

Replumbing the Bengal Basin: A Shift in Recharge Driven by Groundwater Irrigation Revealed by Stable Water Isotopes

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2 ***Groundwater Irrigation Revealed by Stable Water Isotopes***

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17 **Abstract**

18 Groundwater supports agriculture and provides domestic water for over 250 million
19 people in the Bengal Basin. Our analysis of stable water isotope ratios in rain, surface, and
20 groundwater shows that the proportion of groundwater recharge originating from stagnant surface
21 water bodies has increased by about 50% over the last seventy years while the relative contribution
22 from direct infiltration of rain has decreased. This regional shift in the source of groundwater
23 shows how the simultaneous expansion of irrigated rice, excavated ponds and groundwater
24 pumping has changed the hydrologic system by cycling evaporated standing water through the
25 subsurface. Analysis of water isotope data also reveals that most recharge from standing water
26 enters during the latter part of the dry season (February-April), while most rainwater recharge
27 occurs in the early months of the monsoon (June-August) before aquifers fill to capacity and reject
28 additional recharge of rainwater.

29 **Introduction**

30 Groundwater is the main source of domestic and agricultural water for more than 250
31 million people living in the Bengal Basin of Bangladesh and India that lies in the downstream
32 floodplains and deltas of the Ganges, Brahmaputra and Meghna Rivers (1–3). Large-scale
33 abstraction of groundwater over the last six decades for dry-season agriculture has increased
34 recharge to groundwater (4–7), caused the water table to decline further during the dry season and,
35 in some parts of the basin, prevented the water table from rebounding as high during the monsoon
36 (1, 6). Pumping draws more shallow water to deeper aquifers increasing the risk of contaminating
37 relatively low-arsenic deep groundwater aquifers (8, 9) and making coastal aquifers more
38 vulnerable to seawater intrusion (10). Understanding the ramifications of large-scale pumping on
39 groundwater dynamics is important for future sustainability of both the quantity and quality of
40 groundwater, the main source for household and irrigation water in the basin (11).

41 Groundwater in the Bengal Basin has been assumed to be recharged by direct monsoon
42 precipitation (9, 12–15) largely based on the fact that rainfall exceeds potential evapotranspiration.
43 Several site-specific studies have challenged this notion by finding that shallow groundwater
44 aquifers today are recharged by rivers, local ponds and rice field water (16–20). Irrigation pumping
45 induces downward head gradients beneath these surface sources that increase as the water table
46 falls towards its dry-season minimum. Because recharge occurs while the water table is falling
47 annual recharge cannot be estimated by analysis of the rising portion of the groundwater
48 hydrograph alone. Groundwater hydrographs alone do not provide enough information to estimate
49 recharge.

50 The contribution of the different sources in the geomorphically active Bengal Basin can
51 vary spatially and temporally. Older groundwater may contain recharge from surface water bodies
52 that no longer exist because river meandering and avulsion moves channels across the landscape
53 over decades and centuries (21). For instance, the Brahmaputra flowed on the eastern side of the
54 Dhaka until it avulsed about 200 years ago and now flows about 80 km west of Dhaka (22, 23).
55 Hydrograph analyses show that groundwater recharge has increased in the last three decades due
56 to intensive groundwater abstraction for irrigation. These studies argue that the increased recharge
57 is sourced from direct rain infiltration and river water (6, 24) during the monsoon. However, prior
58 regional scale studies have not (1) quantified the contributions of different recharge sources, (2)
59 accounted for recharge from ponds and rice fields (3) analyzed how recharge sources have changed
60 over decades and centuries.

61 Regional water and land use changes suggest that ponds and rice fields could contribute
62 significantly to groundwater recharge primarily during the dry season. In the last 35 years, the
63 number of irrigation wells has increased from < 100,000 to >1.7 million (25) primarily to support
64 dry-season irrigation that has also correspondingly increased from <1 million hectares to 5 million
65 hectares (~20% of the total land) (26, 27). Groundwater abstraction in Bangladesh is estimated to
66 be >33 km³ in 2010, 95% of which is used to support dry season farming (28); a volume that can
67 fill the 5 million hectares with a water depth equaling 650 mm. The volume of groundwater
68 abstraction in Bangladesh and West Bengal (India) has only increased in recent times (25). At
69 least 10 million hand-pumped wells have also been installed in the region, but they account for
70 only a minor fraction of total pumping (29). Man-made ponds cover an estimated 2.6% of the area
71 of Bangladesh, with more than 4 million households owning a pond (30); most of these ponds are
72 recent and have been constructed in the past 70 years (31), although the number of ponds may be
73 stabilizing post-2000 (32).

74 Evidence of rapid cycling of water through shallow aquifers, combined with the recent
75 declines in the water table elevation and the growth of standing water in rice fields and ponds,
76 motivated us to investigate how groundwater recharge may have changed over time. We postulate
77 that: 1) rice fields, pond water and rivers have now become a major contributor to groundwater
78 recharge across the basin and 2) the majority of the recharge from standing water occurs during
79 the latter part of the dry season (February to April) when groundwater levels reach their minima
80 so that large vertical head gradients draw more recharge from standing water into the available
81 aquifer storage.

82 In this work, we rely on radiogenic and stable water isotopes to answer our hypotheses.
83 We first establish the regional groundwater age profile by compiling existing radiocarbon and
84 tritium groundwater data from the region. Subsequently we partition the contribution of different
85 sources recharging shallow groundwater in the Bengal Basin using stable water isotopes that are
86 used extensively for partitioning the contributions of different water sources in urban, natural and
87 ecohydrological water systems (33–36). We compiled the existing groundwater stable water
88 isotope ratios ($\delta^2\text{H}$ and $\delta^{18}\text{O}$) for Bangladesh and the eastern portion of West Bengal (bordering
89 Bangladesh) and model the groundwater stable isotope ratios as a mixture of precipitation,
90 standing water (local rivers, ponds and rice fields), and large rivers (Ganges and Brahmaputra) to
91 estimate the contribution of each source to groundwater. The contrast between isotopically
92 enriched standing water (due to evaporation) in comparison to isotopically light large Himalayan
93 rivers (Ganges and Brahmaputra and their tributaries) and unevaporated precipitation values
94 provide the necessary isotopic separation to quantify the contribution of these sources to
95 groundwater recharge. Finally, we interpret the modeling results alongside long-term groundwater
96 hydrographs from the region to develop a simplified regional recharge model.

97 98 **Data source and conceptual framework**

99 100 **Groundwater age profile of the Bengal Basin**

101 Radioactive isotopes in groundwater indicate that in the upper 100 meters of the aquifer
102 system much of the groundwater was recharged contemporaneously with the onset of groundwater
103 irrigated rice and the expansion of man-made ponds (37). Tritium active water is pervasive up to
104 a depth of 100 m (Figure 1A) indicating modern recharge (post-1953) and highlighting vigorous
105 groundwater circulation. Approximately 60% and 20% of the groundwater samples in shallow (0-
106 50 m depth) and intermediate (50-100 m depth) wells have a tritium activity of >1 TU. However,
107 groundwater with tritium levels below detection limits are also present throughout the vertical
108 profile, indicating that older stagnant water can be found at all depths. Below 100 m, tritium
109 concentrations decline, but tritium-active water is still found where data is available to a depth of
110 300 m indicating young water, less than 70 years old, although some of these wells may be
111 compromised by leaks in their casing (38, 39). Radiocarbon dating suggests that waters from below
112 100m are, on average, thousands of years old. A linear regression of uncorrected dissolved
113 inorganic radiocarbon age on depth suggests that water at 100 m depth is on average 4000 years
114 old and 10,000 years old below 200 m, reflecting recharge that predates human influence and
115 occurred most likely during early Holocene and late Pleistocene (Figure 1B).

118 **Stable water isotope data**

119 We combined stable water isotope data for precipitation, groundwater and surface water
120 (rivers, ponds and puddled water in rice fields) for the Bengal Basin from 31 sources (Table S1)
121 that together provide 580 precipitation, 1918 groundwater and 487 surface water values.
122 Precipitation samples were collected at Barasat and Kolkata in India and at Savar, Barisal, Sylhet,
123 Chittagong and Cox's Bazar in Bangladesh (15, 40–43). The distribution of precipitation stations
124 provided an extensive spatial coverage of precipitation isotope ratios from the basin (Figure 2A).
125 Groundwater samples were evenly distributed across Bangladesh and the bordering region
126 between Bangladesh and West Bengal lying to the East of the Hooghly River (tributary of the
127 Ganges, Figure 2). Some locations had high densities of groundwater isotope data as they have
128 been studied extensively for groundwater arsenic (17, 19, 44–46). The lack of isotopic data from
129 the districts west of the Hooghly River in West Bengal is largely due to the fact that they have
130 been investigated to a lesser extent for groundwater arsenic contamination.

131 For all of the investigated precipitation sampling stations across the Bengal Basin (Figure
132 2) isotope values varied seasonally with the heaviest values during the dry season (January–April),
133 followed by decline isotope ratios at the onset of monsoon and the lightest values during late
134 monsoon (September/October, Figure 3A and 3B). This seasonal pattern in the precipitation
135 isotope ratios is consistent across the Bengal Basin (43, 47). Rainfall is highly seasonal, and the
136 majority of precipitation occurs during the monsoon months of May to October (Figure 3C). The
137 stable isotope ratios of precipitation showed a wide range of values from -120 to 25‰ ($\delta^2\text{H}$) and -
138 15 to 5‰ ($\delta^{18}\text{O}$). The amount-weighted mean $\delta^2\text{H}$ and $\delta^{18}\text{O}$ of annual precipitation is -46.5‰ and -
139 -6.9‰ respectively. We define the local meteoric water line (LMWL, $\delta^2\text{H} = 8.2 \delta^{18}\text{O} + 11.2$, $R^2 =$
140 0.97 , $p < 0.05$) as the best-fit regression line through all the precipitation data.

141 We divided surface water samples in groups representing (a) large rivers (Ganges and
142 Brahmaputra) and (b) standing water (small rivers, ponds and rice fields). The Bengal Basin is a
143 deltaic system with an intricate network of streams and rivers. In this analysis, by large rivers we
144 refer to the main Ganges and Brahmaputra channel as well as the tributaries feeding from the main
145 channels – a large proportion of water in these rivers is derived from Himalayan snowmelt and
146 higher altitude precipitation (48). This water is isotopically lighter (49) than the amount weighted
147 mean precipitation isotope ratios of the Bengal Basin (Figure 4A). We define standing water as
148 waterbodies that are recharged primarily from the local precipitation in the Bengal Basin and
149 subsequently undergo evaporative enrichment during the dry season. This includes: 1) Ponds that
150 are filled up during the monsoon and are depleted by human consumption and evaporation during
151 the dry season, 2) Small rivers that do not contain water from the large rivers and may become
152 stagnant during the dry season, 3) Standing water in irrigated rice fields.

153 Samples from large rivers (Ganges and Brahmaputra) were isotopically lighter than the
154 standing water samples (Figure 4A), falling around the LMWL and clustering around the point -
155 50‰ ($\delta^2\text{H}$) and -7.4‰ ($\delta^{18}\text{O}$). Seasonal data from the Ganges were typically lighter and show a
156 subdued isotopic variation compared to local precipitation suggesting significant contribution from
157 isotopically depleted Himalayan tributaries (48). Multi-year $\delta^{18}\text{O}$ isotope values from the
158 Brahmaputra fell between -12 and -6‰ (mean = -8‰) with heavier values observed during April
159 and depleted values observed during late monsoon due to large contributions from snow and
160 glacier melt and high-elevation precipitation (49). The slope of the river water line is >7 (15, 48,
161 49) suggesting negligible evaporative enrichment of stable water isotopes in these rivers.

162 In contrast to large rivers, standing water samples showed a large range of isotope values
163 (Figure 4A). Seasonal measurements suggests that standing water undergoes progressive

164 evaporative enrichment during the dry season (November to April) and the heaviest isotope values
165 were observed towards the end of the dry season in March/April (13, 19). The evaporation slope
166 obtained by regressing through the standing water data was 5.6 ($\delta^2\text{H} = 5.6 \delta^{18}\text{O} - 7.6$, $R^2 = 0.90$, p
167 < 0.05 , Fig 4A), very similar to the theoretically calculated evaporation slope (between 5 and 6) in
168 the Bengal Basin (50) indicating evaporative enrichment of standing water bodies.

169 The stable isotope ratios of groundwater showed a wide range of values from -64.0 to 5.3‰
170 ($\delta^2\text{H}$) and -9.6 to 0.1‰ ($\delta^{18}\text{O}$). The mean groundwater $\delta^{18}\text{O}$ in shallow aquifers (0-50 m deep) are
171 heavier (mean $\pm \sigma = -4.1 \pm 1.3\%$) than intermediate (50-100 m, $-4.6 \pm 1\%$) and deep (>100 m, $-$
172 $4.7 \pm 1\%$) aquifers. While some shallow groundwater (0-50 m depth) samples cluster near the
173 LMWL, many samples also fall below this line indicating groundwater recharge from both
174 unevaporated and evaporated sources (Figure 4B). Intermediate (50-100 m depth) and deep (>100
175 m depth) groundwater samples cluster near the LMWL suggesting recharge dominantly from
176 unevaporated water (Figure 4C and 4D). Deuterium excess (d-excess = $\delta^2\text{H} - 8 \times \delta^{18}\text{O}$) ranged
177 from -17.5 to 18.6, with a mean value of 6.4‰. Mean d-excess of shallow (5.3‰), intermediate
178 (6.9‰) and deep (7.5‰) aquifers was lower than the mean modern precipitation d-excess (9.9‰).
179

180 **Groundwater isotope mixing models**

181 Groundwater in the Bengal Basin can be recharged by multiple sources including direct
182 rainfall, large rivers (Ganges, Brahmaputra and their tributaries), bodies of standing water
183 (irrigated rice fields and ponds), small rivers, or a combination of these sources. The contribution
184 from the different sources can vary locally depending upon the precipitation, proximity of the
185 aquifer to rivers, the hydrological connectivity between the rivers and the aquifers, and the density
186 of standing water bodies (ponds and rice fields) above the aquifer. Furthermore, these sources and
187 their relative contributions have likely varied over time due to changes in climate, land use, and
188 water use.

189 These sources can be divided into three isotopic endmembers. First, the isotopically light
190 water in Ganges and Brahmaputra and their tributaries (Figure 4A). Second, precipitation isotope
191 ratios ($\delta^2\text{H}$ and $\delta^{18}\text{O}$) that exhibit a large range, however, are strongly correlated and fall on or
192 close to the LMWL (Figure 3A). Third, ponds, local rivers and rice fields isotope ratios that show
193 strong seasonal variation. Samples collected from ponds and local rivers during the monsoon
194 season lie close to the LMWL and undergo evaporative enrichment during the non-monsoon
195 season (November to April) which makes them isotopically heavier than local precipitation (Figure
196 4A). Similarly, standing water in rice fields (pumped from the groundwater) undergo evaporative
197 enrichments during the dry season and gets evaporatively enriched over time. Thus, there is a clear
198 distinction between non-evaporated water sources (large rivers and precipitation) and evaporated
199 standing water sources (local rivers, ponds and rice fields, Figures 4 and S3). With this variation
200 in the isotopic signature ($\delta^2\text{H}$ and $\delta^{18}\text{O}$) of the endmembers, we model the proportion of
201 groundwater isotope ratios as a mixture of precipitation, standing water and large rivers.

202 Several combinations of precipitation, river and standing water isotope values can explain
203 the observed isotope ratios of groundwater samples (Figure S3B and 5). Therefore, we use a
204 Monte-Carlo approach to find non-parametric distributions that describe possible mixes of
205 recharge reflected by groundwater isotope values. We constructed three different mixing models,
206 each with uniquely different assumptions about the source of recharge, to test the sensitivity of the
207 results to these assumptions. Complete details are provided in Methods section.

208 In the first model, groundwater isotope ratios are modeled as a linear mixture of a
209 precipitation and standing water sample (Figure 5A), without any input from large rivers (Ganges

210 and Brahmaputra). This assumes that groundwater is recharged primarily from precipitation and
211 local surface waterbodies. The logic behind this assumption is that for regions not lying in the
212 river floodplain it is unlikely that rivers contribute significantly to shallow groundwater recharge.
213 Additionally, for sites in proximity to the Ganges and Brahmaputra, groundwater flow is toward
214 the river, except during the latter part of the dry season when flow may be reversed as the
215 groundwater head falls below the river water level due to extensive pumping (51, 52). During the
216 early monsoon, some local recharge along the Brahmaputra may take place driven by a rising river
217 stage from snowmelt in the headwaters. Water level in Ganges rises later in July (6), hence it does
218 not contribute to early monsoon recharge.

219 In the second model, we considered a 3-endmember mixing model with large rivers,
220 precipitation, and standing water as the potential sources (Figure 5B) – any combination of these
221 sources could recharge the groundwater. The third model is also a 3-endmember model with a
222 standing water endmember and two precipitation endmembers (Figure 5C), but no river water. The
223 idea behind the third model is to account for the possibility that groundwater samples contain a
224 mix of two, rather than just one, precipitation sources with distinct isotopic compositions. Because
225 precipitation isotope ratios vary systematically during the monsoon and transition from heavier to
226 lighter $\delta^{18}\text{O}$ and $\delta^2\text{H}$ values from early to late monsoon, this model can quantify the relative
227 contributions of each, and in doing so, the timing of recharge by precipitation. This model is similar
228 to model 1 in regard to the source endmembers used in the model but generates an independent
229 measure of contributions from standing water.

230

231 Results

232 We observed a large scatter in the $\delta^{18}\text{O}$ and d-excess profile in shallow wells with values
233 ranging from -7 to 0‰ ($\delta^{18}\text{O}$) and -2 to 12‰ (d-excess, Figure 6C). For the intermediate wells,
234 the range of $\delta^{18}\text{O}$ and d-excess starts to taper with depth and the range of $\delta^{18}\text{O}$ and d-excess values
235 at 100 m is much smaller than the range at 50 m. Indeed below 75 m, $\delta^{18}\text{O}$ ranged from -6 to -3‰
236 and barring a handful of samples d-excess was mostly >6‰. The large range of $\delta^{18}\text{O}$ and d-excess
237 in the shallow wells suggests that groundwater is recharged by a diverse suite of sources and during
238 different time of the season. In contrast, the lower spread in $\delta^{18}\text{O}$ and d-excess for intermediate
239 and deeper wells suggests recharge is dominated by a single source.

240 The moving depth-average groundwater $\delta^{18}\text{O}$ decreased from -3.8‰ to -5.1‰ between 7
241 m and 100 m deep (Figure 6A). In the same depth interval, deuterium-excess increased from 4.6‰
242 to 7.2‰. Between 100 m and 250 m, the moving depth-average groundwater $\delta^{18}\text{O}$ increased from
243 -5.1‰ to -3.8‰. The corresponding d-excess values did not change appreciably, increasing
244 slightly from 7.2‰ to 7.9‰ (Figure 6C).

245 All three models, despite with their different assumptions, result in similar estimates of
246 standing-water contribution. Model 2, which includes river input, shifts some of the estimated
247 rainwater contribution to the river as both are unevaporated sources (i.e., lying on the LMWL) but
248 still yields about the same contributions of the evaporated standing water source over depth.
249 Precipitation is the largest contributor to groundwater recharge in all the mixing models. The
250 fractional contributions of precipitation are 0.70 ± 0.25 , 0.50 ± 0.26 , and 0.74 ± 0.22 (mean $\pm \sigma$) in
251 model 1 (precipitation and standing water), 2 (precipitation, river and standing water) and 3 (two
252 precipitation and standing water), respectively. The mean contribution of standing water is
253 0.30 ± 0.25 , 0.28 ± 0.22 and 0.26 ± 0.24 in models 1, 2 and 3, respectively. The mean contribution of
254 rivers is 0.22 ± 0.18 in model 2.

255 The moving depth-average suggests that the mean contribution of the different sources
256 varies with depth (Figure 7). The contribution of non-evaporated sources (precipitation in model
257 1, precipitation and river in model 2, and the two precipitation sources in model 3) increases
258 moving down from 7 meters to 100 meters depth, with proportions increasing from 0.60 to 0.74
259 (model 1), 0.63 to 0.77 (model 2), and 0.62 to 0.75 (model 3). The contribution from standing
260 water at a depth of 7 m was more than 1.5 times the contribution at 100 m, with values at 7 and
261 100 m respectively of 0.43 and 0.26 (model 1), 0.37 and 0.23 (model 2), and 0.38 and 0.25 (model
262 3). Between depths of 100 and 250 m, the contribution of evaporated and non-evaporated sources
263 exhibited more subdued changes with depth (Figure 7). Contributions of non-evaporated sources
264 increased slightly between depths of 100 to 250 m, with values at 100 and 250 m respectively of
265 0.74 and 0.81 (model 1), 0.77 and 0.82 (model 2), and 0.75 and 0.80 (model 3). Correspondingly,
266 the contributions of standing water decreased between depths of 100 m and 250 m, changing from
267 0.26 to 0.19 (model 1), 0.23 to 0.18 (model 2), and 0.25 to 0.20 (model 3). We did not observe any
268 spatial pattern in the contribution of the different sources (see section “Spatial pattern in fractional
269 contribution of the different sources” in the SM for details).

270 Discussion

271 Temporal changes in groundwater isotope ratios

272 The distinct trend in the depth-average groundwater isotope ratios ($\delta^{18}\text{O}$ and $\delta^2\text{H}$, Figure
273 6) where it decreases from 7 m depth to 100 m depth and subsequently increases from 100 m depth
274 to 267 m depth suggests that either the relative contribution of the isotopically distinct sources
275 recharging groundwater has changed over time or the isotope ratios of the sources recharging
276 groundwater have changed. In the Bengal Basin, it appears that both factors have contributed to
277 variations in groundwater isotope ratios over time.

278 Enriched $\delta^{18}\text{O}$ and $\delta^2\text{H}$ values in shallow aquifers were associated with lower d-excess and
279 vice-versa ($r = -0.6$); the inverse relation between the stable isotopes ($\delta^{18}\text{O}$ and $\delta^2\text{H}$) and d-excess
280 provides evidence that the pattern in moving depth-average groundwater isotope ratios between 7
281 to 100 m depth is primarily due to increase in contribution of unevaporated sources as we move to
282 deeper depths. In other words, more recently recharged water (i.e., shallower depths) have higher
283 contribution from evaporated sources (ponds, rice fields and local rivers during the dry season)
284 and deeper water have lower contribution from evaporated sources. It is very unlikely that these
285 patterns in the top 100 meters of groundwater represent changes in precipitation isotope ratios
286 because most of this water recharged in the last millennium, much more recently than the last
287 changes in rainfall isotope ratios that occurred about 10,000 years ago (45, 53).

288 Between 100 m and 267 meters depth, $\delta^{18}\text{O}$ and $\delta^2\text{H}$ reversed the trend observed between
289 7 and 100 m depth and become progressively enriched. However, the trend in d-excess did not
290 reverse; instead, it increased nominally by 0.8‰ (Figure 6) suggesting that the changes in
291 groundwater recharge are not driven by increase in contribution of evaporated water. Groundwater
292 below 100 m is, on average older than 4000 years and below 200 m is, on average, older than
293 10,000 years (Figure 1B). In the Bengal basin, precipitation isotopes in the Pleistocene were
294 isotopically heavier than in Holocene precipitation (53). Thus, the observed pattern in the depth-
295 average groundwater isotope ratios is most likely due to changes in the precipitation endmember
296 values; the groundwater isotope ratios become progressively heavier with increasing depth as the
297 component of older Pleistocene precipitation water increases.

298

299 **Changes in groundwater recharge sources over time**

300 There is a broad consensus that large-scale pumping of groundwater has altered
301 groundwater dynamics in the Bengal Basin by lowering the water table below large cities and
302 intensively irrigated areas, drawing modern water into deeper aquifers and reversing stream-
303 groundwater exchange during the dry-season (3–5, 8, 18, 51, 54). Although field studies have
304 suggested changes in recharge sources in response to groundwater pumping – mostly increased
305 contribution from rice fields and ponds (16–19, 55) - a basin-wide shift in the sources recharging
306 groundwater has not been reported. Previous and recent basin scale work has mostly focused on
307 diffused and focused recharge during monsoon (6, 24) and have failed to explicitly consider the
308 possible recharge from ponds and rice fields at regional level.

309 The trends in isotopic composition are mirrored by changes in source attribution between
310 shallow, recent groundwater and older, deeper groundwater and provide clear evidence that recent
311 human perturbations have affected groundwater recharge sources on a massive scale. In all the
312 models, we observed a sharp transition in the contribution of standing water at the depth of 75 m
313 (Figure 7). The standing water contribution decreases as we move downwards from 7 m to 75 m.
314 However, between 75-100 m depth, the fractional contribution of the different sources does not
315 change appreciably. These differences in contribution with depth suggest that the proportion of
316 recharge from the different sources recharging the groundwater has changed over time.

317 We interpret our results within the framework of the available groundwater dating (C14
318 and tritium concentration) to suggest that the relative contributions in recent times have shifted to
319 contain more standing water sources that have been subjected to evaporation. Between 7 and 100
320 m depth, where the contribution of evaporated water is greatest, groundwater is dominantly
321 modern, with more than 80% samples containing high tritium levels. The average tritium
322 concentration decreases with depth providing evidence that the more recent the recharge is (i.e.,
323 shallower depths), the higher is the contribution of evaporated water. Recent work have suggested
324 that total recharge has increased in the Bengal Basin in response to widespread pumping, however
325 these studies could not partition the contribution from the different sources (24). Groundwater
326 isotopic data and our mixing model suggests that the increased groundwater abstraction have not
327 only increased the amount of recharge but have also drastically altered the sources recharging
328 shallow groundwater aquifer in the Bengal Basin; evaporated water recharge during the dry season
329 (discussed subsequently) have become a major recharge source of shallow groundwater.

330 Below 100 m depth, changes in $\delta^{18}\text{O}$ and $\delta^2\text{H}$ is not mirrored by corresponding changes in
331 the contribution from the respective sources. Water below 100 m is typically >4000 years old
332 suggesting recharge that predates human influence. Therefore, the proportional contribution of
333 evaporated and non-evaporated sources has remained consistent even though the source water
334 isotope ratios changed over time. Overall, it appears that in the last 70 years the contribution of
335 mean contribution of standing water has steadily been increasing with time. For water recharged
336 before large-scale human perturbation, during late Holocene to Pleistocene, the respective
337 contribution of the sources has not changed appreciably.

338 This shift to increased fractions of standing water about 70 years ago coincides with the
339 shift to groundwater-irrigated dry-season rice farming and the growth of pond excavation (31),
340 both of which reflects the broader economic and population changes in the region. To our
341 knowledge this is the first study to show that large-scale groundwater pumping has altered the
342 source contributions to groundwater recharge in the Bengal Basin, with standing water in rice
343 fields, ponds and local rivers are now acting as major recharge sources.

344 The large decline (~1.6 m) in the annual minimum water table depth across much of Bengal
345 Basin (6) caused by groundwater extraction likely result in large vertical hydraulic gradient that
346 draws recharge in from the surface water bodies. Dry-season irrigation in Bangladesh increased
347 from <1 million hectares to 5 million hectares between 1975 and 2010 (26). Irrigated rice fields
348 now cover 21% of the land area and standing water in these field recharges groundwater as
349 irrigation return flow – groundwater pumped for irrigation that returns back to aquifers. Neumann
350 et al. (27) showed that return flow recharges aquifers by passing through macropores beneath
351 bunds (the raised berms that bound rice fields). This recharge circumvents the low-permeability
352 plough pan, the near-surface layer that restricts flow directly through rice fields. They estimated
353 the water budget for rice fields using an average ratio of field perimeter (bund length) to field area
354 and found that rice fields recharge in excess of 100 cm of water per year. Other field level analysis
355 also suggests significant water loss from rice fields from seepage and percolation. Working in the
356 Barind tract (northwest Bangladesh) Qureshi et al. (58) reported that during the dry season only
357 55% of the total applied water to rice fields was lost to evapotranspiration and the remaining 45%
358 was lost to seepage and percolation (56). On a seasonal basis, the authors estimated that net amount
359 of seepage and percolation were 34 and 60 cm for the dry and monsoon season respectively (56).

360 Similarly man-made ponds (31, 57) can recharge shallow groundwater. Stahl et al., (19)
361 have shown that the presence of terrestrial crab burrows in pond-beds can short-circuit low-
362 permeability surface sediments and provide widespread conduits for groundwater recharge; they
363 estimated the recharge flux to be 223 cm/year at their field site.

364 Although on an annual time scale there is a net discharge from aquifers to the rivers (58),
365 local rivers can also recharge shallow aquifers in response to lowering of groundwater level (24,
366 51) especially in the dry season when groundwater heads fall below the river head leading to flow
367 reversal that can recharge the aquifer or during the early monsoon when river and stream stage rise
368 precedes the rise in groundwater levels. Recharge from riverbeds is more complex than recharge
369 from ponds and rice fields. There are two modes of recharge from small rivers: (1) recharge during
370 the dry season when groundwater levels fall; (2) recharge in the early monsoon when river levels
371 rise faster than groundwater levels. During the dry season small river behave like other standing
372 water sources and have an isotopic signature indicating evaporation. This type of recharge is
373 clearly identified as standing water from its isotope values. During the monsoon, however, small
374 rivers act as conduits for recharge of precipitation and thus the isotopic signature matches that of
375 early monsoon precipitation. This type of recharge, although passing through riverbeds, is
376 identified as coming from precipitation in our model.

377

378 **Timing of recharge from precipitation and standing water**

379 Groundwater isotope ratios in temperate regions are isotopically similar to the precipitation
380 amount-weighted isotope ratios (59–61). In the tropics, groundwater isotope ratios have been
381 shown to be isotopically lighter than the amount-weighted precipitation isotope ratios, which has
382 been interpreted as a recharge bias toward large rainfall events (62). However, unlike the
383 commonly observed pattern in the tropics (62), a majority of shallow and intermediate
384 groundwater isotope ratios (Figure 3) in the Bengal Basin are heavier than the modern amount-
385 weighted mean precipitation isotope ratios (-46.5‰ and -6.9‰ for $\delta^2\text{H}$ and $\delta^{18}\text{O}$ respectively).

386 Our mixing models suggests that the combination of precipitation to recharge is skewed
387 towards heavier values along the meteoric mixing line (Figure 8). For all of the models, 70% of
388 the precipitation isotope ratios estimated to contribute to recharge were heavier than -6.9‰ (for
389 $\delta^{18}\text{O}$) - the amount weighted mean precipitation $\delta^{18}\text{O}$ value. Therefore, our modeling suggest that

390 groundwater is not recharged evenly across the monsoon season and is seasonally biased toward
391 early monsoon rainfall, which is typically heavier than late monsoon precipitation (Figures 3A and
392 3B).

393 For all the models, the contribution from isotopically-light standing water sources ($\delta^{18}\text{O} <$
394 6‰) is low ($<10\%$) and the proportion of isotopically enriched ($\delta^{18}\text{O} > -2\text{‰}$) standing water values
395 is $> 50\%$ (Figures 8B1-B3) suggesting that standing water recharges groundwater mostly during
396 the latter part of the dry season when the standing water bodies have become isotopically enriched
397 due to evaporation. Groundwater levels are lowest during the latter part of the dry season, which
398 leads to an increase in the difference between surface water and groundwater levels. The increased
399 downward head gradient during this period draws more recharge from standing water into
400 available aquifer storage. This process of induced recharge from ponds and canals is widely
401 observed in many regions with intensive groundwater pumping including the High Plains and the
402 Mississippi Alluvial aquifer systems of the US as well as the Indus Basin (63–65).

403 Monthly groundwater hydrograph data from 1230 groundwater wells in Bangladesh from
404 2000-2013 supports the isotopic derived inference on the timing of groundwater recharge.
405 Typically, groundwater levels are the lowest in April-May and highest in September-October.
406 With these groundwater hydrographs, we considered rates of net discharge during the dry season
407 (November to April) and rate of net recharge during the monsoon (April to October). During the
408 dry season, the groundwater hydraulic head decreases rapidly from October to January ($>50\%$ of
409 the total decline in groundwater level) for most of the wells in response to natural discharge and
410 groundwater abstraction (Figure S6A). Even though a large proportion of groundwater abstraction
411 for dry season farming occurs between December and April (28), the decrease in head between
412 February and April is smaller compared to the period from October to January (Figures S6B, S6C
413 and S6D), suggesting that a portion of pumped water might be infiltrating back – a phenomena
414 suggested by field experiments (27, 56). It is important to note that a falling water table does not
415 imply that recharge is not occurring, but rather indicates that discharge is in excess of recharge at
416 the time. Thus, during periods of high levels of discharge (e.g., during the irrigation season) total
417 recharge may actually increase, while groundwater levels will nonetheless fall if discharge exceeds
418 recharge. In fact, in the case of induced recharge from standing bodies of water, a falling water
419 table will increase the downward head gradient and lead to greater recharge from standing water.

420 During the monsoon, the data suggests that most of the well experience their greatest
421 hydraulic head increase between June and early August (Figure S7A and S7B) suggesting that a
422 large proportion of net recharge happens during the early part of monsoon – thus, verifying the
423 results obtained from the isotope mixing models. The head change in September is negligible
424 (Figure S7C) across the country except in some northwestern parts which are experiencing long-
425 term decline in groundwater level due to massive abstraction and reduction in precipitation (66,
426 67). The lack of rise in groundwater level in September suggests little recharge from precipitation.
427 Similarly, in October, heads do not change for several wells suggesting even less recharge from
428 precipitation as compared to the earlier months; for wells lying in the floodplains of Ganges and
429 Brahmaputra the head in fact goes down as groundwater starts discharging to the river (Figure
430 S7D).

431

432 **Regional model of shallow groundwater recharge**

433 Our analysis of the combined isotope data base supports a conceptual model of modern
434 groundwater recharge in the Bengal Basin that divides groundwater recharge into three phases: (1)
435 dry season (November-April), (2) early monsoon (May-August) and (3) late monsoon (September-

436 October). These phases of the seasonal hydrologic cycle are similar to the results of Harvey et al.,
437 small-scale study (18) based on measured heads and water levels in one village. However, Harvey
438 et al did not distinguish between recharge from precipitation and recharge from standing water, an
439 important focus of this regional study.

440 (1) During the dry season, the groundwater head falls because of pumping and discharge
441 to rivers (6, 68, 69) and the resulting increase in vertical head gradients draws more recharge from
442 standing water into aquifers even as the water stored in aquifers decreases (Figure 9A). The
443 proportion of recharge from different sources depends on several factors including water table
444 drawdown, distance from rivers, patterns of hydraulic conductivity, and availability of standing
445 water sources.

446 The depth of the dry season decline in the water table elevation has increased over recent
447 decades because of groundwater extraction. For rice field irrigation, which makes up more than
448 85% of groundwater pumping (70), the amount of dry-season decline in the water table due to
449 irrigation pumping is limited by the potential evaporation of rice fields. Groundwater pumped for
450 irrigation circulates through rice fields, and what is not evapotranspired recharges the aquifer as
451 return flow. Although drawdown is limited by the maximum rate of evapotranspiration, the rate of
452 recharge from irrigated fields that returns to aquifers is not constrained and will increase with
453 increasing rates of irrigation.

454 This circulation process can be formulated with a simple water balance: $Q=RF+ET$, where
455 Q is the extraction rate, RF is return flow, and ET is evapotranspiration, all in units of volume flux
456 of water per unit area of rice field. The decline of the water table is $dH/dt = (Q -FR)/S_y = ET/ S_y$
457 where S_y is the specific yield. This formulation assumes that irrigated fields are not simultaneously
458 drained to rivers. Since rice is grown in standing water, and rice farming techniques used in the
459 Bengal Basin are not the most efficient (71), a surplus of water (i.e. RF) recharges back to the
460 aquifer. An important implication for groundwater isotopes is that enrichment may continue as
461 groundwater is circulated. Over decades, groundwater may become progressively even more
462 enriched in heavy isotopes as it is pumped to the surface, subjected to evaporation, and returned
463 back to aquifers.

464 (2) During the early monsoon the unsaturated zone above the unconfined water table is
465 filled by monsoon precipitation (Figure 9B). During this phase, maximum recharge by
466 precipitation takes place, as evidenced from the distribution of precipitation isotope ratios
467 estimated to be in groundwater by our model (Figure 8) and the rise in groundwater hydrographs
468 (Figure 9). Independent analysis of groundwater hydrographs also suggests a rapid rise in
469 groundwater levels during this period (6, 25). Recharge occurs across the entire basin and in some
470 regions (such as northeast, Figure 9) maximum groundwater levels are observed during this time
471 (6). Estimates of the mass of total water storage from GRACE and groundwater storage from wells
472 during 2002-2010 (72, 73) indicate that on average 85% of the water is recharged by the end of
473 July.

474 (3) In the late monsoon, groundwater recharge varies regionally, but is generally small. In
475 eastern parts of Bangladesh (light green region in Figure S7) where shallow groundwater aquifers
476 have already been recharged to a large extent during phase 2 the remaining space -if any- in the
477 shallow aquifers is recharged in the early part of phase 3 (Figure 9C). Groundwater levels in the
478 shallow aquifers in these regions are highest in August or early September (6). Precipitation and
479 flooding events during this time do not recharge the aquifers and these waters are rejected as the
480 shallow aquifers have already been filled. This region not only receives more precipitation than
481 the western parts (Figure S1) of the basin, but also has a lower proportion of irrigated area that

482 relies on groundwater (25). Other regions, mostly the western parts of Bangladesh and West
483 Bengal, receive lower rainfall (Figure S1), are less prone to flooding, and are intensively farmed
484 and pumped for irrigation (25). In these regions (shown in light gray in Figures S7C and S7D),
485 where annual abstraction exceeds the groundwater recharge (6, 25, 74), the aquifer continues to be
486 recharged even during the late monsoon season.

487 488 **Implications for groundwater quantity, groundwater quality and human** 489 **habitability**

490 Changes in the contribution from different sources suggests that the flow path of water
491 entering the subsurface had changed. Large scale shifts in the flow paths would likely have major
492 effects on the geochemistry of groundwater. Regional increases in contributions from ponds, rice
493 fields, and rivers could threaten the regional quality of shallow groundwater. Circulating irrigation
494 return flow may contaminate groundwater with increasing levels of solutes from rice field. Input
495 of reactive organic carbon has been implicated in arsenic mobilization within aquifers (17, 18, 55,
496 75, 76), although the source of this reactive carbon remains an area of active inquiry. Small-scale
497 field studies have documented pollutants transported from ponds to drinking water wells (77–80),
498 there is a growing concern of nitrate pollution in surface and groundwater (81), and local
499 groundwater quality assessments have documented *E. coli*, major ions, trace elements and organic
500 compounds (82, 83, 92, 84–91).

501 The Indian subcontinent accounts for more than 25% of the total global groundwater
502 withdrawal and several aquifers are experiencing rapid declines in groundwater levels or reduction
503 in groundwater quality (84, 85). A variety of studies in the subcontinent such as (9, 24, 57, 64, 95–
504 97) has focused on understanding the rate of recharge and the chemical load carried in recharge,
505 regional groundwater depletion and sources recharging groundwater, however a systematic
506 analysis on the effect of large-scale pumping and dry-season farming on the timing and sources of
507 recharge across the Indus-Ganges-Brahmaputra Basin is missing. Understanding how intensive
508 groundwater pumping has affected, and will continue to affect, both the sources and timing of
509 groundwater recharge is essential to understanding and managing the regional hydrologic system.

510 Our isotopic analysis provides evidence that across much of the eastern and southern
511 Bengal Basin late monsoon precipitation does not recharge groundwater, likely because aquifers
512 are full by August so that precipitation in September-October is lost as runoff. This “rejected
513 potential recharge” was the focus of Roger Revelle’s seminal paper (98) that promoted the idea of
514 reducing monsoon flooding by pumping groundwater to lower the water table during the dry
515 season so that more precipitation would infiltrate during the monsoon rather than contribute to
516 flood water. Local hydrological modeling and field studies have arrived at a similar conclusion (6,
517 24, 25, 99). Our isotope analysis suggest that increased pumping could be sustained in many parts
518 of the basin where precipitation is sufficient to return the water table to the land surface every
519 monsoon.

520

521 **Methods: Monte-Carlo Simulations**

522 **Model 1: Mixtures of precipitation and standing water**

523 We model groundwater isotope ratios as a mixture of precipitation and standing waters
524 (i.e., ponds and rice fields):

$$525 \delta^{18}O_{gw} = f_{precip}\delta^{18}O_{precip} + f_{stand}\delta^{18}O_{stand}$$

526 $\delta^2H_{gw} = f_{precip}\delta^2H_{precip} + f_{stand}\delta^2H_{stand}$
 527 where $\delta^{18}O_{gw}$ and δ^2H_{gw} are the isotope values of the groundwater, $\delta^{18}O_{precip}$ and δ^2H_{precip}
 528 are the isotope values of precipitation and $\delta^{18}O_{stand}$ and δ^2H_{stand} are the isotope values of the
 529 evaporated standing water bodies. f_{precip} and f_{stand} are the fractional contribution of precipitation
 530 and standing water to the groundwater and $f_{precip} + f_{stand} = 1$.

531 We simulated all pairs of groundwater and standing water samples. We first obtain the
 532 $\delta^{18}O_{precip}$ and δ^2H_{precip} endmember values defined as the intersection between the local meteoric
 533 water line (LMWL) and the line joining the groundwater and the standing water sample in the dual
 534 isotope space (see Figure 5A). Then, f_{precip} and f_{stand} were calculated as:

$$536 \quad f_{precip} = \frac{\delta^{18}O_{precip} - \delta^{18}O_{gw}}{\delta^{18}O_{gw} - \delta^{18}O_{stand}}$$

$$537 \quad f_{stand} = 1 - f_{precip}$$

538 After calculating the contribution of precipitation and standing water endmembers for each
 539 pair (Figure 5A), we calculated the mean to obtain the average contribution of precipitation and
 540 standing water endmembers respectively.

541

542 **Models 2 and 3: Mixtures of three recharge sources**

543 In the 3-endmember second and the third mixing models, we applied a Monte Carlo method
 544 (36, 100–103) to estimate probability distributions of the contributions of different sources
 545 recharging groundwater. This method entails more steps than model 1, that had only the two
 546 endmembers. The Monte Carlo simulations each draw from the large data set of precipitation and
 547 standing water isotope data to fully cover the probabilistic range of values as characterized by the
 548 available data.

549 Below, we describe the isotope values of each endmember. For the river endmember, we
 550 chose a fixed value ($\delta^{18}O_{riv} = -7.4\text{‰}$ and $\delta^2H_{riv} = -50\text{‰}$), as the isotopic variation in the river values
 551 were negligible (Figure 4A). We calculated amount-weighted mean precipitation isotope ratios by
 552 weighting each sample by its proportional contribution to total annual precipitation. We assumed
 553 that precipitation isotope values follow a bivariate normal distribution with values $\{m_{2-H}, m_{18-O},$
 554 $\sigma_{2-H}, \sigma_{18-O}, \rho\}$ of $\{-46.5, -6.9, 31.2, 3.7, 0.9\}$ where $m_{2-H}, m_{18-O}, \sigma_{2-H}, \sigma_{18-O}$ are the amount-
 555 weighted mean and standard deviation of the precipitation isotope values and ρ is the correlation
 556 between the O and H precipitation isotopes. For the standing water endmember, we assumed that
 557 all the standing water samples included in this study ($n = 447$) are representative of the true
 558 distribution of the standing water isotope values and we treated each standing water isotope value
 559 as a potential endmember.

560 For each standing water sample, we simulated 500 random draws of precipitation isotope
 561 values to obtain precipitation endmember values. For each groundwater sample, we repeated the
 562 above process of simulating random draws of precipitation isotope values for each of the 447
 563 standing water samples. In total, our approach generated $>200,000$ (447 standing water values
 564 multiplied by 500 precipitation values) possible “mixing triangles” for each groundwater sample.

565 For each combination of standing water, river and precipitation endmembers, where the
 566 endmember values were solvable ($0 \leq \{f_{precip}, f_{riv}, f_{stand}\} \leq 1$ and $f_{precip} + f_{riv} + f_{stand} = 1$), we
 567 obtained the fractional contribution of the endmembers (Figure 5B). When a groundwater sample
 568 fell outside the triangular domain (“mixing triangle”) defined by the three end members (Figure

569 5B), we discarded this combination and did not calculate f_{precip} , f_{riv} , f_{stand} as they would not fulfill
 570 the criteria: $0 \leq \{f_{precip}, f_{riv}, f_{stand}\} \leq 1$. After calculating the contribution of precipitation, river
 571 and standing water endmembers for each solvable triangle (Figure 5), we calculated the mean to
 572 obtain the average contribution of precipitation, river and standing water endmembers
 573 respectively.

574 We only included groundwater samples that fell below the LMWL and were not highly
 575 evaporatively enriched. For groundwater samples, falling on or above the LMWL (14% of the total
 576 groundwater sample), the contribution of surface water is minimal, and we assumed that the
 577 groundwater sample is completely recharged from precipitation i.e., $f_{precip} = 1$ and $f_{stand} = f_{riv} = 0$.
 578 Similarly, highly evaporated groundwater samples (deuterium excess $< 0\%$, 5% of the total
 579 groundwater sample) also typically fall outside the mixing triangle because the groundwater
 580 isotope ratios will have heavier isotope values than most of the standing water endmember values;
 581 this will result in only a small number of plausible mixing triangles. We assumed these evaporated
 582 samples to be recharged solely from standing water i.e., $f_{stand} = 1$ and $f_{riv} = f_{precip} = 0$. These
 583 assumptions are valid as it is very unlikely that samples falling above the LMWL are recharged by
 584 standing water. Similarly, highly evaporated groundwater samples are likely to be recharged
 585 primarily by standing water.

587 **Model 2: Mixtures of one precipitation event, river and standing water**

588 In the second model, we considered a 3-endmember mixing model with precipitation, large
 589 rivers and standing water as the potential sources (Figure 5B). Groundwater sample isotope ratios
 590 ($\delta^{18}O_{gw}$ and δ^2H_{gw}) were modeled as a mixture of precipitation, river and standing waters
 591 endmembers:

$$592 \quad \delta^{18}O_{gw} = f_{precip}\delta^{18}O_{precip} + f_{riv}\delta^{18}O_{riv} + f_{stand}\delta^{18}O_{stand}$$

$$593 \quad \delta^2H_{gw} = f_{precip}\delta^2H_{precip} + f_{riv}\delta^2H_{riv} + f_{stand}\delta^2H_{stand}$$

594 where $\delta^{18}O_{gw}$ and δ^2H_{gw} are the isotope values of the groundwater, $\delta^{18}O_{precip}$ and δ^2H_{precip}
 595 are the isotope values of precipitation, $\delta^{18}O_{riv}$ and δ^2H_{riv} are the isotope values of the large rivers
 596 and $\delta^{18}O_{stand}$ and δ^2H_{stand} are the isotope values of the evaporated standing water bodies. f_{precip} , f_{riv}
 597 and f_{stand} are the fractional contribution of precipitation, large rivers and standing water to the
 598 groundwater and $f_{precip} + f_{riv} + f_{stand} = 1$. After defining the isotope ratios of each endmember, we
 599 solve the above equations to obtain the fractional contributions of the endmembers (f_{river} , f_{precip} , and
 600 f_{stand}).

602 **Model 3: Mixtures of two precipitation events with standing water**

603 The third model consists of two precipitation and a standing water endmember (Figure 5C).
 604 In this model, groundwater isotope ratios were modeled as:

$$605 \quad \delta^{18}O_{gw} = f_{precip1}\delta^{18}O_{precip1} + f_{precip2}\delta^{18}O_{precip2} + f_{stand}\delta^{18}O_{stand}$$

$$606 \quad \delta^2H_{gw} = f_{precip1}\delta^2H_{precip1} + f_{precip2}\delta^2H_{precip2} + f_{stand}\delta^2H_{stand}$$

607 where $\delta^{18}O_{gw}$ and δ^2H_{gw} are the isotope ratios of the groundwater, $\delta^{18}O_{precip1}$ and $\delta^2H_{precip1}$ and
 608 $\delta^{18}O_{precip2}$ and $\delta^2H_{precip2}$ are the isotope ratios of precipitation endmembers respectively. $f_{precip1}$,
 609 $f_{precip2}$ and f_{stand} are the fractional contribution of precipitation and standing water to the
 610 groundwater and $f_{precip1} + f_{precip2} + f_{stand} = 1$. In this model, for each standing water endmember, we
 611 generate two distinct sets of 500 random precipitation isotope values, and from each set a random
 612 precipitation value is selected to obtain a pair of precipitation isotope ratios (Figure 5C). We then

613 solve the above equations to obtain the fractional contributions of the endmembers (f_{river} , $f_{precip2}$
614 and $f_{stand.}$).

615 References

- 616 1. A. M. MacDonald, H. C. Bonsor, K. M. Ahmed, W. G. Burgess, M. Basharat, R. C. Calow, A.
617 Dixit, S. S. D. Foster, K. Gopal, D. J. Lapworth, R. M. Lark, M. Moench, A. Mukherjee, M. S.
618 Rao, M. Shamsudduha, L. Smith, R. G. Taylor, J. Tucker, F. Van Steenbergen, S. K. Yadav,
619 Groundwater quality and depletion in the Indo-Gangetic Basin mapped from in situ observations.
620 *Nat. Geosci.* **9**, 762–766 (2016).
- 621 2. K. M. Ahmed, P. Bhattacharya, M. A. Hasan, S. H. Akhter, S. M. M. Alam, M. A. H. Bhuyian, M.
622 B. Imam, A. A. Khan, O. Sracek, Arsenic enrichment in groundwater of the alluvial aquifers in
623 Bangladesh: An overview. *Appl. Geochemistry*. **19**, 181–200 (2004).
- 624 3. H. A. Michael, C. I. Voss, Estimation of regional-scale groundwater flow properties in the Bengal
625 Basin of India and Bangladesh. *Hydrogeol. J.* **17**, 1329–1346 (2009).
- 626 4. M. R. Khan, M. Koneshloo, P. S. K. Knappett, K. M. Ahmed, B. C. Bostick, B. J. Mailloux, R. H.
627 Mozumder, A. Zahid, C. F. Harvey, A. Van Geen, H. A. Michael, Megacity pumping and
628 preferential flow threaten groundwater quality. *Nat. Commun.* **7**, 1–8 (2016).
- 629 5. P. S. K. Knappett, B. J. Mailloux, I. Choudhury, M. R. Khan, H. A. Michael, S. Barua, D. R.
630 Mondal, M. S. Steckler, S. H. Akhter, K. M. Ahmed, B. Bostick, C. F. Harvey, M. Shamsudduha,
631 P. Shuai, I. Mihajlov, R. Mozumder, A. van Geen, Vulnerability of low-arsenic aquifers to
632 municipal pumping in Bangladesh. *J. Hydrol.* **539**, 674–686 (2016).
- 633 6. M. Shamsudduha, R. G. Taylor, K. M. Ahmed, A. Zahid, The impact of intensive groundwater
634 abstraction on recharge to a shallow regional aquifer system: Evidence from Bangladesh.
635 *Hydrogeol. J.* **19**, 901–916 (2011).
- 636 7. M. Shamsudduha, R. G. Taylor, Groundwater storage dynamics in the world’s large aquifer
637 systems from GRACE: uncertainty and role of extreme precipitation. *Earth Syst. Dyn.* **11**, 755–
638 774 (2020).
- 639 8. H. A. Michael, C. I. Voss, Evaluation of the sustainability of deep groundwater as an arsenic-safe
640 resource in the Bengal Basin. *Proc. Natl. Acad. Sci. U. S. A.* **105**, 8531–8536 (2008).
- 641 9. D. J. Lapworth, A. Zahid, R. G. Taylor, W. G. Burgess, M. Shamsudduha, K. M. Ahmed, A.
642 Mukherjee, D. C. Gooddy, D. Chatterjee, A. M. MacDonald, Security of Deep Groundwater in the
643 Coastal Bengal Basin Revealed by Tracers. *Geophys. Res. Lett.* **45**, 8241–8252 (2018).
- 644 10. F. L. Naus, P. Schot, K. Groen, K. M. Ahmed, J. Griffioen, Groundwater salinity variation in
645 Upazila Assasuni (southwestern Bangladesh), as steered by surface clay layer thickness, relative
646 elevation and present-day land use. *Hydrol. Earth Syst. Sci.* **23**, 1431–1451 (2019).
- 647 11. M. A. Mojid, M. F. Parvez, M. Mainuddin, G. Hodgson, Water Table Trend—A Sustainability
648 Status of Groundwater Development in North-West Bangladesh. *Water*. **11**, 1182 (2019).
- 649 12. S. Datta, A. W. Neal, T. J. Mohajerin, T. Ocheltree, B. E. Rosenheim, C. D. White, K. H.
650 Johannesson, *Geophys. Res. Lett.*, in press, doi:10.1029/2011GL049301.
- 651 13. S. Sengupta, J. M. McArthur, A. Sarkar, M. J. Leng, P. Ravenscroft, R. J. Howarth, D. M.
652 Banerjee, Do ponds cause arsenic-pollution of groundwater in the Bengal Basin? An answer from
653 West Bengal. *Environ. Sci. Technol.* **42**, 5156–5164 (2008).
- 654 14. UNDP, Bangladesh : ground-water survey : the hydrogeologic conditions of Bangladesh :
655 technical report :DP/UN/BGD-74-009/1 (1982).
- 656 15. R. K. Majumder, M. A. Halim, B. B. Saha, R. Ikawa, T. Nakamura, M. Kagabu, J. Shimada,
657 Groundwater flow system in Bengal Delta, Bangladesh revealed by environmental isotopes.
658 *Environ. Earth Sci.* **64**, 1343–1352 (2011).
- 659 16. A. Biswas, H. Neidhardt, A. K. Kundu, D. Halder, D. Chatterjee, Z. Berner, G. Jacks, P.
660 Bhattacharya, Spatial, vertical and temporal variation of arsenic in shallow aquifers of the Bengal
661 Basin: Controlling geochemical processes. *Chem. Geol.* **387**, 157–169 (2014).
- 662 17. R. B. Neumann, K. N. Ashfaq, A. B. M. Badruzzaman, M. Ashraf Ali, J. K. Shoemaker, C. F.
663 Harvey, Anthropogenic influences on groundwater arsenic concentrations in Bangladesh. *Nat.*
664 *Geosci.* **3**, 46–52 (2010).

- 665 18. C. F. Harvey, K. N. Ashfaq, W. Yu, A. B. M. Badruzzaman, M. A. Ali, P. M. Oates, H. A.
666 Michael, R. B. Neumann, R. Beckie, S. Islam, M. F. Ahmed, Groundwater dynamics and arsenic
667 contamination in Bangladesh. *Chem. Geol.* **228**, 112–136 (2006).
- 668 19. M. O. Stahl, M. H. Tarek, D. C. J. Yeo, A. B. M. Badruzzaman, C. F. Harvey, Crab burrows as
669 conduits for groundwater-surface water exchange in Bangladesh. *Geophys. Res. Lett.* **41**, 8342–
670 8347 (2014).
- 671 20. C. F. Harvey, Response to Comments on “Arsenic Mobility and Groundwater Extraction in
672 Bangladesh.” *Science (80-.)*. **300**, 584d – 584 (2003).
- 673 21. R. Jakobsen, J. Kazmierczak, H. U. Sørensen, D. Postma, *Water Resour. Res.*, in press,
674 doi:10.1029/2018WR023685.
- 675 22. B. Weinman, S. L. Goodbred, Y. Zheng, Z. Aziz, M. Steckler, A. van Geen, A. K. Singhvi, Y. C.
676 Nagar, Contributions of floodplain stratigraphy and evolution to the spatial patterns of
677 groundwater arsenic in Arai-hazar, Bangladesh. *Bull. Geol. Soc. Am.* **120**, 1567–1580 (2008).
- 678 23. J. Fergusson, On recent changes in the delta of the Ganges. *Q. J. Geol. Soc. London.* **19**, 321–354
679 (1863).
- 680 24. S. Nowreen, R. G. Taylor, M. Shamsudduha, M. Salehin, A. Zahid, K. M. Ahmed, Groundwater
681 recharge processes in an Asian mega-delta: hydrometric evidence from Bangladesh. *Hydrogeol. J.*,
682 1–16 (2020).
- 683 25. J. M. Kirby, M. D. Ahmad, M. Mainuddin, W. Palash, M. E. Quadir, S. M. Shah-Newaz, M. M.
684 Hossain, The impact of irrigation development on regional groundwater resources in Bangladesh.
685 *Agric. Water Manag.* **159**, 264–276 (2015).
- 686 26. M. Mainuddin, M. Kirby, National food security in Bangladesh to 2050. *Food Secur.* **7**, 633–646
687 (2015).
- 688 27. R. B. Neumann, M. L. Polizzotto, A. B. M. Badruzzaman, M. A. Ali, Z. Zhang, C. F. Harvey,
689 Hydrology of a groundwater-irrigated rice field in Bangladesh: Seasonal and daily mechanisms of
690 infiltration. *Water Resour. Res.* **45** (2009), doi:10.1029/2008WR007542.
- 691 28. M. Mainuddin, M. Kirby, R. A. R. Chowdhury, S. M. Shah-Newaz, Spatial and temporal
692 variations of, and the impact of climate change on, the dry season crop irrigation requirements in
693 Bangladesh. *Irrig. Sci.* **33**, 107–120 (2015).
- 694 29. N. B. Jamil, H. Feng, K. M. Ahmed, I. Choudhury, P. Barnwal, A. Van Geen, Effectiveness of
695 Different Approaches to Arsenic Mitigation over 18 Years in Arai-hazar, Bangladesh: Implications
696 for National Policy. *Environ. Sci. Technol.* **53**, 5596–5604 (2019).
- 697 30. B. Belton, A. Azad, The characteristics and status of pond aquaculture in Bangladesh.
698 *Aquaculture.* **358–359** (2012), pp. 196–204.
- 699 31. I. Kränzlin, thesis, University of Basel (2000).
- 700 32. N. Huq, Small scale freshwater ponds in rural Bangladesh: Navigating roles and services. *Int. J.*
701 *Water.* **11**, 73–85 (2017).
- 702 33. B. J. Tipple, Y. Jameel, T. H. Chau, C. J. Mancuso, G. J. Bowen, A. Dufour, L. A. Chesson, J. R.
703 Ehleringer, Stable hydrogen and oxygen isotopes of tap water reveal structure of the San Francisco
704 Bay Area’s water system and adjustments during a major drought. *Water Res.* **119** (2017),
705 doi:10.1016/j.watres.2017.04.022.
- 706 34. Y. Jameel, S. Brewer, S. P. Good, B. J. Tipple, J. R. Ehleringer, G. J. Bowen, Tap water isotope
707 ratios reflect urban water system structure and dynamics across a semiarid metropolitan area.
708 *Water Resour. Res.* **52**, 5891–5910 (2016).
- 709 35. J. Evaristo, S. Jasechko, J. J. McDonnell, Global separation of plant transpiration from
710 groundwater and streamflow. *Nature.* **525**, 91–94 (2015).
- 711 36. G. J. Bowen, A. Putman, J. R. Brooks, D. R. Bowling, E. J. Oerter, S. P. Good, Inferring the
712 source of evaporated waters using stable H and O isotopes. *Oecologia.* **187**, 1025–1039 (2018).
- 713 37. S. Fendorf, H. A. Michael, A. Van Geen, Spatial and temporal variations of groundwater arsenic
714 in South and Southeast Asia. *Science (80-.)*. **328** (2010), pp. 1123–1127.
- 715 38. M. O. Stahl, J. B. Ong, C. F. Harvey, C. D. Johnson, A. B. M. Badruzzaman, M. H. Tarek, A. van

- 716 Geen, J. A. Anderson, J. W. Lane, Detecting Well Casing Leaks in Bangladesh Using a Salt
717 Spiking Method. *Groundwater*. **52**, 195–200 (2014).
- 718 39. I. Choudhury, K. M. Ahmed, M. Hasan, M. R. H. Mozumder, P. S. K. Knappett, T. Ellis, A. van
719 Geen, Evidence for Elevated Levels of Arsenic in Public Wells of Bangladesh Due To Improper
720 Installation. *Groundwater*. **54**, 871–877 (2016).
- 721 40. S. Sengupta, A. Sarkar, Stable isotope evidence of dual (Arabian Sea and Bay of Bengal) vapour
722 sources in monsoonal precipitation over north India. *Earth Planet. Sci. Lett.* **250**, 511–521 (2006).
- 723 41. N. C. Munksgaard, N. Kurita, R. Sánchez-Murillo, N. Ahmed, L. Araguas, D. L. Balachew, M. I.
724 Bird, S. Chakraborty, N. Kien Chinh, K. M. Cobb, S. A. Ellis, G. Esquivel-Hernández, S. Y.
725 Ganyaglo, J. Gao, D. Gastmans, K. F. Kaseke, S. Kebede, M. R. Morales, M. Mueller, S. C. Poh,
726 V. dos Santos, H. Shaoneng, L. Wang, H. Yacobaccio, C. Zwart, Data Descriptor: Daily
727 observations of stable isotope ratios of rainfall in the tropics. *Sci. Rep.* **9** (2019),
728 doi:10.1038/s41598-019-50973-9.
- 729 42. B. Kumar, S. P. Rai, U. S. Kumar, S. K. Verma, P. Garg, S. V. V. Kumar, R. Jaiswal, B. K.
730 Purendra, S. R. Kumar, N. G. Pande, Isotopic characteristics of Indian precipitation. *Water*
731 *Resour. Res.* **46** (2010), doi:10.1029/2009WR008532.
- 732 43. M. Tanoue, K. Ichiyangi, K. Yoshimura, M. Kiguchi, T. Terao, T. Hayashi, Seasonal variation in
733 isotopic composition and the origin of precipitation over Bangladesh. *Prog. Earth Planet. Sci.* **5**,
734 1–16 (2018).
- 735 44. J. M. McArthur, U. Ghosal, P. K. Sikdar, J. D. Ball, Arsenic in Groundwater: The Deep Late
736 Pleistocene Aquifers of the Western Bengal Basin. *Environ. Sci. Technol.* **50**, 3469–3476 (2016).
- 737 45. I. Mihajlov, M. Stute, P. Schlosser, B. J. Mailloux, Y. Zheng, I. Choudhury, K. M. Ahmed, A. van
738 Geen, Recharge of low-arsenic aquifers tapped by community wells in Arahazar, Bangladesh,
739 inferred from environmental isotopes. *Water Resour. Res.* **52**, 3324–3349 (2016).
- 740 46. A. Van Geen, E. B. Ahmed, L. Pitcher, J. L. Mey, H. Ahsan, J. H. Graziano, K. M. Ahmed,
741 Comparison of two blanket surveys of arsenic in tubewells conducted 12 years apart in a 25km²
742 area of Bangladesh. *Sci. Total Environ.* **488–489**, 484–492 (2014).
- 743 47. A. H. A. N. Khan, M. A. Q. Bhuyian, M. A. Ahsan, F. Islam, M. M. Karim, M. Moniruzzaman, in
744 *Water, Flood Management and Water Security Under a Changing Climate* (Springer International
745 Publishing, 2020; https://doi.org/10.1007/978-3-030-47786-8_1), pp. 1–14.
- 746 48. A. Kumar, P. Sanyal, S. Agrawal, Spatial distribution of $\delta^{18}\text{O}$ values of water in the Ganga river
747 basin: Insight into the hydrological processes. *J. Hydrol.* **571**, 225–234 (2019).
- 748 49. L. Lambs, K. Balakrishna, F. Brunet, J. L. Probst, Oxygen and hydrogen isotopic composition of
749 major Indian rivers: a first global assessment. *Hydrol. Process.* **19**, 3345–3355 (2005).
- 750 50. J. J. Gibson, S. J. Birks, T. W. D. Edwards, *Global Biogeochem. Cycles*, in press,
751 doi:10.1029/2007GB002997.
- 752 51. A. Mukherjee, S. N. Bhanja, Y. Wada, Groundwater depletion causing reduction of baseflow
753 triggering Ganges river summer drying. *Sci. Rep.* **8**, 1–9 (2018).
- 754 52. P. Aggarwal, A. Basu, J. P. Robert, K. Kulkarni, S. Tarafdar, M. Ali, N. Ahmed, A. Hussain, M.
755 Rahman, S. Reazuddin Ahmed, “A Report on Isotope Hydrology of Groundwater in Bangladesh:
756 Implications for Characterization and Mitigation of Arsenic in Groundwater Work Performed
757 within an IAEA-TC Project (BGD/8/016)” (2000).
- 758 53. P. Ravenscroft, J. M. McArthur, M. S. Rahman, Identifying multiple deep aquifers in the Bengal
759 Basin: Implications for resource management. *Hydrol. Process.* **32**, 3615–3632 (2018).
- 760 54. A. Mukherjee, A. E. Fryar, P. D. Howell, Regional hydrostratigraphy and groundwater flow
761 modeling in the arsenic-affected areas of the western Bengal basin, West Bengal, India.
762 *Hydrogeol. J.* **15**, 1397–1418 (2007).
- 763 55. M. Lawson, D. A. Polya, A. J. Boyce, C. Bryant, D. Mondal, A. Shantz, C. J. Ballentine, Pond-
764 derived organic carbon driving changes in arsenic hazard found in asian groundwaters. *Environ.*
765 *Sci. Technol.* **47**, 7085–7094 (2013).
- 766 56. A. S. Qureshi, Z. U. Ahmed, T. J. Krupnik, “Groundwater management in Bangladesh: An

- 767 analysis of problems and opportunities” (2014), (available at
768 <https://repository.cimmyt.org/bitstream/handle/10883/4273/56862.pdf>).
- 769 57. M. O. Stahl, A. B. M. Badruzzaman, M. H. Tarek, C. F. Harvey, Geochemical transformations
770 beneath man-made ponds: Implications for arsenic mobilization in South Asian aquifers. *Geochim.*
771 *Cosmochim. Acta.* **288**, 262–281 (2020).
- 772 58. H. A. Michael, C. I. Voss, Estimation of regional-scale groundwater flow properties in the Bengal
773 Basin of India and Bangladesh. *Hydrogeol. J.* **17**, 1329–1346 (2009).
- 774 59. I. Clark, P. Fritz, Environmental isotopes in hydrogeology (1997).
- 775 60. S. Jasechko, Global Isotope Hydrogeology—Review. *Rev. Geophys.* **57** (2019), pp. 835–965.
- 776 61. M. O. Stahl, J. Gehring, Y. Jameel, *Hydrol. Process.*, in press, doi:10.1002/hyp.13832.
- 777 62. S. Jasechko, R. G. Taylor, Intensive rainfall recharges tropical groundwaters. *Environ. Res. Lett.*
778 **10**, 124015 (2015).
- 779 63. M. Cherry, T. Gilmore, A. Mittelstet, D. Gastmans, V. Santos, J. B. Gates, Recharge seasonality
780 based on stable isotopes: Nongrowing season bias altered by irrigation in Nebraska. *Hydrol.*
781 *Process.* **34**, 1575–1586 (2020).
- 782 64. S. K. Joshi, S. P. Rai, R. Sinha, S. Gupta, A. L. Densmore, Y. S. Rawat, S. Shekhar, Tracing
783 groundwater recharge sources in the northwestern Indian alluvial aquifer using water isotopes
784 ($\delta^{18}O$, δ^2H and $3H$). *J. Hydrol.* **559**, 835–847 (2018).
- 785 65. M. C. Gratzler, G. R. Davidson, A. M. O’Reilly, J. R. Rigby, Groundwater recharge from an
786 oxbow lake-wetland system in the Mississippi Alluvial Plain. *Hydrol. Process.* **34**, 1359–1370
787 (2020).
- 788 66. K. Alam, Farmers’ adaptation to water scarcity in drought-prone environments: A case study of
789 Rajshahi District, Bangladesh. *Agric. Water Manag.* **148**, 196–206 (2015).
- 790 67. A. S. Rahman, M. Kamruzzama, C. S. Jahan, Q. H. Mazumder, Long-term trend analysis of water
791 table using ‘MAKESENS’ model and sustainability of groundwater resources in drought prone
792 Barind area, NW Bangladesh. *J. Geol. Soc. India.* **87**, 179–193 (2016).
- 793 68. M. R. H. Mozumder, H. A. Michael, I. Mihajlov, M. R. Khan, P. S. K. Knappett, B. C. Bostick, B.
794 J. Mailloux, K. M. Ahmed, I. Choudhury, T. Koffman, T. Ellis, K. Whaley-Martin, R. San Pedro,
795 G. Slater, M. Stute, P. Schlosser, A. Geen, Origin of Groundwater Arsenic in a Rural Pleistocene
796 Aquifer in Bangladesh Depressurized by Distal Municipal Pumping. *Water Resour. Res.* **56**
797 (2020), doi:10.1029/2020WR027178.
- 798 69. P. K. Sikdar, S. Chakraborty, Numerical modelling of groundwater flow to understand the impacts
799 of pumping on arsenic migration in the aquifer of North Bengal Plain. *J. Earth Syst. Sci.* **126**, 29
800 (2017).
- 801 70. S. Bhattacharjee, B. Saha, B. Saha, M. S. Uddin, C. H. Panna, P. Bhattacharya, R. Saha,
802 Groundwater governance in Bangladesh: Established practices and recent trends. *Groundw.*
803 *Sustain. Dev.* **8**, 69–81 (2019).
- 804 71. S. I. Bhuiyan, M. A. Sattar, M. A. K. Khan, Improving water use efficiency in rice irrigation
805 through wet-seeding. *Irrig. Sci.* **16**, 1–8 (1995).
- 806 72. M. S. Steckler, S. L. Nooner, S. H. Akhter, S. K. Chowdhury, S. Bettadpur, L. Seeber, M. G.
807 Kogan, Modeling Earth deformation from monsoonal flooding in Bangladesh using hydrographic,
808 GPS, and Gravity Recovery and Climate Experiment (GRACE) data. *J. Geophys. Res.* **115**,
809 B08407 (2010).
- 810 73. M. Steckler, D. R. Mondal, S. L. Nooner, S. H. Akhter, L. Seeber, S. V. Bettadpur, C. Sazedul
811 Karim, M. Howe, F. Masson, T. Maurin, C. Rangin, Modeling Earth deformation from monsoonal
812 flooding in Bangladesh using hydrographic, GPS, and Gravity Recovery and Climate Experiment
813 (GRACE) data - Steckler - 2010 - Journal of Geophysical Research: Solid Earth - Wiley Online
814 Library. *AGU* (2013), (available at
815 <https://agupubs.onlinelibrary.wiley.com/doi/pdf/10.1029/2009JB007018>).
- 816 74. M. Shamsudduha, R. E. Chandler, R. G. Taylor, K. M. Ahmed, Recent trends in groundwater
817 levels in a highly seasonal hydrological system: the Ganges-Brahmaputra-Meghna Delta. *Hydrol.*

- 818 *Earth Syst. Sci.* **13**, 2373–2385 (2009).
- 819 75. C. F. Harvey, C. H. Swartz, A. B. M. Badruzzaman, N. Keon-Blute, W. Yu, M. A. Ali, J. Jay, R.
820 Beckie, V. Niedan, D. Brabander, P. M. Oates, K. N. Ashfaq, S. Islam, H. F. Hemond, M. F.
821 Ahmed, Arsenic mobility and groundwater extraction in Bangladesh. *Science* (80-). **298**, 1602–
822 1606 (2002).
- 823 76. B. J. Mailloux, E. Trembath-Reichert, J. Cheung, M. Watson, M. Stute, G. A. Freyer, A. S.
824 Ferguson, K. M. Ahmed, M. J. Alam, B. A. Buchholz, J. Thomas, A. C. Layton, Y. Zheng, B. C.
825 Bostick, A. Van Geen, Advection of surface-derived organic carbon fuels microbial reduction in
826 Bangladesh groundwater. *Proc. Natl. Acad. Sci. U. S. A.* **110**, 5331–5335 (2013).
- 827 77. P. S. K. Knappett, V. Escamilla, A. Layton, L. D. McKay, M. Emch, D. E. Williams, R. Huq, J.
828 Alam, L. Farhana, B. J. Mailloux, A. Ferguson, G. S. Sayler, K. M. Ahmed, A. van Geen, Impact
829 of population and latrines on fecal contamination of ponds in rural Bangladesh. *Sci. Total Environ.*
830 **409**, 3174–3182 (2011).
- 831 78. A. Van Geen, K. M. Ahmed, Y. Akita, M. J. Alam, P. J. Culligan, M. Emch, V. Escamilla, J.
832 Feighery, A. S. Ferguson, P. Knappett, A. C. Layton, B. J. Mailloux, L. D. McKay, J. L. Mey, M.
833 L. Serre, P. K. Streatfield, J. Wu, M. Yunus, Fecal contamination of shallow tubewells in
834 Bangladesh inversely related to arsenic. *Environ. Sci. Technol.* **45**, 1199–1205 (2011).
- 835 79. P. S. K. Knappett, L. D. McKay, A. Layton, D. E. Williams, M. J. Alam, M. R. Huq, J. Mey, J. E.
836 Feighery, P. J. Culligan, B. J. Mailloux, J. Zhuang, V. Escamilla, M. Emch, E. Perfect, G. S.
837 Sayler, K. M. Ahmed, A. Van Geen, Implications of fecal bacteria input from latrine-polluted
838 ponds for wells in sandy aquifers. *Environ. Sci. Technol.* **46**, 1361–1370 (2012).
- 839 80. V. Escamilla, P. S. K. Knappett, M. Yunus, P. K. Streatfield, M. Emch, Influence of Latrine
840 Proximity and Type on Tubewell Water Quality and Diarrheal Disease in Bangladesh.
841 <https://doi.org/10.1080/00045608.2013.756257> (2013), doi:10.1080/00045608.2013.756257.
- 842 81. R. K. Majumder, M. A. Hasnat, S. Hossain, K. Ikeue, M. Machida, An exploration of nitrate
843 concentrations in groundwater aquifers of central-west region of Bangladesh. *J. Hazard. Mater.*
844 **159**, 536–543 (2008).
- 845 82. M. A. Akber, M. A. Islam, M. Dutta, S. M. Billah, M. A. Islam, Nitrate contamination of water in
846 dug wells and associated health risks of rural communities in southwest Bangladesh. *Environ.*
847 *Monit. Assess.* **192**, 1–12 (2020).
- 848 83. M. S. Bhuyan, M. A. Bakar, A. Akhtar, M. B. Hossain, M. M. Ali, M. S. Islam, Heavy metal
849 contamination in surface water and sediment of the Meghna River, Bangladesh. *Environ.*
850 *Nanotechnology, Monit. Manag.* **8**, 273–279 (2017).
- 851 84. M. S. Islam, M. K. Ahmed, M. Habibullah-Al-Mamun, M. F. Hoque, Preliminary assessment of
852 heavy metal contamination in surface sediments from a river in Bangladesh. *Environ. Earth Sci.*
853 **73**, 1837–1848 (2015).
- 854 85. M. S. Islam, M. B. Hossain, A. Matin, M. S. Islam Sarker, Assessment of heavy metal pollution,
855 distribution and source apportionment in the sediment from Feni River estuary, Bangladesh.
856 *Chemosphere.* **202**, 25–32 (2018).
- 857 86. M. Rahman, M. Bodrud-Doza, M. Muhib, M. Sikder, M. Shammi, R. Akter, M. Uddin, Human
858 Health Risk Assessment of Nitrate and Trace Metals Via Groundwater in Central Bangladesh.
859 *Pollution.* **6**, 253–266 (2020).
- 860 87. L. Benneyworth, J. Gilligan, J. C. Ayers, S. Goodbred, G. George, A. Carrico, M. R. Karim, F.
861 Akter, D. Fry, K. Donato, B. Piya, Drinking water insecurity: water quality and access in coastal
862 south-western Bangladesh. *Int. J. Environ. Health Res.* **26**, 508–524 (2016).
- 863 88. J. M. McArthur, P. K. Sikdar, M. A. Hoque, U. Ghosal, Waste-water impacts on groundwater:
864 Cl/Br ratios and implications for arsenic pollution of groundwater in the Bengal Basin and Red
865 River Basin, Vietnam. *Sci. Total Environ.* **437**, 390–402 (2012).
- 866 89. M. A. H. Bhuiyan, M. Bodrud-Doza, A. R. M. T. Islam, M. A. Rakib, M. S. Rahman, A. L.
867 Ramanathan, Assessment of groundwater quality of Lakshimpur district of Bangladesh using
868 water quality indices, geostatistical methods, and multivariate analysis. *Environ. Earth Sci.* **75**, 1–

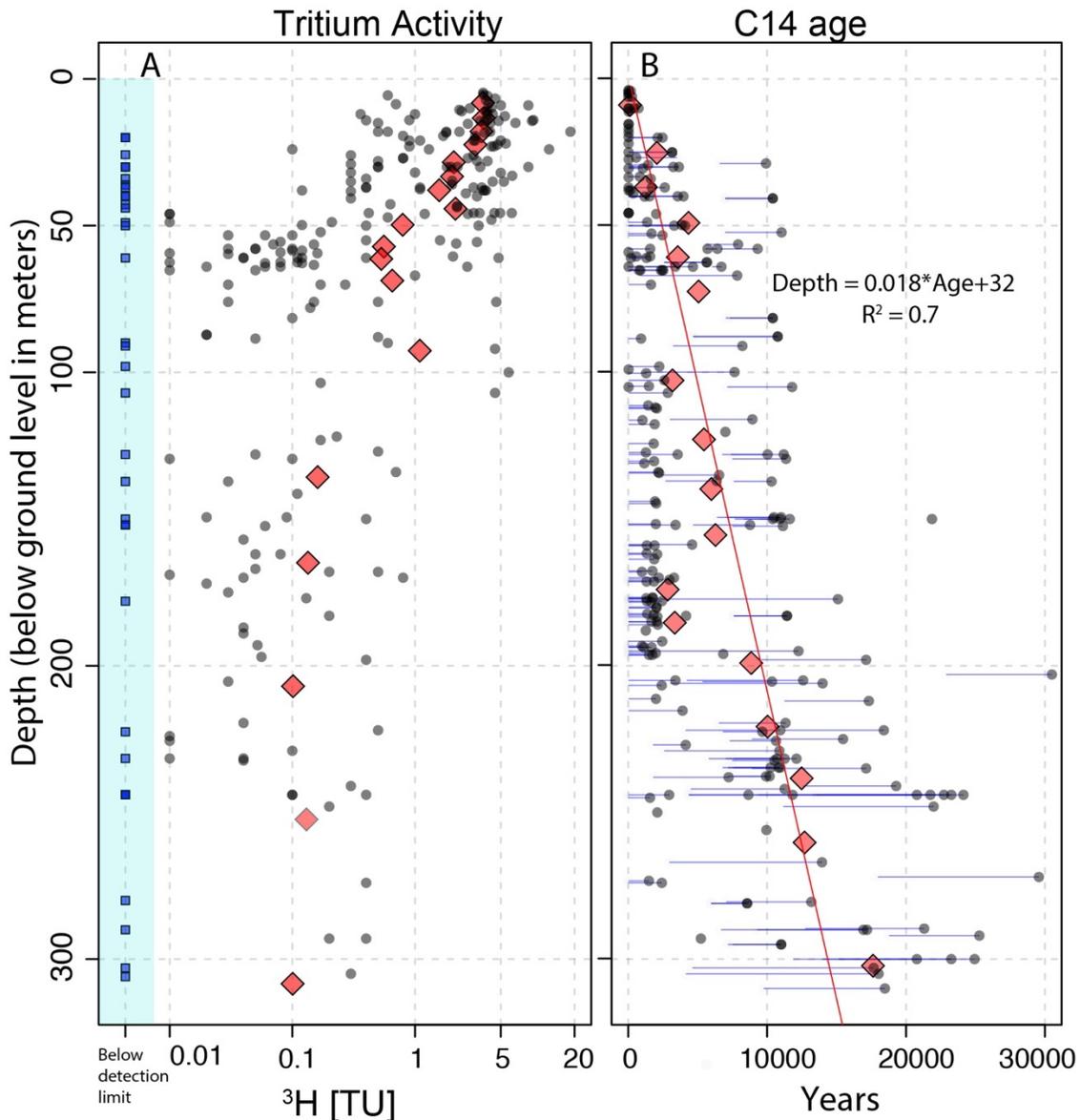
- 869 23 (2016).
- 870 90. A. R. M. T. Islam, N. Ahmed, M. Bodrud-Doza, R. Chu, Characterizing groundwater quality ranks
871 for drinking purposes in Sylhet district, Bangladesh, using entropy method, spatial autocorrelation
872 index, and geostatistics. *Environ. Sci. Pollut. Res.* **24**, 26350–26374 (2017).
- 873 91. A. Das, S. S. Das, N. R. Chowdhury, M. Joardar, B. Ghosh, T. Roychowdhury, Quality and health
874 risk evaluation for groundwater in Nadia district, West Bengal: An approach on its suitability for
875 drinking and domestic purpose. *Groundw. Sustain. Dev.* **10**, 100351 (2020).
- 876 92. S. K. Mandal, S. K. Dutta, S. Pramanik, R. K. Kole, Assessment of river water quality for
877 agricultural irrigation. *Int. J. Environ. Sci. Technol.* **16**, 451–462 (2019).
- 878 93. A. Asoka, T. Gleeson, Y. Wada, V. Mishra, Relative contribution of monsoon precipitation and
879 pumping to changes in groundwater storage in India. *Nat. Geosci.* **10**, 109–117 (2017).
- 880 94. W. Aeschbach-Hertig, T. Gleeson, Regional strategies for the accelerating global problem of
881 groundwater depletion. *Nat. Geosci.* **5** (2012), pp. 853–861.
- 882 95. D. J. Lapworth, A. M. MacDonald, G. Krishan, M. S. Rao, D. C. Goody, W. G. Darling,
883 Groundwater recharge and age-depth profiles of intensively exploited groundwater resources in
884 northwest India. *Geophys. Res. Lett.* **42**, 7554–7562 (2015).
- 885 96. T. Keesari, D. A. Sharma, M. S. Rishi, D. Pant, H. V. Mohokar, A. K. Jaryal, U. K. Sinha, Isotope
886 investigation on groundwater recharge and dynamics in shallow and deep alluvial aquifers of
887 southwest Punjab. *Appl. Radiat. Isot.* **129**, 163–170 (2017).
- 888 97. N. Mushtaq, A. Younas, A. Mashiatullah, T. Javed, A. Ahmad, A. Farooqi, Hydrogeochemical and
889 isotopic evaluation of groundwater with elevated arsenic in alkaline aquifers in Eastern Punjab,
890 Pakistan. *Chemosphere.* **200**, 576–586 (2018).
- 891 98. R. Revelle, V. Lakshminarayana, The Ganges water machine. *Science (80-.).* **188**, 611–616
892 (1975).
- 893 99. M. R. Khan, C. I. Voss, W. Yu, H. A. Michael, Water Resources Management in the Ganges
894 Basin: A Comparison of Three Strategies for Conjunctive Use of Groundwater and Surface Water.
895 *Water Resour. Manag.* **28**, 1235–1250 (2014).
- 896 100. Y. Jameel, S. Brewer, R. P. Fiorella, B. J. Tipple, S. Terry, G. J. Bowen, Isotopic reconnaissance
897 of urban water supply system dynamics. *Hydrol. Earth Syst. Sci.* **22** (2018), doi:10.5194/hess-22-
898 6109-2018.
- 899 101. A. A. Nghiem, M. O. Stahl, B. J. Mailloux, T. T. Mai, P. T. Trang, P. H. Viet, C. F. Harvey, A.
900 Geen, B. C. Bostick, Quantifying Riverine Recharge Impacts on Redox Conditions and Arsenic
901 Release in Groundwater Aquifers Along the Red River, Vietnam. *Water Resour. Res.* **55**, 6712–
902 6728 (2019).
- 903 102. J. J. Follstad Shah, Y. Jameel, R. M. Smith, R. S. Gabor, P. D. Brooks, S. R. Weintraub,
904 Spatiotemporal Variability in Water Sources Controls Chemical and Physical Properties of a
905 Semi-arid Urban River System. *J. Am. Water Resour. Assoc.* **55**, 591–607 (2019).
- 906 103. J. Cable, K. Ogle, D. Williams, Contribution of glacier meltwater to streamflow in the Wind River
907 Range, Wyoming, inferred via a Bayesian mixing model applied to isotopic measurements.
908 *Hydrol. Process.* **25**, 2228–2236 (2011).
- 909 104. M. A. Hoque, W. G. Burgess, ¹⁴C dating of deep groundwater in the Bengal Aquifer System,
910 Bangladesh: Implications for aquifer anisotropy, recharge sources and sustainability. *J. Hydrol.*
911 **444–445**, 209–220 (2012).
- 912

913 Data Availability

914 Data analyzed in this paper are publicly available and the references are provided in the
915 supplementary information. New data included in this analyzed are included in the
916 supplementary information. Groundwater hydrograph data is available upon request.
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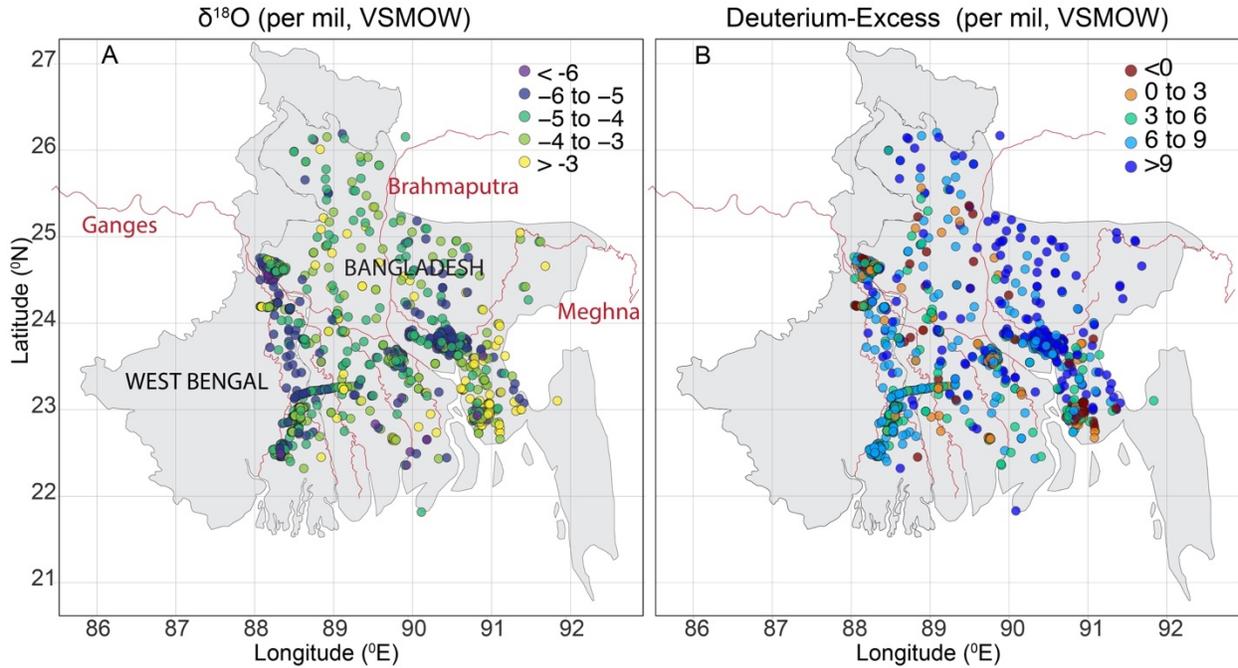
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924 designed the analyses, YJ conducted the analyses and wrote the manuscript. All the authors
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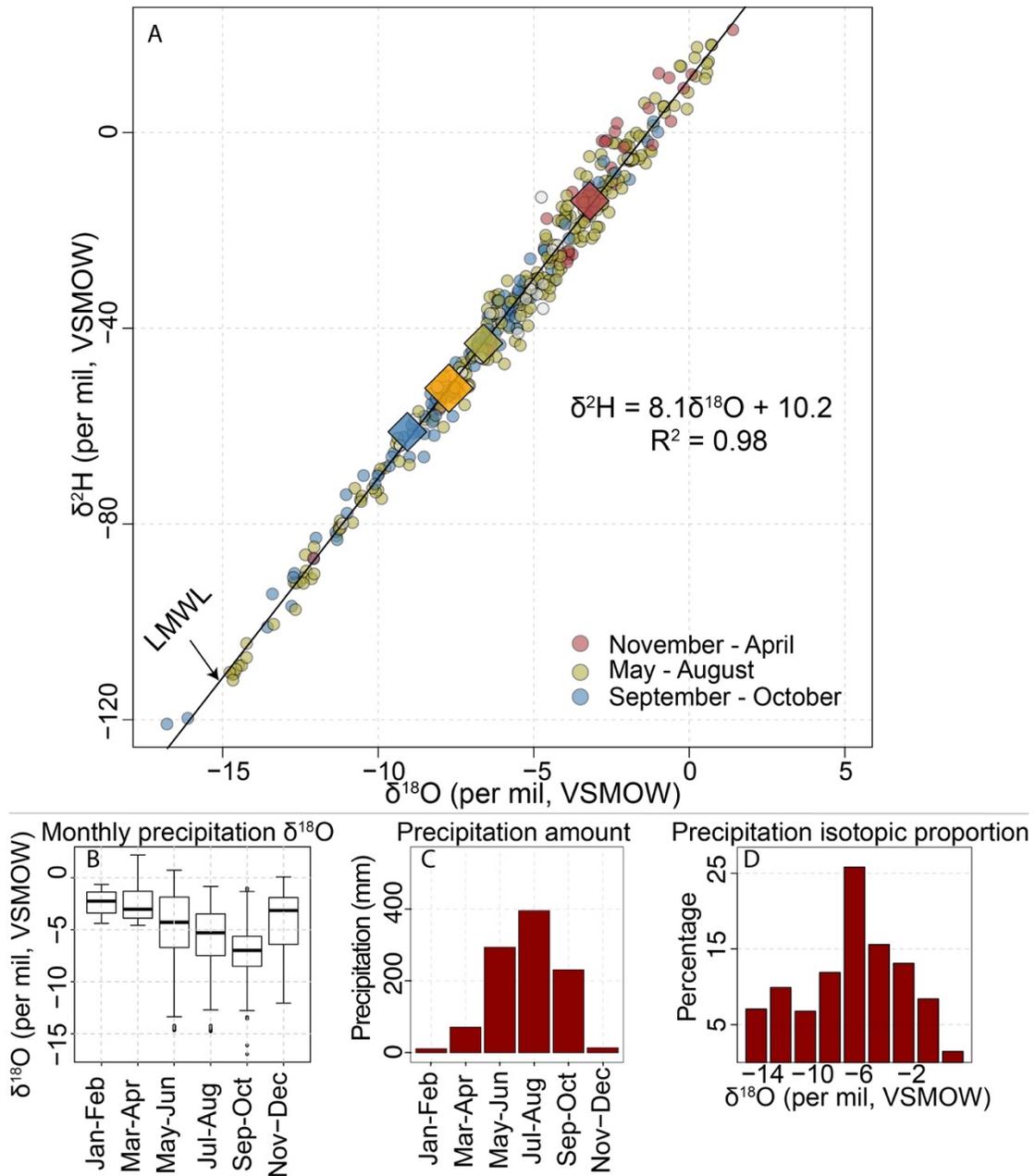
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Figure 1: Tritium activity and ^{14}C ages at different depths from across the Bengal Basin. Tritium activity (A) and uncorrected groundwater ^{14}C ages (B) shown in black circles. The red diamonds represent average over 15 measurements. Blue squares in panel A are values below detection limit, which varies from study to study (Table S2). The horizontal lines in panel B represent corrected ages after accounting for radiocarbon-dead dissolved inorganic carbon using the method described in reference (104). Red line in panel B is the best fit line to the depth averaged ^{14}C ages (red diamonds). See tables S2 and S3 for data source.



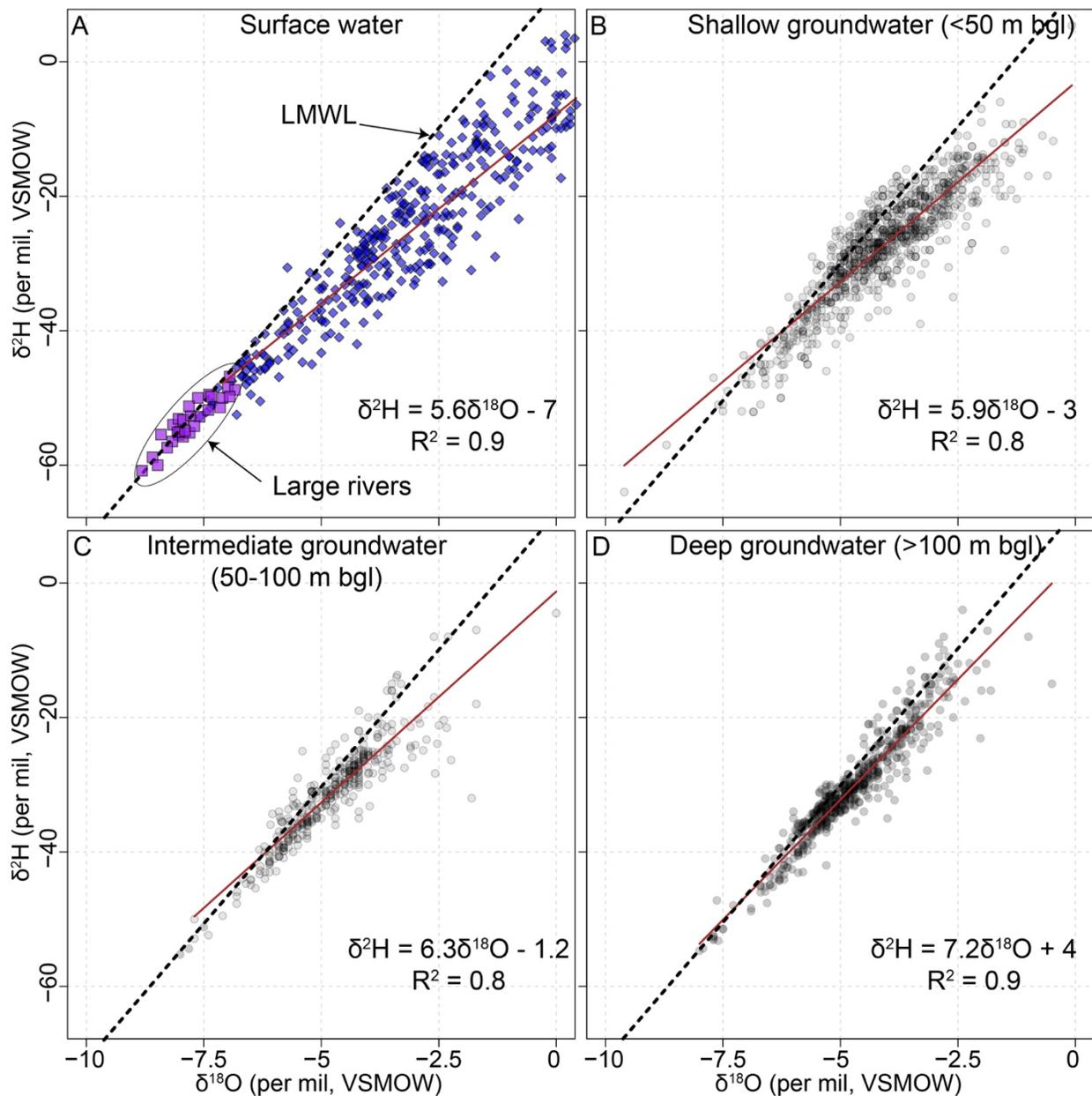
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Figure 2: Spatial distribution of groundwater samples. $\delta^{18}\text{O}$ (A) and d-excess (B) of georeferenced groundwater samples (circles) across the Bengal basin (Bangladesh and West Bengal, India). The three major rivers Ganges, Brahmaputra and Meghna and their tributaries are shown in red lines. Samples that were not georeferenced but included in the mixing model analysis have not been shown here. Groundwater $\delta^{18}\text{O}$ and d-excess values across the basin were not spatially correlated (Moran's I statistics was 0.18 and 0.11, $p < 0.005$ for $\delta^{18}\text{O}$ and d-excess respectively).

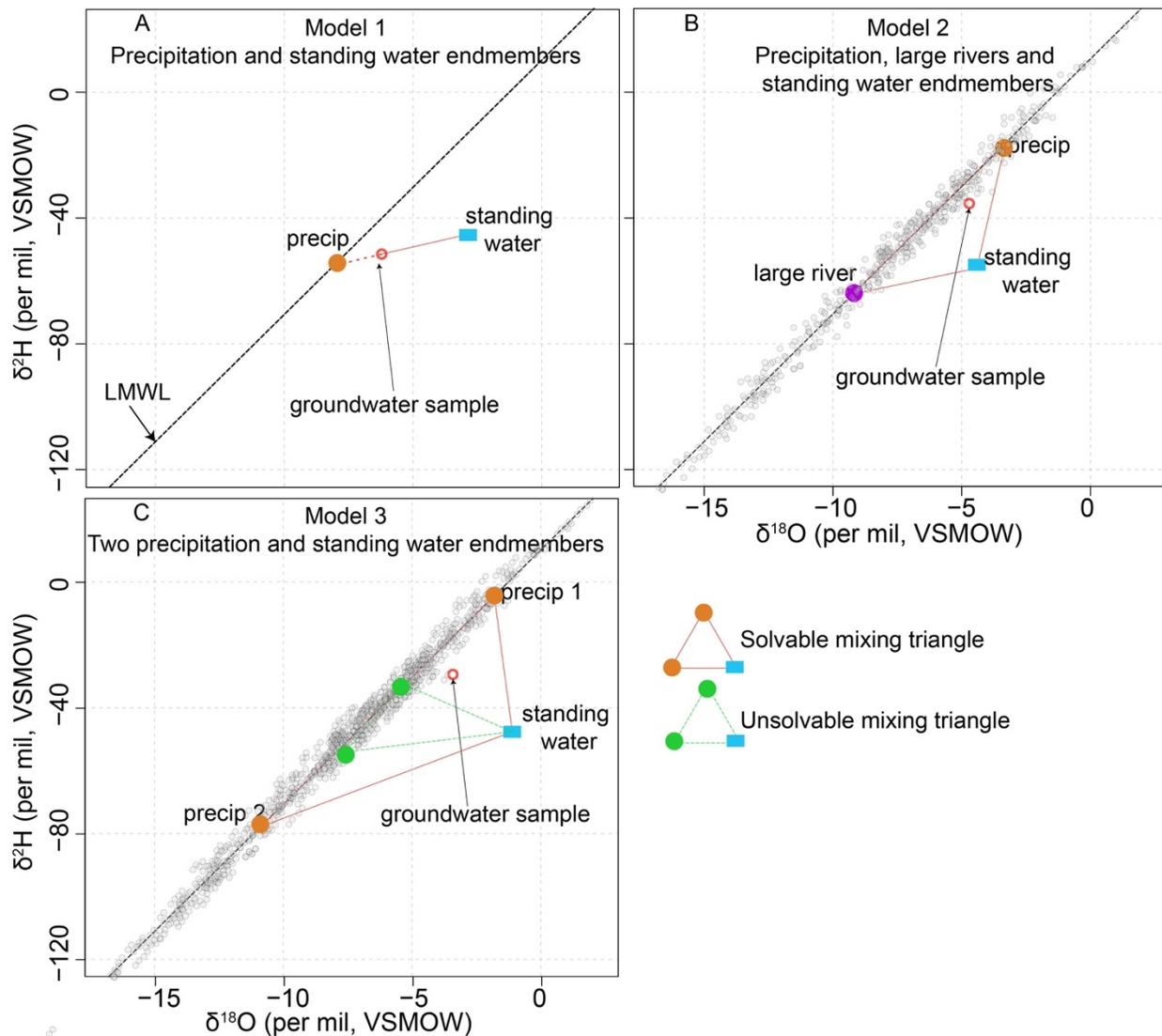


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Figure 3: Precipitation isotope ratios in the Bengal Basin. (A) Precipitation stable isotope values from Bengal basin. Black line is the local meteoric water line (LMWL). Precipitation samples collected in dry months (November - May), early monsoon (June - July) and late monsoon (August - October) are colored red, green and blue respectively. The orange diamond is the amount-weighted mean precipitation isotope value for the Bengal Basin. The red, green and blue diamonds are the amount-weighted mean precipitation isotope values for dry season, early monsoon and late monsoon seasons. (B) Monthly precipitation $\delta^{18}\text{O}$ values. (C) Average precipitation amount. (D) Proportion of rainfall amount grouped by their isotopic composition. Please refer to Figure S1 and S2 for spatial pattern in precipitation amount and precipitation $\delta^{18}\text{O}$ values.



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 959 (A), shallow (B), intermediate (C) and deep (D) wells in the Bengal Basin. The local evaporation
 960 line (LEL) for surface water (A) is shown in brown line. Purple squares and blue diamonds in
 961 panel (A) are the isotope ratios of large rivers and standing water respectively. Dark red line in
 962 panels (A-D) are the respective groundwater lines for shallow, intermediate and deep wells. The
 963 equations are the best fit evaporation line (panel A) and groundwater lines (panels B-D). Black
 964 dashed line in each panel is the LMWL. For clarity, very enriched standing and groundwater
 965 samples ($\delta^{18}\text{O} > 0\text{‰}$) and depleted river samples ($\delta^{18}\text{O} < -10\text{‰}$) have been removed from the
 966 figure.
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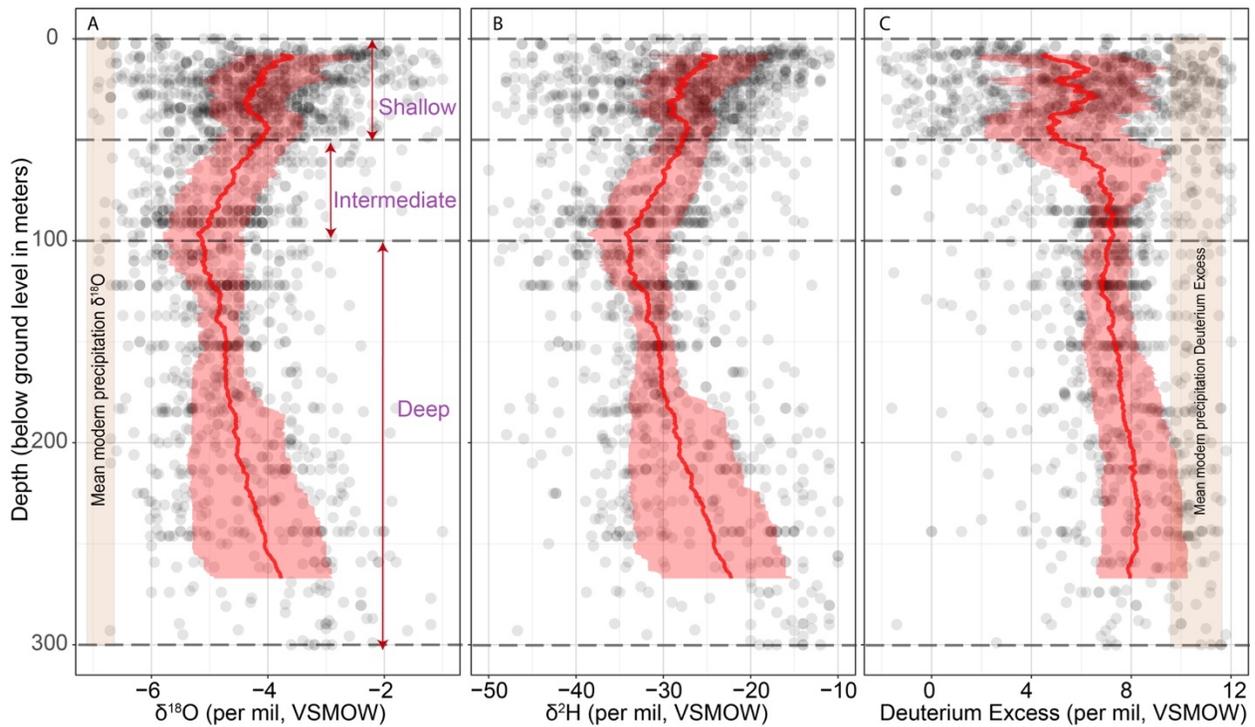
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Figure 5: Graphical illustration of the three mixing models. In model 1 (A), groundwater is modeled as a mixture of precipitation and standing water. The line joining the surface water endmember and the groundwater (solid red) is extended to join the LMWL (dashed red). The point of intersection is the precipitation endmember isotope ratio. In model 2 (B), groundwater is modeled as a mixture of precipitation, large river and standing water endmembers. The small light gray circles are the 500 randomly generated precipitation isotope ratios. The red triangle connecting the large river (purple circle), standing water (blue rectangle) and precipitation (orange circle) endmembers illustrates one of the 500 possible triangles for each standing water sample. In the example shown here the groundwater sample (open red circle) falls within the red triangle and hence the mixing model is solvable. In model 3 (C), groundwater is modeled as a mixture of 2 precipitation endmembers and a standing water endmember. For each standing water endmember, we generate two random distribution of 500 precipitation isotope values and from each of those distributions chose the precipitation endmember values (orange and green circles). The red triangle connecting two orange circles (precipitation endmembers) and blue rectangle (standing water endmember) results in a solvable mixing triangle i.e., groundwater isotope ratios can be modeled as a mixture of these three endmembers. The green triangle connecting two green circles

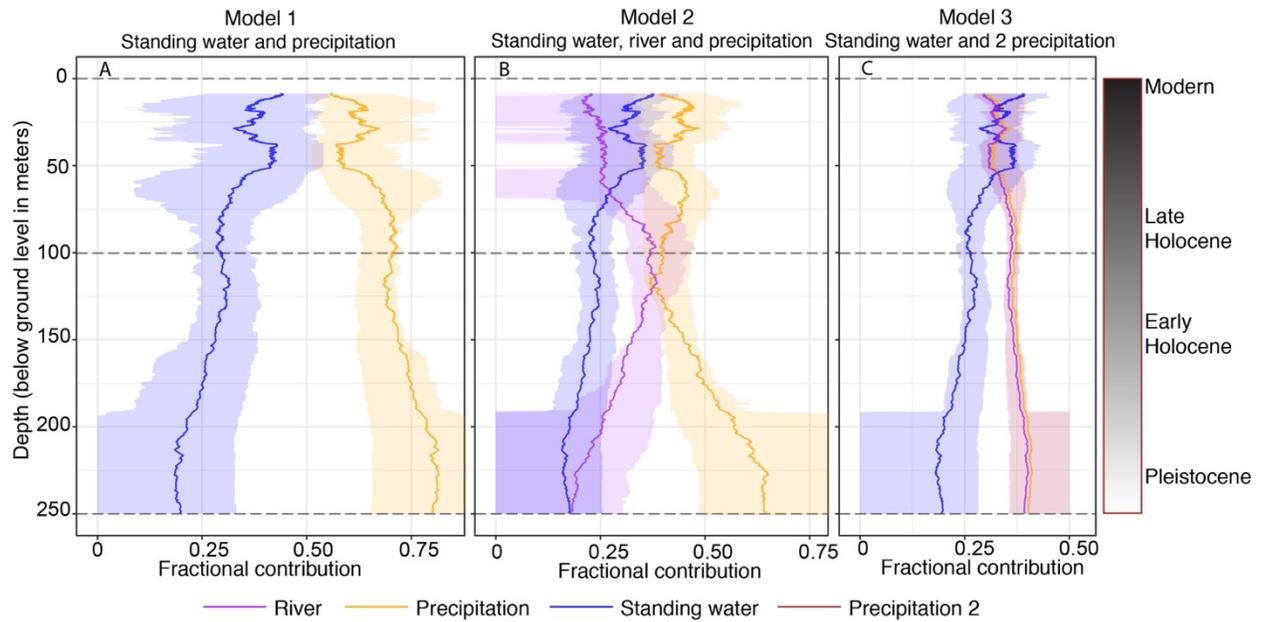
985 (precipitation endmembers) and blue rectangle (standing water endmember) results in an
986 unsolvable mixing triangle i.e., linear mixing between these endmember values cannot explain the
987 observed groundwater isotope ratios. Figure S3 shows the isotopic range of the respective
988 endmembers.

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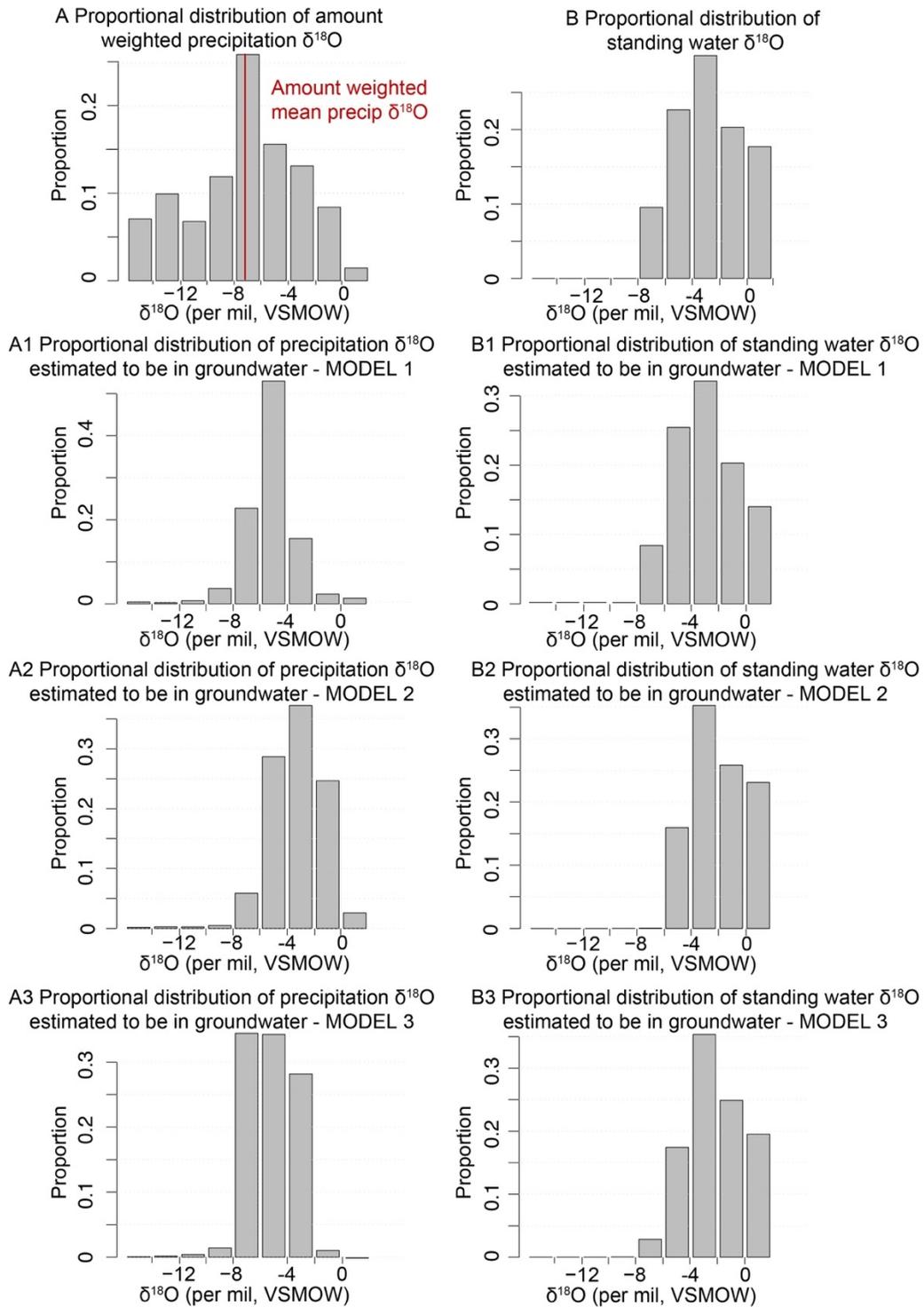
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Figure 6: Depth plot versus groundwater $\delta^{18}\text{O}$ (A), $\delta^2\text{H}$ (B) and d-excess (C). The red line and the light red shaded region are the moving depth average and the interquartile range of 200 groundwater samples starting from 0 meters depth. The black circles in panels (A-C) are the groundwater $\delta^{18}\text{O}$, $\delta^2\text{H}$ and deuterium-Excess values. The light brown rectangles in (A) and (C) are the mean modern precipitation $\delta^{18}\text{O}$ and d-excess.



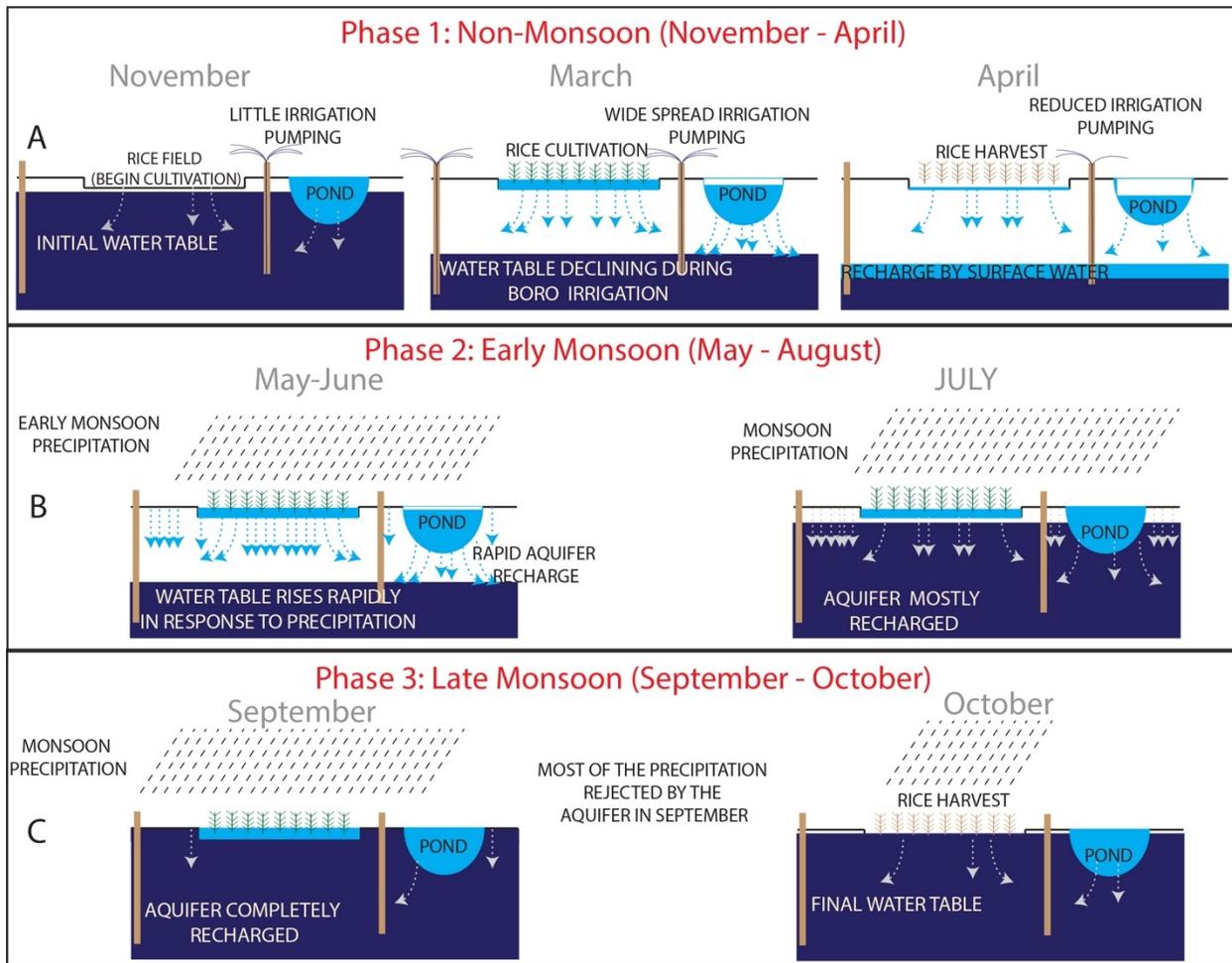
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Figure 7: Proportional contribution of large rivers, surface water and precipitation sources versus depth. (A) Model 1, (B) model 2 and (C) model 3. The purple, blue and orange lines are the moving depth average of 200 groundwater samples. The purple, blue and orange shaded regions are the corresponding interquartile range. The dashed gray line at 100 m roughly differentiates between modern and late Holocene recharged water (0-100 m depth) with early Holocene and late Pleistocene water (100-250 m depth). See section “Limitations of using modern precipitation and river endmember values for deeper groundwater samples” in SM on the limitations of using modern endmember values for late Holocene – Early Pleistocene recharged deep waters



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Figure 8: Distribution of input (A and B) precipitation and standing water $\delta^{18}\text{O}$ values. Distribution of modeled precipitation and standing water $\delta^{18}\text{O}$ values in groundwater (A1 to B3). The red line in panels A is the amount weighted mean precipitation $\delta^{18}\text{O}$ value.



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Figure 9: Conceptual recharge model for a shallow (<50 m deep) aquifer in the Bengal Basin from ponds and rice fields. For simplicity exchange between river and groundwater has been excluded from this figure. (A) In phase 1 (November - April), large amounts of groundwater are extracted for rice irrigation leading to a lowering of the water table. During the same period, the aquifers are also recharged due to lowering of the water table and higher head of ponds and rice fields. (B) In phase 2 (May - July), the shallow groundwater starts to recharge rapidly in response to the incoming monsoon precipitation and by July the aquifer is mostly recharged. (C) During the early part of phase 3 (September), the aquifers are completely replenished and most of the subsequent monsoon precipitation is rejected by the aquifer. By the end of phase 3 (October), monsoon has finished, and the flood waters have receded, beginning a period of groundwater abstraction for irrigation. The blue and gray arrows illustrate the recharge paths.

Figures

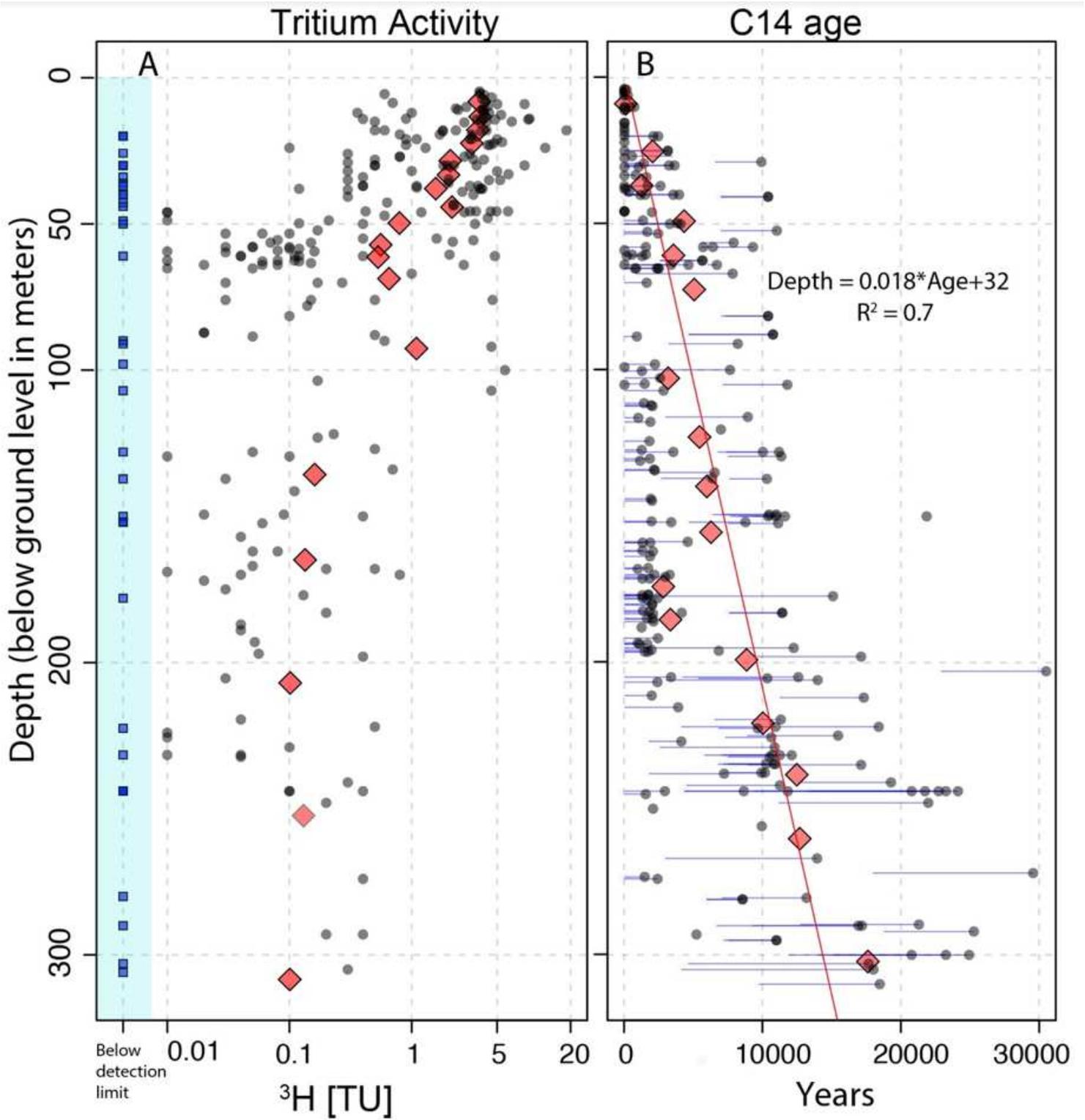


Figure 1

Tritium activity and ^{14}C ages at different depths from across the Bengal Basin. Tritium activity (A) and uncorrected groundwater ^{14}C ages (B) shown in black circles. The red diamonds represent average over 15 measurements. Blue squares in panel A are values below detection limit, which varies from study to

study (Table S2). The horizontal lines in panel B represent corrected ages after accounting for radiocarbon-dead dissolved inorganic carbon using the method described in reference (104). Red line in panel B is the best fit line to the depth averaged C14 ages (red diamonds). See tables S2 and S3 for data source.

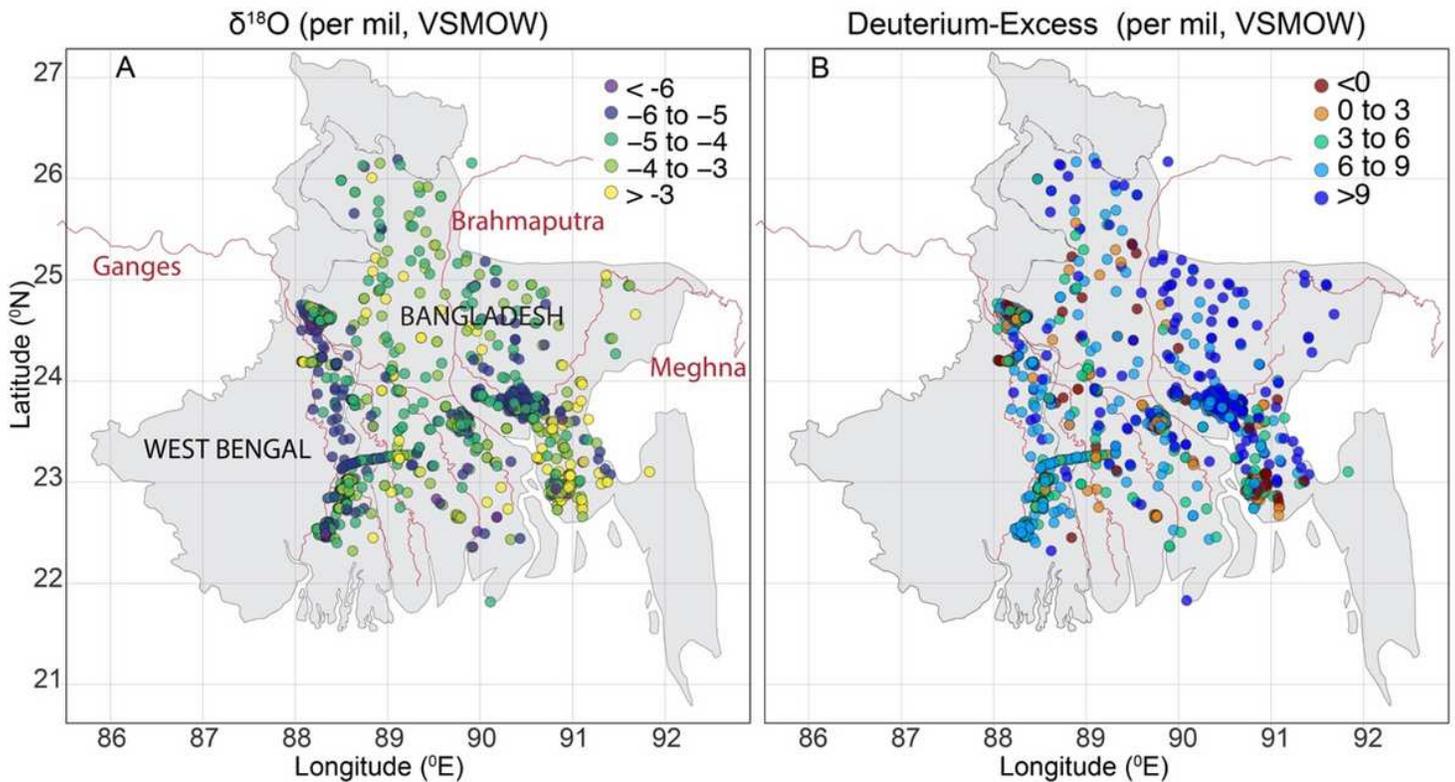


Figure 2

Spatial distribution of groundwater samples. $\delta^{18}\text{O}$ (A) and d-excess (B) of georeferenced groundwater samples (circles) across the Bengal basin (Bangladesh and West Bengal, India). The three major rivers Ganges, Brahmaputra and Meghna and their tributaries are shown in red lines. Samples that were not georeferenced but included in the mixing model analysis have not been shown here. Groundwater $\delta^{18}\text{O}$ and d-excess values across the basin were not spatially correlated (Moran's I statistics was 0.18 and 0.11, $p < 0.005$ for $\delta^{18}\text{O}$ and d-excess respectively). Note: The designations employed and the presentation of the material on this map do not imply the expression of any opinion whatsoever on the part of Research Square concerning the legal status of any country, territory, city or area or of its authorities, or concerning the delimitation of its frontiers or boundaries. This map has been provided by the authors.

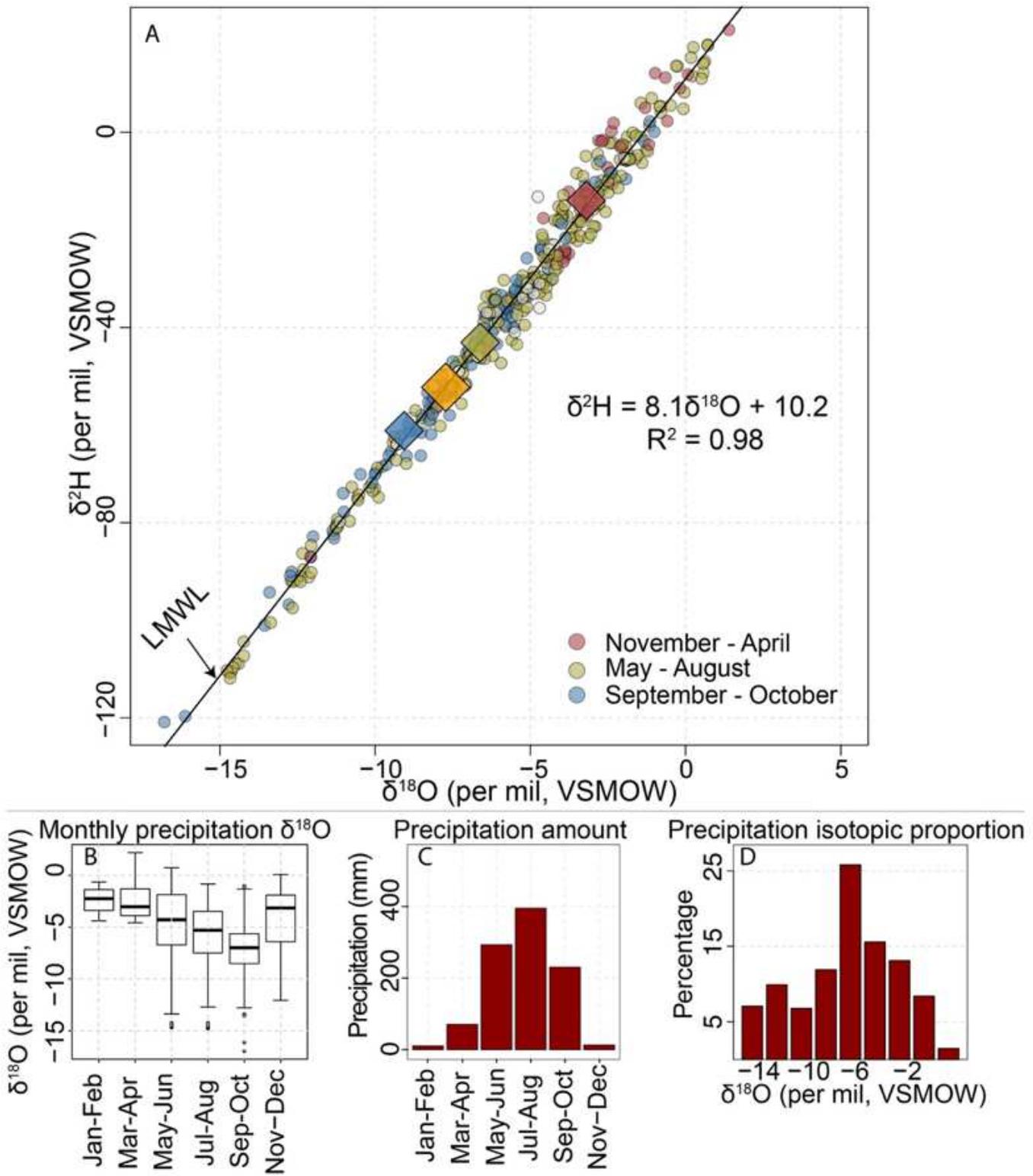


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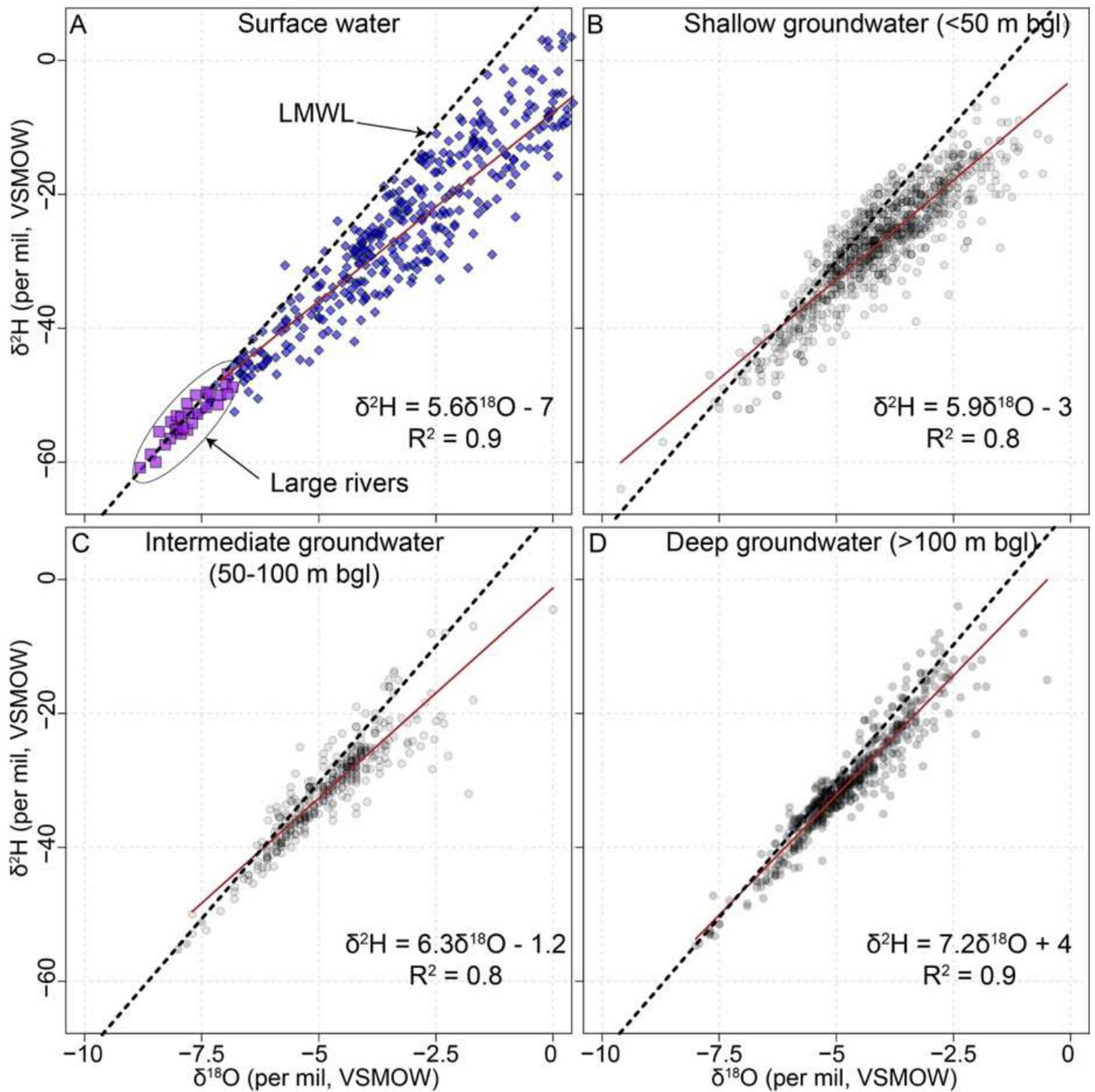


Figure 4

Surface and groundwater isotope ratios. Stable isotope values of surface water (A), shallow (B), intermediate (C) and deep (D) wells in the Bengal Basin. The local evaporation line (LEL) for surface water

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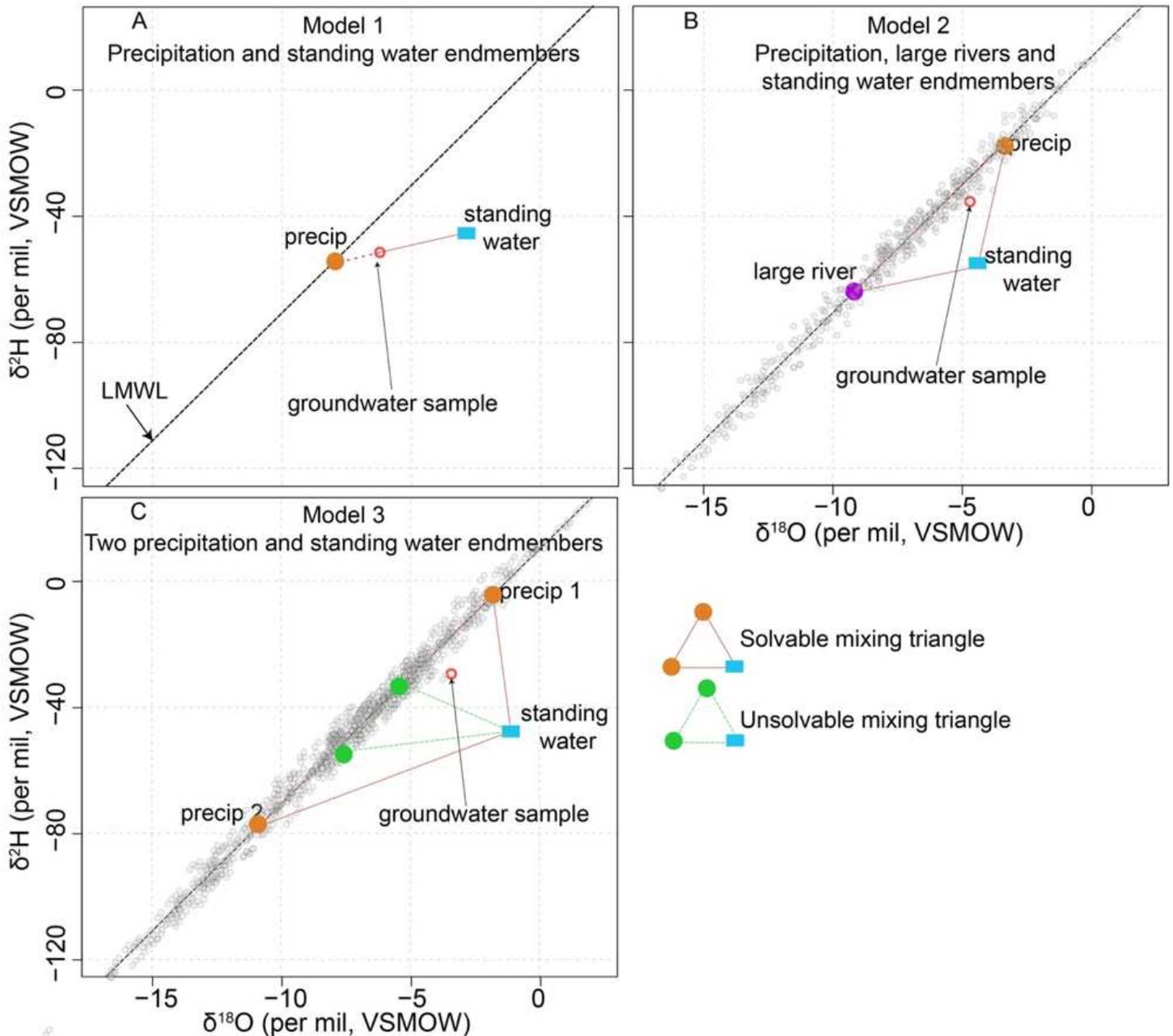


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Graphical illustration of the three mixing models. In model 1 (A), groundwater is modeled as a mixture of precipitation and standing water. The line joining the surface water endmember and the groundwater (solid red) is extended to join the LMWL (dashed red). The point of intersection is the precipitation

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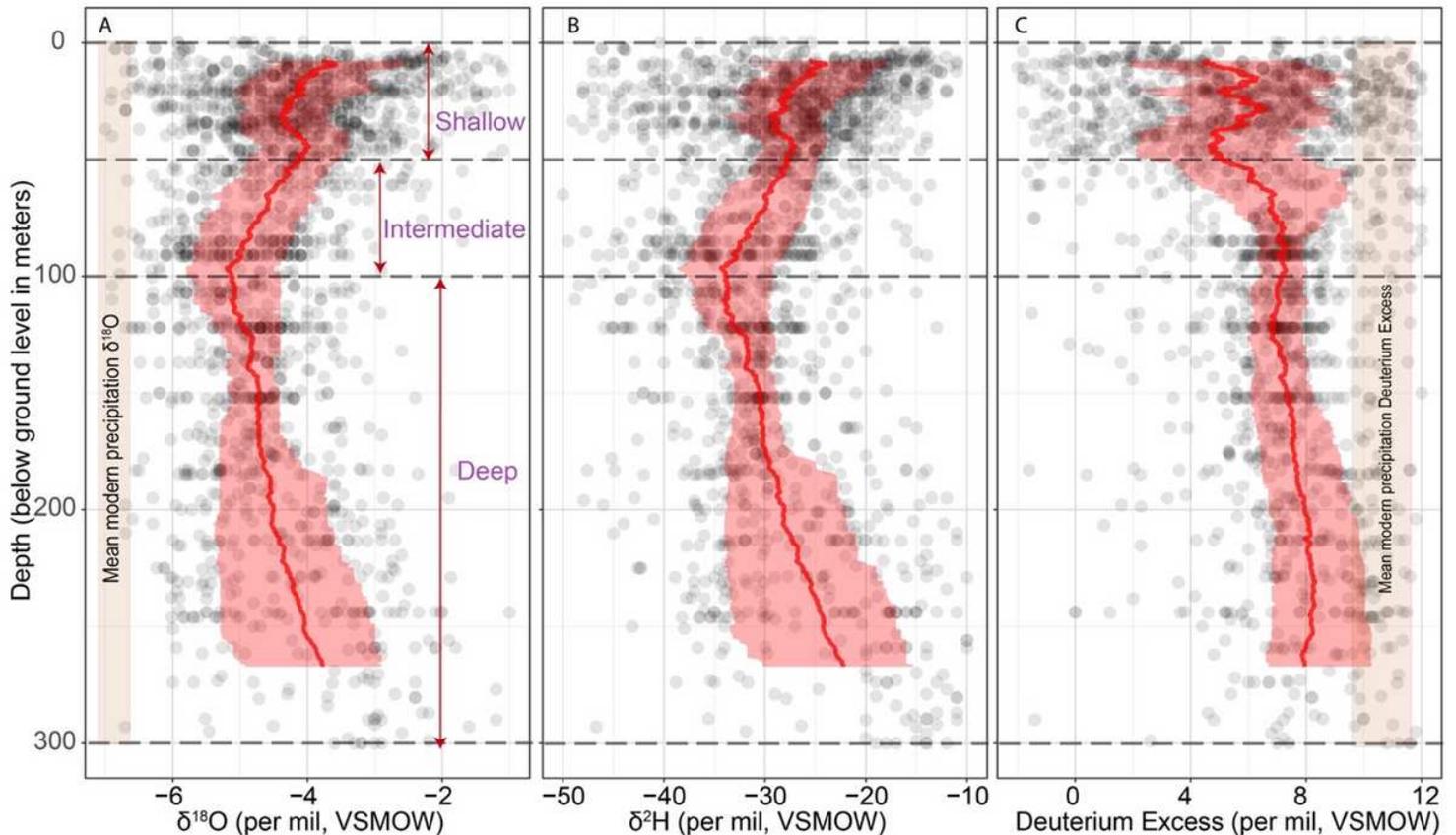


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Depth plot versus groundwater $\delta^{18}\text{O}$ (A), $\delta^2\text{H}$ (B) and d-excess (C). The red line and the light red shaded region are the moving depth average and the interquartile range of 200 groundwater samples starting from 0 meters depth. The black circles in panels (A-C) are the groundwater $\delta^{18}\text{O}$, $\delta^2\text{H}$ and deuterium-

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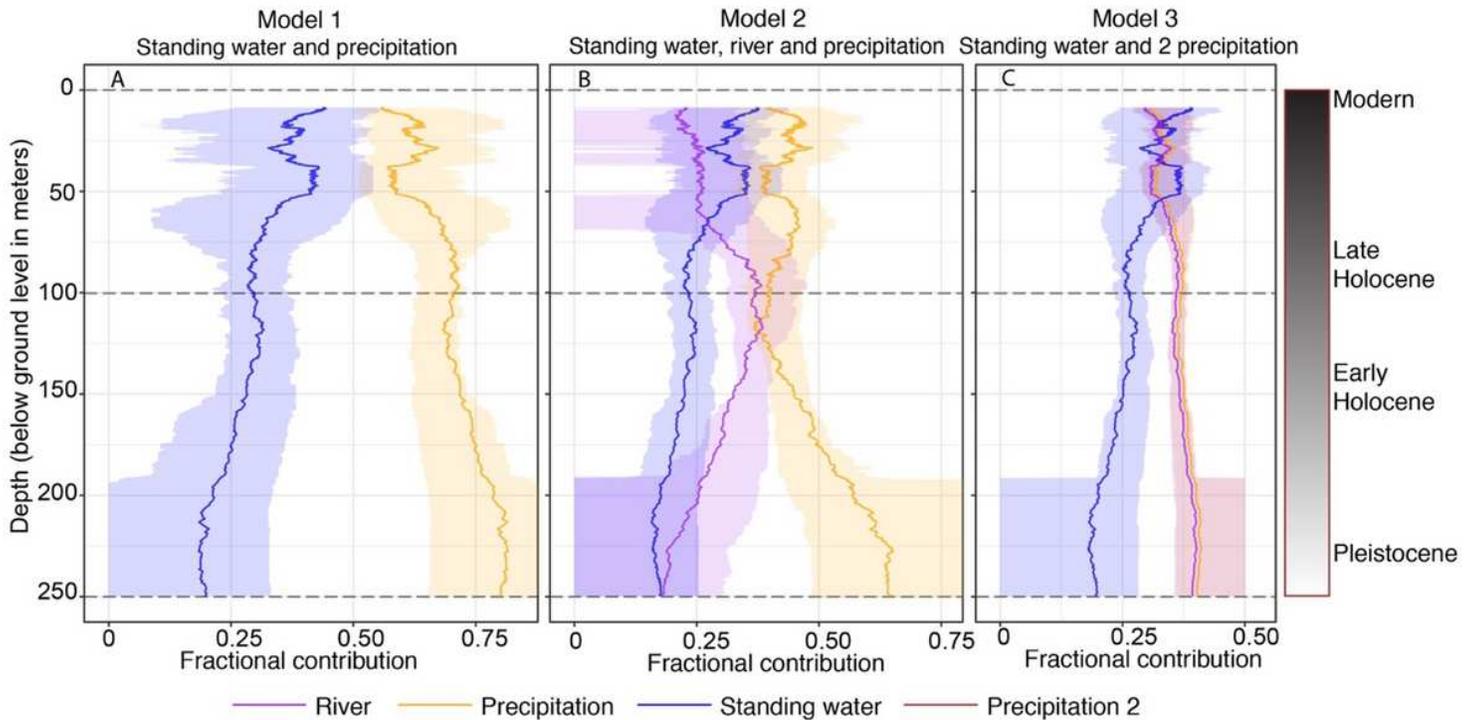


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Proportional contribution of large rivers, surface water and precipitation sources versus depth. (A) Model 1, (B) model 2 and (C) model 3. The purple, blue and orange lines are the moving depth average of 200 groundwater samples. The purple, blue and orange shaded regions are the corresponding interquartile range. The dashed gray line at 100 m roughly differentiates between modern and late Holocene recharged water (0-100 m depth) with early Holocene and late Pleistocene water (100-250 m depth). See section "Limitations of using modern precipitation and river endmember values for deeper groundwater samples" in SM on the limitations of using modern endmember values for late Holocene – Early Pleistocene recharged deep waters

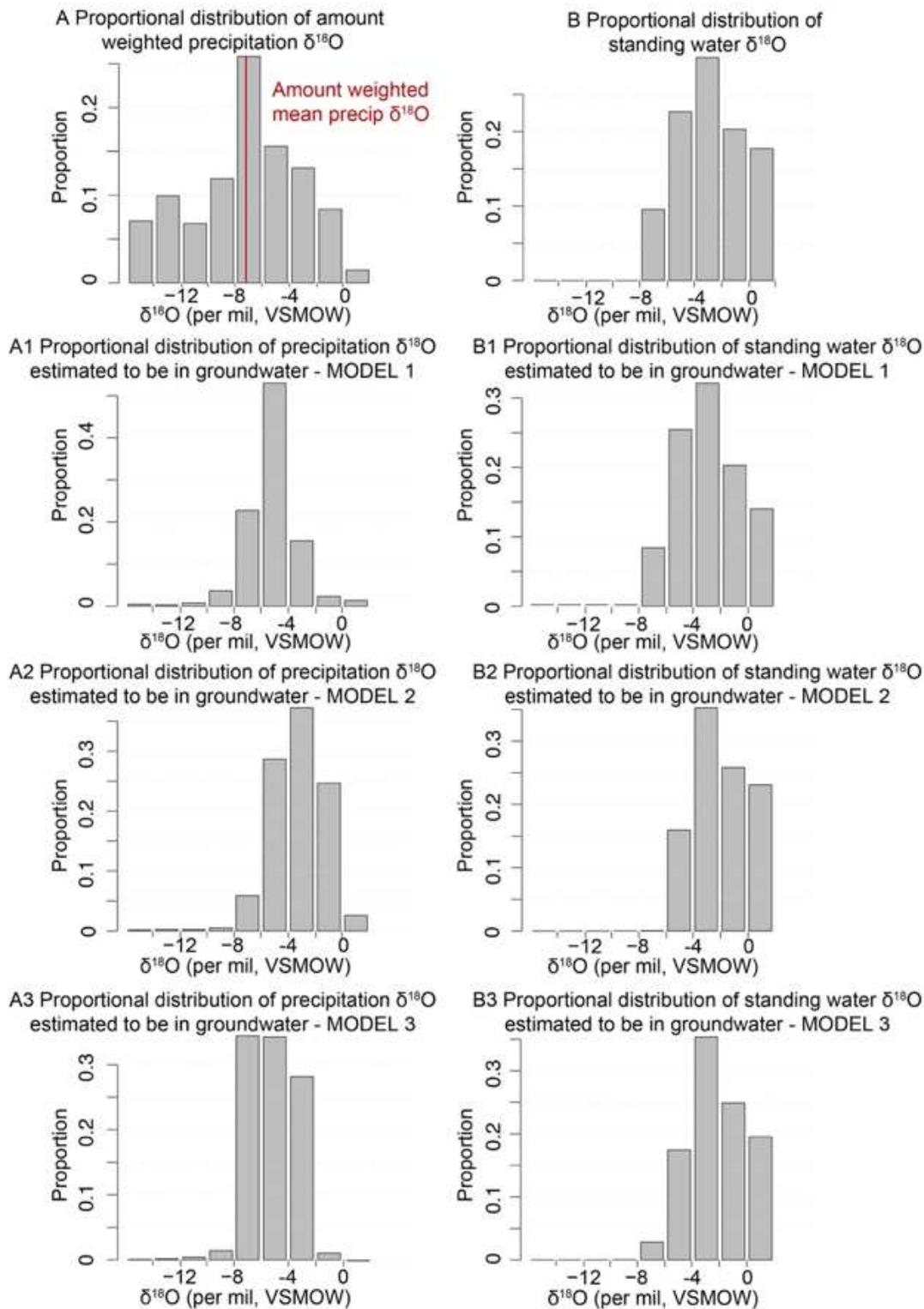


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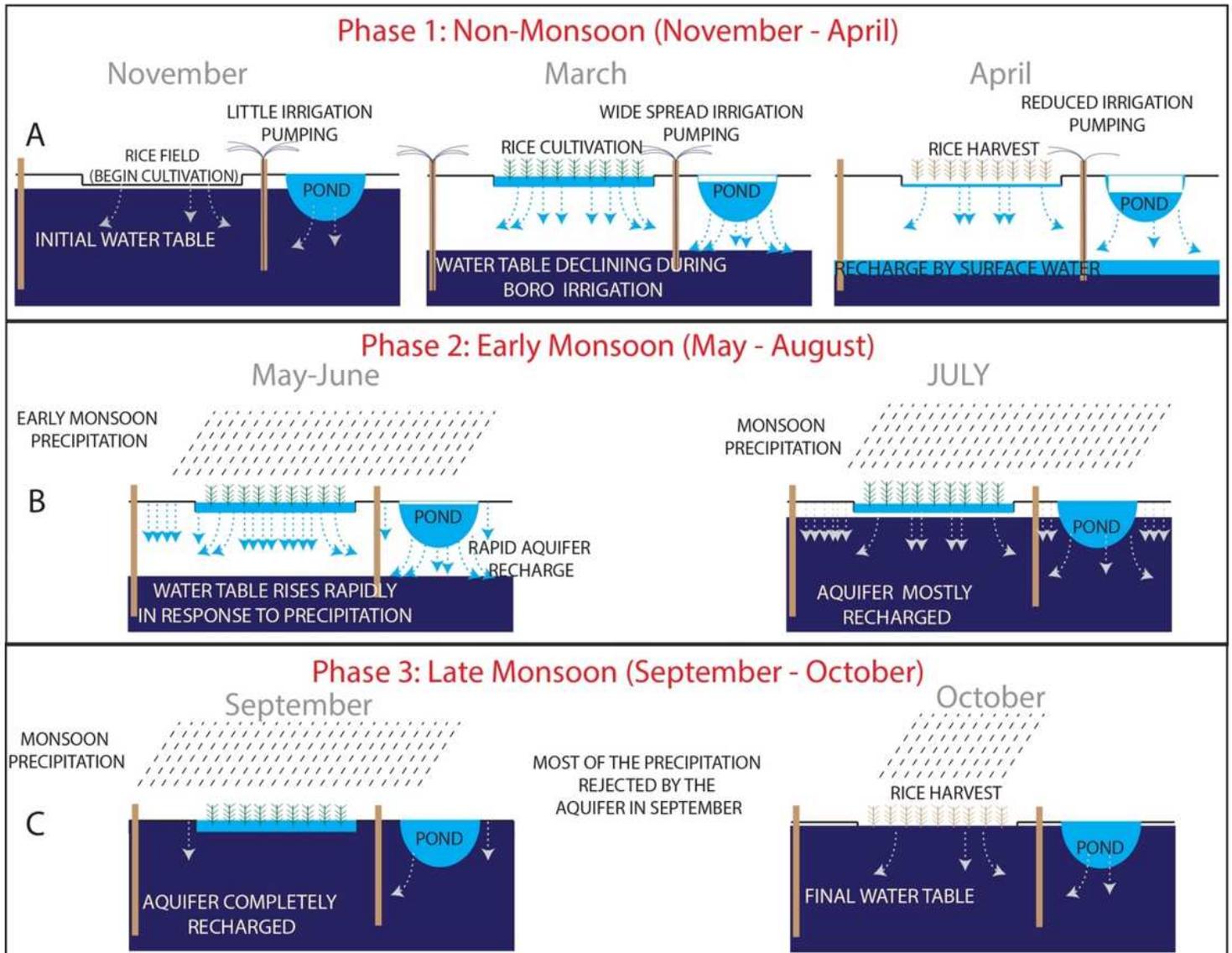


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Supplementary Files

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