areas (protected areas, African
101 communities and indigenous communities) is reducing, supports current findings that these
102 areas may be losing their effectiveness in acting as deforestation stoppers²⁶. Finally, higher
103 parameter values for distance to coca plantations in the most recent scenario (High) relative
104 to the previous three years (Low) supports that the deforestation frontier is moving towards
105 forests located closer to where these plantations are^{5,27}.

106 Forest cover in Colombia reduced from 756,894 km² in 2000 to 731,353 km² in 2018
107 (Figure 1), i.e. an average annual deforestation rate of 1419 km²/year (2000-2018). This rate
108 increased sharply in 2016-2018 to 2597 km²/year, nearly three times higher than in the
109 previous three years (986 km²/year). Such fluctuations in the deforestation rates resulted in
110 marked differences in the projections by 2036. In detail, under the BAU scenario we project
111 that Colombia could lose a further 25,560 km² (± 14 km² [standard error, s.e.]) of forests by
112 2036, compared to 16,402 km² (± 12 km² s.e.) in the Low scenario and 41,062 km² (± 18 km²
113 s.e.) in the High scenario (Figure 1). This means that if an effort had been made to keep the
114 rates at the low levels of 2013-2015, Colombia could have saved 2341 km² (± 7.4 km² s.e.) of
115 forest area by 2018, and could save up to 24,660 km² by 2036, an area almost equivalent to
116 what was observed to be lost between 2000 and 2018 (25,541 km²).

117 Historically, deforestation in Colombia has varied greatly across departments (Table
118 S1). Over the 18 years analysed, there were 46 km² per year of forest being cut down on
119 average per department; with Caquetá, Meta and Antioquia being the departments with the
120 highest annual deforestation rates, respectively 278 km²/year, 204 km²/year and 159 km²/year
121 (Figure 2a). Furthermore, between 2013-2015 and 2016-2018 there was an increase in the

122 annual deforestation rate in all departments of Colombia, from an average 32 km²/year to 84
123 km²/year (Table S1). Such increase was mostly concentrated in Caquetá and Meta where
124 there was a sharp acceleration (509 km²/year and 356 km²/year, respectively) but also in
125 Guaviare (289 km²/year) (Table S1).

126 The three scenarios resulted in different amounts of forest lost for each department
127 (Figure 2a). We projected that deforestation would mostly continue to occur alongside areas
128 that have previously been deforested (Figure 3). We do however project that the expansion
129 into new territories occurs at a more accelerated pace in the High scenario than in the Low
130 (Figure 3d). At the department level, we found that on average, each department is projected
131 to lose 824 km² (or 4.2% of its forests), 529 km² (2.9 %) and 1324 km² (7.2%) of forest from
132 2018 through 2036, under the BAU, Low and High scenarios, respectively (Figure 2a). By
133 comparison, in the previous eighteen years, there was an average total loss of 823 km² or
134 4.3% of forest. This means that ensuring the rates at the level of the Low scenario would
135 allow to keep the deforestation rates significantly lower than the previous 18 years whereas
136 the High scenario would increase these by more than 50%. For instance, in terms of absolute
137 area, Caquetá was projected to lose 4918 km² of forest from 2018 through 2036 under the
138 High scenario compared to less than half of that under the Low scenario (2150 km²),
139 perpetuating the trends from the previous years (Table S1). Proportionally, though, other
140 departments emerge as more problematic (Table S1). For example, under the High scenario,
141 and with a projected loss of 1034 km² of forest by 2036, Arauca is expected to lose 17% of its
142 2018 forest cover, Bolivar is projected to lose 16% (2227 km²) and Sucre 14% (348 km²).

143 Such losses are expected to have had (and will continue to have) extensive
144 environmental and ecological impacts on the biodiversity- and carbon-rich forests of
145 Colombia. To assess such impacts, we determined a variety of indicators: loss of
146 aboveground biomass and growing stock volume (Figure 3b and c); loss in biodiversity -

147 species richness (as a whole), forest-specific species, endemic species and species that are
148 both endemic and forest-specific (Figure 4 a-d), and per species impacts (Figure 5).

149 We estimated that historically (2000-2018) Colombia has lost 376.42 Mt of
150 aboveground biomass (AGB) and 440.62 M m³ of growing stock volume (GSV). On average,
151 each department lost 673,560 t/ year (s.e. 194,112 t/year, 3.29% of the value in 2000), and
152 788,409 m³/year (s.e. 226,797 m³/year, 3.27% of the value in 2000). Since Caquetá, Meta and
153 Guaviare were the departments with greatest forest losses, these were also the departments
154 with greater absolute losses of AGB and GSV (Figure 2b and c). Proportionally, however,
155 Meta, Bolivar and Sucre emerge as those departments with highest losses, i.e. 6-7% of the
156 AGB and 6-8% of GSV present in 2000 (Tables S2 and S3).

157 Under the High scenario, we projected that Colombia may lose up to 807.06 M m³ of
158 GSV by 2036, i.e. 4.45% of the remaining GSV in the forests of the country. By contrast in
159 the Low scenario, 312.30 M m³ would be lost, or 1.72% of the remaining GSV, saving up to
160 494.76 M m³ if the rates of deforestation stayed as low as in the 2013-2015 period. Similarly,
161 in terms of AGB, we projected losses between 267.10 and 688.32 Mt by 2036, or between
162 1.75 and 4.5% of the remaining AGB of the country's forests. The effort to keep rates at the
163 Low level would have saved nearly the same amount as the loss of AGB observed between
164 2000 and 2018. While in absolute terms these losses are expected to continue in the same
165 departments as previously, i.e. Meta, Guaviare and Caquetá, proportionally, we found that
166 under the High scenario new departments emerge with rapid losses, such as Guainía, Vichada
167 and Valle del Cauca (Tables S2-S3, Figure 2b, c), thus highlighting the need for a
168 coordinated regional policy to prevent further deforestation as to avoid leakages across
169 departments.

170 Future deforestation will also lead to important impacts on biodiversity²⁸. We found
171 that on average, between 2000 and 2018, the number of overall species in each 1 km² pixel

172 reduced by 3.06% in Colombia. For forest-specific species, the reduction was from 335
173 species to 324 (3.28%), whereas the endemic species reduced by 2.78% and forest-specific
174 endemic species by 2.63%. By department, we found that on average 21 species were
175 impacted between 2000-2018, 11 of which were forest-specific (Tables S4-S7).
176 Proportionally, we estimated that departments have lost 3.69% of its biodiversity, 3.81% of
177 the forest-specific species, 2.44% of its endemic as well as forest-specific and endemic
178 species. Unlike the AGB, GSV results, when it comes to biodiversity, the focus is placed on
179 other departments: Meta, Putumayo and Atlántico had the highest proportional losses (Figure
180 4a-b) of overall and forest-specific species (Tables S4-S5). However, when it comes to
181 endemic species (Figure 4c-d), the highest losses were observed in Atlántico, Bolivar and
182 Santander (Tables S6-S7). Such results reiterate the mismatch between carbon and
183 biodiversity in tropical regions²⁹, and support the need to use multiple indicators when
184 assessing environmental impacts of forest loss. These different impacts may play a different
185 role in decision making and thus help support more targeted policies towards conservation.

186 Under the pre-signing scenario, between 2018-2036 Colombia would reduce the
187 average impact on the range of forest-specific species by nearly 50% relative to 2000-2018,
188 whereas under the post-signing scenario it would increase by 33% (1.48% and 3.92% vs
189 2.93%). Moreover, we projected further losses across departments with an average of 12 to
190 31 species being impacted (Low to High), of which nearly 50% are forest-specific and thus
191 less likely to persist once the trees are removed. By 2036, for each department, this represents
192 an average loss of between 2.5% and 6.3% of the species present in 2018 (Tables S4-S5),
193 with departments such as Arauca, Bolivar e Sucre having the highest proportional losses by
194 2036 (Figure 4a-b), ranging from 6 to 15% of overall and forest-specific species losses
195 relative to 2018. For the endemic species, the proportional losses are expected to range from

196 an average 1.7% (Low) to 4.6% (High) (Tables S6-S7), with Santander, Sucre and Bolivar
197 being projected to have the highest losses (Figure 4c-d).

198 While aggregated species indicators may be useful to highlight hotspots of change,
199 species-specific analyses can provide additional information to help design more targeted
200 conservation actions, i.e. to incorporate the ecology of the species and what is known of the
201 behaviour into policy including information about thresholds of species loss³⁰. Therefore, our
202 final set of results concern species-specific indicators to highlight species at under greater
203 risk of being impacted by future deforestation in Colombia. On average, 71% of these
204 species' potential distribution (Figure S1) overlapped with forests in 2000, reducing by 3% in
205 2018. Between 2000 and 2018, each species lost on average 2.6% of their habitat, reaching a
206 maximum of 19% loss in the case of *Alchornea castaneifolia*; *Xiphorhynchus elegans* forest-
207 specific species with highest loss of 18%, and *Ortalis garrula* endemic species with highest
208 loss (-12%). From the three higher IUCN threat categories, we found that the Near
209 Threatened *Egretta rufescens* (-11%), the Endangered *Saguinus leucopus* (-7.4%), which is
210 also an endemic species and the Critically Endangered *Crocodylus intermedius* (-3.2%) were
211 the most impacted species in terms of range loss.

212 Importantly, endemic species are widely impacted (Figure 5), but which ones depend
213 on the scenario. For instance, the critically endangered and endemic species *Lipaugus weberi*
214 would be more impacted than in the past in two out of three scenarios (except Low, Figure 5),
215 whereas the Endangered endemic species *Saguinus leucopus* is projected to be only as
216 impacted as in the past in the BAU scenario (Figure 5c). Such variations are even higher
217 when we analyse impacts at the department level (Table S9). One important limitation is that
218 we did not consider whether the species could survive in highly transformed habitats.
219 Nonetheless, this information is crucial for policy and decision-makers to develop effective
220 measures as there is no one-size-fits-all when it comes to species' conservation. This informs

221 not only on the amount of forest lost, the potential carbon and stock volume losses, the
222 overall species impacted, and the specific species impacts. Such information allows policy
223 makers to assess the trade-offs of allowing deforestation to follow certain scenarios at the
224 department and region levels.

225 With our analyses we firstly aimed to provide policy- and decision-makers in
226 Colombia a better understanding of the changes in drivers of deforestation (at the national
227 scale) across the three different periods analysed, and then provide a combined estimate of
228 potential future environmental impacts (in terms of species losses, aboveground biomass, and
229 growing stock volume) in Colombia. While some of the drivers did not change over time (e.g.
230 roads), importantly, we demonstrated that conservation areas may be losing their
231 effectiveness in preventing deforestation and that new frontiers are being opened close to
232 areas previously under control of armed groups, a finding supported by recent research³¹⁻³³.
233 An understanding of where deforestation is likely to occur and the drivers leading to such
234 changes is useful for decision making given the current and expected trends of deforestation
235 risking the biodiversity of one of the most important hotspots across the globe. Further, with
236 our department level analyses, we hope to attribute a sense of responsibility not only to the
237 government of Colombia but also to the regional authorities to ensure that they act before it is
238 too late.

239 While we obviously do not advocate for a ‘gunpoint conservation’ approach, and the
240 current pandemic might also play an important role, it is important to recognise that the years
241 leading to the signing of the agreement were beneficial for the country’s natural resources.
242 Suggesting that formal governance is essential for Colombia to protect its forests³⁴, since
243 higher state presence and judicial capacity seem to reduce deforestation³⁵. Our analyses are in
244 line with previous research that raised concerns about the future of the ecosystems in
245 Colombia in a post-peace agreement era^{5,31,33}, and support calls for the Colombian

246 government to strengthen environmental research and engage scientists in decision-making
247 processes³⁶. As Colombia ratified several international commitments in terms of biodiversity
248 conservation as well as climate change action, highlighting potential consequences of
249 investment in infrastructure, and allowing deforestation to progress without major counter-
250 action, is hopefully important to raise policy- and decision-makers on the need to act now.

251

252

253 **Methods**

254 *Model Input Data Sources*

255 To analyse forest cover change in Colombia we downloaded the freely available tree
256 cover (0-100%) and tree loss (0-18) dataset from Hansen et al. (2013), version 1.6 (2000-
257 2018). The tree cover dataset was transformed into a binary forest/non-forest map (1/0) using
258 a conservative threshold higher than the forest definition of the Food and Agriculture
259 Organisation (10% tree cover)³⁷, i.e. a pixel was classified as forest if tree cover was above
260 50%. The full time series (2000-2018) was then split into three distinct time periods for all
261 subsequent analyses.

262 As highlighted by Baptiste et al.⁵, Aguilar et al.³⁸ and others, forest cover change and
263 the armed conflict in Colombia have historically been quite interrelated. It has been argued
264 that forests were used to protect coca cultivation, and that areas where the armed conflict was
265 more intense have been protected by the lack of socio-economic development. Therefore, we
266 split the time series into three periods related to the peace agreement signing: *post-signing*
267 corresponding to the last three years (2016-2018) when deforestation rates increased rapidly
268 (High scenario); *pre-signing* (2013-2015) when deforestation rates were at the lowest levels
269 (Low scenario), and the full time series (2000-2018) as a business-as-usual scenario (BAU
270 scenario). These three periods were used to calibrate three different deforestation models,
271 thus assessing the strength and direction of impact of the different drivers across time

272 periods, and then used to project deforestation up until 2036. We then compared the
273 outcomes not just in terms of forest change under each scenario, but also potential impacts on
274 aboveground biomass and local biodiversity.

275 Alongside the forest maps, we collected a myriad of variables representing the main
276 direct and indirect drivers of forest cover change in Colombia^{20,39}, and on tropical
277 deforestation more broadly²². For instance, Armenteras et al.²⁰ and Ayram et al.³⁹ identified
278 some of the ‘traditional’ drivers (roads, population, elevation, slope, etc.), whereas Murad
279 and Pearse⁴⁰ as well as Davalos et al.⁴¹ highlighted the role of agriculture (cattle ranching),
280 and others identified more Colombia-specific drivers such as coca cultivation^{9,41}, illegal
281 mining and oil pipelines⁴². Conversely, protected areas have been found to reduce
282 deforestation³² although their effectiveness seems to be reducing since the peace agreement,
283 with increased fires taking place inside these areas³³. As a result of this literature search, we
284 included predictors that reflected accessibility, such as the Euclidean distance to the nearest
285 road (in m), to the nearest trail (in m), to the nearest river (in m), as well as the slope (in
286 degrees) and elevation (in m); predictors related to other socio-economic activities such as
287 soil fertility (low to high), Euclidean distance to extraction wells (in m), and to mining
288 operations (in m), as well as population density (number of people per km²). Two variables
289 were directly related to the conflict: armed conflict density (number of armed conflict actions
290 per km²), and Euclidean distance to coca plantation (in m). Finally, three land tenure related
291 variables: protected areas, indigenous areas and African communal lands (all as binary
292 variables, in/out). The data were all freely available upon request from the Alexander von
293 Humboldt Institute in Bogotá, Colombia. All data layers were pre-processed to ensure spatial
294 overlap at 1 km² resolution using the MAGNA Colombia coordinate system, and are shared
295 publicly in the author’s GitHub repository (see Supplementary Materials).

296

297 *Spatially-explicit deforestation model*

298 In this study we used a widely validated spatially-explicit tropical deforestation
299 model^{21,29,30,43,44} (code provided in GitHub, see Supplementary Materials) to: (1) understand
300 the fluctuations in the effect of the drivers of deforestation in Colombia in the three different
301 time periods (2000-2018, 2013-2015, and 2016-2018); and (2) project the future rates and
302 location of deforestation using the three models as three different scenarios (Low, BAU and
303 High). In practice, the model projects the cumulative probability of each 1 km² pixel
304 changing from forest to non-forest by 2036, and it allows us to understand what is driving
305 deforestation in each of the time periods analysed.

306 The model is based around $P_{defor,x,t}$, i.e. the probability that pixel x becomes deforested
307 in a set interval of time t . This probability is defined as a logistic function (Eq. 1), so
308 as $k_{x,t}$ goes from minus infinity to plus infinity, $P_{defor,x,t}$ goes from 0 to 1. Simple linear models
309 can then be tested for $k_{x,t}$ as a function of the driver variables affecting location x at time t .

310
$$P_{defor,x,t} = \frac{1}{(1 + \exp^{-k_{x,t}})} \quad \text{Eq. (1)}$$

311 To ensure comparability, each model was initiated for each time period with all
312 variables (representing the drivers of deforestation) described before. Then, using an internal
313 model selection procedure based on maximum likelihood (see Rosa et al.²¹ for details) the
314 variables that mostly contributed to explain the rate and patterns of deforestation in Colombia
315 in each time period were identified. Once the ‘best’ model (the one where the combination of
316 variables led to the highest test likelihood) for each time period was identified, each models’
317 area under the receiver operating characteristic (ROC) curve, a measure of goodness-of-fit of
318 the model, was determined for each time period. All three models showed good fitness to the
319 data, i.e. AUC of 0.89 (BAU), 0.90 (Low) and 0.90 (High).

320 The calibrated models were then used to project future deforestation by 2036 under
321 each of the three scenarios (Low, BAU and High). In fact, to allow for the uncertainty of the

322 parameter values to be propagated through the projections, we randomly sampled the
323 Gaussian distribution of each parameter for 100 iterations to ensure that each iteration would
324 lead to a slightly different outcome. Further, the model does not impose a rate of change,
325 rather this emerges from the probabilities determined by the model itself, by comparing the
326 probability with a random uniform number for each pixel (in each year) and then selecting
327 that pixel to be deforested if the probability is larger than the random number. The procedure
328 is repeated for the same 100 iterations to allow for quantification of uncertainty posed by the
329 sampling of the posterior distribution of the parameters as well as the random selection
330 procedure. In practice, for each year and each scenario, we produced 100 forest/non-forest
331 maps, as well as an accumulated probability of deforestation map by 2036. The probability
332 maps are the result of combining all iterations into a final probability of each pixel being
333 deforested, i.e. a pixel that was selected 80 times out of 100 to be deforested would have an
334 80% chance of being deforested by 2036. For a more in-depth description of the modelling
335 approach, see Rosa et al.²¹, where the model was originally described.

336

337 *Estimating impacts on local biodiversity*

338 To assess potential impacts on biodiversity in Colombia (as a country and per
339 department), we overlapped the historical forest maps (2000, 2018) and our projections of
340 deforestation by 2036 with the BioModelos dataset²⁴ (see Data availability in Supplementary
341 Materials). This dataset contains expert-validated species distribution maps at 1 km², and
342 includes 5808 species for Colombia, of which, according to the International Union for
343 Conservation of Nature (IUCN) 2313 are forest-specific, 99 are endemic and 84 are both
344 endemic and forest-specific. These were the four subsets of data that we used in our analyses
345 (Figure S1) for the aggregate analyses, alongside species-specific analyses were also
346 performed.

347 On average, we found that an individual 1 km² pixel contained 908 species, 464
 348 forest-specific species, three endemic species and two species that were simultaneously
 349 endemic and forest-specific. Of the forest-specific species in the Biomodelos dataset, four
 350 were considered Critically Endangered (CR), two of them endemic, i.e. *Crocodylus*
 351 *intermedius*, *Hyloxalus abditaurantius*, *Lipaigus weberi* and *Oophaga histrionica*. The
 352 seventeen species classified by IUCN as Endangered (EN) are all forest-specific species,
 353 eight of which are endemic. Eighty are Near Threatened (NT), 62 are Vulnerable (V) and
 354 2142 are of Least Concern (LC), with the remainder either Data Deficient or Unavailable.
 355 Departments varied greatly in terms of number of species present with Casanare, Caquetá and
 356 Meta being the richest in terms of forest-specific species, and Antioquia, Córdoba and
 357 Risaralda the richest in terms of endemic forest species (Table S8).

358 Our analyses focused on observed impacts (2000-2018) as well as projected impacts
 359 (2018-2036) under the three scenarios: Low, BAU and High, both for Colombia as a whole
 360 and per department (Table S10). In particular, we overlapped the forest maps in 2000, in
 361 2018 and the 100 forest/non-forest maps projected in 2036 under each scenario to estimate
 362 the loss in range (in %) of the four groups of species mentioned above (total richness, forest-
 363 specific richness, endemic richness and forest and endemic richness), as well as for each
 364 individual species (Eqs. 2 and 3). Results for the whole of Colombia and at the department
 365 level are provided as an average of the impacts on each group of species (Eq 4., Tables S4-7)
 366 as well as the total impact on each species individually (Eq. 5). Finally, the impacts per
 367 species were ranked and the top-impacted species per department were also identified (Table
 368 S9). By providing species-specific information across two spatial scales we hope to inform
 369 more targeted conservation actions.

$$370 \quad SR_{2000,2018} = Forest_{2000,2018} \times SR_0 \quad \text{Eq. (2),}$$

$$371 \quad SR_{2036,s} = \frac{(\sum_i^I Forest_{2036,s} \times SR_0)}{I} \quad \text{Eq. (3),}$$

372 where species richness in time t equals the forest map in time t (2000, 2018 or each i
 373 of the $I = 100$ iterations in 2036, for each scenario s) multiplied by the species richness base
 374 map (SR_0). The species richness base map can be for all species, forest-specific species,
 375 endemic species, forest and endemic species or the individual species range, depending on
 376 the analysis being carried out.

377 To aggregate the results per department or for the whole of Colombia:

$$378 \quad \overline{SR}_{d,t,s} = \frac{\sum_n^{N_d} SR_{n,t,s}}{N_d} \quad \text{Eq. (4),}$$

379 where, $\overline{SR}_{d,t,s}$ (in number of species) corresponds to the overall species richness,
 380 forest-specific species, endemic species, or forest-specific and endemic species; d is the
 381 department or Colombia as a whole, and N the total number of pixels ($N_d =$ number of pixels
 382 in department d). Repeated for each scenario (s) and each time step (t).

383 For each individual species, we determined the total range reduction by calculating
 384 $SPP_{d,t,s}$, i.e. the species range in time t under scenario s (Eq. 5), and then calculating the %
 385 change relative to a previous t (2000 or 2018). For example, to calculate the reduction in
 386 range of the species *Saguinus leucopus*, in department d , between 2018 and 2036 under the
 387 High scenario, we applied the following equations:

$$388 \quad SPP_saguinus_{d,t,s} = \sum_n^{N_d} SPP_saguinus_{n,t,s} \quad \text{Eq. (5),}$$

$$389 \quad \%SPP_saguinus_{d,18-36,HIGH} = \left(\frac{SPP_saguinus_{d,2036,HIGH} - SPP_saguinus_{d,2018}}{SPP_saguinus_{d,2018}} \right) \times 100 \quad \text{Eq. (6).}$$

390 The same logic was applied for all other species, all other departments, and all other
 391 time steps under consideration in our analyses. These temporal ratios were also determined
 392 for the groups of species.

393

394 *Estimating impacts on aboveground biomass and growing stock volume*

395 To assess potential impacts on aboveground biomass (AGB in t/ha) as well as
396 growing stock volume (GSV in m³/ha) in the whole of Colombia as well as per department,
397 we used the data provided by Santoro et al.²⁵ (see Data availability in Supplementary
398 Materials). As per the FAO definition, adopted by the authors, GSV represents the “volume
399 of all living trees more than 10 cm in diameter at breast height measured over bark from
400 ground or stump height to a top stem diameter of 0 cm, excluding smaller branches, twigs,
401 foliage, flowers, seeds, stump and roots”; whereas AGB is “the mass, expressed as oven-dry
402 weight of the woody parts (stem, bark, branches and twigs) of all living trees excluding
403 stump and roots”. In detail, we downloaded the two 40° x 40° tiles that covered the whole of
404 Colombia (N40W100 and N00W100), and processed the data to only include values within
405 the boundaries of the country (Figure S2). We downloaded the data for 2010 as well as the
406 error estimates on both AGB and GSV as provided by Santoro et al.²⁵. These latter datasets
407 were used to allow for uncertainty to be propagated in the analysis.

408 Afterwards, we followed a similar approach as the one described above for the
409 biodiversity datasets, i.e. we overlapped the forest maps in 2000, in 2018 and the 100
410 forest/non-forest maps projected in 2036 under each scenario with the AGB, GSV and
411 respective error maps (Eqs. 7-8). We then calculated the total AGB and GSV per pixel and
412 aggregated these values to the whole of Colombia as well as per department using a simple
413 sum (the same for the error maps) as stated in Eq. 9. Relative % changes in AGB/GSV over
414 time were then determined for each scenario and each department in a similar fashion as
415 described above in the biodiversity analyses. Although we found that AGB and GSV were
416 unsurprisingly highly correlated (Pearson’s correlation [r] = 0.99, p < 0.001), we still chose to
417 present both results as they provide estimates that might be relevant for different policy and
418 decision makers.

419
$$AGB_{2000,2018} = Forest_{2000,2018} \times AGB_0 \quad \text{Eq. (7),}$$

420
$$AGB_{2036,s} = \frac{(\sum_i^I Forest_{2036,s} \times AGB_0)}{I} \quad \text{Eq. (8),}$$

421 where AGB in time t equals the forest map in time t (2000, 2018 or each i of the $I =$
 422 100 iterations in 2036, for each scenario s) multiplied by the AGB base map (AGB_0). To then
 423 aggregate for the whole of Colombia or per department d :

424
$$AGB_{d,t,s} = \sum_n^{N_d} AGB_{n,t,s} \times 100 \quad \text{Eq. (9),}$$

425 where $AGB_{d,t,s}$ is the total aboveground biomass (tonnes) in each department or
 426 Colombia as a whole (d) for each time step t (2000, 2018, and 2036), under each scenario s
 427 (observed, Low, BAU and High). The same procedure was adopted for total GSV (m^3), as
 428 well as for the error maps for both AGB (tonnes) and GSV (m^3). N_d is the number of pixels in
 429 each department d .

430 On average, we found that an individual 1 km^2 forested pixel in 2000 contained
 431 137.91 t/ha of aboveground biomass (76.13 t/ha error) and $163.37 \text{ m}^3/\text{ha}$ of growing stock
 432 volume ($85.76 \text{ m}^3/\text{ha}$ error). The average AGB and GSV in Colombia were reduced by 2.40%
 433 and 2.37%, respectively, by 2018, with strong variations across departments (Tables S2 and
 434 S3). Finally, it is important to notice that when correlating AGB (and GSV) values with
 435 species richness, we observed that pixels with higher AGB (and GSV) also had more species
 436 in them, although this was a weak positive correlation ($r = 0.17$ for AGB and $r = 0.19$ for
 437 GSV , $p\text{-value} < 0.001$), which strengthen slightly with forest-specific richness data ($r = 0.24$
 438 for AGB and $r = 0.25$ for GSV , $p\text{-value} < 0.001$). Interestingly, pixels with higher endemism,
 439 overall or of forest-specific species, seem to have less AGB and GSV ($r = -0.09$ for AGB and
 440 $r = -0.08$ for GSV , $p\text{-value} < 0.001$, for all endemic species; and $r = -0.10$ for AGB and $r = -$
 441 0.09 for GSV , $p\text{-value} < 0.001$, for forest-specific endemic species).

442

443

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554

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558

559 **Author contributions**

560 IMDR and JMOQ designed the analysis and IMDR developed the model and run the
561 analyses. ED prepared the datasets and helped with data analysis. EK prepared the species
562 data and helped with analysis. MVP revised code. All authors contributed to the writing.

563 **Tables**

564 **Table 1** – Model calibration results for best model under each scenario (BAU: 2000-2018,
 565 Low: 2013-2015 and High: 2016-2018). Posterior distribution mean and 95% confidence
 566 interval used for the model projections. Model goodness of fit shown inside round brackets
 567 next to the scenario name (Area under the ROC curve).

568
 569

| Parameter | BAU (0.89) | | Low (0.90) | | High (0.90) | |
|------------------------------|----------------------------|---------------------------|----------------------------|---------------------------|----------------------------|---------------------------|
| | <i>mean</i> | $\pm 95\%$ <i>C.I.</i> | <i>mean</i> | $\pm 95\%$ <i>C.I.</i> | <i>mean</i> | $\pm 95\%$ <i>C.I.</i> |
| Intercept | -1.92 | 0.58 | -3.74 | 0.16 | -2.41 | 0.24 |
| Previous deforestation | 1.96 | 1.47 | 0.64 | 0.37 | - | - |
| Distance to all roads | -1.5 x 10 ⁻⁵ | 1.1 x 10 ⁻⁵ | -3.9 x 10 ⁻⁵ | 5.5 x 10 ⁻⁶ | -6.4 x 10 ⁻⁵ | 1.3 x 10 ⁻⁵ |
| Distance to type 1 roads | - | - | - | - | - | - |
| Distance to rivers | - | - | - | - | - | - |
| Elevation | -7.8 x 10 ⁻⁴ | 1.6 x 10 ⁻⁵ | -8.8 x 10 ⁻⁴ | 2.0 x 10 ⁻⁴ | -9.3 x 10 ⁻⁴ | 9.2 x 10 ⁻⁵ |
| Distance to mining | - | - | - | - | - | - |
| Slope | - | - | -0.03 | 0.01 | -0.02 | 0.008 |
| Distance to trails | - | - | - | - | - | - |
| Armed actions density | 1.6 x 10 ⁻⁵ | 2.9 x 10 ⁻⁴ | 4.1 x 10 ⁻⁴ | 3.9 x 10 ⁻⁴ | - | - |
| Distance to coca plantations | -3.0 x 10 ⁻⁵ | 3.5 x 10 ⁻⁶ | -1.2 x 10 ⁻⁵ | 3.5 x 10 ⁻⁶ | -1.3 x 10 ⁻⁵ | 1.1 x 10 ⁻⁵ |
| Population density | - | - | - | - | - | - |
| Distance to Extraction Wells | - | - | - | - | - | - |
| Protected Areas | -0.59 | 0.25 | -0.99 | 0.32 | -0.54 | 0.15 |
| African communities | -1.42 | 0.38 | -1.90 | 0.10 | -1.18 | 0.21 |
| Indigenous communities | -1.31 | 0.18 | -1.23 | 0.26 | -1.24 | 0.18 |

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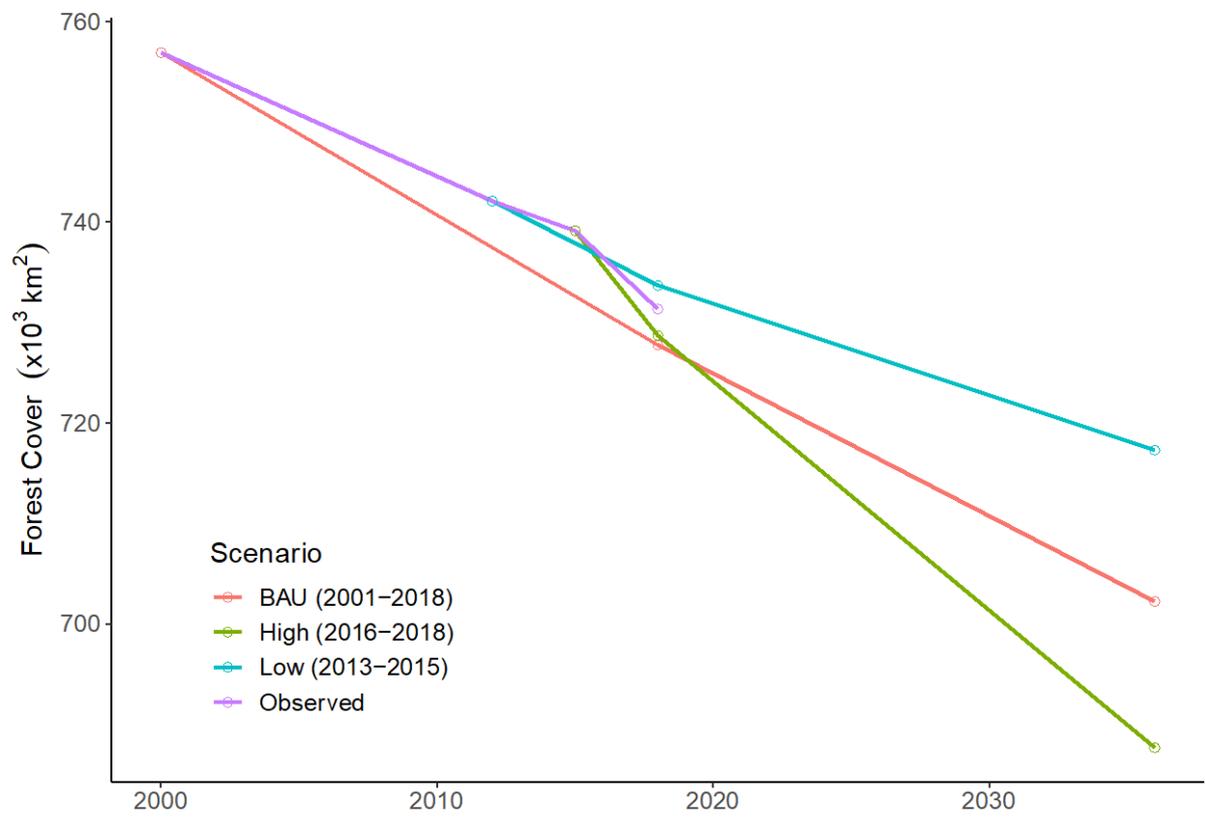
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574 **Figures**

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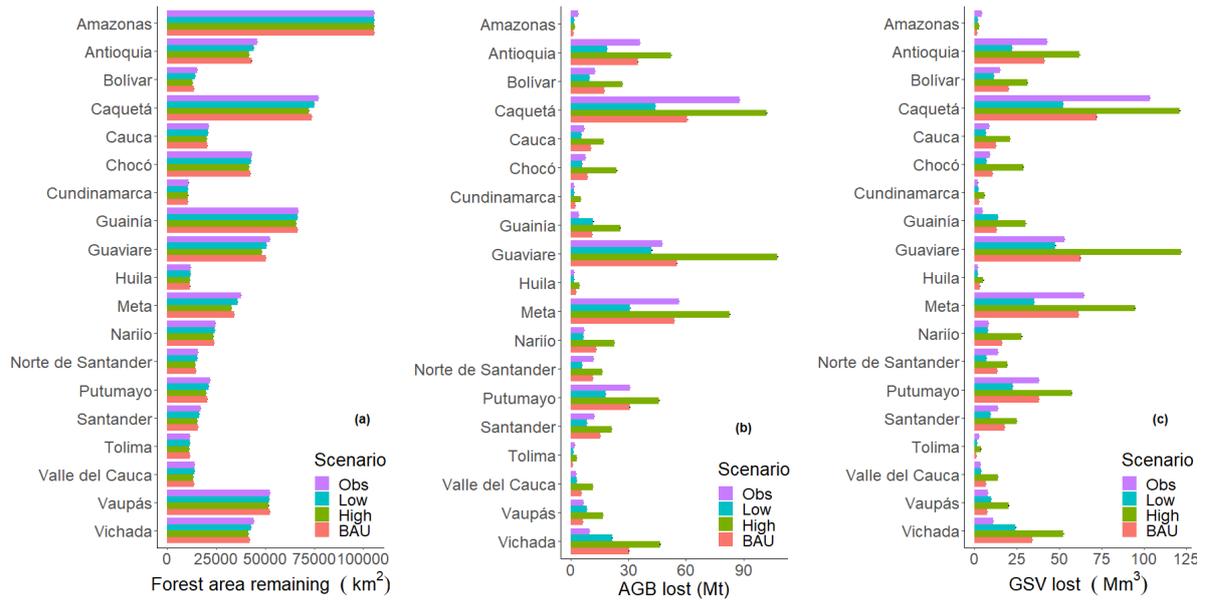
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578 **Figure 1** – Forest cover loss from 2000 through 2036 under three scenarios. Error bars
579 represent 95% confidence interval (100 iterations of the model).

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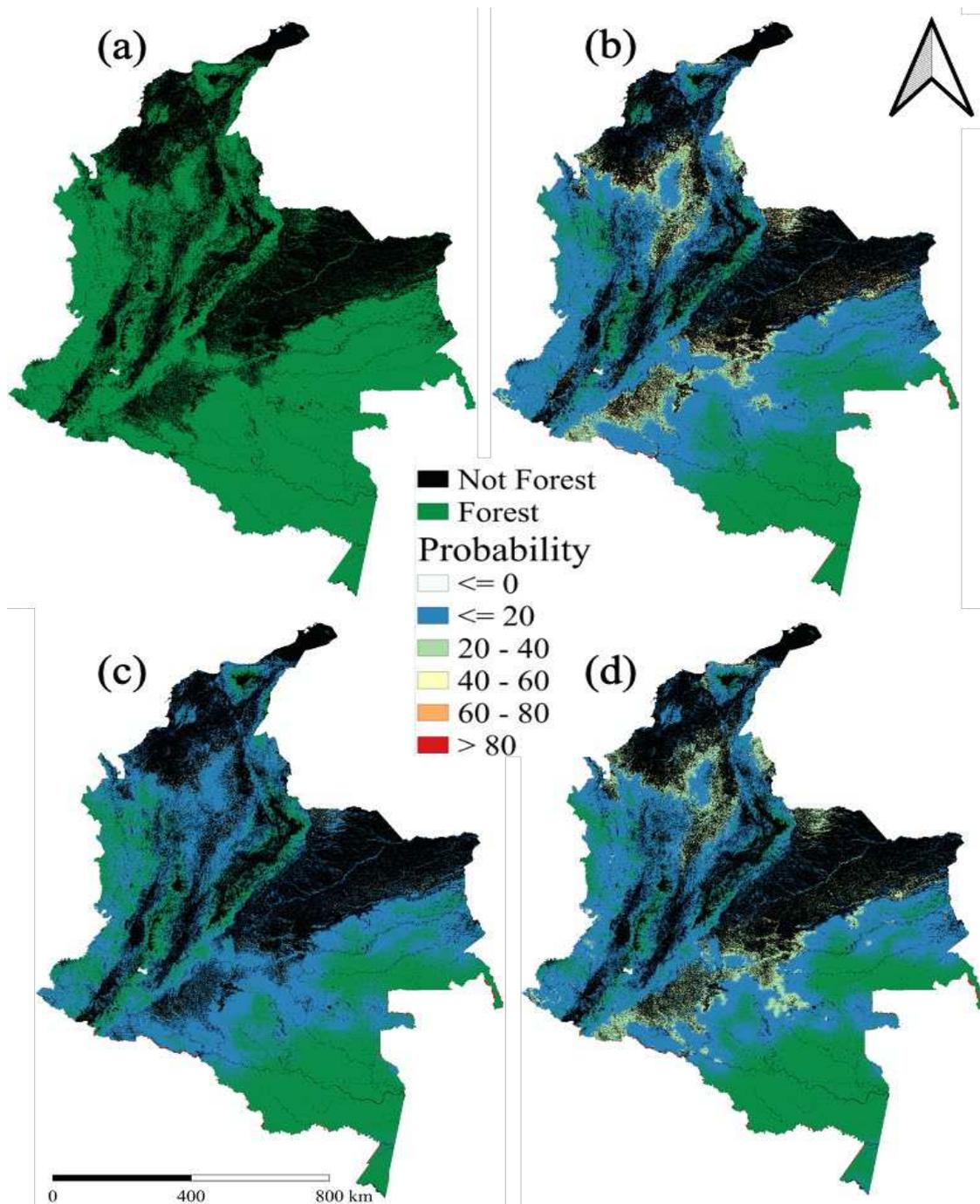
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584 **Figure 2** – (a) Forest area in 2018 (Obs) and remaining forest by 2036 under the three
 585 scenarios (BAU, Low and High), per department in Colombia. Aboveground biomass (AGB)
 586 (b) and growing stock volume (GSV) (c) lost between 2001-2018 (Obs), under the same three
 587 scenarios, per department in Colombia. Error bars represent 95% confidence interval (100
 588 iterations of the model). Showing only departments with at least 10.000 km² of forest area in
 589 2018 to improve readability (all departments in Tables S2-S4).

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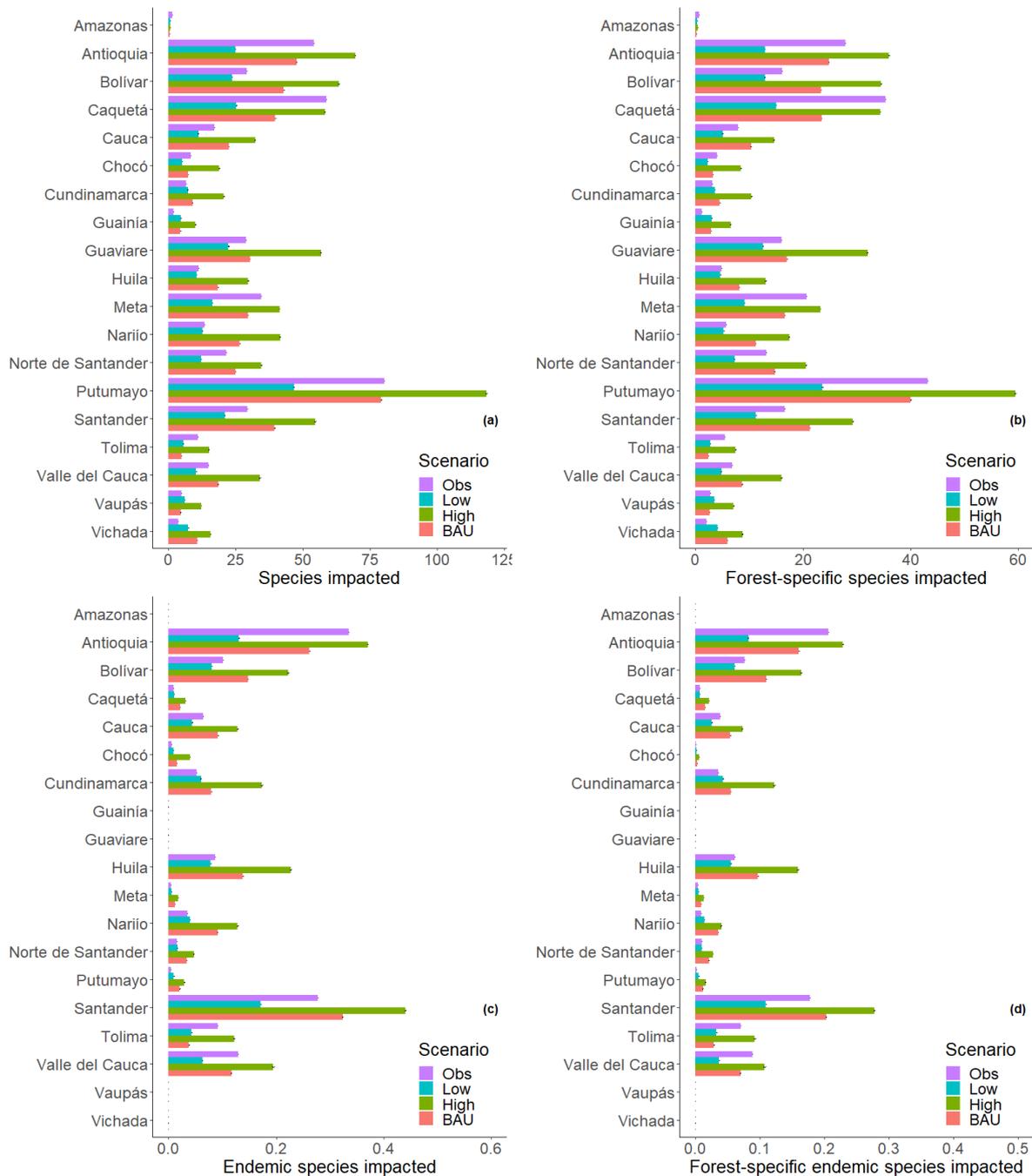
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595 **Figure 3** – Observed forest cover (a) in 2018, and projected probability of loss of forest by

596 2036 under the three scenarios (b) BAU, (c) Low, and (d) High.

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Figure 4 – Observed (2001-2018) and projected (2018-2036) average impact on the number

603

of (a) all species, (b) forest-specific species, (c) endemic species and (d) forest-specific

604

endemic species, per department in Colombia, under three different scenarios (Obs, High,

605

Low and BAU). Error bars represent 95% confidence interval (100 iterations of the model).

606

Showing only departments with at least 10.000 km² of forest area in 2018 to improve

607

readability (all departments in Tables S5-S8).

Supplementary Files

This is a list of supplementary files associated with this preprint. Click to download.

- [ColombiaSpeciesScenariosImpacts.csv](#)
- [ColombiaperDepartmentSpeciesScenariosImpacts.csv](#)
- [manuscriptSM.docx](#)