

1 **Genomics Facilitates Evaluation and Monitoring of McCloud River Redband Trout (*Oncorhynchus***
2 ***mykiss stonei*)**

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19 **Abstract**

20 The McCloud River Redband Trout (MRRT; *Oncorhynchus mykiss stonei*) is a unique subspecies of
21 rainbow trout that inhabits the isolated Upper McCloud River of Northern California. A major threat to MRRT
22 is introgressive hybridization with non-native rainbow trout from historical stocking and contemporary
23 unauthorized introductions. To help address this concern, we collected RAD-sequencing data on 308 total
24 individuals from MRRT and other California *O. mykiss* populations and examined population structure using
25 Principal Component and admixture analyses. Our results are consistent with previous studies; we found that
26 populations of MRRT in Sheepheaven, Swamp, Edson, and Moosehead creeks are nonintrogressed.
27 Additionally, we saw no evidence of introgression in Dry Creek, and suggest further investigation to determine
28 if it can be considered a core MRRT conservation population. Sheepheaven Creek was previously thought to be
29 the sole historical lineage of MRRT, but our analysis identified three: Sheepheaven, Edson, and Dry creeks, all
30 of which should be preserved. Finally, we discovered diagnostic and polymorphic SNP markers for monitoring
31 introgression and genetic diversity in MRRT. Collectively, our results provide a valuable resource for the
32 conservation and management of MRRT.

33 **Keywords:** conservation genetics, introgression, trout, SNP assay

34 **Introduction**

35 The McCloud River Redband Trout (MRRT; *Oncorhynchus mykiss stonei*, Jordan 1894) is a subspecies of
36 rainbow trout (*O. mykiss ssp.*) that has been isolated in the Upper McCloud River for tens of thousands of years
37 by a series of waterfalls that arose through volcanic activities (Figure 1; Legendre et al. 1972; Miller 1972;
38 Moyle et al. 2008). The long isolation of MRRT is indicated by their unique ancestral characteristics compared
39 to other subspecies of *O. mykiss*, such as the lowest number of gill rakers, greater number of scales along and
40 above the lateral line, and the frequent presence of vestigial basibranchial teeth (Behnke 1992, 2002). These
41 ancestral characteristics have led to the suggestion that MRRT is a distinctive lineage descended from an early
42 invasion of ancestral trout into the McCloud River headwater system (Behnke 1992, 2002).

43 The Upper McCloud River and MRRT have been heavily impacted by anthropogenic disturbance
44 over the last hundred years, beginning with grazing and logging. Intense cattle grazing in the first half of the
45 20th century eliminated streamside vegetation, created shallower and wider streams with warmer temperatures,
46 and reduced water quality (Moyle et al. 2008). Logging began in the late 1800s and rapidly expanded with
47 railroad construction through World War II. Logging degraded stream habitat through removal of shade canopy,
48 further increasing in water temperatures, sedimentation, and peak storm flows, while lowering fish habitat
49 diversity (Bolda and Meyers 1997; Moyle et al. 2008). Road construction throughout the Upper McCloud Basin
50 also provided easy access to streams, increasing inputs of sediment and pollutants. However, in the late 20th
51 century, both private and public land managers began to limit logging and grazing in the area (Moyle et al.
52 2008).

53 Recurring drought cycles also threaten MRRT, especially given the porous volcanic soil in the region.
54 Even in average or above average water years, many streams in the Upper McCloud region have dry reaches
55 (Pittman 2011). Several times over the past several decades managers have responded to drought conditions and
56 declining MRRT census sizes by translocating fish from drying pools as a conservation measure. Sheepheaven
57 Creek is the most isolated and morphologically distinct population of MRRT (Behnke 2002), and for many
58 years was thought to be the only non-introgressed MRRT population (see below), due to its remote location. In
59 the 1970s, a series of droughts led managers to translocate MRRT from Sheepheaven Creek into presumably
60 fishless Trout and Swamp creeks as a safeguard (Nielsen et al. 1999; Simmons et al. 2010). Sheepheaven Creek
61 experienced another bottleneck during the 1990-1994 drought when the population declined to fewer than 200

62 individuals (Nielsen et al. 1999). More recently, severe drought conditions from 2013-2015 spurred the
63 California Department of Fish and Wildlife (CDFW) to “rescue” MRRT from three populations (Edson,
64 Swamp, and Moosehead creeks) and place them in the Mount Shasta Hatchery to prevent local extirpation
65 (CDFW 2017). These creeks were prioritized for rescue because genetic analyses (Simmons et al. 2010)
66 indicated that these three populations, along with Sheepheaven Creek had not become introgressed (i.e., were
67 genetically “pure” MRRT) with hatchery rainbow trout (*O. mykiss ssp.*) (see below). The rescued MRRT were
68 spawned across populations in captivity before they could be returned to the wild, founding a new captive
69 MRRT broodstock. Since then, their progeny have been used to stock the Upper McCloud River and McCloud
70 Reservoir with MRRT for recreational angling and overall population reinforcement. As the risk of severe
71 drought increases with climate change, continued management efforts to limit grazing and logging, combined
72 with captive propagation aim to make MRRT more resilient (Moyle et al. 2008).

73 Another historical and ongoing major threat to MRRT is the introduction of both distantly and closely
74 related nonnative trout species. Brown trout (*Salmo trutta*) and brook trout (*Salvelinus fontinalis*) were
75 introduced as early as late 1800s. Brown and brook trout impact MRRT through a combination of factors such
76 as predation, competition, and disease introductions (Fausch et al. 2006; McGrath and Lewis 2007; Moyle et al.
77 2008). Yet an even greater threat to MRRT was the introduction of closely related *O. mykiss ssp.*, leading to
78 introgressive hybridization (i.e., incorporation of alleles from one species into the gene complex of another
79 (Anderson 1949)), which can result in loss of both genetic identity and locally adapted alleles (Allendorf and
80 Leary 1988; Wolf et al. 2001; White et al. 2018). Following introgressive hybridization with a conspecific, the
81 phenotype and genotypes of introduced trout tend to become pronounced over time, leading to the erosion of
82 the distinctive characteristics of the native taxon (Wilde and Echelle 1992; Moyle et al. 2008; Seehausen et al.
83 2008).

84 Due to increasing recognition of the threat of introgressive hybridization – both from authorized and
85 unauthorized introductions, the stocking of hatchery rainbow trout was discontinued by CDFW in 1994, after
86 Berg (1994) reported introgression in extant MRRT populations (Nielsen et al. 1999; M. Dege, CDFW, pers.
87 comm.). Subsequently, several rounds of genetic studies were conducted to discern levels of introgression in the
88 extant MRRT populations (Nielsen 1999; Simmons et al. 2010). To be conservative, the population in
89 Sheepheaven Creek was considered to be the only non-introgressed population of MRRT until genetic analysis

90 by Simmons et al. (2010). Nielsen et al. (1999) used microsatellite data and found unique genetic characteristics
91 (e.g., number of alleles, heterozygosity, genetic distance) of Sheepheaven MRRT in comparison to other
92 MRRT. The most recent genetic study was by Simmons et al. (2010), who expanded the sampling effort of
93 Nielsen and used ten SNPs (one mitochondrial and nine nuclear SNPs). Simmons found no evidence of
94 introgression in four out of nine sampled MRRT populations (Sheepheaven, Moosehead, Edson, and Swamp
95 creeks) and only low levels of introgression in several other MRRT populations. Based on Simmons et al.
96 (2010), these four non-introgressed populations are currently classified by CDFW as the “core MRRT
97 conservation populations”, and due to the complex geology and hydrology of the Upper McCloud Watershed,
98 are almost entirely isolated except during very rare high flow events (Nielsen et al. 1999; Moyle et al. 2008). In
99 addition, CDFW wrote a genetic monitoring plan as part of a strategy to manage and protect MRRT, including
100 preventing, detecting, and monitoring introgression from any unauthorized stocking. For MRRT, the threats of
101 small population sizes, hatchery propagation, and unauthorized stocking make ongoing management actions
102 and genetic monitoring necessary.

103 This project aims to develop genetic tools to aid in monitoring of introgression and genetic diversity
104 in MRRT populations. Specific goals are to: 1) test previous “core MRRT conservation population”
105 designations with genomic data; 2) evaluate additional putatively non-introgressed populations for possible
106 designation as core conservation populations; and 3) develop two sets of genetic markers for rapid genetic
107 monitoring of MRRT: markers that are diagnostic between MRRT and introduced *O. mykiss ssp.* to monitor
108 introgression, and markers that are polymorphic within MRRT to monitor genetic diversity (e.g., overall
109 diversity and inbreeding) in both the wild and captivity. Managers can use these markers to identify and adapt
110 to changes in introgression levels and genetic diversity to make informed decisions for conservation actions.

111 **Materials and methods**

112 **Sample collection and RAD library preparation**

113 We compiled DNA from 308 individuals from a variety of MRRT and other rainbow trout subspecies in
114 California (Table 1). We included every California rainbow trout subspecies for two reasons. First, the
115 placement of MRRT in phylogenetic analysis has not been consistent among previous studies (Berg 1987;
116 Nielsen et al. 1999). Second, the hatchery stocking and translocation records of rainbow trout into the Upper

117 McCloud River watershed are incomplete and do not include unauthorized introductions (Simmons et al. 2010;
118 M. Dege, CDFW, pers. comm.). The sample collection includes individuals from either archived (some of
119 which were also used in Simmons et al. [2010]) or newly collected samples (Table 1). Our study included three
120 general groups of samples. In addition to the MRRT sample group, we included a group of various rainbow
121 trout samples which we refer to as the “Rainbow Trout Group” (RBTG). The RBTG group includes wild
122 rainbow trout from Surprise Valley, Goose Lake, Pit River, and Warner Valley (also commonly called “redband
123 trout”, but here referred to these as “Other Rainbow Trout” because redband trout does not refer to a
124 monophyletic group), hatchery rainbow trout strains (Coleman, Pit, Hot Creek, Mt. Whitney, Mt. Shasta, and
125 Eagle Lake hatcheries), and wild rainbow trout from North Fork American River, Eagle Lake, and Yuba River.
126 The other sample group is called the “Golden Trout Complex” (GTCX) and includes a representative sample of
127 *O. mykiss ssp.* fish from Golden Trout Creek, South Fork Kern River, Little Kern River, and Kern River
128 watersheds (Table 1; Figure 1).

129 We extracted DNA using the DNeasy extraction kit according to the manufacturer’s protocols
130 (Qiagen). After extraction, we prepared libraries for Restriction Site Associated DNA sequencing (RAD-seq)
131 with the SbfI enzyme based on the protocol described in Ali et al. (2016). For sequencing, all libraries were
132 pooled into a single lane for paired end 150 bp sequencing on an Illumina HiSeq 4000 at UC Davis Genome
133 Center.

134 **Alignment, filtration, and population genetic analysis**

135 After sequencing, we de-multiplexed the data into individual samples and aligned them to the rainbow trout
136 reference genome (Pearse et al. 2019) using the MEM algorithm implemented in the software program BWA
137 (Li and Durbin 2009) to generate Sequence Alignment Map (SAM) files for each individual. The SAM files
138 were then converted to Binary Alignment Map (BAM) files using SAMTOOLS (Li and Durbin 2009; Li 2011).
139 We then used SAMTOOLS to sort, filter for proper pairs, remove PCR duplicates, and index the BAM files. At
140 this stage, 19 individuals were removed due to low read numbers, with 289 individuals remaining. We
141 conducted population genetic analyses in ANGSD (Korneliussen et al. 2014), which analyzes BAM files using
142 on a probabilistic framework in the form of genotype likelihoods. For the analyses, we used the SAMTOOLS

143 genotype likelihood model (-GL 1) with a minimum base quality of 10 (-minQ 10) and minimum mapping
144 quality of 20 (-minMapQ 20).

145 We performed hierarchical population structure analyses: 1) on all samples to identify potential
146 sources of introgression with MRRT, 2) on MRRT and potential introgression sources to estimate introgression
147 levels as accurately as possible, and 3) on just MRRT to examine within-group population structure. For each
148 analysis, we used PCA plots and admixture analyses. For PCAs, we used PCAngsd (Meisner and Albrechtsen
149 2018) on all samples that passed initial quality filtering to produce a covariance matrix. Our PCAngsd
150 parameters were: SNP_pval = 1e-6, -doMajorMinor, and -doMaf 1. In addition, sites had to have a -minMaf
151 0.05 and be present in at least 50% of the individuals. We then used admixture analyses to test various
152 population groupings based on genetic structure and shared ancestry. To do this, we used NgsAdmix (Skotte et
153 al. 2013) to estimate admixture proportions of individuals with various cluster (K) values based on genotyped
154 likelihoods calculated in ANGSD (Korneliussen et al. 2014). We conducted 10 runs for each K value.

155 To assess genetic diversity of each population, we used theta statistics in ANGSD (-doThetas) and
156 thetaStat (-do_stat) programs (Korneliussen et al. 2013, 2014). To do this, Site Frequency Spectrum (SFS) (-
157 doSaf) was used as a prior to calculate Tajima's Θ (Θ_π) (Tajima 1983). Tajima's Θ estimates theta ($\Theta = 4N\mu$)
158 based on the average number of pairwise nucleotide differences, and when genomic data is used, accurate
159 estimates can be made with even a small sample size (Nelson et al. 2012; Subramanian 2016).

160 **Discovery of candidate diagnostic and polymorphic loci for MRRT populations**

161 Using the results from our admixture and PCA analyses, we designed two sets of candidate SNP loci: one to
162 detect levels of introgression between rainbow trout and MRRT, and a second to monitor within-MRRT genetic
163 diversity. To design the introgression markers (also referred to as MRRT diagnostic markers), we used two
164 overlapping groups of MRRT: MRRT_A and MRRT_p. MRRT_A included all MRRT sampled, even those known to
165 have low levels of introgression, and MRRT_p only includes MRRT that are putatively “pure” (i.e., no detectable
166 introgression) by our population structure analyses (the four core populations, Swamp, Sheepheaven, Edson,
167 and Moosehead creeks, plus Dry Creek; see below). We then used these two sample groups to find loci with
168 substantial allele frequency differences between MRRT and a reference rainbow trout (RBT) group (a subset of
169 RBTG) that includes wild and hatchery rainbow trout (Eagle Lake, North Fork American River, Lower Yuba

170 River, Lower Stanislaus River, Coleman Strain, Eagle Lake Strain, Hot Creek Strain, Mt. Shasta Strain, Pit
171 Strain).

172 We performed genotype calling in ANGSD to find candidate SNPs with alleles fixed or nearly-fixed
173 in MRRT_A or MRRT_P but not present or present at very low frequency in the RBT group. Since MRRT_P is not
174 introgressed, we expected a higher number of MRRT-unique markers to be captured compared to MRRT_A;
175 therefore, for MRRT_P, we increased our cutoff to 99% to discover loci with a higher degree of fixation for pure
176 MRRT. After genotype calling, we calculated the minor allele frequency of each locus in each group and
177 selected loci at two levels: with a frequency of $\geq 90\%$ and $\geq 99\%$ in the MRRT group, and conversely a
178 frequency of $< 10\%$ and $< 1\%$ in the RBT group. The $< 10\%$ filter was used for both MRRT_A and MRRT_P, and
179 the $< 1\%$ filter was used for MRRT_P, specifically. Markers specific to MRRT (both groups) are appropriate for
180 monitoring the entire MRRT population including pure and introgressed, and markers specific to MRRT_P are
181 appropriate for monitoring pure MRRT populations.

182 In a separate analysis, we investigated variable SNP loci suitable for monitoring the genetic attributes
183 of both MRRT_A and MRRT_P groups. We performed genotype calling using the same pipeline applied for
184 introgression markers, except we used a Hardy Weinberg test filter (-doHWE 1) to remove paralogs. Paralogous
185 loci can cause the misidentification of heterozygous/homozygous genotypes. After removing paralogs, we
186 selected polymorphic loci with minor allele frequencies between 0.25 to 0.45 in each of MRRT groups. We
187 applied this range of allele frequencies to capture loci that are polymorphic with a moderate frequency: alleles
188 with a lower frequency (< 0.25) are not present in all individuals, and alleles with a higher frequency (0.45 -0.5)
189 can be duplicates (i.e., paralogous loci) so this filter further reduces paralogs. We performed the process
190 separately for MRRT_A and MRRT_P with the same allele frequency (between 0.25- 0.45). Furthermore, to avoid
191 linkage disequilibrium (LD) between selected loci in both polymorphic and diagnostic, we chose loci that are at
192 least 5,000,000bp apart.

193 **Validation analysis**

194 We conducted a validation analysis to determine whether the MRRT_A diagnostic markers identified above were
195 truly diagnostic for the MRRT lineage. We designed SNPTYPE (Fluidigm Corp.) genotyping assays for a set of
196 44 candidate SNP loci from MRRT_A loci. We obtained dried fin clip samples from fish collected between 2002

197 and 2007 from three putatively pure MRRT populations (Sheepheaven, Swamp, and Edson creeks) and three
198 populations of hatchery rainbow trout from Crystal Lake (Coleman and Pit strains) and Darrah Springs (Eagle
199 Lake strain) hatcheries. All samples used for validation analyses were taken from the collection of samples used
200 by Simmons et al. (2010) to ensure consistency and allow for comparison across studies. We extracted DNA
201 from the fin clips on the Hamilton Microlab NIMBUS® HD (Hamilton Company) using the Omega Bio-tek
202 Mag-Bind® Blood & Tissue DNA HDQ 96 Kit (Omega Bio-tek, Inc.) according to the manufacturer's
203 instructions. We then amplified the 44 candidate MRRT_A diagnostic SNP loci using Fluidigm® SNP Type
204 Assays and the Juno 96.96 Genotyping Integrated Fluidic Circuit (IFC) on the Juno™ instrument (Fluidigm
205 Corporation) following the manufacturer's protocol. The Fluidigm SNP genotyping method first uses a locus-
206 specific primer (LSP) and a specific target amplification (STA) primer to enrich for DNA sequences containing
207 the SNP of interest. After enrichment, the LSP and two fluorescently labeled allele specific primers (ASP1 and
208 ASP2) amplify the two possible SNP alleles at a particular locus. We collected fluorescent end-point reads on
209 the Biomark™ HD instrument and used the Fluidigm® SNP Genotyping Analysis Software (version 4.5.1) for
210 SNP allele scoring.

211 **Results**

212 **Overall population structure and admixture**

213 Three clear clusters were distinguishable in the PCA that corresponded to the Upper McCloud River redband
214 trout (MRRT), rainbow trout (RBTG), and golden trout (GTCX) groups (Fig. 2a; Table 1). The first principal
215 component (PC1) separates MRRT from GTCX and RBTG, and the second principal component (PC2)
216 separated MRRT from GTCX (Fig. 2a). The horizontal distribution of MRRT along PC1 towards RBTG
217 suggests introgression with this group rather than GTCX. The results in the PCA are supported by the
218 admixture analyses with all samples included, using K=2 and K=3 (Fig. 2b). At K=2, MRRT is differentiated
219 from RBTG and GTCX, and at K=3, GTCX, RBTG, and MRRT are three distinct clusters, but there is about
220 20-25% RBTG ancestry within much of the MRRT group. After confirming that all of our Upper McCloud
221 River samples clustered together and were genetically separated from the other trout subspecies, we next
222 attempted to distinguish putatively pure and introgressed populations of MRRT. From the previous PCA and
223 admixture result (Fig. 2), it is clear that MRRT are an independent cluster that does not group with GTCX or

224 RBTG. However, the MRRT distribution on PC1 trends towards a subset of RBTG (a small cluster at the right
225 end of the RBTG main cluster), which suggests that introgression in MRRT is with a subset of RBTG (Fig. 2).
226 This RBTG subset includes wild rainbow trout from Eagle Lake and the American River, Steelhead trout from
227 the Yuba and Stanislaus Rivers, hatchery strains (Shasta, Coleman, Eagle Lake, Hot Creek strains) and three
228 creeks from the “Other Rainbow Trout” group: Lincoln, Lost, and Nelson creeks (Table 1). Hence, we
229 performed PCA and admixture analyses on MRRT and this subset of RBTG to most accurately identify the
230 introgression level in MRRT populations (Fig. 3).

231 The PCA performed with MRRT and the RBTG subset cluster, first suggests that both hatchery and
232 wild Eagle Lake rainbow trout populations are not the likely source of introgression. Even though on the all
233 sample PCA (Fig. 2a) they are clustered separately as a probable introgression source (small circle within
234 RBTG) (Fig. 3A, top right cluster). Second, the MRRT individuals’ horizontal and linear distribution along the
235 PC1 towards the RBTG cluster suggests introgression of some MRRT populations (Fig. 3a). An admixture
236 analysis on the same MRRT and RBTG subset sample groups strongly supports the PCA result (Fig. 3b). The
237 correlation coefficient of 99% ($r = 0.995$) between PC1 position and the proportion of rainbow trout ancestry
238 (%RBT) on each individual further supports this concordance.

239 Our admixture plot showed no evidence of introgression in individuals tested in five populations from
240 the Upper McCloud watershed: Swamp, Edson, Sheepheaven, Dry, and Moosehead creeks (Fig. 3b). However,
241 there is apparent introgression in individuals from Trout, Blue Heron, Raccoon, Cow, Bull, Tate, McKay,
242 Shady Gulch creeks. Furthermore, six sampling locations in the mainstem Upper McCloud River show varying
243 levels of introgression (Fig. 3b). The levels of introgression vary among the locations (Table 3). For example,
244 Bull, Cow, and Trout creeks have the lowest introgression levels (mean %RBT of 0.0205, 0.0547, and 0.0980,
245 respectively), with other locations showing higher levels of introgression (> 0.1910).

246 **Within MRRT population structure**

247 To assist managers in prioritizing MRRT populations for further population genetic investigation, we examined
248 population structure within MRRT. To do this, we conducted admixture analyses with only the MRRT group
249 (Fig. 4). We found three main ancestry groups at $K=4$: red representing Swamp and Sheepheaven creeks

250 ancestry group, blue representing Edson and Dry creeks ancestry group, and purple representing Bull Creek
251 ancestry group (Fig. 4, K=4).

252 We calculated theta statistics (Θ_{π}) (Table 4) to quantify the overall genetic diversity of the pure
253 MRRT populations. Dry and Edson had the highest (0.000699) and lowest (0.000474) Θ_{π} values, respectively,
254 although Moosehead Creek's Θ_{π} value (0.000691) is almost as high as Dry Creek's. Sheepheaven and Swamp
255 creeks had intermediate Θ_{π} estimates (0.000592 and 0.000561, respectively).

256 **Diagnostic loci for MRRT**

257 We identified putatively non-introgressed MRRT populations based on our admixture and PCA results and
258 identified individuals for two MRRT groups (MRRT_A and MRRT_P). We used N=57 individuals from the
259 MRRT_A population group and N=22 for the MRRT_P group (Sheepheaven, Edson, Swamp, Moosehead, and Dry
260 creeks). The RBT group contained 35 wild and hatchery rainbow trout individuals (Eagle Lake, North Fork
261 American River, Lower Yuba River, Lower Stanislaus River, Coleman Strain, Eagle Lake Strain, Hot Creek
262 Strain, Mt. Shasta Strain, Pit Strain).

263 We discovered 44 differentially fixed SNPs between the broader MRRT_A group and RBT. We found
264 2,649 diagnostic SNPs at a frequency of $\geq 90\%$ between the MRRT_P and RBT (Table 2). When we increased the
265 minimum frequency of diagnostic alleles to 99% in the MRRT_P, the number of SNPs was reduced to 574. All of
266 these 574 fixed SNPs had the frequency of 100% in MRRT_P and zero percent in RBT before validation (Table
267 2). Of the MRRT_A loci, 80% were found in MRRT_P with the 90% cutoff, and 50% were found in MRRT_P with
268 the 99% cutoff. For polymorphic markers, we found 6,639 loci in MRRT_A and 7,316 in MRRT_P (Table 2).

269 **Validation results**

270 We successfully amplified and visualized distinct heterozygote and homozygote clusters in 27 of the 44
271 candidate MRRT_A diagnostic SNP loci using the Fluidigm SNP Type assays (Tables S1 and S2). There were
272 two loci (omy01_36537055 and omy15_57867903) that failed to amplify in more than 90% of individuals in
273 specific MRRT_P and the tested hatchery rainbow trout populations; this data should therefore be interpreted
274 with caution (denoted by asterisks in Table 5). We obtained genotypes for 90 MRRT from putatively non-
275 introgressed locations and 87 hatchery rainbow trout individuals for at least 90% of these two loci; we discarded
276 data from one individual from Edson Creek because it did not meet this threshold. Of the 27 remaining MRRT_A

277 diagnostic loci, 16 were completely fixed between the groups of MRRT_A and hatchery rainbow trout that we
278 tested (Table 5). The frequency of MRRT alleles was greater than 95% for all historically non-introgressed
279 MRRT populations that we tested (Swamp, Sheepheaven, and Edson creeks) in 21 of the candidate diagnostic
280 loci (Table 5); we propose that these loci should be used in assessing introgression between MRRT_A and
281 MRRT_P with hatchery rainbow trout as part of future monitoring efforts.

282 **Discussion**

283 Using population structure analysis, we identified potential MRRT introgression sources from RBTG and we
284 confirmed former findings of four non-introgressed MRRT locations (Edson, Swamp, Sheepheaven,
285 Moosehead), and potentially one more (Dry). In addition, we identified three main historical lineages with
286 MRRT: Sheepheaven and Swamp creeks, Edson and Dry creeks, and Bull Creek. Using these results, we
287 discovered and validated diagnostic and polymorphic SNP loci specific to the MRRT_P (Swamp, Edson,
288 Sheepheaven, Moosehead, and Dry creeks) and MRRT_A (Swamp, Edson, Sheepheaven, Dry, Moosehead, Blue
289 Heron, Bull, Cow, Tate, McCloud River, Raccoon, Trout, McKay, and Shady Gulch Creek) groups to monitor
290 their introgression and genetic diversity both in the wild and the captive breeding program.

291 **Ancestral history of MRRT**

292 Previous research has identified morphological similarities (Schreck and Behnke 1971; Hoopagh 1974, Gold
293 1977) and a shared karyotype (2n=58; Thorgaard 1983) between GTCX and MRRT. Behnke (1981) reconciled
294 their morphological similarities and in light of their geographic separation by suggesting that GTCX may have
295 resulted from multiple invasions of a primitive redband trout via the Sacramento River and Tulare Lake, and
296 that all ancient redbands were subsequently extirpated except for those in the Upper McCloud River. However,
297 Gall (1981) found that the trout in the Upper McCloud River differ substantially from golden trout and rainbow
298 trout by meristic, chromosomal, and electrophoretic characteristic traits. Furthermore, subsequent microsatellite
299 data found that the golden trout complex (except Little Kern golden trout) is significantly different from the
300 MRRT (Nielsen et al. 1999), but the authors suggested that inbreeding and introgression could cause significant
301 genetic difference when two stocks with a common ancestry are isolated. Later, Stephens (2007) also reported a
302 more distant common ancestor of MRRT and golden trout complex from AFLP data.

303 In our analysis we found three main groups (MRRT, RBTG, and GTCX) using PCA and admixture
304 analyses in our samples (Fig. 2a, b). Despite the fact that the level of introgression in the MRRT group found by
305 the admixture analysis was not very high, at $K=2$, the introgression source in MRRT appears to be from the
306 RBTG and GTCX ancestry group because these two groups are combined at $K=2$. However, at $K=3$, when
307 GTCX and RBTG are split, the introgression source is RBTG (Fig. 2b). Our assessment that MRRT is separate
308 from both RBTG and GTCX supports Behnke's (1992) claim that MRRT should have subspecific status. Thus,
309 we did not find evidence of recent shared ancestry between MRRT and the golden trout complex.

310 **Potential sources of introgression**

311 Most hatchery strains of rainbow trout (primarily *O. m. irideus*) used in California were derived from rainbow
312 trout in the upper Sacramento River (Nielsen et al. 1999). The Eagle Lake rainbow trout (*O. m. aquilarum*) was
313 also stocked extensively in California for several decades and is still in use by CDFW. Given the long history of
314 trout stocking (hatchery records indicating stocking of Mt. Shasta hatchery trout in the Upper and Lower
315 McCloud River (M. Dege, CDFW, pers. Comm.)) and the geographic proximity to the McCloud River, we
316 expected the Mt. Shasta Hatchery (Mt. Shasta, California) to be the most probable source of planted fish in the
317 McCloud River. Prior to Mt. Shasta Hatchery, there was also extensive stocking of trout beginning in the late
318 1800s from Baird Hatchery, which was located near the confluence of the McCloud River and the Pit River but
319 covered by Lake Shasta following the construction of Shasta Dam in the 1940s. Indeed, Berg (1994) reported
320 hatchery introgression of the Upper McCloud fish by assessing several protein loci, ultimately leading to the
321 end of the hatchery stocking in 1994.

322 Our results show that only a small subset of the rainbow trout that we examined (Fig. 2a, bottom right
323 of top left cluster) can be sources of introgression. This small group includes wild and hatchery rainbow trout,
324 and three populations from the "Other Rainbow Trout" group. Among the hatchery strains used in this study,
325 Mt. Shasta, Coleman, and Hot Creek represent a potential source of introgression, but Pit Strain and possibly
326 Eagle Lake strain are not; Pit Strain is markedly different from other hatchery trout strains used in California;
327 Pit strain rainbow trout were originally sourced from the Pit River, which falls within the geographic range of
328 the general redband trout designation. It is notable that Nielsen et al. (1999) assessed the same hatchery groups
329 as our study but only found significant genetic association (low genetic distance, R_{ST}) between Sheepheaven

330 Creek MRRT and Eagle Lake hatchery strain rainbow trout, but not other hatchery strains. The conflict may be
331 explained by the power of RAD sequencing with a much greater coverage and thousands of markers in
332 comparison to highly variable but limited number of microsatellite markers for discovering and genotyping
333 polymorphic loci. In addition, mainstem Upper McCloud River has been extensively stocked with a variety of
334 rainbow trout strains and individual hatcheries often stock multiple different strains of domesticated rainbow
335 trout. For example, Mt. Shasta Hatchery has used Shasta, Coleman and Eagle Lake strain trout for various
336 stocking activities (M. Dege, CDFW, pers. comm.). Additionally, incomplete records of stocking from the past
337 century and ad hoc crosses between hatchery strains can explain observing multiple hatchery strains as a source
338 of introgression. Identifying this small group as the source of introgression allows us to more accurately
339 quantify introgression level.

340 **Levels of Introgression**

341 Our results are generally consistent with Nielsen et al. (1999) and Simmons et al. (2010) in supporting the
342 classification of Edson, Swamp, Sheepheaven, and Moosehead creeks as non-introgressed MRRT core
343 conservation streams. As part of marker discovery, we also included samples from five other creeks in the
344 Upper McCloud watershed: Dry, Bull, Cow, Shady Gulch, and Blue Heron. We only used 3-5 samples from
345 each of these creeks for MRRT diagnostic marker discovery, but our introgression analyses suggest that further
346 genetic investigation of Dry Creek as a candidate core conservation stream is warranted because we detected no
347 rainbow trout introgression in our samples. We also found only minor admixture influence in Bull and Cow
348 creeks and recommend that additional genetic sampling be performed to discern whether these creeks should be
349 considered core conservation streams. Furthermore, in future analyses it is important to include samples from
350 different reaches of these creeks; introgression is believed to vary with hydrological connectivity, accessibility,
351 and the potential for unauthorized stocking (M. Dege, CDFW, pers. comm.).

352 The unique hydrology and connectivity patterns in the Upper McCloud River could explain the
353 introgression patterns we observed. In general, we expect lower stream connectivity and isolation from the
354 public to correlate with lower levels of introgression. The northern streams (Swamp, Edson, Sheepheaven, and
355 Trout creeks) are the least introgressed and the most isolated. These locations are mostly disconnected from the
356 mainstem Upper McCloud except during rare high-flow events (Nielsen et al. 1999) and are largely on private

357 timber properties (M. Dege, CDFW, pers. comm.). The only non-introgressed population in the south is
358 Moosehead Creek. Although Moosehead Creek is regularly connected by surface flows, a concrete fish barrier
359 near its confluence with the Upper McCloud River prevents fish from ascending into the creek (M. Dege,
360 CDFW, pers. comm.). The tributaries that are mainly south of the Upper McCloud River are expected to be
361 introgressed, because they are connected to the mainstem Upper McCloud River and may have been stocked
362 with hatchery rainbow trout in the past. For example, Raccoon and Tate creeks showed introgression consistent
363 with Simmons et al. (2010). Raccoon and Tate creeks lack fish barriers between the mainstem Upper McCloud
364 River and are easily accessible to the general public, which may explain the observed higher levels
365 introgression.

366 We expected that Swamp and Trout creeks, which were recipients of MRRT from Sheepheaven Creek
367 in the 1970s, to be genetically similar to Sheepheaven Creek. However, in the admixture analysis, Trout Creek
368 shows low levels of introgression (0.098), but Swamp does not (Fig. 3). Prior to stocking MRRT, there was a
369 rainbow trout eradication action in Trout Creek which may have been incomplete, or rainbow trout may have
370 been unlawfully introduced via a public campground in the Lower Trout Creek. However, Pittman (2011)
371 reported MRRT as the only salmonid species observed in Trout Creek. Although Trout Creek has no
372 discernable barriers to fish passage, the levels of introgression are highly dependent on sampling location, with
373 lower Trout Creek showing patterns of introgression while upper Trout Creek appears to be more “pure” MRRT
374 (Stephens et al. 2013). In addition to Trout Creek, Bull and Cow creeks had low levels of introgression (0.0205,
375 0.0547, and 0.098 respectively). These creeks are hydrologically isolated from Upper McCloud River most of
376 the year but may be connected during seasonal or higher flow events, especially in the lower section of Bull
377 Creek (S. Plemons and M. Dege, CDFW, pers. comm.) leading to possible introgression with RBT.

378 **Pure MRRT have three ancestral groups**

379 We observed three historical lineages with MRRT using admixture analyses: 1) Sheepheaven and Swamp
380 creeks, 2) Edson and Dry creeks, and 3) Bull Creek (Fig. 4, K=4). Sheepheaven and Swamp creeks’ common
381 ancestry on the admixture plots is consistent with the fact that Swamp Creek was historically fishless prior to a
382 translocation of MRRT from Sheepheaven Creek. Similar to Simmons et al. (2010), our genetic diversity
383 analysis shows that Swamp Creek has a slightly lower genetic diversity ($\Theta_{\pi} = 0.000561$) than Sheepheaven (Θ_{π}

384 = 0.000592) which is expected for a derived population and consistent with founder effects (Table 4). Similarly,
385 Trout Creek was founded by a translocation from Sheepheaven Creek and shows a higher proportion of
386 common ancestry with Sheepheaven and Swamp creeks than with other introgressed populations.

387 Interestingly, the Edson and Dry creek ancestry seems to be the most common ancestry group (Fig. 4),
388 as this ancestry group was present in most MRRT populations. However, there is less Sheepheaven Creek
389 ancestry in MRRT_A populations than we expected; Sheepheaven was previously thought to be the “sole
390 representative” of MRRT (Behnke 1992, 2002). Simmons et al. (2010) also found private alleles specific to
391 Edson and Moosehead which were not present in Sheepheaven. For the same reason, it is also surprising that
392 Bull Creek, which has the same ancestry group with Edson, Dry, and Moosehead at K=3, shows a unique
393 MRRT signature at K=4.

394 Our results have significant implications for the management of MRRT. For example, our results
395 suggest that Sheepheaven Creek is not the only representative of ancestral MRRT lineages. Edson and Dry
396 creeks may also represent distinct historical genetic lineages; additional genetic testing with a broader sampling
397 distribution within each tributary is warranted to further refine MRRT conservation efforts for each individual
398 creek. In addition, our analyses showed that Edson Creek has the lowest genetic diversity ($\Theta_{\pi} = 0.000474$). This
399 indicates that Edson Creek may benefit from targeted conservation management such as supplementation from
400 other pure MRRT populations within the same lineage (e.g. Moosehead). Simmons et al. (2010) reported low
401 allelic richness for Edson and Sheepheaven, and significant inbreeding and a bottleneck in Sheepheaven based
402 on microsatellite data. However, to prioritize the populations for conservation purpose based on their adaptive
403 potential, neutral genetic diversity is not enough and other factors must be considered such as balancing
404 selection on particular segregating sites, effective population size, rate of mutation, and the populations'
405 adaptive diversity (Messer and Petrov 2013; Kardos and Luikart 2021; Teixeira and Huber 2021).

406 **Diagnostic and polymorphic markers**

407 We identified SNP loci that are both diagnostic for identifying non-introgressed and introgressed MRRT and
408 loci that are polymorphic in MRRT. The development of new SNP type assays will facilitate rapid, consistent
409 genetic typing of individual fish from both pure MRRT and introgressed populations to help inform adaptive
410 conservation strategies and actions. We found 44 diagnostic loci (21 after validation) for the entire MRRT

411 population and more than 500 SNP loci for the non-introgressed MRRT, reflective of their unique, divergent
412 lineage from RBTG. Although only MRRT_A diagnostic markers were validated, we would expect a similar
413 validation efficiency for the MRRT_P markers since the marker discovery methods used for MRRT_P and MRRT_A
414 were similar.

415 **Management implications and conclusion**

416 The major goal of genetic monitoring in fish and wildlife populations is to help maintain adaptive capacity in
417 unique genetic lineages, providing populations with resiliency against environmental changes into the future.
418 Introgression between threatened native and introduced populations can alter the genetic diversity of native
419 populations and impact their long-term viability (White et al. 2018; Frankham et al. 2009). The historical and
420 ongoing potential for unauthorized introduction of other rainbow trout has exposes the naturally isolated,
421 genetically unique MRRT to introgressive hybridization with nonnative rainbow trout subspecies. Furthermore,
422 MRRT is especially prone to further loss of genetic diversity through processes inherent to small, isolated
423 populations, such as genetic drift and an accumulation of inbreeding effects. Currently, MRRT only exists in a
424 handful of small, isolated populations and also in a conservation hatchery program as a captive broodstock, thus
425 making an adaptive management strategy a necessity.

426 Admixture, population structure, and genetic diversity analyses conducted in this study corroborate
427 previous work that showed that Edson, Swamp, Sheepheaven, and Moosehead creeks are not introgressed with
428 hatchery rainbow trout. However, we suggest further investigation of Dry, Bull, Trout, and Cow creeks with
429 larger sample sizes to further characterize patterns of introgression and to determine which, if any should be
430 considered for core conservation stream status.

431 The monitoring and management of introgressive hybridization in MRRT core conservation
432 populations should be a critical objective of an MRRT management strategy. Currently, most of the recent
433 hatchery-produced MRRT have been stocked in the Upper McCloud River. Hatchery-produced F₁ juveniles,
434 which were the result of the spawning of rescued wild adult MRRT were stocked into both Trout and
435 Moosehead creeks in a one-time event to help restart the populations following the extreme drought conditions.
436 Currently, no authorized stocking is occurring in the creeks with pure MRRT populations, and further stocking
437 activities will likely be limited to the Upper McCloud River between Algoma (Upper McCloud River) and

438 McCloud Reservoir (M. Dege, CDFW, pers. comm.). Overall, the results from this study will be a valuable
439 resource to assist with future genetic evaluation and monitoring of MRRT.

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448

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452 **Availability of data and material:**

453 **Author contributions:** Resources: E.H., M.R.M., J.R., M.S., and A.J.F .; investigation: E.H., M.R.M., and
454 A.J.F.; formal analysis: E.H., M.R.M., and A.J.F ; validation: D.G., L.S., and J.R.; visualization: E.H., M.R.M.,
455 and A.J.F; writing of original draft: E.H., M.R.M., and A.J.F.; writing, review, and editing: E.H., M.R.M., D.G.,
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457 **Code availability:** Not applicable **Ethics approval:** Not applicable **Consent to participate:** Not applicable

458 **Consent for publication:** Not applicable

459

460

461 **Figure and Table captions:**

462 **Fig. 1.** Sampling locations of all samples in this study. The left map shows sampling locations of all the samples
463 by watershed, including wild, hatchery rainbow trout, “Other Rainbow Trout”, and Kern River golden trout and
464 rainbow trout. Inset area is the Upper McCloud River (UMCR) watershed. The red box shows sampling
465 locations within the Upper McCloud River watershed, the area above the Middle Falls. Highlighted river in
466 purple is the mainstem McCloud River.

467 Golden Trout Creek (GTCR), South Fork Kern River (SFKR), Kern River (KRNR), Eagle Lake (EGLK), North
468 Fork American River (NFAR), Lower Stanislaus River (LSTN), Lower Yuba River (LYBA), Coleman
469 Hatchery (COLE), Eagle Lake Hatchery (EGLH), Hot Creek Strain (HTCS), Mt. Shasta Hatchery (MTSH), Pit
470 Strain Hatchery (PITS), Warner Valley (WARV), Goose Lake (GOSL), Surprise Valley (SPRV), North Fork
471 Pit River (NFPR), South Fork Pit River (SFPR), Upper Pit River (UPIT), Lower Pit River (LPIT), Yuba North
472 Fork (YUBA), Upper McCloud River (UMCR): Swamp Creek (SWPC), Edson Creek (EDSN), Sheepheaven
473 Creek (SHPN), Dry Creek (DRYC), Moosehead Creek (MOHD), Bull Creek (BLLC), Cow Creek (COWC),
474 Trout Creek (TRTC), Shady Gulch Creek (SHGU), Raccoon Creek (RCCN), McCloud River (MCLD), McKay
475 Creek (MCKY), , Blue Heron Creek (BLUN), Tate Creek (TATE)

476

477 **Fig. 2.** Population Structure of all samples. **a.** All samples PCA, color represents watershed. Three main groups
478 are distinguishable: Golden Trout Complex (GTCX), Rainbow Trout Group (RBTG), and McCloud River
479 Redband Trout (MRRT). PC1(8.7% variance explained) / PC2(7.65% variance explained). **b.** All samples
480 admixture plots at K=2 (top admixture plot) and K=3 (bottom admixture plot). Blue represents MRRT ancestry
481 group which is different from GTCX (green) and RBTG (red).

482 Golden Trout Creek (GTCR), South Fork Kern River (SFKR), Kern River (KRNR), Eagle Lake (EGLK), North
483 Fork American River (NFAR), Lower Stainslaus River (LSTN), Lower Yuba River (LYBA), Coleman
484 Hatchery (COLE), Eagle Lake Hatchery (EGLH), Hot Creek Strain (HTCS), Mt. Shasta Hatchery (MTSH), Pit
485 Strain Hatchery (PITS), Warner Valley (WARV), Goose Lake (GOSL), Surprise Valley (SPRV), North Fork
486 Pit River (NFPR), South Fork Pit River (SFPR), Upper Pit River (UPIT), Lower Pit River (LPIT), Yuba North
487 Fork (YUBA), Upper McCloud River (UMCR)

488

489 **Fig. 3.** PCA and admixture analyses of MRRT with a group of a potential source of introgression – a small
490 subset of RBTG. **a.** PCA of MRRT with the RBTG small subset, color represents watershed. The RBTG subset
491 includes: wild rainbow trout from Eagle lake (EGLK) and North Fork American River (NFAR), Steelhead from
492 Lower Yuba River (LYBA) and Stanislaus River (LSTN), hatchery strains from Mt. Shasta (MTSH), Coleman
493 (COLE), Eagle Lake (EGLH), Hot Creek (HTCS), and three from the “Other Rainbow Trout” group: Lincoln
494 and Lost from Lower Pit River watershed (LPIT) and Nelson creeks from Yuba watershed (YUBA),
495 PC1(11.6% variance explained) / PC2(3.44%, variance explained). **b.** admixture plot of MRRT and the RBTG
496 subset cluster. Five pure populations are identified within the MRRT population: Swamp creek (SWPC), Edson
497 Creek (EDSN), Sheepheaven Creek (SHPN), Dry Creek (DRYC), and Moosehead Creek (MOHD)

498
499 **Fig. 4.** Population structure within MRRT group. MRRT group admixture plot. Three major ancestry groups are
500 distinguishable: red represents Swamp and Sheepheaven creeks’ ancestry group, blue represents Edson and Dry
501 creeks’ ancestry group, and purple represent Bull creek’s ancestry group. The population’s order is based on
502 increasing rainbow trout ancestry.

503 Swamp Creek (SWPC), Edson Creek (EDSN), Sheepheaven Creek (SHPN), Dry Creek (DRYC), Moosehead
504 Creek (MOHD), Bull Creek (BLLC), Cow Creek (COWC), Trout Creek (TRTC), Shady Gulch Creek (SHGU),
505 Raccoon Creek (RCCN), McCloud River (MCLD), McKay Creek (MCKY), , Blue Heron Creek (BLUN), Tate
506 Creek (TATE)

507
508 **Fig. S1.** Admixture analysis of all the samples from K=4-6. X-axis shows grouping by watershed. There is no
509 common ancestor between MRRT group (UMCR watershed) and the California Golden Trout Complex
510 (GTCX).

511 Golden Trout Creek (GTCR), South Fork Kern River (SFKR), Kern River (KRNR), Eagle Lake (EGLK), North
512 Fork American River (NFAR), Lower Stanislaus River (LSTN), Lower Yuba River (LYBA), Coleman
513 Hatchery (COLE), Eagle Lake Hatchery (EGLH), Hot Creek Strain (HTCS), Mt. Shasta Hatchery (MTSH), Pit
514 Strain Hatchery (PITS), Warner Valley (WARV), Goose Lake (GOSL), Surprise Valley (SPRV), North Fork
515 Pit River (NFPR), South Fork Pit River (SFPR), Upper Pit River (UPIT), Lower Pit River (LPIT), Yuba North
516 Fork (YUBA), Upper McCloud River (UMCR)

517

518 **Table 1.** Additional information for all samples included in the analysis and samples removed after the
519 sequencing and alignment qualifying filtering. Three groups of California native trout were used in this study:
520 Rainbow Trout Group (RBTG), Golden Trout Complex (GTCX), Upper McCloud River Redband Trout
521 (MRRT)

522

523 **Table 2.** Number of loci discovered for monitoring introgression (N loci differentiating MRRT from RBT) and
524 genetic diversity monitoring (N Polymorphic Loci)

525

526 **Table 3.** Percentage of Rainbow Trout ancestry in each MRRT population. The third and fourth columns show
527 minimum and maximum percentage of Rainbow Trout (RBT) ancestry and the last column shows the average
528 across the samples of each population.

529

530 **Table 4.** Theta estimate of the five pure MRRT populations. Θ_{π} = Tajima's Θ , Θ_w = Watterson's Θ . Dry and
531 Edson, respectively, have the highest and the lowest estimate of the genetic diversity using theta statistic.

532

533 **Table 5.** Frequencies of MRRT alleles by locus for selected MRRT and hatchery rainbow trout populations.
534 Numbers in bold denote MRRT allele frequencies that are less than 95% for a specific population or grouping. *
535 = loci that failed to amplify in more than 90% of individuals tested in a specific population or grouping. n =
536 sample size.

537

538 **Table S1.** Diagnostic MRRT SNP marker sequence information. Sequences are shown in the 5' to 3' direction
539 with brackets denoting the two possible SNP alleles. The chromosome name and position refer to the NCBI
540 *Oncorhynchus mykiss* genome assembly accession GCA_002163495.1.

541

542 **Table S2.** Diagnostic MRRT SNP marker Fluidigm assay information. The SNP assay name refers to the
543 *Oncorhynchus mykiss* chromosome number and position from the NCBI genome assembly accession
544 GCA_002163495.1. ASP1 SNP = allele associated with allele-specific primer 1; ASP2 SNP = allele associated

545 with allele-specific primer 2; ASP1 = allele-specific primer 1 sequence; ASP2 = allele-specific primer 2
546 sequence; LSP = locus-specific primer sequence; STA = specific target amplification primer sequence; AMP
547 GC = proportional GC content.

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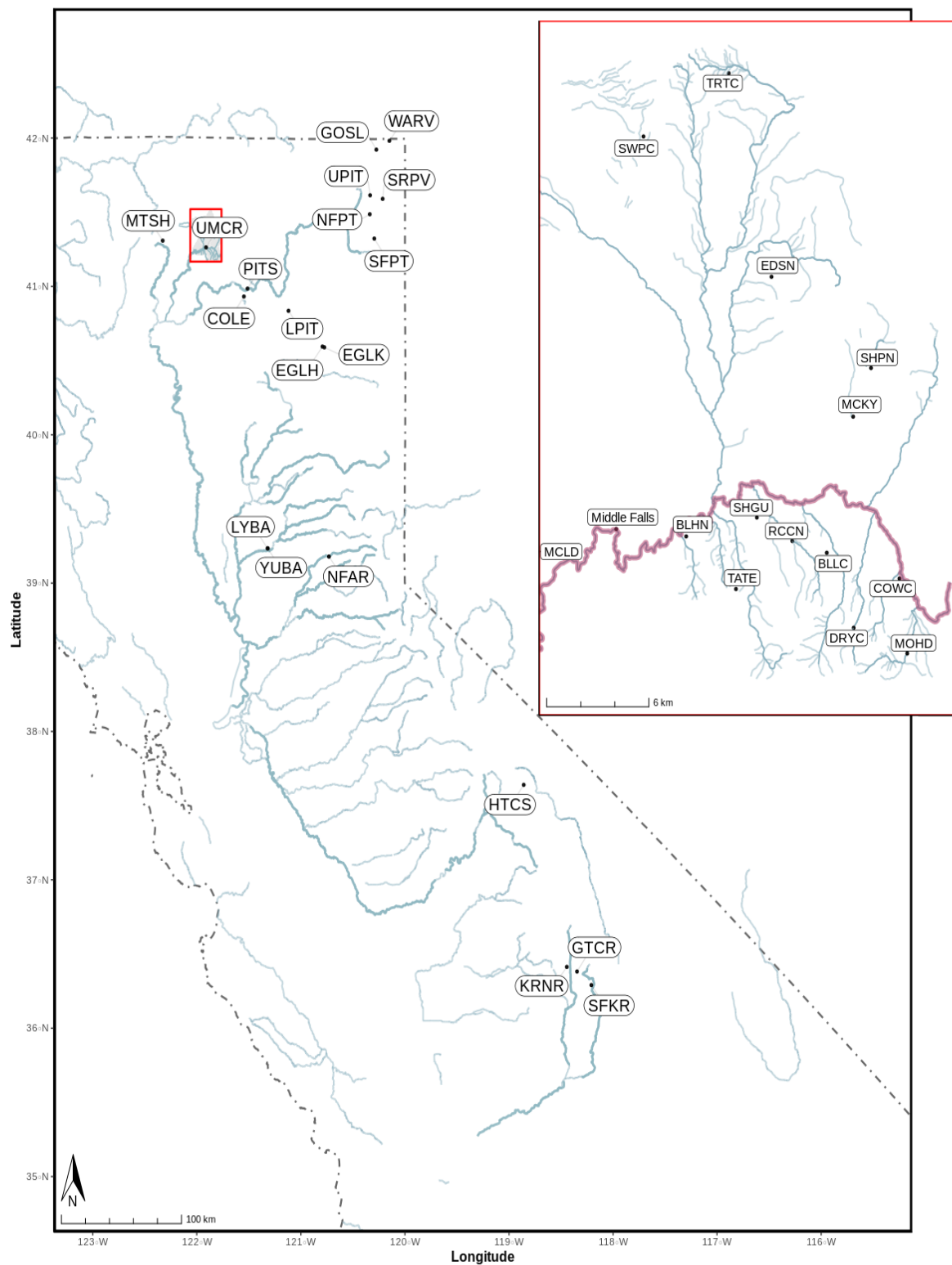
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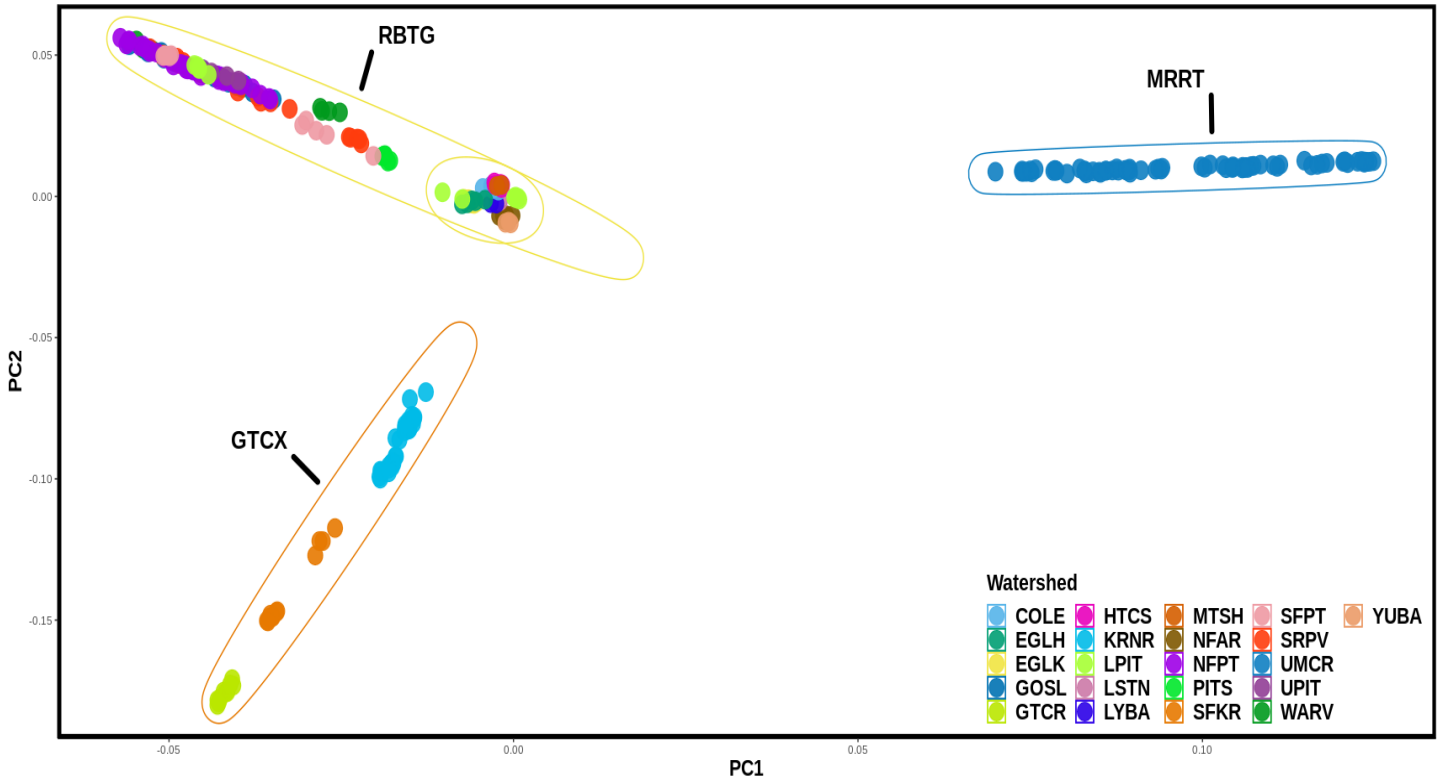
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Figure 1.

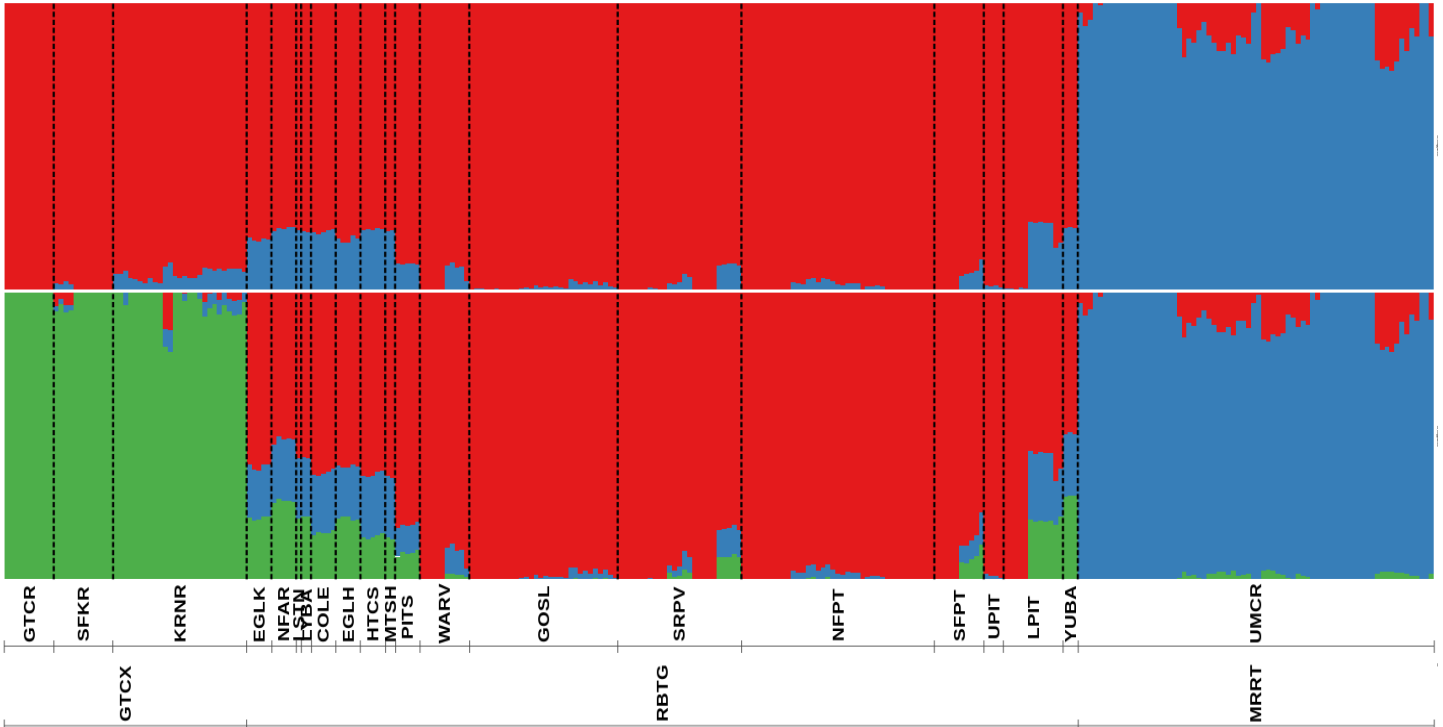


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Figure 2.

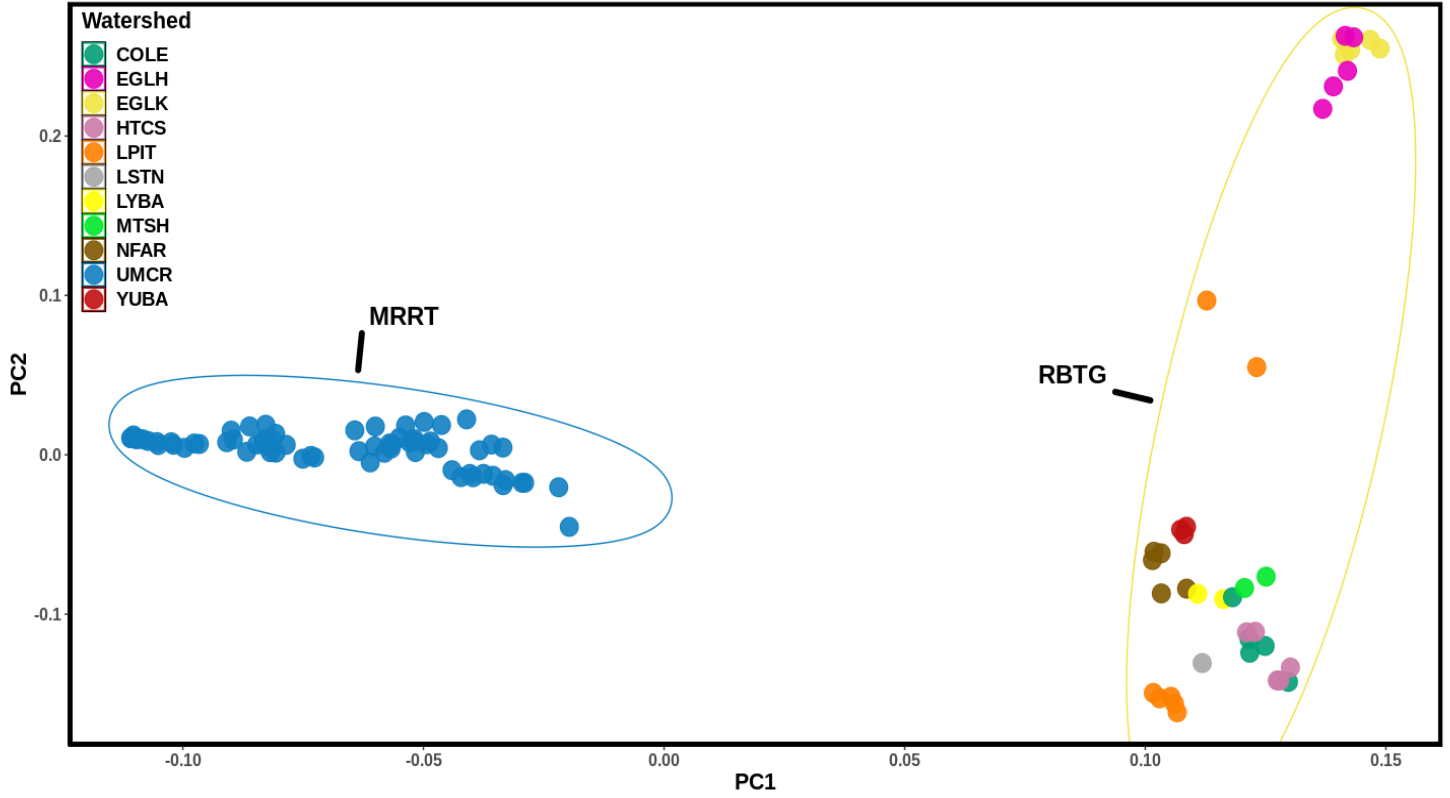


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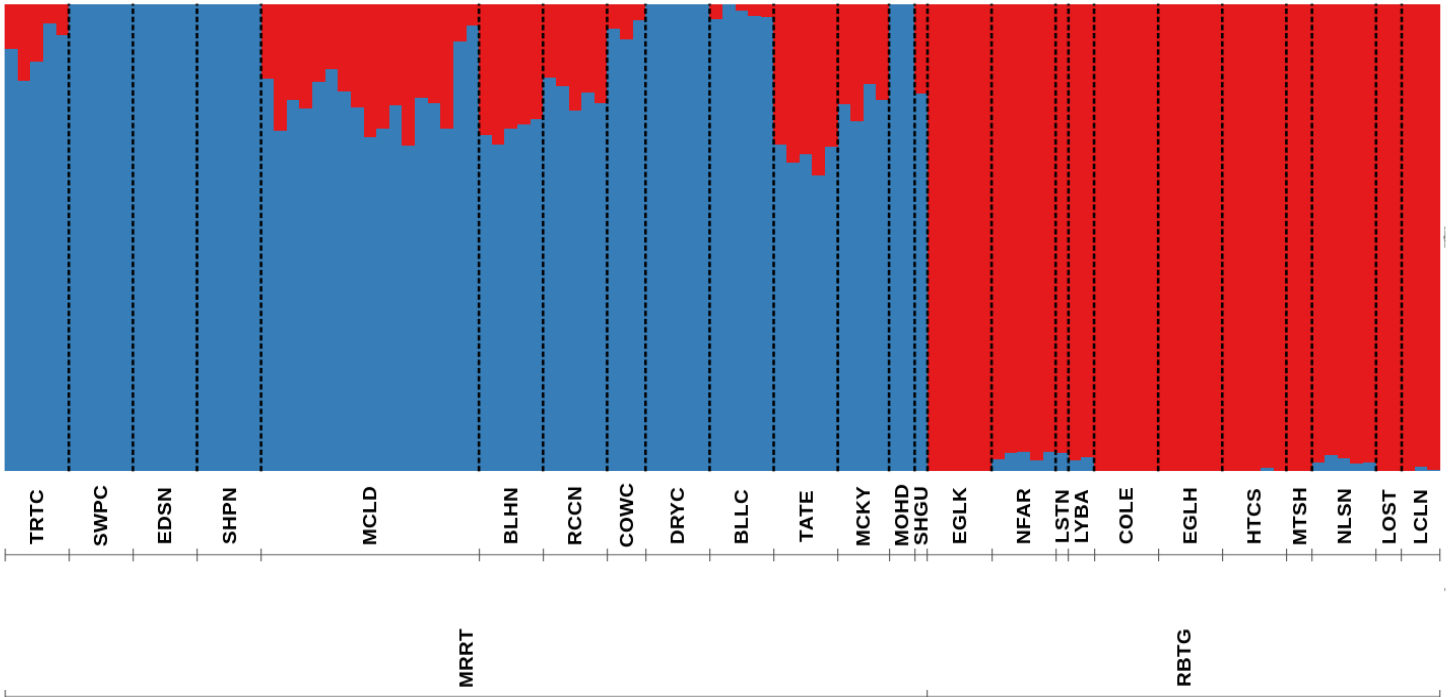


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Figure 3.



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Figure 4.

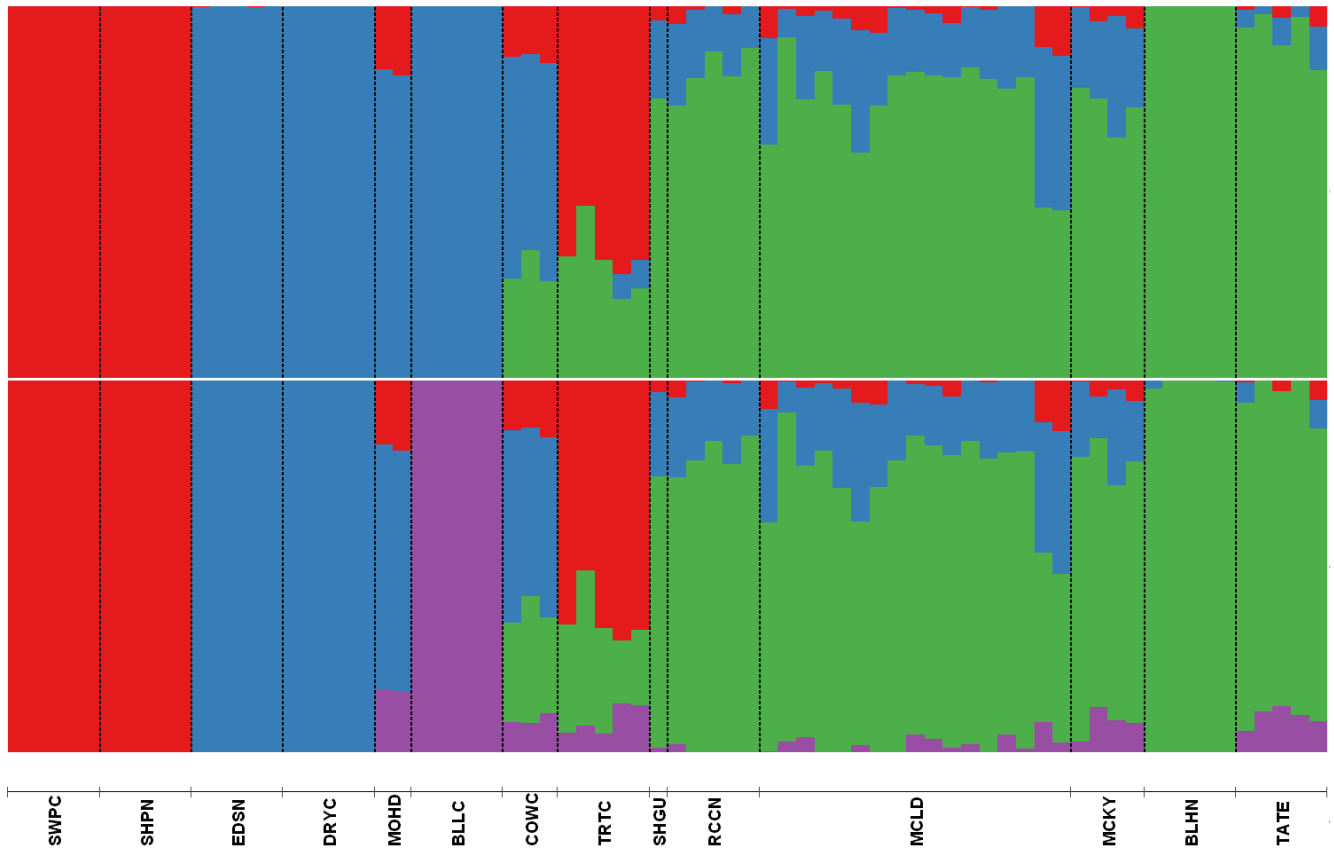


Table 1.

Group	Location	Watershed	Sample Year	Watershed code	N Samples	N Samples removed
Rainbow Trout Group (RBTG)	Lassen Cr (LASN)	Goose Lk	2004	GOLE	5	
	Cottonwood Cr (CTWD)	Goose Lk	2004	GOLE	5	
	Buck Cr (BUCK)	Goose Lk	2009	GOLE	5	
	Davis Cr-Goose Lk (DVGL)	Goose Lk	2009	GOLE	5	
	Willow Cr (WILW)	Goose Lk	2009	GOLE	5	
	Upper Pine Cr (UPNE)	Goose Lk	2003	GOLE	5	
	Lost Cr (LOST)	Lower Pit R	2009	LPIT	5	3
	East Fork Nelson Cr (NLSN)	Lower Pit R	2004	LPIT	5	
	Davis Creek (Pit R)	Lower Pit R	2008	LPIT	5	
	Joseph Cr (JSPH)	NF Pit R	2009	NFPT	5	
	Spring Canyon Cr (SPRC)	NF Pit R	2009	NFPT	5	
	Parker Cr (PARK)	NF Pit R	2003	NFPT	5	
	Shields Cr (SHLD)	NF Pit R	2009	NFPT	5	
	Couch Cr (COUC)	NF Pit R	2003	NFPT	5	1
	N Fork Pit R (NFPT)	NF Pit R	2009	NFPT	5	
	East Cr (EAST)	NF Pit R	2009	NFPT	5	
	Franklin Cr (FRKN)	NF Pit R	2009	NFPT	5	
	Thoms Cr (THMS)	NF Pit R	2009	NFPT	4	
	Parsnip Cr (PSNP)	SF Pit R	2009	SFPT	5	
	Fitzhugh Cr (FTZH)	SF Pit R	2003	SFPT	5	
	Emerson Cr (EMRN)	Surprise V	2009	SRPV	5	
	Mill Cr (MILL)	Surprise V	2009	SRPV	5	
	Cedar Cr (CEDR)	Surprise V	2003	SRPV	5	
	Bidwell Cr (BDWL)	Surprise V	2009	SRPV	5	
	Deep Cr (DEEP)	Surprise V	2009	SRPV	5	
	Dismal Cr (DISM)	Warner V	2002	WARV	5	
	Upper 12 Mile Cr (UTMC)	Warner V	2002	WARV	5	
	Lincoln Cr (LCLN)	Yuba N Fork	2007	YUBA	3	
	Hot Creek Strain (HTCS)	Hatchery	2011	HTCS	5	
	Mt. Whitney Strain (MTWS)	Hatchery	2002	MWTS	2	2
	Mt. Shasta Strain (MTSH)	Hatchery	2002	MTSH	4	2
	Coleman Strain (COLE)	Hatchery	2011	COLE	5	
	Pit Strain (PITS)	Hatchery	2004	PITS	5	
Eagle Lake (EGLH)	Hatchery	2004	EGLH	5		
N Fork American R (NFAR)	American R	2006	NFAR	5		
Eagle Lk (EGLK)	Eagle Lk	2009	EGLK	5		
Lower Yuba R (LYBA)	Yuba	2009	LYBA	2		
Battle Creek CNFH	Battle Cr	2009	BATC	1	1	
Lower Stanislaus R (LSTN)	Stanislaus R	2009	LSTN	1		
Total					177	
Upper McCloud River Redband Trout (MRRT)	Sheepheaven Cr (SHPN)	Upper McCloud	2002	UMCR	5	
	Bull Cr (BLLC)	Upper McCloud	2008	UMCR	5	
	Dry Cr (DRYC)	Upper McCloud	2008	UMCR	5	
	McKay Cr (MCKY)	Upper McCloud	2007	UMCR	4	
	Raccoon Cr (RCCN)	Upper McCloud	2007	UMCR	5	

	Swamp Cr (SWPC)	Upper McCloud	2007	UMCR	5	
	Tate Cr (TATE)	Upper McCloud	2008	UMCR	5	
	Blue Heron Cr (BLHN)	Upper McCloud	2008	UMCR	5	
	Trout Cr (TRTC)	Upper McCloud	2007	UMCR	5	
	Shady Gulch Cr (SHGU)	Upper McCloud	2007	UMCR	1	
	Edson Cr (EDSN)	Upper McCloud	2007	UMCR	5	
	Cow Cr (COWC)	Upper McCloud	2007	UMCR	4	1
	Moosehead Cr (MOHD)	Upper McCloud	2016	UMCR	3	1
	McCloud River (MCLD)	Upper McCloud	2007	UMCR	17	
Total					74	2
Golden Trout Group (GTCX)	Volcano Cr Left String (VCLS)	Golden Trout Cr	2005	GTCR	2	
	Salt Lick Cr (STLC)	Golden Trout Cr	2005	GTCR	2	
	Groundhog Cr (GDHG)	Golden Trout Cr	2005	GTCR	2	
	Little Whitney Cr (LTWY)	Golden Trout Cr	2005	GTCR	2	
	Mouth Barrigan (MBNS)	Golden Trout Cr	2005	GTCR	2	
	Upper SFork Kern R (SFKR)	S Fork Kern R	2006	SFKR	2	
	Mulkey Creek (MLKY)	S Fork Kern R	2006	SFKR	2	
	Above Shaefer (ASHF)	S Fork Kern R	2006	SFKR	2	
	Above Ramshaw Barrier (RmsB)	S Fork Kern R	2006	SFKR	2	
	Above Templeton Barrier (TMPB)	S Fork Kern R	2006	SFKR	2	
	SFK Below Snake Creek (BSNK)	S Fork Kern R	2006	SFKR	2	
	Wind R (WIND)	Wyoming	2005	WYOM	2	2
	Upper N Fork Clicks Cr (UNFC)	Little Kern R	2005	KRNR	2	2
	Fish Cr (FISH)	Little Kern R	2011	KRNR	2	
	Silver Lake (SILV)	Little Kern R	2007	KRNR	2	
	Rifle Cr (RIFL)	Little Kern R	2006	KRNR	2	
	Sheep Cr (SHPC)	Little Kern R	2001	KRNR	2	1
	Lion Cr (LION)	Little Kern R	2006	KRNR	2	
	Upper Willow Cr (UWLC)	Little Kern R	2011	KRNR	2	
	Tamarack Cr (SMNT)	Little Kern R	2006	KRNR	2	
	S Mountaineer Cr (SMNT)	Little Kern R	2006	KRNR	2	
	Upper Wet Meadow Cr (UWMC)	Little Kern R	2002	KRNR	1	
	Upper Soda Spring Cr (USSC)	Little Kern R	2002	KRNR	2	1
	Nine Lakes North (NLKN)	Kern R	2007	KRNR	2	1
Chagoopa Cr (CHGC)	Kern R	2007	KRNR	2		
Chagoopa Lake (CHGL)	Kern R	2006	KRNR	2		
Kern-Kaweah Cr (KKWC)	Kern R	2007	KRNR	2		
Upper Chagoopa Cr (UCHG)	Kern R	2007	KRNR	2		
Picket Cr (PIKT)	Kern R	2007	KRNR	2	1	
Total					57	
Total All Samples					308	19

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Table 2.

Group	N individuals	N loci differentiating MRRT from RBTG		N Polymorphic Loci
		Cutoff 90%	Cutoff 99%	
MRRT _A	72	46	-	6639
MRRT _p	20	2,649	574	731

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Table 3.

Location	N	Minimum RBT %	Maximum RBT %	Mean
Swamp Creek	5	0.0000	0.0000	0.0000
Sheepheaven Creek	5	0.0000	0.0000	0.0000
Edson Creek	5	0.0000	0.0000	0.0000
Dry Creek	5	0.0000	0.0000	0.0000
Moosehead Creek	2	0.0000	0.0000	0.0000
Bull Creek	5	0.0000	0.0331	0.0205
Cow Creek	3	0.0356	0.0758	0.0547
Trout Creek	5	0.0415	0.1640	0.0980
Shady Gulch Creek	1	0.1910	0.1910	0.1910
Raccoon Creek	5	0.1572	0.2283	0.1924
McCloud River	17	0.0456	0.3034	0.2035
McKay Creek	4	0.1726	0.2507	0.2107
Blue Hearon Creek	5	0.2457	0.3016	0.2709
Tate Creek	5	0.3003	0.3678	0.3272

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Table 4.

Location	$\Theta\pi$
Dry Creek	0.000699
Edson Creek	0.000474
Moosehead Creek	0.000691
Sheepheaven Creek	0.000592
Swamp Creek	0.000561

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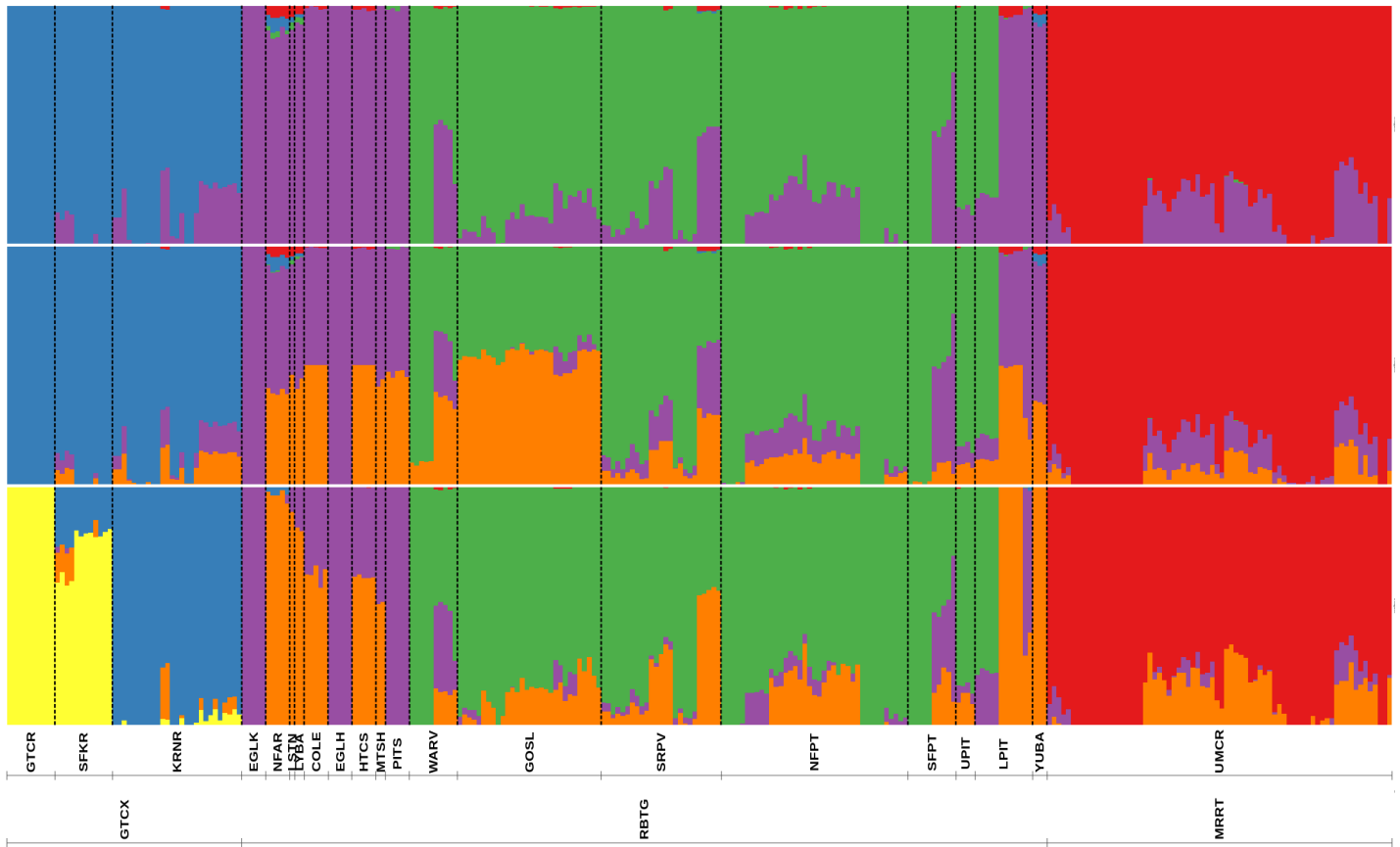
Table 5.

Locus	McCloud River Redband Trout				Hatchery Rainbow Trout			
	Swamp (n = 16)	Sheepheave n (n = 20)	Edson (n = 54)	Overall (n = 90)	Crystal: Coleman (n = 20)	Crystal: Pit (n = 17)	Darrah: Eagle Lake (n = 50)	Overall (n = 87)
omy01_36537055	1.00	1.00	1.00	1.00	0.00*	0.00	0.00	0.00
omy01_36542230	1.00	1.00	1.00	1.00	0.00	0.00	0.00	0.00
omy01_75717390	1.00	1.00	1.00	1.00	0.00	0.00	0.00	0.00
omy04_23890713	1.00	1.00	1.00	1.00	0.00	0.00	0.00	0.00
omy04_57267157	1.00	1.00	1.00	1.00	0.00	0.00	0.00	0.00
omy04_74937939	1.00	1.00	1.00	1.00	0.08	0.00	0.09	0.07
omy06_1274949	1.00	1.00	1.00	1.00	0.00	0.00	0.00	0.00
omy07_12066898	1.00	1.00	0.98	0.99	0.00	0.35	0.02	0.08
omy07_12545308	1.00	1.00	1.00	1.00	0.00	0.00	0.00	0.00
omy07_13121873	1.00	1.00	1.00	1.00	0.00	0.00	0.00	0.00
omy07_1624310	1.00	1.00	1.00	1.00	0.05	0.00	0.15	0.10
omy07_9878739	1.00	1.00	1.00	1.00	0.00	0.00	0.00	0.00
omy11_65719977	1.00	1.00	1.00	1.00	0.00	0.00	0.00	0.00
omy14_22294317	1.00	1.00	1.00	1.00	0.05	0.00	0.00	0.01
omy14_22294535	1.00	1.00	0.98	0.99	0.03	0.00	0.00	0.01
omy15_57867903	1.00	1.00	0.80*	0.89*	0.03	0.00	0.00	0.01
omy15_58335356	1.00	1.00	1.00	1.00	0.15	0.00	0.00	0.03
omy18_32367886	1.00	1.00	1.00	1.00	0.00	0.00	0.00	0.00
omy18_57370236	1.00	1.00	1.00	1.00	0.00	0.00	0.00	0.00
omy21_13749597	1.00	1.00	1.00	1.00	0.00	0.00	0.00	0.00
omy24_12391438	1.00	1.00	1.00	1.00	0.00	0.00	0.00	0.00
omy24_13095037	1.00	1.00	1.00	1.00	0.00	0.03	0.00	0.01
omy24_16290480	1.00	1.00	1.00	1.00	0.05	0.00	0.00	0.01
omy24_19143772	1.00	1.00	1.00	1.00	0.00	0.00	0.00	0.00
omy24_26165023	1.00	1.00	1.00	1.00	0.05	0.00	0.00	0.01
omy26_12749297	1.00	1.00	1.00	1.00	0.00	0.00	0.00	0.00
omy27_9625126	1.00	1.00	1.00	1.00	0.08	0.00	0.05	0

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670 **Figure S1.**

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SNP Assay Name	Sequence	Chromosome	Position
Omy01_36537055	GCTATATGCTACCTGCATCTATACTGAACAAAAATATAAACGAGTTCATATAATTTAATTAATTAATAGGCCTAATCTATGAGTTTCACATGACTG GGCATGGATGCAGTTCATGGGTGAGCCTGGGAGGGCATAGGCCATCCACATGGCAGCCAAGCCTACCCACTGGGGAGCCAGGCCAGCTAATCAGAA TGAGT/	Omy01	36537055
Omy01_36542230	GCGCACAATTGGCCAGCATCGTTCAAGTAAGGGAAGGGTATGGCCGTGAGGGCTTACCAGGCTCATTGCACTCTAGCGACTCTTGTGGGCGGG CGCCTGACGGCTGACTTCGGTCATCAGTTGACAGGGTAAGAGCGTCCAGTGTGTGTCTGTAACGCCCCAGGGAGAACTCTGAATTTAAAAAT GGAC[A/ G]ATCTGACTCTGGCACTCCAGATTGAATTTAAGAACACACANAAANTCTTACAATGCTAGCTAGGAATTTGTTATGCTAGCAAGAGGCTGCATAGTA ACAGCACTCACTCAGGTAGAGAGGCGAAGCGTTAGTACCGTCAACTGAAAGGATACCGTTTACAGCGGGATGGCAATTCCAGTCTCGAGGGCT GACTGG	Omy01	36542230
Omy01_75717390	ATACITTCGGAAGGCCTGGCATACTCTCACTTAAGTATTGAAAGTGCAGTGCACACATCTGATGTACGGATGTTGACATTGCTAGTATCAAA TGCATTTGAATTTCTACATACATAGCACATGCTGTCTCTCTGTGTGGAGATATGCTTGACTTTAATGATGGCAGTGGCACTGTGACATGACATTT GA[A/ G]AAAACGCTGTTCATATCAACCTTAATGGGAACTCAATTAGTTAGGGGCTATGATTCTTAATCCAGCCCTGCAGGTGCACATCTGTAGCAT TTTGAAGGAGAGGGGACATTTGGTTCATAGCCTTGCTGAACTTCAGAGAGTTTCCGTATAGGGGGGGTNGAGATTTACAGTTCGAGTTAGTA TCTTT	Omy01	75717390
Omy04_23890713	TGACCGAGTTGGTATGTCCTGTTAGNGTGTGACCGAGTTGGTATGTCCTGTTAGGGGTGTGACCGAGTTGGTATGTCCTGTTAGGGGTGTA TCGAGTTGGTATGTCCTTCTGACGGTGTACCCCTCTGTTGTTGTCCTGTTACGCTCCCCCGGGCGGGGATGGTTCATCTAAAGGTTANGAGG CA[G/ C]JAGCTGTATGGACATCGGAGCGTATGGAGAACCAGGTTAAACACTTCAGGCTAAAGCCTCGTGTAGACAGCTGGAGCAGTTTGTGAGAGAGN TTACCGGTTGAGTTTCTTTACACACACACACACACACACCGCACCCCAACAGGGGCTATAGAGAGTAGGGAGTTAGGGTTACAGTTGATGTGTGTG TGCTC	Omy04	23890713
Omy04_57267157	AGGAACCCCTTTGTCTTCAATAATCCAAGCCAGATGGGTCATTACATGCCATTTGGGAATATCTAAGTATATCTAAATTCACGATTGTGTGCATATTT TCCTGATATTTAGGAAAATGTTGGCCTACTTCTACCTCATGGCTAAATGTAACCTCATGTTTATCTTGTTCCTATGGGGTCAAACCTAATGCCCG TT/	Omy04	57267157
Omy04_74937939	ACTCATTTTTCATGATGAGGTGATCATATCCTTCCATTTGATGATGCCCTGGTGGCCAGCCATATCATAGACTTCTGCAGGTGTCCACATAGACT TCACGCTGTGCTTCCAGCCAGAGATCAACTCAGTGTGTTACTCCAGTGGATGGGGCTGGCTGATATTTGGTCAAATAAAATACTTGGCAA AT/	Omy04	74937939
Omy06_1274949	TGTTGGAAGGTGAACCTTACCCAGTCTGAGGTCCTGAGGGCTCTGGAGCAGGGTTTCATCAAGGATCTCTGTAATTTGCTCCGTTTCATCTTCTCT CGATCTGACTAGACTCCCTTCCCTGCCGCTGTAAAAACATCCCCACAGGATGATGCTGCCACCANCATGCTTTAGCATAAGGATGCTGCCAGGTTTA G[A/ G]CATGTAGCTAGGTATNAACCCGCATAATCCCACTCATACTACTACCAATACAACTGATTGTCATAGCTGTAGTATGAACCTGCAGGTAGCTAAA GCTAACCAACTAGGTTCAATGTTAGTCAATAGTCAACATGAGGCTTAACTAGCAATGCAAAATGGCTCTGAGATATGAATAATATTACTACACAGAT CATACA	Omy06	1274949
Omy07_12066898	AGAGTTAGAATGGAGGACTACCACTGATCCACAAAACCATCTAACAGAGACTTGCTCATGAATGATGTTGAAGTCTCGTTAGATCAACATGAT GAGAAAAGCGGTCAATAAATCTAGAACGCCAAGTGTCTTCTCTGTCGGCGGACCACTATAGGAGGACGGAGTGGTGTCCGTTCTTCTGTTCACT TTG[A/C/ G]AAATGAACAGCATCTCCGGCCTGCAGCAGGTTTCTGTCTCTATCTATTAATAGCCTGGTGGGGACAGGATCTCTGCAGGCTGCAGAACACTGA TGTAAGAGGGCAGAGAGAGCTTCCCGCAAGATTAAATCTCCCAAAATGCTAACCTTTGGCATAGGTGTATATCCAGGATCTAACCTTGG AAAGA	Omy07	12066898
Omy07_12545308	AGGCAGTCTTCGACCACCACCCAGGGTAGAGAGGCGAGTGAATGCTTTTCCATCTGAGGAGGGGACGAGGAGTCCCAGGAGTTGCAACCTG GCTGGCGGCGAGNGATGGAACGAAAGGGCCCTGATCGACCACTATCACTGCACTTTGTTGAGGAGCTCTCCGAGAGTTGACCTGCAGGGACACCA ACCTCA[C/ T]CTTCGACCAGTGTGGACCTGTCCATTCGGTTGGAATACCTGTGGCTACCCCGGATGTCAGAGCGGGTCTGTGGTTCCATCCNACAGCAC ACCGCTCTGATGCCATGGAGCTGGAGGTGTGCGCTCAGGGGACCGGAGCAGGTTCCGTTCTATGCACCATCTGTGGCTGCAGAGGTCACTGCG CGGTC	Omy07	12545308
Omy07_13121873	GTTTCAAAAATAGGCAGGAAACTGACACAAGAATACAGCTGACACTACAAGGGACAAGGACACTAAGGTATCAACAGAAAGTTTCTTAAAGAG NACGGTCTTTAGGCTTGTTTTAAATGAGCCTAGAGTCTTTGGGGCTGCAGGTGGTCCGGGAGGGCATTCCAGAGGCTGGGAGGCTTGTTCATTTTT AAAAAT/	Omy07	13121873
Omy07_9878739	CAAGGGCAGATCGACATATTCTCACCTAGTCGGCTCAGGGATTCAACCTTTCGGTGTGGACCACTCTTAAGTCTAGGCTACCTGCCGCCAAA GCTTTAAATAAGGCTCTATAGTCTGACTTCTGATGTAAGTCGAGCGCAGGCCTTTTTCATCTTAAACCAGCTGCAGACTGGCTTCATAGAGCGATAA AT[A/ G]AAGGGGATCTCCCACTGGCGTGTGANACTCAGGACCCCTCATATCAGGACCACGAACTATGAGAGTTCTCTGGTGCAGCCTGGCCTGC AGGCCGGATTACAGGACTGAGGTGTGTAGCTCCAGGGCCTGCCCTGACCAGGCTGTAGTACAGCTCTCTGTATCTTATGAAGCACCCT AGTCCA	Omy07	9878739
Omy11_65719977	NN AACACTTGGAGCCCTGCAGGAGTATACCACAANTAAAGAATTGCTGTTTGTGACCCACGCAAGGGAAAAGGGGCACTGGGGGACGGCATGGCA GGTGTCTGTG[G/ A]GCTTATTGAATCTCTTCTCTCNATAGTCAAAACAGTCAAGGCAAGATGGCTGAGTGCCTGGAATCTATCTCAATCTCGTAATGCAAAAGAAA CATCCGTCAGGCAACCCAGAAATAAAACAGCTTTGGAGTGTGACGACTTGACGCTTATTCCAAGTAAACAAAGGAAAGAAAGATTAACAACA ACAACA	Omy11	65719977
Omy14_22294317	CAACACTTTTATGAAACAGGGCAAAATGGATATATGCCTATAACAGTTTGGATCAGCTTGTATCTCCCTTTAAATAAAGGACACACCGTGGCTGCCT TCCAAGCAATGGAAACCCTCAGAGAGGAGAGACAGTTAAAAAGGTCAGAGATAGGCTTGGAGATGATATGGGCTGCAACCCTAAAGAAAGAAA GGACTTT/	Omy14	22294317
Omy14_22294535	GTTTTGTTGTTAGTTAAGGAGCTCTTTAGCACCTCAGACTCAGTGCAGCCTGCAGGGGAGAACTTTGTAGCAGGGGAAATAGAGGGAGGAGGAT CGGGGCTAGTCGATTAGAAGGGTGGGAGATCAGGAAATGTTGACGGGCTATGAGGCATGGCTGTGCAAAATAGGAATTCGNACTTAATGAAGT GTGATTA	Omy14	22294535

Table S2.

Fluidigm Assay ID	SNP Assay Name	ASP1 SNP	ASP2 SNP	AMP GC	Primer sequence (5' to 3')	
					ASP1:	ASP2:
GTA0251450	omy01_36537055	T	G	0.39	LSP:	GGCCCAGCTAATCAGAATGAGT
					ASP2:	GGCCCAGCTAATCAGAATGAGG
					LSP:	LSP: TGTCTATAATAAAGCCCTTTTGTGGCA
					STA:	STA: GCCAAGCTACCCACTG
GTA0251401	omy01_36542230	A	G	0.48	ASP1:	CTGGAGTGCCAGAGTCAGATT
					ASP2:	CTGGAGTGCCAGAGTCAGATC
					LSP:	ACGCCCCCAGGGAGAAA
					STA:	TGTGTGTCTTAAATTCATCTGGAG
GTA0251420	omy01_75717390	A	G	0.44	ASP1:	CATTAAGGGTTGATATGAACAGCTGTTTT
					ASP2:	ATTAAGGGTTGATATGAACAGCTGTTTC
					LSP:	GGCAGTGGCACTGTGACAT
					STA:	CCCCTAACCTAATTGAGTTTCCC
GTA0251458	omy04_23890713	G	C	0.59	ASP1:	TCCGATGTCCACATACAGCTC
					ASP2:	TCCGATGTCCACATACAGCTG
					LSP:	ATGCTCCCCCGGGC
					STA:	CCGGGTTCTCCATACGCT
GTA0254186	omy04_57267157	T	A	0.46	ASP1:	ACACACATAGTCAACTACATATGCCA
					ASP2:	ACACACATAGTCAACTACATATGCCT
					LSP:	CCATGGGGTCAAACCTAATGCC
					STA:	GAGGCGTTCATTTGAGCCA
GTA0251409	omy04_74937939	G	T	0.42	ASP1:	CACAGTGAATGCAGCACTGAC
					ASP2:	CACAGTGAATGCAGCACTGAA
					LSP:	GGCTGGCCTGATATTTGGTCAA
					STA:	GGCCATTTAGCTGAAAATAAAACACA
GTA0251392	omy06_1274949	A	G	0.40	ASP1:	ATAAGGATGCTGCCAGGTTTAGA
					ASP2:	AGGATGCTGCCAGGTTTAGG
					LSP:	ACAGCTATGACAATCAGTTTGTATTGGTAGT
					STA:	CATGCTTTAGCATAAGGATGCTG
GTA0251403	omy07_12066898	C	G	0.50	ASP1:	CCGTGTCTTCTGTTCAAGTTTGAC
					ASP2:	TCCGTGTCTTCTGTTCAAGTTTGAG
					LSP:	TGCTGCAGGCCGGAGA
					STA:	ACCACTATAGGAGGACGGAGT
GTA0251422	omy07_12545308	C	T	0.62	ASP1:	GCAGGGACACCAACCTCAC
					ASP2:	GCAGGGACACCAACCTCAT
					LSP:	CCGAATGGACAGGTCCACCA
					STA:	GGACGTCTGCCGAGAGT
GTA0251400	omy07_13121873	T	G	0.44	ASP1:	AGCAGTCAACCCCCAAAAGATA
					ASP2:	GCAGTCAACCCCCAAAAGATC
					LSP:	CCAGAGGCTGGGAGGCTT
					STA:	CAATGTAAGTGATCAAACCCGAGA
GTA0251417	omy07_9878739	A	G	0.51	ASP1:	CAGACTGGCTTCATAGAGCGATAAATA
					ASP2:	AGACTGGCTTCATAGAGCGATAAATG
					LSP:	TCACAACGCCAGTGGGGA
					STA:	GCCTTTTCATCTTTAACCCAGC

GTA0251414	omy11_65719977	G	A	0.47	ASP1:	TGGCAGGTGTCTGCTGG
					ASP2:	ATGGCAGGTGTCTGCTGA
					LSP:	GCCATCTTGGCCTGACTGTTT
					STA:	GCAGGCATGGCAGGT
GTA0251444	omy14_22294317	T	G	0.45	ASP1:	GGAGCTCCTTAAACTAGACCACAAAAA
					ASP2:	GAGCTCCTTAAACTAGACCACAAAAC
					LSP:	TGGGCTGCAACCCTAAAGAAGA
					STA:	TCACTGAGTCTGAGGTGCTAAA
GTA0251452	omy14_22294535	C	T	0.43	ASP1:	TGATTAAGAGCTCAGCCATTTCG
					ASP2:	GTGATTAAGAGCTCAGCCATTTCG
					LSP:	ACAGCTGCCCATGTCCCT
					STA:	TGGCTGTGTCAAATAGGAATTCCG
GTA0251388	omy15_58335356	A	T	0.56	ASP1:	TGCTAGATCCCCAGCCAGA
					ASP2:	GCTAGATCCCCAGCCAGT
					LSP:	TGCCTTCCCTGAGATGTGTCA
					STA:	TCTATAGACACACACACACACT
GTA0251464	omy18_32367886	C	G	0.42	ASP1:	ACGTGATTATTGTCTAGTCATATGGTCAAG
					ASP2:	ACGTGATTATTGTCTAGTCATATGGTCAAC
					LSP:	TGGCCAGCTGGAGCTTA
					STA:	CTGCAGGCTCTAGTTTTGTCTT
GTA0251441	omy18_57370236	G	A	0.49	ASP1:	CCCTATCCCTACACAGTGCAC
					ASP2:	CCCTATCCCTACACAGTGCAT
					LSP:	TGAGTCCCAAATGGCACCCT
					STA:	AGATTACATCTCAAATGGCACCC
GTA0251454	omy21_13749597	T	G	0.41	ASP1:	CTTTTTCATGTCTTCACTTGAAAGTCA
					ASP2:	TTTTTCATGTCTTCACTTGAAAGTCC
					LSP:	GCCAGAGCGATGAGTATTGCC
					STA:	GTCGTTGGATAATCCTCCCATTC
GTA0251396	omy24_12391438	C	A	0.31	ASP1:	GGTGCCTTGGAGTAACATGTCTTAG
					ASP2:	GGTGCCTTGGAGTAACATGTCTTAT
					LSP:	TGCATTTGTTTTCAGTAAACTTAACATGTTTAT
					STA:	GAAGTATCCGGAACTTGCAG
GTA0251462	omy24_13095037	G	A	0.58	ASP1:	GCAGACCACATGTGACCTCG
					ASP2:	GCAGACCACATGTGACCTCA
					LSP:	GACGATCCACAGGTGAGGA
					STA:	CCAACAGGCCTTACAACCG
GTA0251437	omy24_16290480	C	A	0.54	ASP1:	GTTTGACGTTTGTGTTTAGCTTTCCC
					ASP2:	GGTTTGACGTTTGTGTTTAGCTTTCCA
					LSP:	TCTGGTCAGTTCAGAGACCGA
					STA:	ACAAGAAAACAGAAGAGAAGCAGAA
GTA0251391	omy24_19143772	C	T	0.51	ASP1:	GTTTAAGGAGCACCTTTAGCACATC
					ASP2:	GTTTAAGGAGCACCTTTAGCACATT
					LSP:	TGCAGGCAGTCACTGAGTCC
					STA:	GTCTAAACCAACTGACCCAGATG
GTA0254169	omy24_26165023	G	A	0.46	ASP1:	CTGTGCTACTAATCACTTCTCTCTC
					ASP2:	CTGTGCTACTAATCACTTCTCTCTT
					LSP:	ACAAAACAGGTTTACACCTGGAGGA
					STA:	AGTCACCTGGGCTCTAAGTATTTA

GTA0251451	omy26_12749297	A	T	0.38	ASP1:	TGTAGAACACACACACACAAGCATA
					ASP2:	GTAGAACACACACACACAAGCATT
					LSP:	CACATCACCCCTCCAAATTGGCA
					STA:	TCGCTCACA CTCTTAGCTGT
GTA0251460	omy27_9625126	C	T	0.57	ASP1:	CGCAAAC TGGGTCAATAAACCG
					ASP2:	CGCAAAC TGGGTCAATAAACCA
					LSP:	GGAGACAGACAGGCCAACAGA
					STA:	CCATGCGACCTGAGCC

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