

# Biosorption and Biodegradation of Anthraquinone Based Dye (Indanthrene Blue RS) by the Immobilized Microbial Consortium in a Continuous Packed bed Bioreactor Using Corn-Cob Biochar

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## Research Article

**Keywords:** Anthraquinone vat dye, Biodegradation, Biosorption, Consortium, Corn-cob biochar, Packed bed bioreactor

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1 **Biosorption and biodegradation of anthraquinone based dye (Indanthrene Blue RS) by the immobilized**  
2 **microbial consortium in a continuous packed bed bioreactor using corn-cob biochar.**

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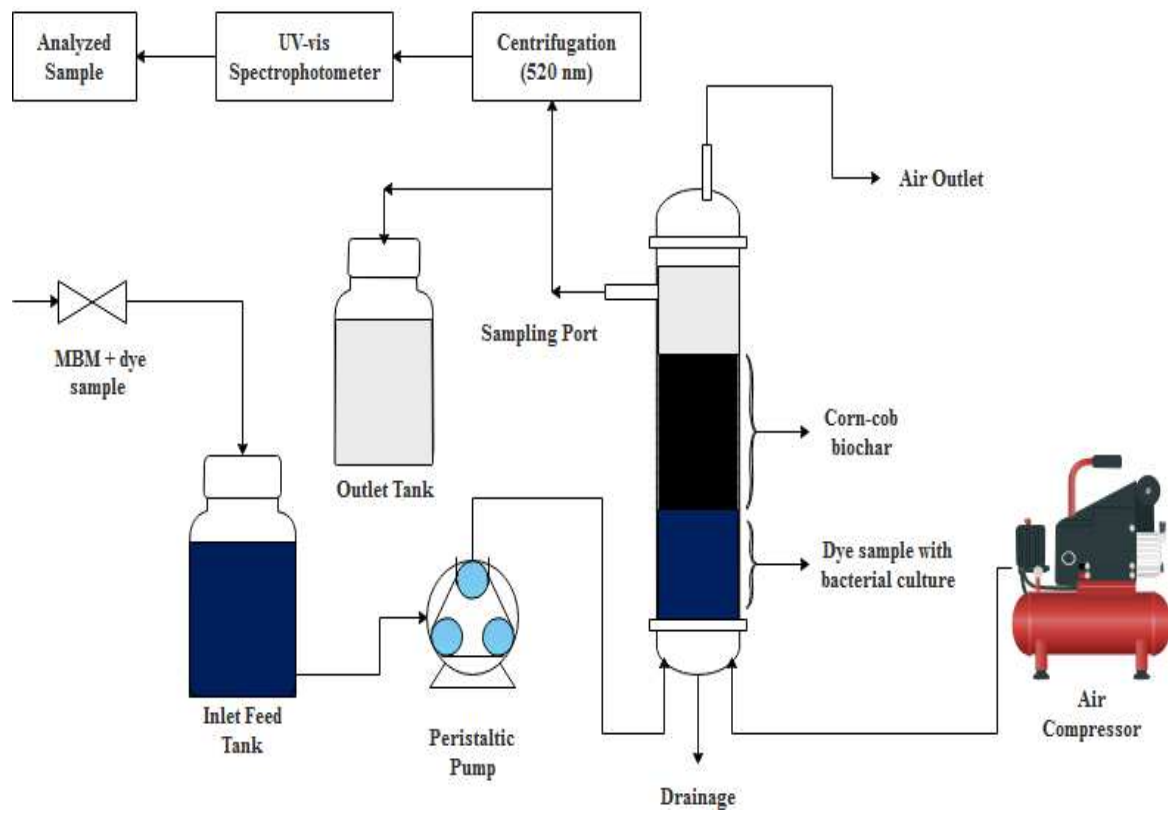
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### Graphical Abstract

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## Abstract

16 The current study describes the aerobic biodegradation of Indanthrene Blue RS dye by a microbial consortium  
17 immobilized on corn-cob biochar in a continuous up-flow packed bed bioreactor. The adsorption experiments  
18 were performed without microbes to monitor the adsorption effects on initial dye decolorization efficiency. The  
19 batch experiments were carried out to estimate the process parameters, and the optimal values of pH, temperature,  
20 and inoculum volume were identified to be 10.0, 30 °C, and  $3.0 \times 10^6$  CFU mL<sup>-1</sup>, respectively. During the  
21 continuous operation, the effect of flow rate, initial substrate concentration, inlet loading rate of Indanthrene Blue  
22 RS on the elimination capacity, and its removal efficiency in the bioreactor was studied. The continuous up-flow  
23 packed bed bioreactor was performed at different flow rates (0.25 to 1.25 L h<sup>-1</sup>) under the optimal parameters.  
24 The maximum removal efficiency of 90% was observed, with the loading rate varying between 100 to 300 mg L<sup>-1</sup>  
25 d<sup>-1</sup>. The up-flow packed bed bioreactor used for this study was extremely useful in eliminating Indanthrene Blue  
26 RS dye using both the biosorption and biodegradation process. Therefore, it is a potential treatment strategy for  
27 detoxifying textile wastewater containing anthraquinone based dyes.

28 **Keywords:** *Anthraquinone vat dye; Biodegradation; Biosorption; Consortium; Corn-cob biochar; Packed bed*  
29 *bioreactor*

## 30 1. Introduction

31 Rapid urbanization increases the demand for fast industrialization, thereby causing problems for the environment.  
32 The release of immense amounts of wastewater from the industrial sector, decrease the freshwater availability <sup>[1,2]</sup>,  
33 thus leading to pollution issues worldwide if not appropriately treated. Industries such as textile, printing, tannery,  
34 paint, plastic, cosmetics enormously used synthetic dyes because of their intense and solid colors. O'Neill et al.  
35 1999 stated the recognizable amount of colors in water to be higher than  $1 \text{ mg L}^{-1}$  <sup>[3]</sup>. Generally, 100,000 synthetic  
36 dyes and dyestuffs are utilized by dyeing industries <sup>[4]</sup>, and nearly 10-15% of the unfixed dyes are released to the  
37 water stream directly without being treated. The entry of untreated textile effluents into the water bodies alter the  
38 pH, BOD, COD, as well as reduce the sunlight penetration <sup>[5]</sup> and thereby disturb the biodiversity of that system.

39 Based on the classification of dyes, their disposal into the water stream ranges from 2% (basic dyes) to 50%  
40 (reactive dyes) of the initial dye concentration <sup>[3,6]</sup>. Among these reactive dyes, azo and anthraquinone dyes  
41 represent a significant group of synthetic dyes <sup>[7]</sup>. Their complex structure makes them highly toxic and resists  
42 degradation <sup>[8,9]</sup>. Anthraquinone dyes are among the major groups of colored pollutants highly resistant to  
43 degradation due to the presence of the fused aromatic rings, which help them to remain colored for a longer time  
44 <sup>[10,11,12]</sup>. They represent about 15% of total dyes are water-insoluble due to the existence of the chromophoric  
45 group designed by the conjugation of C=O and C=C <sup>[13,14]</sup>. Among the anthraquinone dyes, "Indanthrene Blue  
46 RS," are commonly used for cotton and silk dyeing, is primarily observed in the wastewater of textile industries.

47 Various physicochemical methodologies such as adsorption, oxidation, coagulation, precipitation, and membrane  
48 filtration <sup>[15]</sup> employed for decolorization of textile effluents are cost-intensive, less efficient and result in the  
49 production of large aggregates of secondary pollutants <sup>[16,17]</sup>. Therefore, there is a burgeoning demand to establish  
50 cost-effective methods to decolorize these polluting dyes. Biological processes using bacteria, algae, yeast, and  
51 fungi are excellent alternative methods as compared to physicochemical methods in decolorizing the textile  
52 effluents. They have less operational cost, are eco-friendly, and under optimal operating conditions produce less  
53 sludge compared to other methodologies <sup>[18,19]</sup>. This biodegradation process has turned out to be a promising  
54 method as it completely decolorizes the dye and transforms them into a non-toxic chemical form <sup>[20,21]</sup>. Under  
55 optimized, aerobic or anaerobic conditions using microorganisms, a significantly high percentage of  
56 decolorization and degradation of dyes can be achieved <sup>[22,23,24]</sup>.

57 The bioremediation of textile wastewaters is continuously expanding within the area of environmental  
58 biotechnology as it is fast and efficient. Pure bacterial strains generally are incapable of degrading the dyes

69 completely, producing carcinogenic aromatic amines as intermediates, which further needs to be decomposed <sup>[25]</sup>.  
60 It's crucial to scale up and maintain large scale pure cultures for wastewater treatment systems <sup>[26]</sup>. In the recent  
61 past, many consortiums having enhanced degradation abilities have been studied. The microbial consortium has  
62 been used over pure cultures for dye degradation as it possesses a high degree of mineralization and biodegradation  
63 <sup>[7]</sup> due to the synergistic interaction in the metabolism of the bacterial community. In a consortium, the bacterial  
64 strains metabolize the molecular structure of dye by attacking at various positions of the aromatic rings or by  
65 using metabolites formed by the dominant strains to degrade further <sup>[27,28]</sup>. Nowadays, trials on bacterial-bacterial  
66 synergism are being utilized to develop new environmentally-friendly abatement technologies to degrade the  
67 textile dye wastes without producing toxic metabolites.

68 Bioreactors are the basis of all biotransformation processes, such as the production of vaccines, enzymes,  
69 nutrients, etc., as well as biodegradation activities <sup>[29]</sup>. Various biodegradation assays of dye were performed  
70 using reactors in batch mode, but continuous reactors are recognized as being more efficient and appropriate for  
71 real-time applications <sup>[30]</sup>. Simultaneous processes of biosorption and biodegradation (BB) were established to  
72 remove high concentration dyes, involving adsorption followed by biological treatment. Using this advanced  
73 treatment strategy effectively removes the contaminants that will be converted into harmless compounds in the  
74 bioremediation process. The BB process is an economically feasible, highly effective, and environmentally  
75 friendly process, and thus has advantages in practical applications compared to other conventional techniques  
76 <sup>[31]</sup>. The efficiency of a continuous reactor can be increased under optimal conditions by using packing materials  
77 <sup>[32,33]</sup>. A continuous method for the treatment of effluent needs to be developed by taking into consideration the  
78 large-scale elimination of anthraquinone dyes, mainly from the textile industry into the waterways. Continuous  
79 mode bioreactor operations for the effluent degradation under aerobic conditions require immobilized biofilms  
80 suspended to biological systems. Biofilms can be developed by encapsulating the microorganisms into the  
81 packing material, which supports growth. This immobilization system has been reported to be better than the  
82 free cell system as it offers a high loading rate as well as inhibits biomass washout. The other advantages of  
83 using this technology include higher stability, reduced land usage, lower operational cost, and recycling of waste  
84 <sup>[34]</sup>. Therefore, it is difficult to find efficient treatment technologies for the dyeing industries that can meet the  
85 environmental sustainability requirement. Immobilized cell technology has now become an accepted method for  
86 treating dye wastewaters and thereby have achieved increasing attention. Studies on dye biodegradation have  
87 been reported to be mostly done in batch reactors.

88 Moreover, continuous reactors are effective and appropriate for real-time operation <sup>[30]</sup>. The continuous reactor

89 efficiency could be improved by optimal operating conditions in packed bed mode <sup>[32]</sup>. Over several years,  
90 activated carbon was used for wastewater treatment. It provides both a support matrix for biofilm development  
91 and a large surface area to adsorb substances passing through the reactor <sup>[35,36]</sup>.  
92 Pyrolyzed biomass, commonly referred to as biochar, is an alternative to activated carbon, which is explored as  
93 a low-cost strategy for treating wastewater. Corn-cob (CC) biochar has been identified as a suitable matrix for  
94 the immobilization of the bacterial cells in the continuous flow bioreactor. It facilitates the absorption of a large  
95 number of bacterial cells and biofilm formations, both inside and on the surface of the biochar <sup>[37]</sup>. To date, no  
96 work has been reported on the aerobic bioremediation of Indanthrene Blue RS using CC-biochar in continuous  
97 up-flow packed-bed bioreactors. In the present study, Indanthrene Blue RS dye biodegradation was performed  
98 by immobilizing cells on CC-biochar. In an up-flow packed bed bioreactor (UFPBBR), the continuous studies  
99 were performed to evaluate its effect on biodegradation using consortium-BP. Process parameters, including  
100 initial dye concentration, inoculum volume, and flow rate, were optimized for maximum degradation efficacy.  
101 The performance of the reactor efficiency was evaluated with immobilized cells operated in batch and continuous  
102 mode. The UV-vis spectrophotometer was used to measure the percentage of decolorization. Cells that were  
103 immobilized on CC-biochar continuously removed a high concentration of Indanthrene Blue RS (500 mg L<sup>-1</sup>)  
104 by using the BB process. The findings suggest that this BB process is more effective than conventional treatment  
105 approaches and has higher effectiveness.

## 106 **2. Materials and methods**

### 107 **2.1 Dyes and chemicals**

108 The textile anthraquinone vat dye, Indanthrene Blue RS, was procured from Sigma-Aldrich, India. The  
109 composition of liquid mineral-base medium (MBM) includes NaNO<sub>3</sub> (0.3%), KCl (0.05%), K<sub>2</sub>HPO<sub>4</sub> (0.1%),  
110 MgSO<sub>4</sub> (0.05%), and yeast extract (0.02%) with glucose (1%). The alkaline pH of the media was maintained by  
111 using sterilized NaOH (1M). The chemicals and media components used in these experiments were of analytical  
112 grade.

### 113 **2.2 Microorganisms and culture conditions**

114 The isolation of a pure culture of *Bacillus flexus* TS8, *Proteus mirabilis* PMS, and *Pseudomonas aeruginosa* NCH  
115 were carried out from the textile wastewater under laboratory conditions. Under aerobic conditions, the culture  
116 was grown in 250 mL flask, each containing 100 mL of nutrient broth, and was incubated at 30 °C for 24 h. The

117 consortium-BP was prepared by aseptically transferring 24 h grown culture (50 mL) of each strain into 250 mL  
118 flasks to retain the equal number of cells in both the pure culture and the consortium, respectively.

### 119 **2.3 Evaluation of Corn-cob Biochar as a packing material**

120 The corn-cob was pyrolyzed at 400 °C to prepare biochar and was used as supporting material for microbial  
121 immobilization. The biochar was sterilized before use by autoclaving for 15 min at 15 psi (121 °C) to ensure that  
122 microbes were not involved during the adsorption process. FT-IR analysis by KBr pellet technique was performed  
123 to assess the presence of functional groups in biochar before and after adsorption using the Thermofisher  
124 Scientific, Nicolet IS10 in ATR mode (with the 16-scan speed in the mid-IR range of 400–4000 cm<sup>-1</sup>). CHNS  
125 analysis was carried out to determine the composition of existing elements in biochar, such as carbon, hydrogen,  
126 nitrogen, and sulfur (Elementar Analysen Systeme, Germany/Vario EL). SEM analysis (JEOL JSM-6084LV)  
127 determines the surface morphology of the biochar. Biochar porosity, surface area, and adsorption volume were  
128 evaluated using a BET analyzer (Quanta chrome/AUTOSORB-1).

129 The CC-biochar has also been examined for the adsorption and desorption of the Indanthrene Blue RS dye packed  
130 in the column. For adsorption experiments, 2.5 g of CC-biochar was added with Indanthrene Blue RS (100 mg L<sup>-1</sup>)  
131 to 100 mL of MBM. The flasks were incubated in an isothermal shaker (120 rpm) at room temperature. The  
132 samples were withdrawn at a regular time interval (30 min) for 4 h. The supernatant was then centrifuged at 8000  
133 rpm for 10 min. The percentage of dye removal was measured spectrophotometrically at 520 nm. The desorption  
134 experiments were carried out by subsequently drying and transferring the spent CC-biochar sample into NaCl  
135 solution (100 mL) (Voudrias et al., 2002). The flasks were incubated in an isothermal shaker, and the study was  
136 carried out as specified in adsorption experiments. The final concentrations of desorbed dye were determined  
137 spectrophotometrically at 520 nm.

### 138 **2.4 Adsorption Study**

139 In Erlenmeyer flasks of 250 mL, this study was conducted by subsequently changing the initial dye concentrations  
140 (25 to 150 mg L<sup>-1</sup>), contact time (5 to 240 min), adsorbent dose (0.2 to 2 g L<sup>-1</sup>), and temperature (25 to 50 °C). A  
141 UV-vis spectrophotometer was used to measure the dye concentration in the supernatant at 520 nm. The quantity  
142 of dye adsorbed by the unit weight of CC-biochar at time *t*, *q<sub>t</sub>* (mg/g) and the percentage of removal of dye, *R* was  
143 determined as follows:

$$144 \quad q_t = \frac{(C_0 - C_t) \times V}{w} \quad (1)$$



145 
$$R\% = \frac{(C_0 - C_t)}{C_0} \times 100 \quad (2)$$

146 Where  $C_0$  and  $C_t$  ( $\text{mg L}^{-1}$ ) represents dye concentrations at the initial and any time ( $t$ , min),  $V$ ,  $q_t$ , and  $W$   
 147 represents the volume of the solution (L), the adsorbed amount ( $\text{mg g}^{-1}$ ) at any time ( $t$ , min), and the mass of  
 148 adsorbent (g), respectively.

149 **2.4.1 Adsorption Isotherms**

150 The adsorption isotherms were determined by adding  $1\text{ g L}^{-1}$  of CC-biochar in 100 mL dye solution with varying  
 151 concentrations (25, 50, 100, and  $150\text{ mg L}^{-1}$ ). pH 10.0 was maintained, and the experiment was performed for 240  
 152 min with continuous shaking at  $30\text{ }^\circ\text{C}$  to validate the equilibrium time. After 90 min incubation, the solution  
 153 reached equilibrium. The difference between the two concentrations determined the quantity of adsorbed dye ( $\text{mg}$   
 154  $\text{g}^{-1}$ ) on the surface of the adsorbent.

155 The Freundlich and Langmuir isothermic equations were used to evaluate the experimental equilibrium data for  
 156 Indanthrene Blue RS adsorption. The Langmuir isotherm model is established on the presumption that the  
 157 adsorbent has monolayer coverage on the outer surface of the adsorbent can be defined in a linearly as:

158 
$$\frac{1}{q_e} = \frac{1}{q_m} + \frac{1}{K_L q_m C_e} \quad (3)$$

159 Where  $K_L$ : adsorption energy-related Langmuir constants,  $q_m$ : maximum adsorption capacity. The linear plot of  
 160  $1/q_e$  Vs.  $1/C_e$  can be used to derive the values of  $K_L$  and  $q_m$ .

161 The Freundlich isotherm refers to the heterogeneous adsorption surfaces. The following equation represents the  
 162 linear form of Freundlich isotherm as:

163 
$$\ln q_e = \ln K_F + \frac{1}{n} \ln C_e \quad (4)$$

164 Where  $K_F$  ( $\text{mg g}^{-1}$ ): Freundlich isotherm constant related to the adsorbent's adsorption capacity,  $n$ : Freundlich  
 165 isotherm constant associated with the adsorbent affinity.

166 **2.4.2 Adsorption Kinetics**

167 The Indanthrene Blue RS adsorption on the surface of the CC-biochar was evaluated using the basic kinetic models  
168 to study the adsorption process. In this study, kinetic models of the pseudo-first-order and pseudo-second-order  
169 are evaluated to identify the best fit model for the observed data.

#### 170 **2.4.2.1 Pseudo-First-Order Kinetic Model**

171 The kinetic model of the pseudo-first-order equation stated that over time, the solute absorption rate was related  
172 to the variation in the concentration of saturation and the adsorbed quantity. In most situations, the adsorption  
173 mechanism followed by diffusion over a boundary obeys the pseudo-first-order kinetic rate equation. The linear  
174 relation between the adsorbed dye  $q_t$  (mg/g) and time  $t$  is described as:

$$175 \ln(q_e - q_t) = \ln q_e - k_1 t \quad (5)$$

176 The pseudo-first-order kinetic rate constants  $q_e$  (mg g<sup>-1</sup>) and  $k_1$  (min<sup>-1</sup>) are calculated from the intercept and the  
177 slope of the graph obtained from  $\ln(q_e - q_t)$  vs.  $t$ .

#### 178 **2.4.2.2 Pseudo-Second-Order Kinetic Model**

179 The kinetic model of the pseudo-second-order equation fits the adsorption mechanism, with chemisorptions  
180 becoming the rate-control. The pseudo-second-order model may define the linear form of adsorption kinetics as:

$$181 \frac{t}{q} = \frac{1}{K_2 q_e^2} + \frac{1}{q_e} t \quad (6)$$

182 The  $t/q_t$  vs.  $t$  plot provides a straight line with  $1/q_e$  as slope and  $1/K_2 q_e$  as intercept. The  $K_2$  value is computed  
183 from the intercept slope using  $q_e$  estimated from the slope.

### 184 **2.5 Microbial growth and degradation of Indanthrene Blue RS in immobilized cell batch experiments**

185 The microbial growth was obtained by inoculating consortium-BP in the MBM medium. The degradation studies  
186 were carried out by adding a 5% (v/v) inoculum volume to various Indanthrene Blue RS concentrations. An aliquot  
187 (2 mL) was withdrawn at a regular interval and was harvested by centrifuging at 8000 rpm for 10 min for  
188 separation of cell biomass. The decolorization efficiency was measured spectrophotometrically by observing the  
189 culture supernatant at 520 nm (Shimadzu, UV-1800). The percentage of decolorization was defined as:

$$190 \% \text{ Decolorization} = \frac{(\text{Initial absorbance} - \text{Final absorbance})}{\text{Initial Absorbance}} \times 100 \quad (7)$$

### 191 **2.6 Parameter optimization**

192 Batch studies were carried out to optimize the process parameters such as pH (6.0 to 12.0), temperature (20 to 40  
193 °C), and inoculum volume ( $1.0 \times 10^6$  to  $5.0 \times 10^6$  CFU mL<sup>-1</sup>). The parameters were varied at a fixed concentration  
194 of Indanthrene Blue RS (200 mg L<sup>-1</sup>). All the studies were performed in triplicates.

### 195 **2.7 Effect of Dye Concentration in the Immobilized Cell Batch Reactor**

196 Indanthrene Blue RS dye decolorization was performed at varying dye concentrations using immobilized  
197 consortium cells in batch bioreactors at the optimal process parameters. The CC-biochar was used as the  
198 immobilization media. Such studies were performed to determine the maximum tolerated level up to which the  
199 microbes can actively degrade the dye without inhibition of the substrate as well as other possible toxic effects.  
200 The Indanthrene Blue RS dye decolorization studies were performed at concentrations ranging from 100 to 700  
201 mg L<sup>-1</sup>.

### 202 **2.8 Up-Flow Packed Bed Bioreactor Set Up and Operation**

203 An up-flow packed bed bioreactor (PBBR) has been designed with a flat base using Perspex glass with 60 cm  
204 long and 8 cm in diameter, respectively. The PBBR consists of an inlet feed tank, peristaltic pump, air compressor,  
205 outlet tank, CC-biochar, and a reactor with a sampling port. The PBBR was operated under aerobic conditions by  
206 providing purified air in the bed. The biochar was packed across two tube grooves up to a height of 20 cm and  
207 was covered at the base by sieves (metal) of < 0.5 µm size and cotton (saturated with dye). The bioreactor was  
208 loaded with 400 g of CC-biochar with 3016 mL and 1005 mL of the total volume and working volume,  
209 respectively. The microbial culture in the liquid mineral-base medium was circulated to immobilize the bed  
210 adequately with consortium cells through the peristaltic pump. The medium was supplied to the bioreactor in an  
211 upward direction to prevent channeling effects and increased retention time. The airflow rate of 0.1 LPM was  
212 maintained and supplied by an air compressor at  $30 \pm 3$  °C (room temperature).

213 The bioreactor was subsequently operated with 500 mg L<sup>-1</sup> of Indanthrene Blue RS dye solution in continuous  
214 mode from a feed tank of 25 L, and at varying feed rate (0.25 to 1.25 L h<sup>-1</sup>) was maintained for proper distribution  
215 using a peristaltic pump. The bioreactor inoculum was 10% of the volume having  $3.0 \times 10^6$  CFU/mL. The optimal  
216 parameters achieved from the batch experiments were considered for the bioreactor operation. The Indanthrene  
217 Blue RS decolorization was performed over 20 days. The samples were withdrawn on complete decolorization,  
218 centrifuged at 7168 ×g for 10 min, and then were quantified using UV-vis spectroscopy.

### 219 **2.9 Degradation Analysis and Performance Equations**

220 After complete decolorization, the supernatant was separated by centrifuging 50 mL of the culture medium at  
221 7168 ×g for 10 min. The extent of decolorization was determined at 520 nm using a UV–vis spectrophotometer.  
222 The PBBR performance in the continuous mode was calculated under optimal conditions and at varying inlet  
223 loading rates in the form of elimination capacity (EC) and removal efficiency (RE) and defined by:

$$224 \quad \% \text{ Removal Efficiency (RE)} = \frac{C_{in} - C_{out}}{C_{in}} \quad (8)$$

$$225 \quad \text{Elimination Capacity (EC)} = Q \frac{C_{in} - C_{out}}{V} \quad (9)$$

$$226 \quad \text{Inlet Loading Rate (ILR)} = \frac{C_{in}}{V} Q \quad (10)$$

227 Where  $C_{in}$  and  $C_{out}$  represents the inlet and outlet concentrations of Indanthrene Blue RS. Q, V represents the  
228 volumetric feeding flow rate and the working volume, respectively.

### 229 3. RESULTS AND DISCUSSION

#### 230 3.1 Biochar Characterization and Analysis (FT-IR, CHNS, SEM, and BET Surface Area)

231 The efficiency of Biochar's sorption depends mainly on the functionality's chemical reactivity to the biochar  
232 surface and its porosity. Therefore, a detailed understanding of the functional groups on the surface of the biochar  
233 allows for a better interpretation of the sorption process. When the biomass is heated to 350 to 650 °C, the chemical  
234 bonds break and rearrange, forming new functional groups [38]. The FT-IR analysis of the biochar indicates the  
235 participation of the stretching vibration of C-C (quinones) and N-H (amine) at 1101.1 cm<sup>-1</sup> and 3471.2 cm<sup>-1</sup>,  
236 respectively, correspond to the N-H (amine) bending vibration at 1602.6 cm<sup>-1</sup> before biosorption [39,40]  
237 (Supplementary Fig. 1).

238 Pyrolysis may cause biosorption bands to disappear, which are characteristic of raw material, and the new band  
239 appears typical for the biochar samples. The presence of the functional group after sorption results in the stretching  
240 vibrations of =C-O (ketones, aldehydes, and esters), C-N (nitrile), C-H (alkyl), O-H (alcoholic and phenolic), and  
241 N-H (amine) at 1051.0 and 1708.6 cm<sup>-1</sup>, 1261.2 cm<sup>-1</sup>, 1386.5 cm<sup>-1</sup>, 3235.9 cm<sup>-1</sup>, and 3398.1 cm<sup>-1</sup>, respectively,  
242 correspond to the bending vibrations of –C-H (alkyl), N-H (amine) at 1386.5 cm<sup>-1</sup> and 1579.4 cm<sup>-1</sup> (Supplementary  
243 Fig. 1).

244 The spectral analysis of CHNS (after sorption) resulted in an increase in the significant components of the  
245 respective samples: Carbon (70.97%), Hydrogen (1.22%), and Nitrogen (14.96%), and thus indicating its  
246 adsorption. On increasing the pyrolysis temperature, there is an increase in the carbon content of the biochar [41,42].

247 The high carbon content of the biochar possibly suggests that it still contains a specific quantity of plant organic  
248 residues from plants (cellulose) <sup>[43]</sup>. Supplementary Table 1 illustrates the elemental analysis of corn-cob biochar  
249 before and after treatment.

250 The surface morphology of the CC-biochar sample revealed an adequate number of free pores or absorption sites  
251 on the surface using a scanning electron microscope (SEM) (Supplementary Fig. 2). All these pores represent the  
252 efficiency of dye biodegradation by microbial strain on the biochar surface. After pyrolysis, the morphology of  
253 CC-biochar was observed, and it revealed that biochar has a porous surface that facilitates bacterial cell growth  
254 and biofilm formation. SEM images captured at 20000x displayed the biofilm formation on the corn-cob biochar  
255 surface. Also, higher bacterial cell density was observed within the corn-cob biochar when thin biochar cross-  
256 sections were examined under SEM (Supplementary Fig. 3).

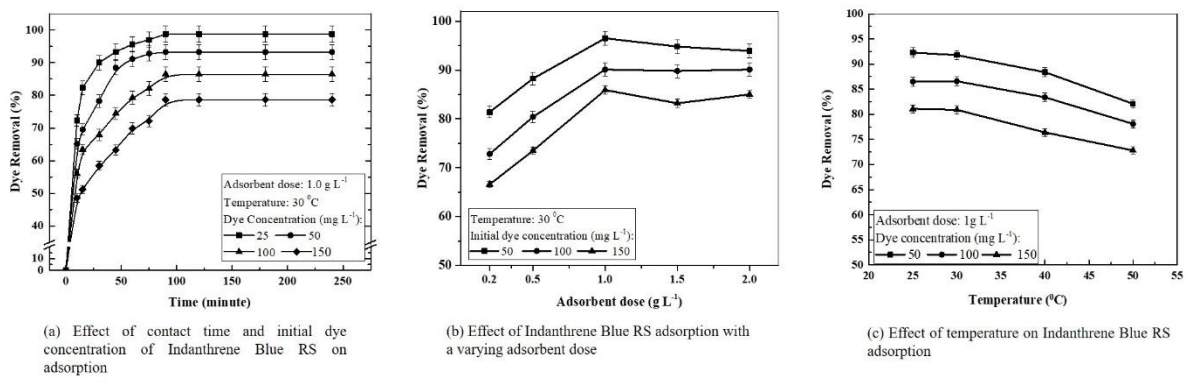
257 BET analysis concluded the surface area of 26.17 m<sup>2</sup> g<sup>-1</sup>, an average pore diameter of 5.25 nm, and a total pore  
258 volume of 0.08 cm<sup>3</sup> g<sup>-1</sup>, which was higher than that of the previous studies reported for biochar. The surface area  
259 is one of the most significant properties of biochar and depends on feedstock type <sup>[44,45]</sup>. Furthermore, the  
260 degradation of aliphatic alkyls and ester groups, and the exposure of the aromatic lignin core to higher pyrolysis  
261 temperatures, would result in an increased surface area <sup>[46]</sup>. The absorption of organic contaminants into the  
262 biochar depends on the total volumes of the micropores and mesopores. The absorption is further accelerated as  
263 the ionic radius is small, leading to the increased adsorption capacity of the biochar <sup>[47,48]</sup>. The biochar surface is  
264 generally negatively charged owing to the dissociation of functional groups containing oxygen, which causes  
265 electrostatic attraction between biochar and positively charged molecules <sup>[48,49]</sup>. Rafiq et al. 2016 reported that an  
266 increase in temperature increased the biochar surface area <sup>[50]</sup>. It is because the pore-blocking compounds are  
267 washed off or thermally damaged with rising pyrolysis temperature, thus increasing the surface area, which is  
268 readily accessible. As the temperature rises, the porosity of biochar increases because of the decomposition of  
269 lignin, the rapid production of H<sub>2</sub> and CH<sub>4</sub>, and the aromatic condensation reaction <sup>[51,52]</sup>. Supplementary Fig. 4  
270 presented the BET surface area plot for corn-cob biochar.

## 271 **3.2 Stand-alone adsorption Studies of the Packing Material in Batch Mode**

### 272 **3.2.1 Effects of Contact Time and Initial Dye Concentration, Adsorbent Dose, and Temperature**

273 Before using the corn-cob biochar as the packing material, adsorption studies were performed without microbes  
274 to evaluate the adsorption efficiency in the removal of the dye. Indanthrene Blue RS dye adsorption studies were

275 conducted at different concentrations using corn-cob biochar (Fig. 1a). At varying initial dye concentrations, an  
 276 increase in the decolorization efficiency was observed with a fixed-dose ( $1\text{ g L}^{-1}$ ) and temperature ( $30\text{ }^{\circ}\text{C}$ ). The  
 277 study showed that equilibrium was attained within 90 mins (1.5 h). Afterward, no significant changes in the degree  
 278 of adsorption were found. Indanthrene Blue RS concentration decreases significantly as the incubation time  
 279 increases. This finding suggested that the use of CC-biochar alone could partially remove Indanthrene Blue RS.  
 280 More than 80% adsorption was measured for an initial concentration of up to  $50\text{ mg L}^{-1}$ , while the dye removal  
 281 percentage was reduced to 78% and 69% for 100 and  $150\text{ mg L}^{-1}$ , respectively. In the above observation, it is  
 282 apparent that the adsorption is very fast for the lower initial dye concentration. The adsorption capacity of corn-  
 283 cob biochar for an initial dye RS concentration of 25, 50, 100, and  $150\text{ mg L}^{-1}$  was found to be 7.28, 9.81, 4.78, and  
 284  $2.22\text{ mg g}^{-1}$ , respectively. The percent removal of dye reduces with a rise in initial concentration. This result was  
 285 in agreement with that reported by Zheng et al., 2017 [31]. It needs a longer time to achieve equilibrium, and with  
 286 increased concentration of dyes, competition for active adsorption sites will increase, and the adsorption process  
 287 will slow down further [30,53]. It can be defined better by an adsorption mechanism where the dye molecules first  
 288 enter a boundary layer, then diffuse to the adsorbent surface from the boundary layer film, and lastly diffuse  
 289 through the adsorbent's porous structure [41].



290

291 **Fig. 1.** Effects of Contact Time and Initial Dye Concentration, Adsorbent Dose, and Temperature on  
 292 Indanthrene Blue RS adsorption.

293 Fig. 1b shows the effect of adsorbent dose (ranging from 0.2 to  $2.0\text{ g L}^{-1}$ ) on the dye absorption capacity at three  
 294 varying initial dye concentrations 50, 100, and  $150\text{ mg L}^{-1}$ . At first, the significant rise in adsorption with the  
 295 increased dose can be due to the increased surface area and availability of more adsorption sites. After the critical  
 296 dose of  $1\text{ g L}^{-1}$ , the adsorption rate is slowing increasingly. The percentage of removal of dye is increased as there  
 297 is rapid superficial adsorption on the adsorbent surface at a higher adsorbent dose, resulting in a lower

308 concentration of the solute in the solution than when the adsorbent dose is low. Similarly, as the initial dye  
309 concentration for constant adsorbent dose increases, the absorption of dye increases due to the existence of more  
300 dye molecules in the solution <sup>[54]</sup>.

301 The effect of temperature on the adsorption of Indanthrene Blue RS is presented in Fig. 1c. The studies were  
302 performed at three varying concentrations of dye (50, 100, and 150 mg L<sup>-1</sup>) and four separate temperatures (25,  
303 30, 40, and 50 °C) utilizing 1 g L<sup>-1</sup> of adsorbent dose. It was observed that as the temperature rises from 25 to 30  
304 °C, there was an increase in the adsorption rate. When the temperature rose from 35 to 50 °C, no significant  
305 difference in the removal percentage was observed <sup>[55]</sup>. Hence, the effect of temperature on Indanthrene Blue RS  
306 decolorization is negligible.

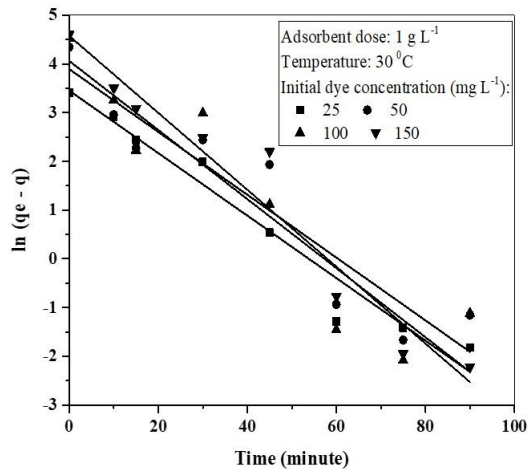
307 The above results, therefore, suggest that the use of corn-cob biochar as an adsorbent is not an effective method  
308 since the adsorbent will soon become saturated due to the low adsorbent capacity.

### 309 3.2.2 Adsorption Kinetics and Isotherms

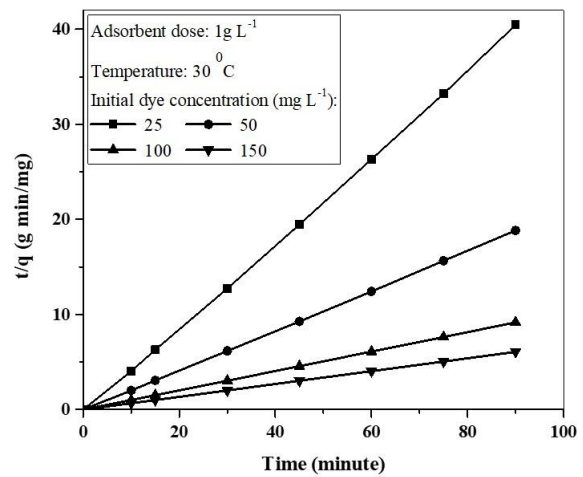
310 The kinetic parameters were evaluated by using linear plots of pseudo-first-order, pseudo-second-order kinetic  
311 models (Fig. 2a, Fig. 2b). The parameters were determined, as presented in Table 1. The plot of  $\ln(q_e - q_t)$  vs. t  
312 will present a straight line with  $-k_1$  and  $\ln q_e$  as slope and intercept, allowing to evaluate the adsorption rate  
313 constant  $k_1$  and equilibrium adsorption capacity  $q_{ecal}$  (Fig. 2a). It was found that the experimental data point  
314 does not fit a straight line and the values determined for  $k_1$  and  $q_{ecal}$  are given in Table 1. Hence, it can be  
315 inferred from the findings that the kinetics of Indanthrene Blue RS adsorption on corn-cob biochar is not likely to  
316 follow the pseudo-first-order kinetic model and thus not a diffusion-controlled phenomenon.

317 The  $t/q_t$  vs. t plot gives a straight line with  $1/q_e$ , and  $1/(k_2 q_e^2)$  as slope and intercept are presented in Fig. 2b.

318 The  $k_2$  value is estimated from the intercept using  $q_{ecal}$  value determined from the slope. Table 1 shows the  
319 estimated value of  $k_2$ ,  $q_{ecal}$ , and their respective regression coefficient ( $R^2$ ) values. The  $R^2$  value is 1.00 for 25  
320 and 50 mg L<sup>-1</sup> and almost unity (0.99) for 100 and 150 mg L<sup>-1</sup> of Indanthrene Blue RS, indicating that the  
321 Indanthrene Blue RS adsorption kinetics follows a pseudo-second-order kinetic model. It may also be observed  
322 from Table 1 that  $q_{ecal}$  values are very close to  $q_{exp}$  values that were obtained experimentally. Therefore, it  
323 can be stated that the pseudo-second-order kinetic model can describe the Indanthrene Blue RS adsorption on CC-  
324 biochar better than the pseudo-first-order kinetic model, and the mechanism is regulated by chemisorption. A  
325 similar result was observed by Zheng et al., 2017 for the adsorption of AO10 molecules onto MHSA-AC <sup>[31]</sup>.



(a) Pseudo-first-order kinetic model for Indanthrene Blue RS adsorption on corn-cob biochar



(b) Pseudo-second-order kinetic model for Indanthrene Blue RS adsorption on corn-cob biochar

326

327

**Fig. 2.** Kinetics for Indanthrene Blue RS adsorption.

328

329

**Table 1** Kinetic model parameters at 30 °C

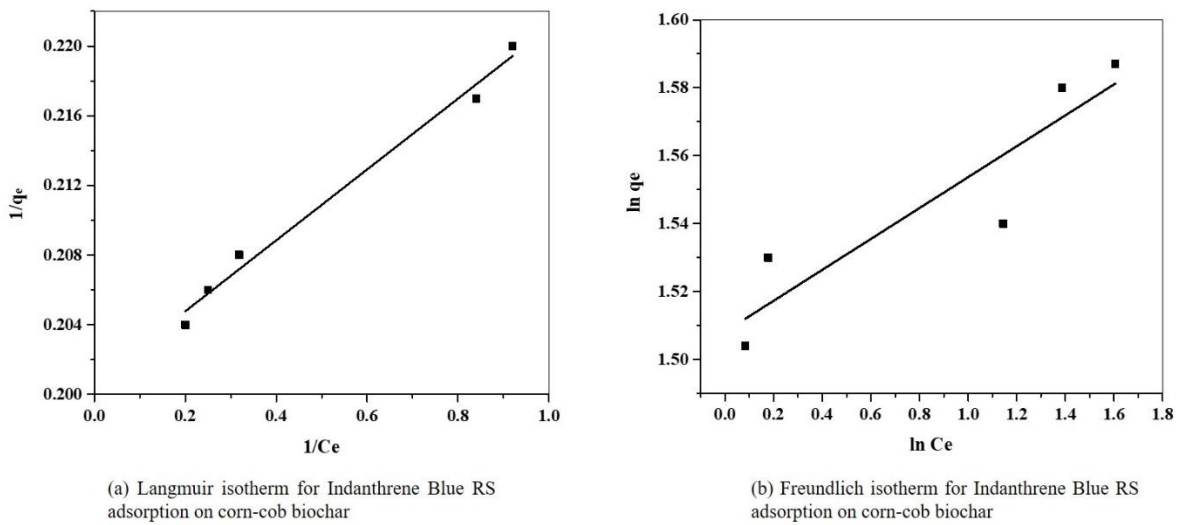
Indanthrene Blue RS concentration (mg L <sup>-1</sup> )	Pseudo-first order				Pseudo-second order		
	q <sub>e, exp</sub> (mg g <sup>-1</sup> )	q <sub>e, cal</sub> (mg g <sup>-1</sup> )	k <sub>1</sub> (min <sup>-1</sup> )	R <sup>2</sup>	q <sub>e, cal</sub> (mg g <sup>-1</sup> )	k <sub>2</sub> (g mg <sup>-1</sup> min)	R <sup>2</sup>
25	7.28	4.68	0.08	0.95	7.41	0.67	1.00
50	9.81	5.79	0.07	0.87	9.81	0.69	1.00
100	4.78	1.48	0.06	0.88	4.78	0.55	0.99
150	2.22	0.14	0.06	0.96	2.22	0.45	0.99

330 The experimental equilibrium data were plotted (Fig. 3a, Fig. 3b) by using the Langmuir and Freundlich isotherm

331 model at 30°C. The coefficient of correlation for Langmuir and Freundlich adsorption isotherm is determined by



332 applying the experimental adsorption equilibrium results and are given in Table 2. The Langmuir isotherm  
 333 correlation coefficient ( $R^2 = 0.98$ ) is higher than the Freundlich isotherm obtained value ( $R^2 = 0.84$ ).  
 334 As the dye concentration increases, the adsorption capacity increases, suggesting that the available binding sites  
 335 were saturated immediately as more dye was present in the aqueous solution <sup>[56]</sup>. In an aqueous solution, the dye  
 336 molecule and the corn-cob biochar exist as cation and anion, respectively. Consequently, adsorption is considered  
 337 to be the chemisorption in nature and describes that the adsorption is taking place on strongly homogeneous  
 338 surfaces.



339

340

**Fig. 3.** Isotherms for Indanthrene Blue RS adsorption.

341

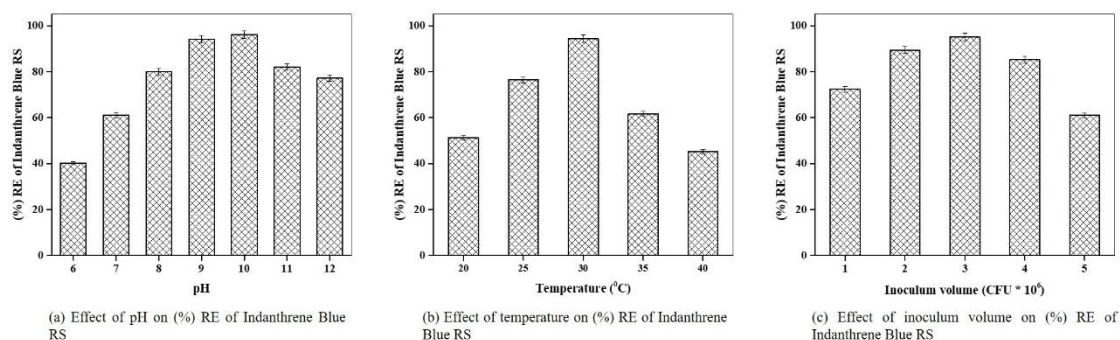
**Table 2** Isotherm constants for Indanthrene Blue RS adsorption onto corn-cob biochar.

Adsorbent	Langmuir constants				Freundlich constants		
	$Q_{0, cal}$ (mg g <sup>-1</sup> )	$Q_{0, exp}$ (mg g <sup>-1</sup> )	$K_L$	$R^2$	$K_f$	$n$	$R^2$
Corn-cob biochar	4.55	4.98	4.92	0.98	4.52	21.97	0.84

342 The equilibrium data for the Indanthrene Blue RS adsorption onto CC-biochar are best fitted with the Langmuir  
 343 isotherm model, indicating that the monolayer coverage of the dye molecule occurs over the adsorbent surface. In  
 344 the Freundlich isotherm model, the value for  $n$  greater than 1 represents the favorable adsorption, and the adsorbent  
 345 is efficient for the entire range of dye concentrations [57]. Therefore, the Langmuir isotherm model would reflect  
 346 the adsorption mechanism better physically than the Freundlich isotherm model (Table 2).

### 347 3.3 Batch Biodegradation

348 Batch experiments were carried out by varying the optimum parameters such as pH, temperature, and inoculum  
 349 volume to evaluate the optimal values for the Indanthrene Blue RS decolorization. It was observed that the  
 350 removal percentage of Indanthrene Blue RS increases rapidly with a pH increases from 9.0 to 10.0 with an  
 351 optimum pH as 10.0 at which the RE of 96.2% was found (Fig. 4a). The temperature range of 20 to 40 °C was  
 352 evaluated. It was found that the RE increased at 30 °C to 94.3% and then decreased drastically (Fig. 4b) similar  
 353 to the findings of Das et al [58]. As the inoculum volume increased from  $1.0 \times 10^6$  to  $5.0 \times 10^6$  CFU mL<sup>-1</sup>, the RE  
 354 increased rapidly to 95.2% at an inoculum volume of  $3.0 \times 10^6$  CFU mL<sup>-1</sup> (Fig. 4c). The optimal value of pH,  
 355 temperature, and inoculum volume were found to be 10.0, 30 °C, and  $3.0 \times 10^6$  CFU mL<sup>-1</sup>, respectively.



356 (a) Effect of pH on (%) RE of Indanthrene Blue RS (b) Effect of temperature on (%) RE of Indanthrene Blue RS (c) Effect of inoculum volume on (%) RE of Indanthrene Blue RS

357 **Fig. 4.** Parameter optimization in Up-flow packed bed bioreactor.

### 358 3.4 Biodecolorization of Indanthrene Blue RS in Continuous Up-Flow Packed Bed Bioreactor

359 UFPBBR was performed within the concentration range of 100 to 500 mg L<sup>-1</sup> to assess the impact of the initial  
 360 dye concentration on the percentage of removal using optimal parameters achieved from batch experiments. The  
 361 percentage removal of Indanthrene Blue RS at an initial dye concentration of 500 mg L<sup>-1</sup> increases by about 90%.  
 362 It decreases with a subsequent rise in dye concentration owing to the inhibitory effects of substrates.  
 363 Supplementary Fig. 5 illustrates the effect of different initial Indanthrene Blue RS concentration on the adsorption

364 capacity of CC-Biochar and bacteria-immobilized CC-Biochar. It was, therefore, evident that the immobilization  
365 significantly improved the ability of the microbial cells to degrade the dye effectively.

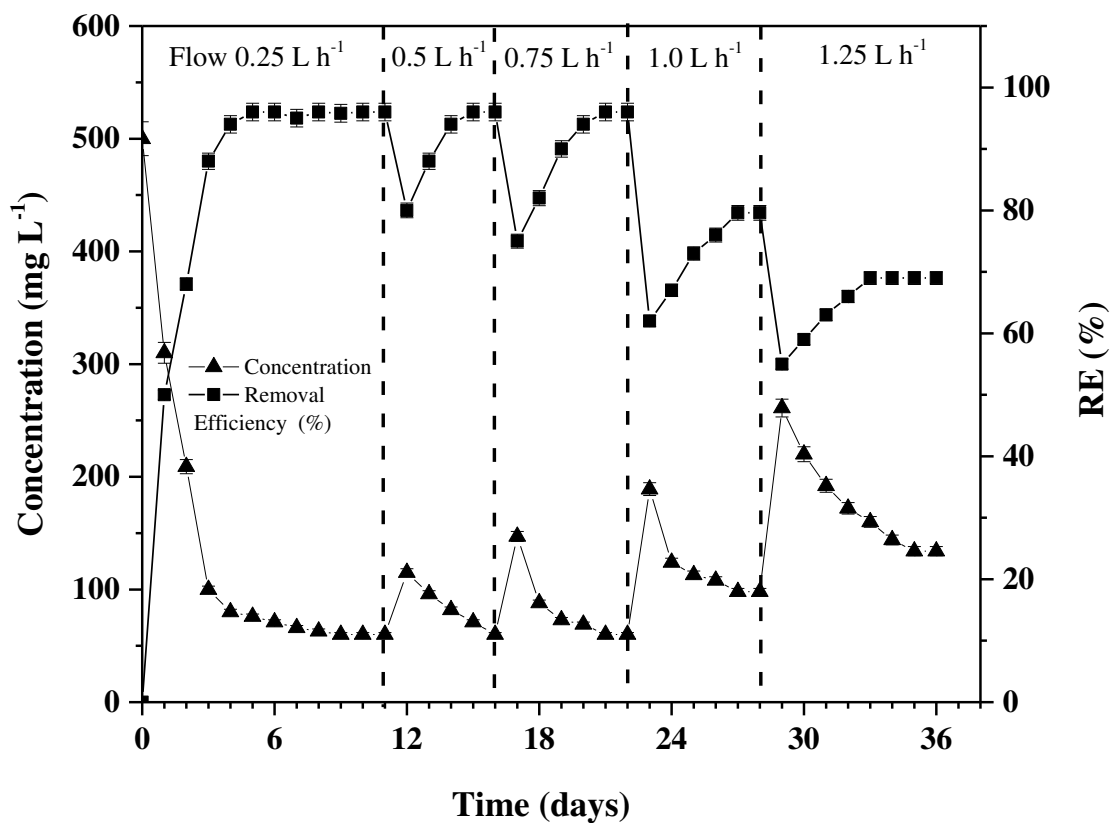
366 The increased surface area of CC-biochar enabled the interactions between both the dye molecules and bacterial  
367 cells at higher concentrations to achieve a higher decolorization percentage. The dye molecules were adsorbed by  
368 the pores and the active sites found on the surface of the CC-biochar and thus established active target sites for  
369 the immobilization of the bacterial cells. The active sites on the CC-biochar transform the dye molecules into  
370 environmentally friendly by-products with the help of the bacteria that are immobilized on the biochar and thus  
371 renew the porous structure further for adsorption <sup>[59,60]</sup>. A synergistic effect was established by the subsequent  
372 degradation of Indanthrene Blue RS dye by the bacterial cells together with active sites on the surface of the  
373 biochar that attributed to the significant dye degradation at higher concentrations. Similar findings have been  
374 documented in earlier studies showing the effectiveness of immobilized cell systems for dye degradation <sup>[30,61]</sup>.

#### 375 **3.4.1 Effect of Flow Rate on Removal Efficiency**

376 The UFPBBR efficiency was evaluated in continuous mode (CUFPBBR) using optimal parameters acquired from  
377 batch experiments by differing the feed flow rate (0.25, 0.5, 0.75, 1.0, and 1.25 L h<sup>-1</sup>) and Indanthrene Blue RS  
378 concentration in the range as in UFPBBR. In a continuous process, the flow rate determines the dyes  
379 decolorization within the effluent. Initially, the bioreactor was maintained at the feed flow rate of 0.25 L h<sup>-1</sup> to  
380 enable adequate bacterial growth, with glucose as a carbon source, and to maintain a steady-state condition. On  
381 the 6<sup>th</sup> day of operation, the steady-state was established that is apparent with the almost constant removal  
382 efficiency (95%) (Fig. 5). It was observed that about 6 days are required for the acclimation of consortium-BP for  
383 effective decolorization of Indanthrene Blue RS.

384 Moreover, a significant reduction in decolorization was reported at high flow rates. The flow rate was increased  
385 on the 11<sup>th</sup> day to 0.50 L h<sup>-1</sup>. On the 12<sup>th</sup> day, a rapid decrease in RE was found the following, which the RE  
386 amended and then became stable at 95% on the 15<sup>th</sup> day. The flow rates were improved to 0.75 L h<sup>-1</sup> on the 16<sup>th</sup>  
387 day. The bioreactor performance was similar, which is a significant decline led by a 95% recovery in removal  
388 efficiency and stabilization. On the 22<sup>nd</sup> day, the flow rate increased to 1.0 L h<sup>-1</sup>, and after a sudden decline, the  
389 RE stabilized at 79.6%. The steady-state RE value dropped significantly at a flow rate of 1.25 L h<sup>-1</sup> and maintained  
390 at about 69%. The decline in RE value was sharper above the flow rate of 0.75 L h<sup>-1</sup>, indicating a change in the  
391 bioreactor's control mechanism <sup>[53]</sup>. The interaction of the dye molecule with cells immobilized on CC-biochar  
392 requires longer retention time. Chen et al. 2005 reported that the dye molecules required more retention time to

393 interact with the bacteria cell than the time necessary for the intraparticle diffusion <sup>[62]</sup>. Though biosorption took  
 394 place during the process, and there is a possibility that the consortia cells used the adsorbed dye onto CC-biochar  
 395 for degradation. It was observed that nearly 6% to 8% of the decolorization was due to the biosorption of the dye  
 396 onto the biochar, with most of the decolorization attributed to the dye utilization by the bacterial consortia <sup>[63]</sup>.  
 397 Biodegradation is attributed to the dye degradation process in the present work since the adsorption process only  
 398 occurs during the initial stage of reactor operation. The continuous BB process can be divided into two phases. In  
 399 the first phase, the concentration of Indanthrene Blue RS gradually decreases, suggesting that biosorption on CC-  
 400 biochar plays a leading role and that much of the Indanthrene Blue RS molecules are adsorbed. The high  
 401 adsorption capacity of CC-Biochar, thus eliminates the toxic risk to the immobilized bacterial cells. CC-biochar  
 402 reaches the highest adsorption value at equilibrium in the second phase, and biodegradation occurs considerably  
 403 until the Indanthrene Blue RS is completely eliminated. Throughout the entire continuous BB process, the  
 404 continuous removal of Indanthrene Blue RS significantly differs from the usual CC-Biochar adsorption process.



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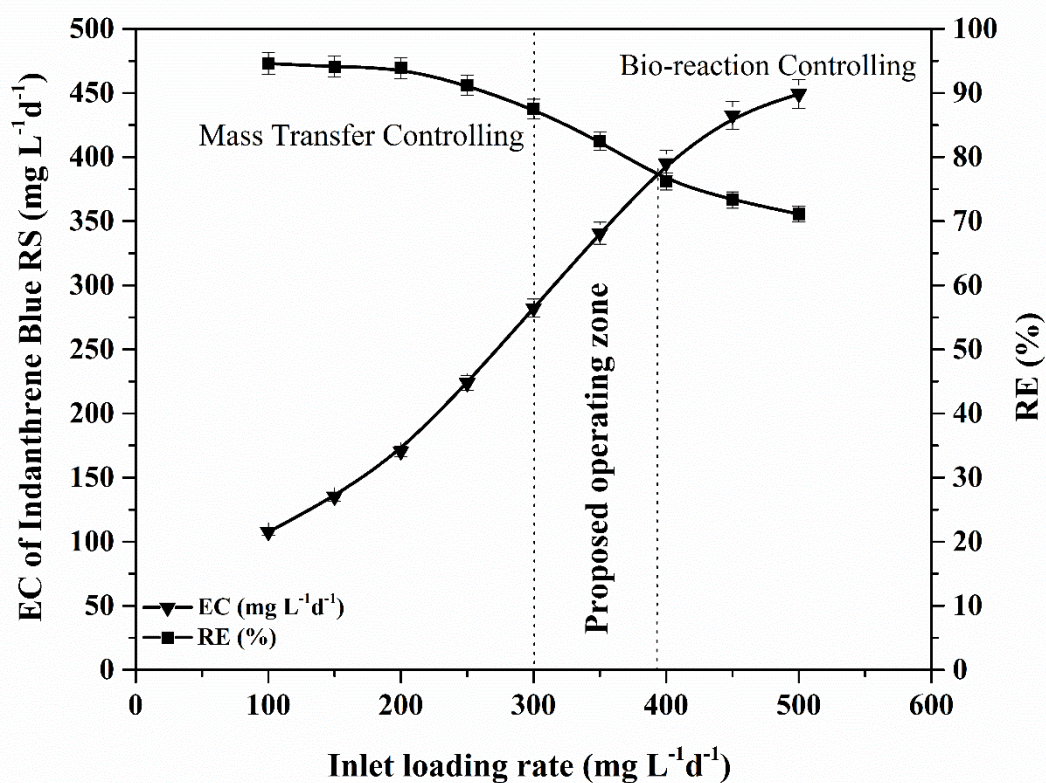
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Fig. 5. Bioreactor performance with changing inlet Indanthrene Blue RS feed flow rate.

### 3.4.2 Effect of Indanthrene Blue RS Inlet Loading Rate on Elimination Capacity and Removal Efficiency

409 The difference in the removal efficiency and elimination capacity of the Indanthrene Blue RS at varying inlet  
 410 loading rate is illustrated in Fig. 7. The Indanthrene Blue RS inlet loading plot displays two separate regions for  
 411 the mass transfer zone and bio-reaction zone. With an increment in the inlet loading rate of Indanthrene Blue RS  
 412 from 100 to 500 mg L<sup>-1</sup> d<sup>-1</sup>, the RE was found to continuously increase above 90% to the loading rate of 300 mg  
 413 L<sup>-1</sup> d<sup>-1</sup> after which it continually decreased. EC linearly increased with an increasing inlet loading rate of  
 414 Indanthrene Blue RS and reached its highest value of 446.1 mg L<sup>-1</sup> d<sup>-1</sup> at the loading rate of 500 mg L<sup>-1</sup> d<sup>-1</sup>.  
 415 Similarly, the EC keeps increasing with an increase in Indanthrene Blue RS loading, which is evident from a slight  
 416 variation in the EC vs. Inlet loading rate plot (Fig. 7). Such results thus confirm the difference in the reactor  
 417 controlling mechanism throughout the entire biodegradation process [53,64,65].



418

419 **Fig. 6.** Effect of inlet Indanthrene Blue RS load on elimination capacity and removal efficiency.

420 The diffusional flux through the biofilm would be reduced at low loading rates, resulted in mass transfer  
 421 limitations. The inner part of the biofilm at the packing media surface remains substrate deficiency and is never  
 422 used by microbes for effective biodegradation. Whereas high loading rates resulted in higher diffusional flux,  
 423 thereby changing the process between the mass transfer zone to the bio-reaction controlling zone. The contaminant  
 424 rapidly approaches the innermost layers of biofilm, and microbes use the adequate substrate available through the

425 biodegradation control mechanism. At the initial loading rate of  $300 \text{ mg L}^{-1} \text{ d}^{-1}$ , it was observed that the process  
426 moved from the mass-transfer zone to the bio-reaction controlling region (Fig. 7).

427 A rise in the inlet loading rate from  $100$  to  $300 \text{ mg L}^{-1} \text{ d}^{-1}$  resulted in increased use of the biofilm corresponding  
428 to nearly constant RE with a significant increase in EC value. The substrate inhibition with higher concentrations  
429 of Indanthrene Blue RS may lead to decreased RE [53]. The operation of bioreactors with an appropriate value of  
430 RE in the bio-reaction controlling region is always satisfactory. The optimal operating range for this study lies  
431 around  $300$  to  $395.4 \text{ mg L}^{-1} \text{ d}^{-1}$ . The inlet loading rate where the process in the bioreactor moves from a mass  
432 transfer zone to the bio-reaction controlling zone and the RE and EC plot intersection value can be used for  
433 industrial applications as an estimated calculation of the bioreactor's operating concentration range (Fig. 6).  
434 Therefore, the continuous immobilized UFPBBR showed promising results that can be used for practical  
435 applications.

#### 436 **Conclusion**

437 The UFPBBR immobilized with CC-biochar significantly increased the ability of the bacterial cells to degrade  
438 the dye effectively. The increased surface area of CC-biochar enabled the interactions in between the dye  
439 molecules and bacterial cells to achieve a higher decolorization percentage at higher concentrations. The  
440 adsorption kinetics showed that the Indanthrene Blue RS adsorption on CC-biochar followed the pseudo-second-  
441 order kinetic model. Similarly, the adsorption of Indanthrene Blue RS on CC-biochar followed the Langmuir  
442 isotherm model with the maximum adsorption capacity of  $4.55 \text{ mg g}^{-1}$  at  $30 \text{ }^\circ\text{C}$ . An increase in the inlet loading  
443 rate of Indanthrene Blue RS from  $100$  to  $500 \text{ mg L}^{-1} \text{ d}^{-1}$ , the RE is continually increasing above  $90\%$  to the loading  
444 rate of  $300 \text{ mg L}^{-1} \text{ d}^{-1}$ , after which; it begins to decrease. The EC value continued to increase linearly with  
445 Indanthrene Blue RS inlet loading rate and reached the highest value of  $446.1 \text{ mg L}^{-1} \text{ d}^{-1}$  with RE of  $75.2\%$  at  $500$   
446  $\text{mg L}^{-1} \text{ d}^{-1}$  loading rate. The present study combined the biosorption and biodegradation processes to develop an  
447 efficient treatment strategy. The above results suggest that the developed consortium-BP and the continuous  
448 immobilized UFPBBR have shown promising results that combine the advantages of both biosorption and  
449 biodegradation processes. The UFPBBR can be used for practical applications at industrial scale in decolorizing  
450 and simultaneously reducing the toxicity of textile effluents containing anthraquinone dyes.

#### 451 **Ethical Approval**

452 'Not Applicable'

453 **Consent to Participate**

454 'Not Applicable'

455 **Consent to Publish**

456 'Not Applicable'

457 **Competing Interest**

458 The authors declare that they have no competing interests.

459 **Authors Contributions**

460 All authors contributed to the study conceptualization and design. Material preparation, data collection and  
461 analysis were performed by Swati Sambita Mohanty. The first draft of the manuscript was written by Swati  
462 Sambita Mohanty and review and edited by Arvind Kumar. The study was carried out under the supervision of  
463 Arvind Kumar. All authors read and approved the final version of the manuscript.

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468 **Availability of data and materials**

469 Data are however available from the authors upon reasonable request and with permission.

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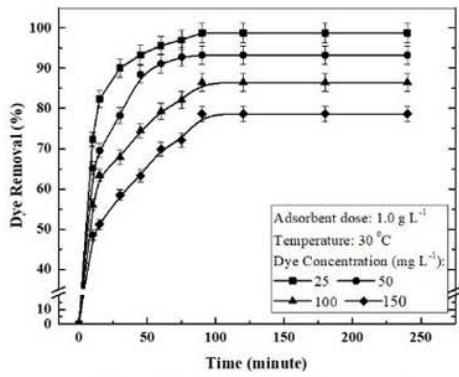
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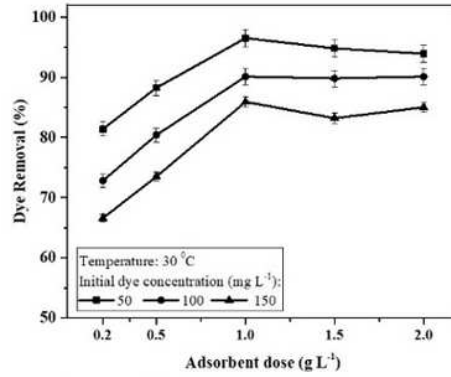
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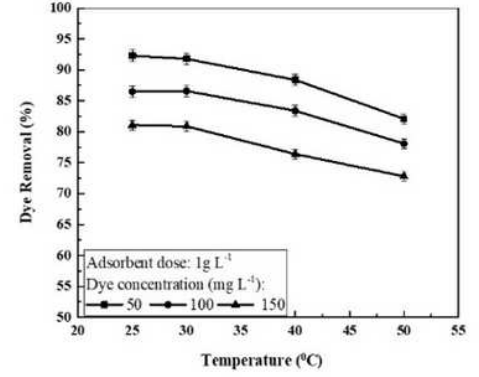
# Figures



(a) Effect of contact time and initial dye concentration of Indanthrene Blue RS on adsorption



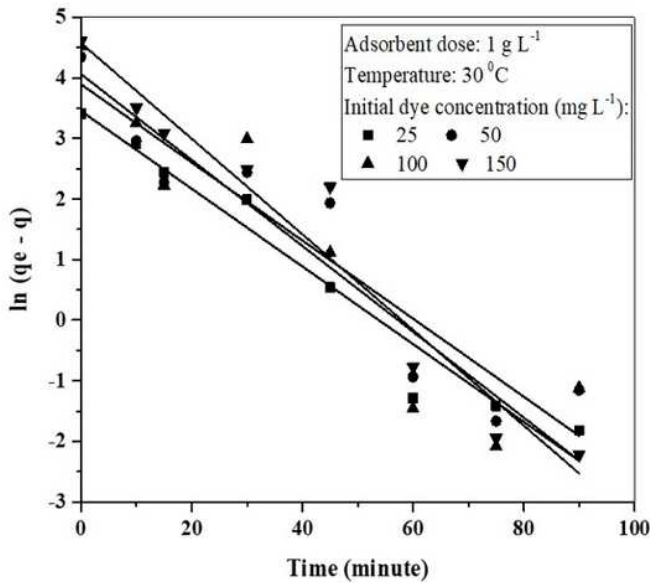
(b) Effect of Indanthrene Blue RS adsorption with a varying adsorbent dose



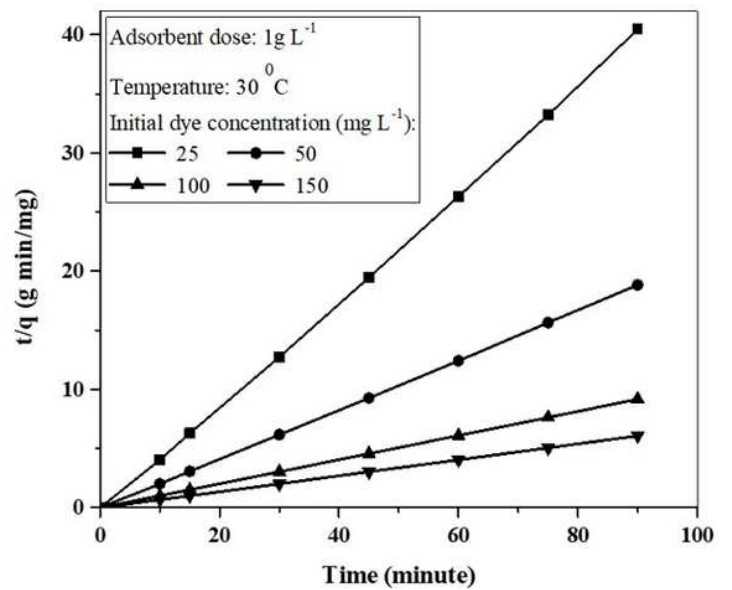
(c) Effect of temperature on Indanthrene Blue RS adsorption

## Figure 1

Effects of Contact Time and Initial Dye Concentration, Adsorbent Dose, and Temperature on Indanthrene Blue RS adsorption.



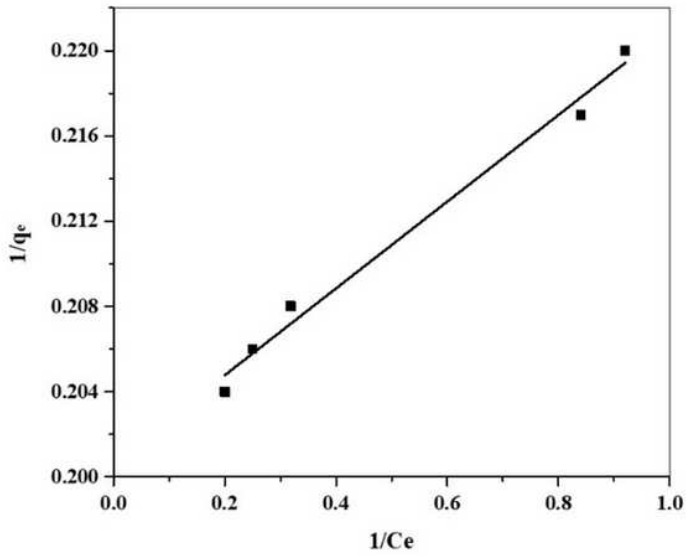
(a) Pseudo-first-order kinetic model for Indanthrene Blue RS adsorption on corn-cob biochar



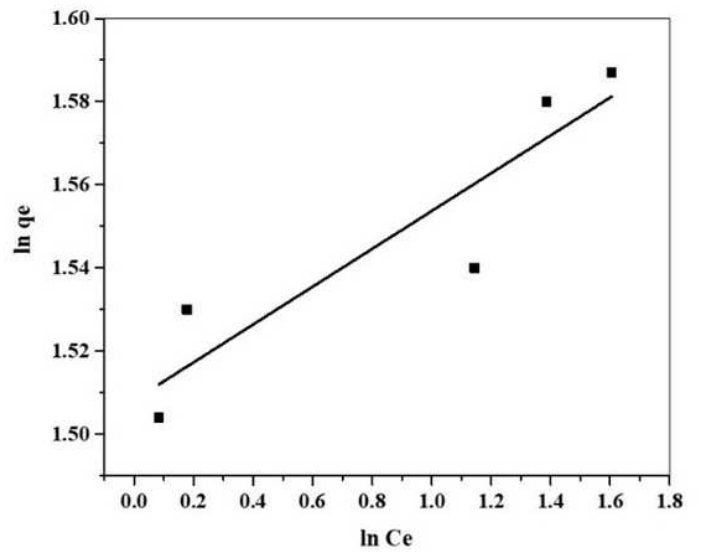
(b) Pseudo-second-order kinetic model for Indanthrene Blue RS adsorption on corn-cob biochar

## Figure 2

Kinetics for Indanthrene Blue RS adsorption.



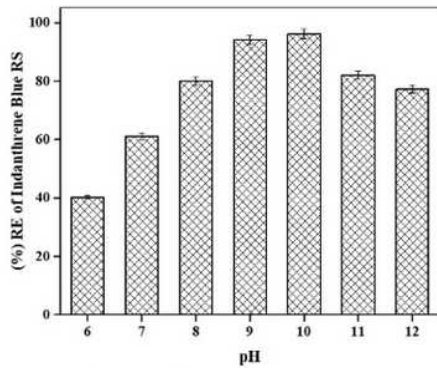
(a) Langmuir isotherm for Indanthrene Blue RS adsorption on corn-cob biochar



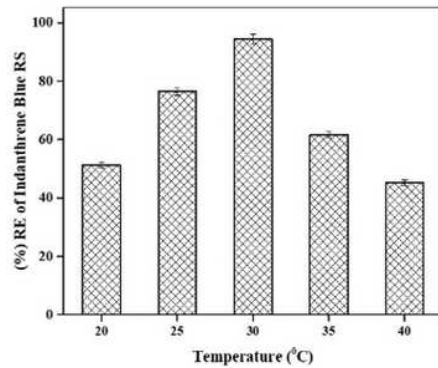
(b) Freundlich isotherm for Indanthrene Blue RS adsorption on corn-cob biochar

### Figure 3

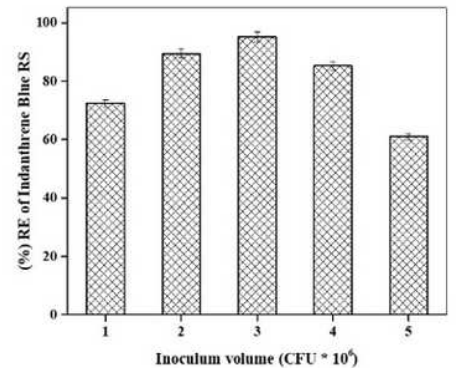
Isotherms for Indanthrene Blue RS adsorption.



(a) Effect of pH on (% RE) RE of Indanthrene Blue RS



(b) Effect of temperature on (% RE) RE of Indanthrene Blue RS



(c) Effect of inoculum volume on (% RE) RE of Indanthrene Blue RS

### Figure 4

Parameter optimization in Up-flow packed bed bioreactor.

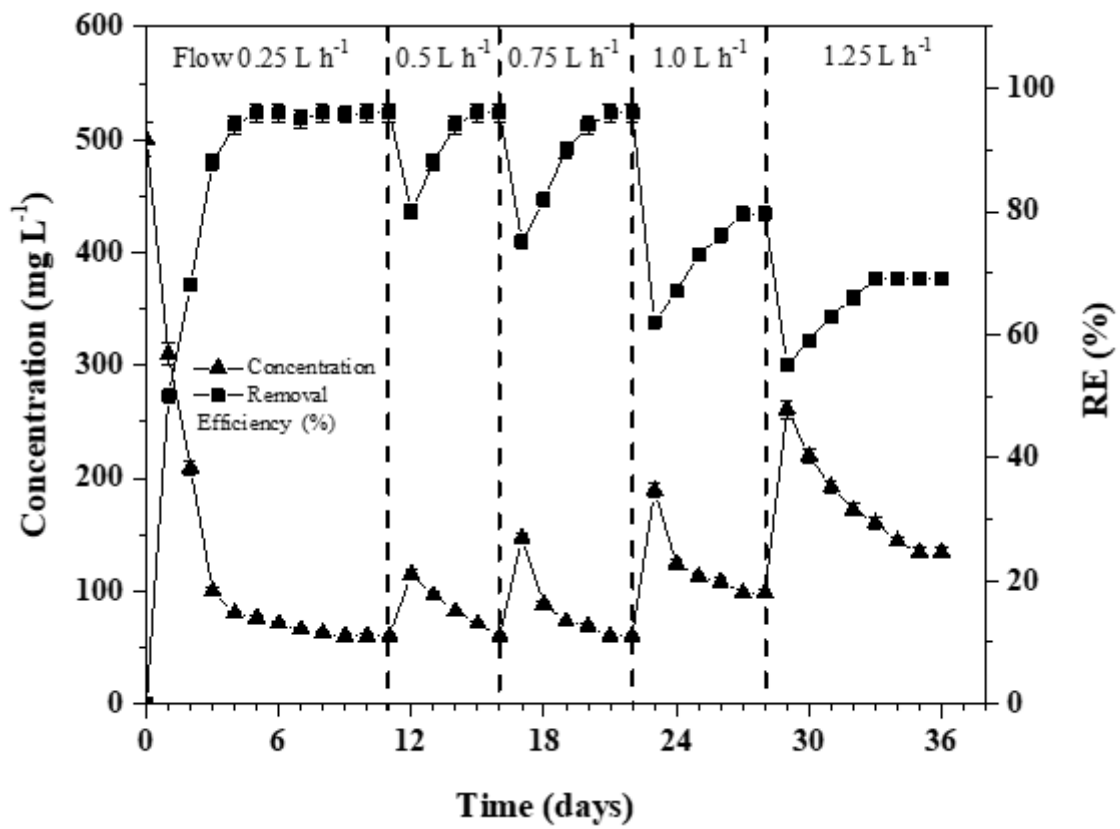


Figure 5

Bioreactor performance with changing inlet Indanthrene Blue RS feed flow rate.



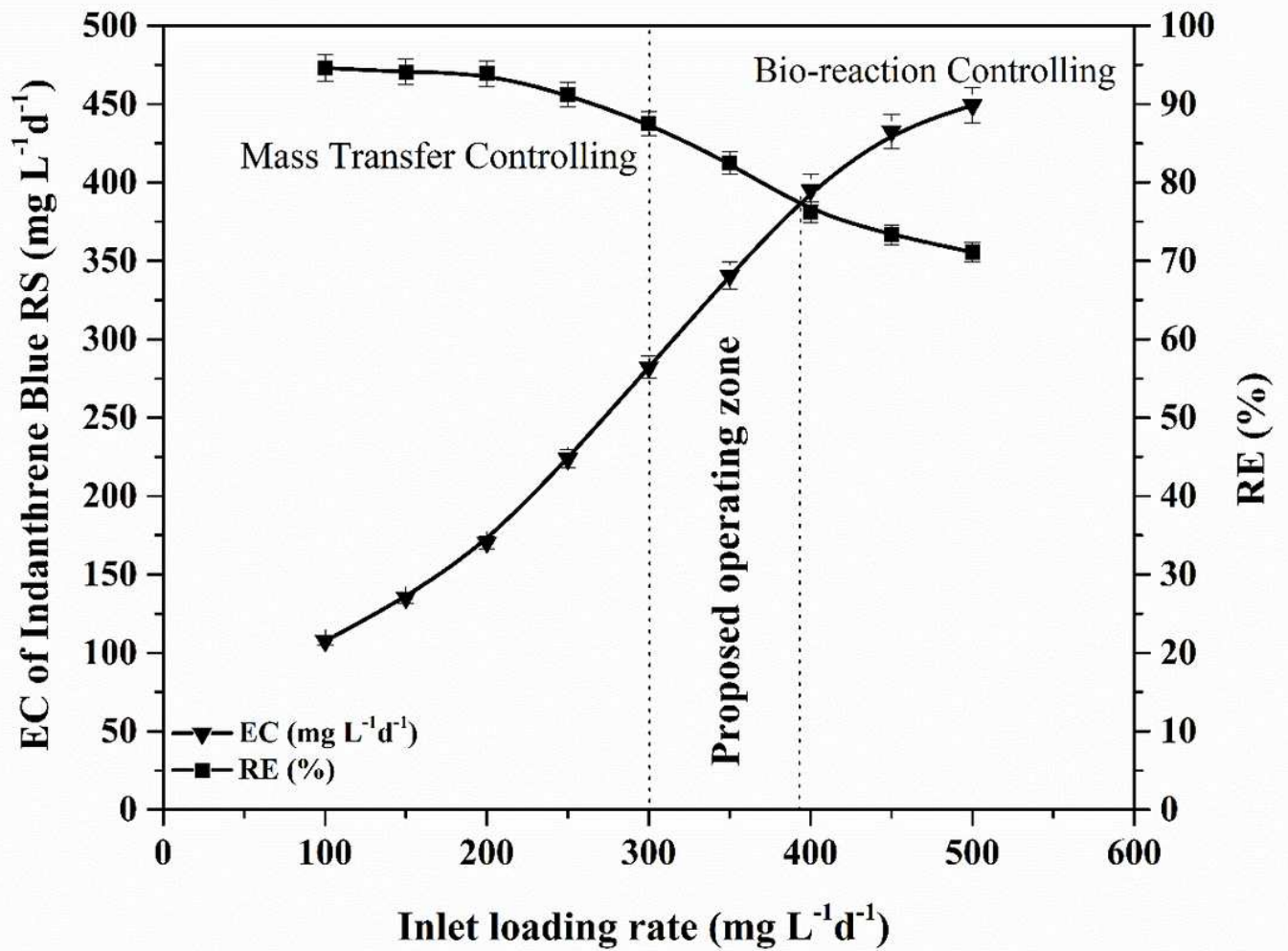


Figure 6

Effect of inlet Indanthrene Blue RS load on elimination capacity and removal efficiency.

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