

# Oyster As, Cd, Cu, Hg, Pb and Zn Levels in the Northern South China Sea: Long-term Spatiotemporal Distributions, Interacting Effects, and Risk Assessment to Human Health

Lifei Wang (✉ [lifei.wang@utoronto.ca](mailto:lifei.wang@utoronto.ca))

University of Toronto <https://orcid.org/0000-0003-2530-6703>

Xuefeng Wang

Guangdong Ocean University

Haigang Chen

Chinese Academy of Fishery Sciences South China Sea Fisheries Research Institute

Zenghuan Wang

Chinese Academy of Fishery Sciences South China Sea Fisheries Research Institute

Xiaoping Jia

Chinese Academy of Fishery Sciences South China Sea Fisheries Research Institute

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## Research Article

**Keywords:** Environmental conservation , Hazard index , Health risk assessment , Heavy metal pollution , Metal pollution index , Oyster *Crassostrea rivularis* , Principal component analysis , Seafood safety , Spatiotemporal distributions , Target hazard quotient

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2 China Sea: long-term spatiotemporal distributions, interacting  
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5 Lifei WANG <sup>\*1,2</sup>, Xuefeng WANG <sup>\*1,3</sup>, Haigang CHEN <sup>4</sup>, Zenghuan WANG <sup>4</sup>, Xiaoping JIA <sup>4</sup>

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7 <sup>1</sup> College of Fisheries, Guangdong Ocean University, Zhanjiang, Guangdong 524088, China

8 <sup>2</sup> Department of Biological Sciences, University of Toronto Scarborough, Toronto, Ontario, M1C 1A4, Canada

9 <sup>3</sup> Southern Marine Science and Engineering Guangdong Laboratory (Zhanjiang), Zhanjiang, Guangdong 524025, China

10 <sup>4</sup> Key Laboratory of Fishery Ecology and Environment, South China Sea Fisheries Research Institute, Chinese Academy of Fishery  
11 Sciences, Guangzhou, Guangdong 510300, China

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\* Corresponding author: Lifei WANG (lifei.wang@utoronto.ca), Xuefeng WANG (xuefeng1999@126.com)

13 Abstract

14 Estuarine and coastal ecosystems are often considered vulnerable due to the complex biogeochemical processes and  
15 the human disturbances through a variety of pollution. Among environmental contaminants, heavy metals in  
16 estuarine and coastal ecosystems have been of increasing concern in environmental conservation. Long-term  
17 exposure to heavy metal contamination, mainly through food and water, could be harmful to human health. It is  
18 therefore critical to understand the quantitative comparisons and interacting effects of different heavy metals in  
19 common seafood species, such as oysters. This work studied the long-term spatiotemporal trends and health risk  
20 assessment of oyster arsenic (As), cadmium (Cd), copper (Cu), mercury (Hg), lead (Pb) and zinc (Zn) levels in the  
21 coastal waters of northern South China Sea. Cultured oysters (*Crassostrea rivularis*) from 23 estuaries and harbors  
22 in the coastal areas of northern South China Sea in 1989-2015 were analyzed for the spatiotemporal trends of the six  
23 heavy metal levels. Metal pollution index (MPI), target hazard quotient (THQ) and hazard index (HI) were used for  
24 quantifying the exposure of the six heavy metals to human health through oyster consumption. Principal component  
25 analysis (PCA) was used for assessing the relative importance of the six metals in oyster heavy metal distribution  
26 patterns in the northern South China Sea. Overall, the As, Cd, Cu, Hg, Pb and Zn levels in oysters from the northern  
27 South China Sea generally declined from 1989 to 2015, stayed relatively high (MPI = 2.42-3.68) during 1989-2000,  
28 gradually decreased since 2000, and slightly increased after 2010. Oyster heavy metal levels were highest in the  
29 Pearl River Estuary (MPI = 1.20-5.52), followed by west Guangdong and east Guangdong, Guangxi, and Hainan  
30 coastal waters. This pattern is probably because economics and industry around the Pearl River Estuary have been  
31 growing faster than the other areas of this work in the recent two decades, and it should be taken as a hotspot for the  
32 monitoring of seafood safety in southern China. Principal component analysis indicated that Cu, Zn and Cd were  
33 the most important metals in the long-term distributions of oyster heavy metal levels in the northern South China  
34 Sea. Health risk assessment suggested that the risk of the six heavy metals exposure through oyster consumption  
35 were relatively high during 1989-2005 (THQ = 1.01-5.82), significantly decreased since 2005 (THQ < 1), and  
36 slightly increased after 2010.

37 **Keywords** Environmental conservation · Hazard index · Health risk assessment · Heavy metal pollution · Metal  
38 pollution index · Oyster *Crassostrea rivularis* · Principal component analysis · Seafood safety · Spatiotemporal  
39 distributions · Target hazard quotient

## 40 Introduction

41 Natural fishery resources have declined for more than ten times since commercial fishing became popular  
42 worldwide (Myers and Worm 2003). As the world population and living condition increasing nowadays, the need  
43 for high-quality protein, such as seafood, increases dramatically. Under the current global trade of seafood, seafood  
44 safety has drawn increasing concerns, especially when huge amounts of seafood imported to developed countries  
45 (e.g., German and Japan) are from developing countries (Kent 2003; Pauly and Jacquet 2019). Estuarine and coastal  
46 ecosystems are often taken as vulnerable due to the complex biogeochemical processes and the human disturbances  
47 through a variety of pollution (Bendell et al. 2020). Among environmental contaminants, heavy metals in estuarine  
48 and coastal ecosystems have been of increasing concern in environmental conservation (Jahan and Strezov 2019;  
49 Suami et al. 2019; Pinzón-Bedoya et al. 2020). Long-term exposure to heavy metal contamination, mainly through  
50 food and water, could be harmful to human health. The Bulletin of Marine Ecological Environment of China  
51 indicated that in 2016 the sediments from 25 out of 81 estuaries did not meet the environmental requirement of their  
52 respective marine functional zones (Chen 2018). Therefore, the monitoring of heavy metal pollution in marine  
53 organisms is critical for seafood safety, environmental evaluation, and aquaculture development (Azevedo et al.  
54 2019; Djedjibegovic et al. 2020; Esilaba et al. 2020).

55 Marine bivalves and planktivorous fish have been serving as major prey for predators (e.g., sharks and marine  
56 mammals) in marine ecological systems. The rapid development of marine bivalves aquaculture has played an  
57 important role in the sustainable development of fisheries in China and the global trade and supply of seafood (Pauly  
58 and Jacquet 2019). China's marine aquaculture production grows from 1.58 million tons in 1989 to 18.76 million  
59 tons in 2015. Oysters have become one of the most popular seafood in coastal or even inland regions nowadays,  
60 because they are of high nutrients, affordable, and widely farmed (García-Rico and Tejada-Valenzuela 2020). The  
61 oyster aquaculture production from the northern South China Sea took around 36.3% of China's total production of  
62 bivalves aquaculture in 2000, 2010, and 2015. The oyster aquaculture production from three provinces Guangdong,  
63 Guangxi and Hainan in the coastal areas of the northern South China Sea increases from 12.5 thousand tons in 1989  
64 to 1.66 million tons in 2015. By 2015, the oyster aquaculture area in the three southern China provinces has been up  
65 to 49.9 thousand hectares, accounting for 35.2% of China's total area of oyster aquaculture.

66 Heavy metal accumulation in bivalve mollusks depends on not only the biological processes within organisms,  
67 but also the concentrations and bioavailability of heavy metals in ambient environment. Due to their limited  
68 dispersal ability, mollusks easily accumulate heavy metals from environment, and the heavy metal concentrations in  
69 their tissues could indicate the bioavailability of heavy metals and the heavy metal levels in the environment.  
70 Studies based on double-box kinetic model suggested that as heavy metal concentrations in ambient water increased,  
71 oysters' bioconcentration factor (BCF) for heavy metals decreased and the heavy metal levels in their tissues  
72 significantly increased at the state of equilibrium, which made oysters a good indicator species for cadmium, lead  
73 and mercury contamination (Wang et al. 2004). Principal component analysis (PCA) provides an ideal tool for  
74 identifying the sources and relations of multiple heavy metals (Zhou et al. 2007).

75 Heavy metal levels in marine invertebrates such as oysters are usually several orders of magnitude higher than  
76 those in seawater (Rainbow 1997). Different species have different accumulation and degradation capacities to  
77 heavy metals. Oysters typically have a stronger ability to accumulate zinc than mussels, and can store zinc as  
78 detoxified granules in tissues, whereas mussels often discharge the zinc they absorb (George et al. 1978). The heavy  
79 metal levels and accumulation processes in oysters depend on the marine environment and the type and species of  
80 heavy metals (Rainbow et al. 1993). There are various paths for heavy metals to be transferred from environment to  
81 organisms, and the absorption efficiency of different paths varies among species and environmental conditions  
82 (Rainbow et al. 1993). Heavy metal ions usually get into organisms through major ion channels, and as salinity  
83 decreases, the function of ion pumps increases, which increases the possibility for heavy metals being absorbed by  
84 organisms. Rainbow (1997) categorized  $\text{Cu}^+$  and  $\text{Hg}^{2+}$  as Class B, and  $\text{As(III)}$ ,  $\text{Cd}^{2+}$ ,  $\text{Cu}^{2+}$ ,  $\text{Pb}$  and  $\text{Zn}^{2+}$  as borderline  
85 ions. Copper and zinc are essential vector elements for organisms, and Class B and borderline ions both have  
86 potential toxicity, but Class B is typically considered as more toxic than borderline ions. An organic ligand can  
87 more easily and strongly bind with Cu than with  $\text{Zn}^{2+}$  (Rainbow 1997). George et al. (1978) examined the heavy  
88 metal species in green-colored oysters, and found that copper and zinc bound with sulfur and phosphorus  
89 respectively, and the formed Cu-S and Zn-P were stored in vesicles. Their study indicated that the copper in  
90 oysters' blood serum could be absorbed by granular amoebocytes and stored in vesicles, which significantly reduced  
91 the toxicity of copper in oysters. Some heavy metals absorbed by organisms could be taken into biological  
92 processes such as metabolisms, but any heavy metals that are nonessential or essential but beyond normal levels  
93 should go through temporary or permanent detoxification, or would harm normal biological activities. Studies

94 suggested that infants and children's blood lead levels were significantly correlated with their intelligence quotient  
95 (IQ), and the IQ value would decrease for about 1-3 when the blood lead level increased for 100  $\mu\text{g/L}$  (Wang et al.  
96 2010; Taylor et al. 2017).

97 Oysters can easily accumulate pollutants in ambient waterbodies because of their limited dispersal ability and  
98 filter-feeding strategy, and the pollutants can be further accumulated through food chain and harm consumers'  
99 health, making oysters one of the seafood of most concern. Although there have been studies on oyster arsenic  
100 (Wang et al. 2017), cadmium (Wang et al. 2012), copper (Wang et al. 2020) and mercury (Zong and Jia 1996) levels  
101 in the northern South China Sea, they usually focused on one single heavy metal and seldom considered the  
102 quantitative comparisons or interacting effects of multiple heavy metals in cultured oysters from this area. Given  
103 that oysters' accumulation and degradation capacity varies between heavy metals and the potential toxicity of heavy  
104 metals differs a lot, it is important to understand the interacting effects and health implications of multiple heavy  
105 metals in popular seafood species such as oysters. This work studied the long-term spatiotemporal trends and health  
106 risk assessment of oyster arsenic (As), cadmium (Cd), copper (Cu), mercury (Hg), lead (Pb) and zinc (Zn) levels in  
107 the coastal waters of the northern South China Sea from 1989 to 2015. The objectives were to investigate the  
108 spatiotemporal patterns and interacting effects of the six heavy metal levels in cultured oysters in this area in 1989-  
109 2015, and to assess the health risk of the six heavy metals exposure through oyster consumption. Results of this  
110 work could help further understand the past, current and future status of the six heavy metal levels in the northern  
111 South China Sea, provide guidance for pollution recovery and monitoring programs, and support sustainable  
112 development of fisheries and human well-being in this area.

## 113 Material and methods

### 114 Sampling and analytical procedures

115 Based on the major farming areas of cultured oysters (*Crassostrea rivularis*) in the three provinces (i.e., Guangdong,  
116 Guangxi, and Hainan) in the coastal areas of the northern South China Sea and the habitat characteristics of coastal  
117 key fishery waterbodies, oyster samples were collected from 23 estuaries and harbors in March and April from 1989  
118 to 2015 (Fig. 1). Among the 23 sampling sites, 17 sites were in Guangdong and sampled in 1989-2015, and two and  
119 four sites were in Guangxi and Hainan respectively and sampled in 2006-2015 because of logistic constraints in

120 early years of this work. The oyster samples were all at the age of two to three, and were made sure to meet the size  
121 requirement of southern China seafood markets. At each sampling location, the number of oysters collected were  
122 made sure to be more than 30. The oysters were rinsed using sea water *in situ* and dissected. The soft tissues were  
123 stored at -20 °C including body fluids. The entire sampling procedure referred to the China National Specification  
124 for Marine Monitoring (GB 17378-1998; GB 17378-2007).

125 Prior to analyzing heavy metal concentrations, the collected oysters were naturally defrosted to around 20 °C.  
126 Samples were homogenized and measured for wet weight ( $W_w$ ), and then oven-dried for about 72 hours and  
127 measured for dry weight ( $W_d$ ). Samples ( $0.500 \pm 0.005$  g) were ashed around 450 °C and digested using a solution  
128 of 1 mL  $H_2O_2$  and 8 mL  $HNO_3$ . Samples were then filtered using spring filters (0.45  $\mu m$ ) and reconstituted to 10  
129 mL in 2% HCl (Wang et al. 2020). Arsenic, cadmium, copper, mercury, lead and zinc concentrations of the samples  
130 were measured in HITACHI atomic absorption spectrophotometry (Model Z-2000). The measurement was  
131 conducted twice in the same environment for every sample, and were made sure their absolute difference no more  
132 than 10% of their arithmetic mean (Wang et al. 2020). For quality assurance, standard reference materials were  
133 analyzed regularly (China Standard - Oyster tissues GSBZ 19002-88). The analytical procedure referred to the  
134 China Methods of Food Hygienic Analysis – Physical and Chemical Section (GB/T 5009-2003).

### 135 Metal pollution assessment

136 Arsenic, cadmium, copper, mercury, lead and zinc concentrations of the collected oysters were expressed as  
137 geometric mean (Han et al. 1998). The locally weighted least squares regression scatterplot smoothing (LOWESS)  
138 was used to examine the temporal variation of the six heavy metal levels in sampled oysters. Metal pollution index  
139 (MPI) was calculated for assessing the heavy metal pollution status of cultured oysters from different area or time:

$$140 \quad MPI = (MC_1, MC_2, \dots, MC_i)^{1/n} \quad (1)$$

141 where MPI is the metal pollution index, and  $MC_i$  is the concentration of metal  $i$  in the soft tissues of oysters (Denil  
142 et al. 2017). The criterion for evaluating heavy metal pollution status according to MPI was summarized in Table 1  
143 (Jamil et al. 2014). Analysis was conducted using R software and the package of ggplot2 (R Development Core  
144 Team 2020).

## 145 Health risk assessment

146 Target hazard quotient (THQ) was calculated for quantifying the health risk of arsenic, cadmium, copper, mercury,  
147 lead and zinc exposure through oyster consumption:

$$148 \quad THQ = \frac{Efr \times ED_{tot} \times SFI \times MCS}{RfD \times BW_a \times AT_n} \times 10^{-3} \quad (2)$$

149 where EFr is the exposure frequency (182.5 days per year); ED<sub>tot</sub> is the total exposure duration (30 years), which  
150 means the exposure time of consuming oysters; SFI is the seafood ingestion which is taken as 26.8 g/person/day  
151 (Tang et al. 2009; Wang et al. 2012), and we assumed that all the seafood consumed was oyster soft tissues; MCS is  
152 the metal contents in edible portion of seafood; RfD is the reference dose by oral consuming in µg/g/day (Table 2);  
153 BW<sub>a</sub> is the average body weight 70 kg; and AT<sub>n</sub> is the average exposure time for noncarcinogenic (ED<sub>tot</sub> × 365  
154 days = 10950 days).

155 Hazard index (HI) was used to analyze the total potential risk to consumers' health for the six heavy metals,  
156 which is the sum of THQ:

$$157 \quad HI = \sum_{i=1}^n THQ_i \quad (3)$$

158 where HI is the hazard index, THQ<sub>i</sub> is the target hazard quotient of metal i, and n is the number of target metals in  
159 oyster soft tissues.

## 160 Relative impact of multiple heavy metal levels

161 The relative impact and interacting effects of arsenic, cadmium, copper, mercury, lead and zinc in sampled oysters  
162 were assessed using principal component analysis (PCA), which maximizing the total variance of loading values for  
163 different factors through varimax rotation (Abdi and Williams 2010).

## 164 Results

### 165 Spatiotemporal distributions of six heavy metal levels in oysters

166 Heavy metal analysis of cultured oysters sampled from 23 sites in the coastal areas of the northern South China Sea  
167 in 1989-2015 suggested that the spatiotemporal distributions of arsenic, cadmium, copper, mercury, lead and zinc  
168 varied a lot between metals (Table 3, Fig. 2). The copper and zinc levels were 53.85  $\mu\text{g/g}$  and 149.23  $\mu\text{g/g}$   
169 respectively (expressed as geometric mean, wet weight), and were three orders of magnitude higher than the  
170 cadmium (0.57  $\mu\text{g/g}$ ), arsenic (0.33  $\mu\text{g/g}$ ) and lead (0.31  $\mu\text{g/g}$ ) levels, and four orders of magnitude higher than the  
171 mercury level (0.01  $\mu\text{g/g}$ ). Overall, the six heavy metal levels in sampled oysters declined during 1989-2010 (e.g.,  
172 arsenic, cadmium, copper, and lead; Fig. 2), and slightly increased in 2011-2015 (e.g., arsenic, copper, mercury, and  
173 lead; Fig. 1), especially that the arsenic and lead levels increased from 0.10  $\mu\text{g/g}$  and 0.20  $\mu\text{g/g}$  in 2006-2010 to 0.52  
174  $\mu\text{g/g}$  and 0.34  $\mu\text{g/g}$  in 2011-2015, with an increase for 420% and 70% respectively. The temporal variation of the  
175 cadmium levels were less than the other five metals, with an increase to 1.58  $\mu\text{g/g}$  in 1996-2000, a decrease for  
176 50.6% (to 0.78  $\mu\text{g/g}$ ) in 2001-2005, and a slight increase in 2006-2015 (0.84  $\mu\text{g/g}$  in 2006-2010 and 0.86  $\mu\text{g/g}$  in  
177 2011-2015; Table 2).

178 Heavy metal analysis of cultured oysters sampled from five zones consisting of the Pearl River Estuary, west  
179 Guangdong, east Guangdong, Guangxi and Hainan in 1989-2015 indicated that the heavy metal levels in the Pearl  
180 River Estuary were significantly higher than those in the other four zones (Table 3, Fig. 2). Heavy metal levels in  
181 the Pearl River Estuary were generally higher than the overall levels in the northern South China Sea, except that the  
182 arsenic and mercury levels in the Pearl River Estuary were similar to those in the whole study area. The overall  
183 levels of arsenic, cadmium, copper, mercury, lead and zinc in the Pearl River Estuary in 1989-2015 were 0.32  $\mu\text{g/g}$ ,  
184 1.49  $\mu\text{g/g}$ , 122.35  $\mu\text{g/g}$ , 0.01  $\mu\text{g/g}$ , 0.42  $\mu\text{g/g}$  and 210.49  $\mu\text{g/g}$  respectively. The six heavy metal levels in the Pearl  
185 River Estuary were relatively high in 1989-2005, the arsenic, copper and cadmium levels decreased for 69.3%,  
186 68.7% and 19.3% from 2001-2005 to 2006-2010, and the arsenic, copper and lead levels increased to 0.40  $\mu\text{g/g}$ ,  
187 84.75  $\mu\text{g/g}$  and 0.32  $\mu\text{g/g}$  in 2011-2015 compared to those in 2006-2010. Heavy metal levels in west and east  
188 Guangdong were very similar to each other, and were relatively high in 1989-1990, 1991-1995 and 1996-2000. The  
189 heavy metal levels in west Guangdong were generally higher than those in east Guangdong during 2001-2005,  
190 except for the arsenic and lead levels. Due to logistic constraints, oyster heavy metal data from Guangxi and Hainan  
191 were not available until 2006, and heavy metal levels in the coastal waters of Guangxi were generally higher than  
192 those in Hainan. However, the heavy metal levels in Guangxi were higher in 2006-2010 than in 2011-2015, whereas  
193 the heavy metal levels in Hainan were lower in 2006-2010 than in 2011-2015. Heavy metal levels in Guangxi and

194 Hainan were close to the overall levels in the northern South China Sea, lower than those in the Pearl River Estuary,  
195 and similar to those in west and east Guangdong.

#### 196 Metal pollution assessment of six heavy metal levels in oysters

197 The MPI analysis suggested that the oyster heavy metal levels in the northern South China Sea during 1989-2000  
198 were in very low contamination (MPI = 1.6; Table 3). Heavy metal levels in this area declined in 1989-2015, with  
199 the MPI value decreasing for 75.1% from 3.82 in 1989-1990 to 0.95 in 2011-2015. The MPI value in this area was  
200 3.82 in 1989-1990 and 3.68 in 1996-2000, both higher than 2.42 in 1991-1995. The degree of oyster heavy metal  
201 pollution in the northern South China Sea was considered as not impacted since 2001. Across the five zones and  
202 different time periods, oyster heavy metal levels in the Pearl River Estuary in 1996-2000 were the highest (MPI =  
203 5.52) and in 1989-1990 were the second highest (MPI = 4.85). The MPI value in the coastal waters of west  
204 Guangdong was 4.17 in 1989-1990, and was 2.45, 3.40 and 2.62 in 1991-1995, 1996-2000 and 2001-2005  
205 respectively, all higher than that in east Guangdong. The degree of oyster heavy metal pollution in the coastal  
206 waters of Guangxi and Hainan was assigned as not impacted according to MPI since 2006.

#### 207 Health risk assessment of six heavy metal levels in oysters

208 The health risk of arsenic, cadmium, copper, mercury, lead and zinc exposure through oyster consumption in the  
209 coastal areas of the northern South China Sea in 1989-2015 according to THQ and HI was summarized in Table 4.  
210 Typically the health risk of heavy metal exposure is considered as high when  $HI > 1$ . The overall health risk of the  
211 six heavy metals exposure in the northern South China Sea was highest in 1989-1990 ( $HI = 3.336$ ), gradually  
212 decreased since 1991 but still relatively high in 1991-1995 ( $HI = 1.309$ ) and 1996-2000 ( $HI = 1.177$ ), and became  
213 low during 2001-2015 ( $HI < 1$ ) but slightly increased in 2011-2015 compared to 2006-2010.

214 For the health risk of individual heavy metal exposure, the risk of copper exposure in the study area in 1989-1990  
215 was the highest ( $THQ_{Cu} = 2.055$ ), due to the high risk of copper exposure in the Pearl River Estuary ( $THQ_{Cu} =$   
216  $4.438$ ) and west Guangdong ( $THQ_{Cu} = 2.454$ ) during this time period. The risk of arsenic exposure in the study area  
217 in 1989-1990 was the second highest ( $THQ_{As} = 0.919$ ), due to the high risk of arsenic exposure in east Guangdong  
218 ( $THQ_{As} = 1.091$ ), west Guangdong ( $THQ_{As} = 0.900$ ) and the Pearl River Estuary ( $THQ_{As} = 0.804$ ) during this time

219 period. The risk of cadmium exposure in the study area in 1996-2000 was the third highest ( $THQ_{Cd} = 0.302$ ), due to  
220 the high risk of cadmium exposure in the Pearl River Estuary ( $THQ_{Cd} = 0.496$ ) and west Guangdong ( $THQ_{Cd} =$   
221  $0.327$ ) during this time period. The risk of zinc exposure was ranked as the fourth, and the risk of lead and mercury  
222 exposure were relatively low.

223 For the health risk of heavy metal exposure in the five zones at different time periods, the risk of heavy metal  
224 exposure in the Pearl River Estuary was highest overall ( $HI = 1.236$ ) except lower in 2006-2010 ( $HI < 1$ ). The  
225 health risk in west Guangdong increased in 2011-2015 compared to 2001-2010. The risk of heavy metal exposure in  
226 east Guangdong became relatively low since 2001. The health risk in Guangxi and Hainan was low overall, but  
227 slightly increased in 2011-2015 compared to 2006-2010.

228 Overall, the HI value of the six heavy metals in the coastal areas of the northern South China Sea in 1989-2015  
229 was 0.747; the health risk gradually decreased and became relatively low since 2001, but slightly increased in 2011-  
230 2015. The risk of copper exposure was the highest among the six heavy metals, followed by arsenic. The health  
231 risk of the six heavy metals exposure was highest in the Pearl River Estuary, relatively low in west Guangdong and  
232 east Guangdong, and low in Guangxi and Hainan.

### 233 Relative impact of six heavy metal levels in oysters

234 The PCA results based on the arsenic, cadmium, copper, mercury, lead and zinc levels in oysters from the northern  
235 South China Sea during 2006-2015 were summarized in Table 5 and Fig. 3. The first and second principal  
236 components explained 37.46% and 25.97% of the heavy metal variation in sampled oysters. The heavy metals  
237 contributed most to the first principal component were copper (29.78%), lead (23.76%) and mercury (22.77%). The  
238 heavy metals contributed most to the second principal component were zinc (46.60%), cadmium (29.02%) and  
239 arsenic (10.77%).

## 240 Discussion

### 241 Sources of heavy metals and pollution status in the northern South China Sea

242 Coastal waters are the key habitat for many important commercial fishery species such as oysters, and a critical  
243 component of marine ecosystems. As the industrialization and urbanization processes continue to accelerate in  
244 coastal areas, there have been increasing marine fishing pressure and environmental pollution, and coastal waters are  
245 impacted most (González-Macías et al. 2006). Due to their high toxicity, bioaccumulability and persistence, heavy  
246 metals can harm human health through food chains and therefore have drawn increasing attention in environmental  
247 science (Zhou et al. 2007).

248 This work analyzed six heavy metal levels in cultured oysters sampled from 23 sites in the coastal areas of the  
249 northern South China Sea in 1989-2015, which allowed us to be able to examine the status and trend of multiple  
250 heavy metals and their interacting effects at higher spatial and temporal scales, whereas previous studies often  
251 focused on one particular heavy metal (Wang et al. 2012; Wang et al. 2017). This work suggested that oysters from  
252 the northern South China Sea were most contaminated by copper and arsenic. Prior studies indicated that cultured  
253 oysters from the Jiaozhou Bay (Yellow Sea) were most contaminated by arsenic and mercury, followed by copper,  
254 cadmium, zinc and lead, probably caused by the discharge of industrial wastewaters (Chen 2018); cultured oysters  
255 from the Gulf of Yueqing (East China Sea) were most contaminated by copper, lead, zinc, cadmium and mercury.  
256 Different heavy metals typically have quite different potential sources; copper, lead and zinc usually come from the  
257 discharge of electroplating industry; zinc and lead are also highly related to the construction, maintenance and  
258 anchoring of vessels; cadmium and mercury are mainly from the discharge of coal-fired power stations (Tanner et  
259 al. 2000; Zhou et al. 2007). Therefore, anthropogenic sources are becoming the major sources of heavy metal  
260 pollution in coastal waters.

261 This work indicated that oysters from the northern South China Sea were most contaminated by copper. Pan et al.  
262 (2014) suggested that copper contamination has become an increasing problem for the coastal and estuarine  
263 environmental protection in China. Copper widely exists in the tissues of organisms, is an important structural  
264 element and a component of many proteins and enzymes in cells, and critical for metabolisms. However, copper  
265 above normal levels can be harmful for organisms. High concentrations of copper can directly or indirectly  
266 influence the oxidation-reduction state of cellular environment and assist produce reactive oxygen species, which  
267 could adversely affect cell structures, functions and metabolisms, or even damage cells. Nowadays copper is  
268 commonly used in electronics, construction, mechanics, electroplating, telecommunication, energy, petrochemistry

269 and aquaculture. China's copper consumption has grown from 2.2 million tons in 2001 to 7.8 million tons in 2011,  
270 and its percentage in global copper consumption has increased steadily (Pan et al. 2014). The appearance of  
271 blue/green-colored oysters is a complex biological and ecological consequence of high copper levels, and suggests  
272 the problem of copper contamination in the coastal waters of China. In recent years, copper contamination started to  
273 lessen because of improved monitoring and management strategies. Wind power industry has increasingly  
274 developed as a substitute for coal-fired power, which also help reduce the discharge of heavy metals.

275 Heavy metal pollution in the Pearl River Estuary is of the most concern in this work. Oyster heavy metal levels  
276 and the MPI values indicated that the heavy metal levels in the Pearl River Estuary were higher than the other zone  
277 of the study area in 1989-2015. One reason would be because oyster farming areas are mainly in coastal waters with  
278 depth < 10 m. The amount of sand brought through water flow to the Pearl River Estuary is  $8.366 \times 10^7$  tons per  
279 year, about 20% of which remains in the Pearl River Delta and the other 80% sedimentates in the coastal area  
280 outside the delta (Lan 1989). Lots of sand and heavy metals are brought through the Pearl River flow to the  
281 sediments in coastal waters with depth < 10 m, whereas waters with depth > 10 m are less affected by land-source  
282 heavy metal input (Ma et al. 2014). Other reasons would be because sediments in the Pearl River Estuary are  
283 mainly land source and biological source; human activities around the estuary and the Pearl River strongly increase  
284 heavy metal input; and the hydrological and geographical characteristics of the Pearl River Estuary are suitable for  
285 the sedimentation of heavy metals (Ma et al. 2012). The copper, lead and zinc levels in the sediments of the Pearl  
286 River Estuary were  $(16.1 \pm 4.0) \mu\text{g/g}$ ,  $(21.9 \pm 2.1) \mu\text{g/g}$  and  $(94.3 \pm 11.2) \mu\text{g/g}$  (Ma et al. 2014); their levels in  
287 Guangdong coastal waters were  $15.5 \mu\text{g/g}$ ,  $30.0 \mu\text{g/g}$  and  $63.3 \mu\text{g/g}$  (He and Wen 1982); and those round the South  
288 China Sea Shelf area were  $7.43 \mu\text{g/g}$ ,  $15.6 \mu\text{g/g}$  and  $54.4 \mu\text{g/g}$  (Zhang and Du 2005). As a consequence of the water  
289 pollution in the Pearl River Estuary, nowadays there has been no more oyster farming area in the city of Shenzhen,  
290 which is located by the estuary. The current major areas for oyster farming are in east and west Guangdong, mostly  
291 in the cities of Yangjiang and Zhanjiang.

## 292 Risk of heavy metal exposure to human health through oyster consumption

293 Large amounts of heavy metals from the discharge of industrial and agricultural wastewaters can directly or  
294 indirectly harm human health (Sheehan et al. 2014). Coastal residents are usually more susceptible to heavy metal  
295 exposure from higher sea food consumption than other populations (Juric et al. 2017). There have been a variety of

296 investigations on the health risk and toxicology of heavy metals (Morais 2012; Wang et al. 2018). With the  
297 application of new technologies in ecological toxicology in recent years, great achievement has been made in studies  
298 on the interacting effects of multiple heavy metals on oyster life history and the complex relationships between the  
299 chemical properties of different heavy metals, which could provide valuable guidance for oyster safety, safe  
300 consumption, heavy metal exposure risk and protecting oyster consumers' health (Wang et al. 2018).

301 Health risk assessment is to quantify the probability of adverse effects on human health due to a specific toxic  
302 agent exposure. The parameters of THQ through oyster consumption in this work were mainly from experiment and  
303 survey data. The assessment results depend mostly on the toxicity and exposure (frequency and duration) of heavy  
304 metals. Risk assessment of heavy metals is typically comprised of hazard assessment, dose-response assessment,  
305 exposure assessment, and risk assessment. The diet structure of seafood consumers varies a lot between different  
306 areas, and health risk assessment of heavy metal exposure should consider the heavy metal type and level and the  
307 diet structure (exposure frequency and duration). Oyster intake used in this work was based on a previous survey of  
308 1626 samples from 12 coastal cities in Guangdong examining pollutant exposure through food consumption (Tang  
309 et al. 2009). Their survey indicated that residents around the coastal areas of the northern South China Sea  
310 consumed more crustaceans and mollusks such as mussels, oysters and scallops than fish, especially for males, and  
311 the intake of crustaceans and mollusks could be 76.1 g/day, which was 22.1 g/day more than fish intake. It has been  
312 reported that pollution cases such as cadmium contamination in Beijiang River and arsenic poisoning in the city of  
313 Yingde happened in Guangdong (Chen 2004; Tang et al. 2009). According to the risk assessment of diet structure  
314 (e.g., vegetables, meat, fish, crustaceans, mollusks, rice, and flour) for Guangdong residents, heavy metal pollution  
315 in seafood is of substantial concern, and the risk of cadmium exposure is high. This work suggested that the health  
316 risk of the six heavy metals exposure through oyster consumption varied between different zones of the study areas  
317 and different time periods, and the risk of copper exposure was the highest among the six heavy metals examined.  
318 The risk of the six heavy metals exposure to human health was relatively low during 2001-2015, but slightly  
319 increased in 2011-2015 compared to that in 2006-2010, and therefore oyster heavy metal levels should still be a key  
320 part in the monitoring of oyster safety. The health risk of the six heavy metals exposure was most due to copper,  
321 arsenic and cadmium, followed by lead and mercury; zinc has the lowest toxicity among the six heavy metals, and is  
322 an essential element for human body.

323 It is recommended that future research on the toxicity and toxicology effects of oysters should pay more attention  
324 to the interacting toxicity and ecological risk assessment of multiple heavy metals exposure, because heavy metal  
325 exposure not only puts oysters or oyster consumers at risk, but also has interacting and combined effects on many  
326 other aquatic species, and could provide information support for the improvement of water quality standard and  
327 ecological risk assessment. Studies based on species sensitivity distributions (SSDs) typically set the target of  
328 protecting the majority (usually 95%) of species in the environment when investigating the levels of chemicals such  
329 as heavy metals (van Straalen and van Rijn 1998). Future exploration combining SSDs and quantitative structure  
330 activity relationships (QSARs) to develop ecological risk assessment methods of chemical species and structure,  
331 especially to be applied around the coastal areas of China, could serve as a basis for the management of aquatic  
332 ecological environment and seafood safety (Feng et al. 2013).

### 333 Conclusions

334 With the increasing of world population and the need for seafood safety, how to improve the ecosystem service  
335 function of limited fishery waterbodies and the sustainable supply of high-quality seafood, is crucial for a healthy  
336 development of fisheries and human well-being. This work examined the long-term spatiotemporal trends and  
337 health risk assessment of oyster arsenic, cadmium, copper, mercury, lead and zinc levels in the northern South China  
338 Sea in 1989-2015. Unlike previous studies usually focusing on one single heavy metal level in cultured oysters from  
339 this area (Zong and Jia 1996; Wang et al. 2017; Wang et al. 2020), this work considered the quantitative  
340 comparisons, interacting effects and health implications of multiple heavy metals, and the results could serve as a  
341 basis for ongoing monitoring of mollusks safety and future development and design of oyster farming.

342 Results indicated that heavy metal pollution in oysters from the northern South China Sea was most due to  
343 copper, arsenic and cadmium, followed by lead and mercury. Across different time periods, oyster heavy metal  
344 levels were highest in 1989-1990, 1991-1995 and 1996-2000 (MPI = 2.42-3.68), and gradually decreased since  
345 2001, but slightly increased in 2011-2015, which deserves further attention. Among the five zones, oyster heavy  
346 metal levels were generally highest in the Pearl River Estuary (MPI = 1.20-5.52), followed by the coastal waters of  
347 west and east Guangdong, Guangxi and Hainan. Health risk assessment according to the THQ and HI values  
348 indicated that the risk of the six heavy metals exposure was relatively high in 1989-2005, and decreased since 2006,

349 but slightly increased in 2011-2015. The quantitative comparisons and time series analysis of multiple heavy metals  
350 in cultured oysters from the study area suggested that copper contamination in the coastal waters of the northern  
351 South China Sea should draw more attention. The Pearl River Estuary generally showed the highest heavy metal  
352 levels, probably because economics and industry around the Pearl River Estuary have been growing faster than the  
353 other areas of this work in the recent two decades, and it should be taken as a hotspot for the monitoring of seafood  
354 safety in southern China. The heavy metal pollution caused by increasing industrialization and urbanization  
355 processes in some areas might outcompete the ongoing environmental protection and recovery programs in the past  
356 decade. Therefore, the environmental protection and management of aquaculture and seafood safety should be a  
357 long-term focus for the sustainability of fishery ecosystems in China.

358

359 **Declarations**

360 **Ethics approval and consent to participate** Not applicable.

361 **Consent for publication** Not applicable.

362 **Availability of data and materials** All data generated or analyzed during this study are included in this article.

363 Source data of the South China Sea Mussel Watch Program are available from the corresponding authors upon  
364 reasonable request.

365 **Competing interests** The authors declare no competing interests.

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467

468 **Table 1** Values of metal pollution index (MPI) and criterion for pollution status

Index value	Degree of Pollution
$MPI \leq 2$	Not impacted
$2 < MPI \leq 5$	Very low contamination
$5 < MPI \leq 10$	Low contamination
$10 < MPI \leq 20$	Medium contamination
$20 < MPI \leq 50$	High contamination
$50 < MPI \leq 100$	Very high contamination
$MPI > 100$	Extreme contamination

469

470 **Table 2** Reference dose (Rfd) of arsenic (As), cadmium (Cd), copper (Cu), mercury (Hg), lead (Pb) and zinc (Zn)\*

Metal	Cu	Zn	Pb	Cd	As	Hg
Rfd ( $\mu\text{g/g/day}$ )	0.0400	0.3000	0.0035	0.0010	0.0003	0.0005

471 \*The Rfd values referred to US EPA (2013), United States Environmental Protection Agency, Integrated Risk  
472 Information System, <http://www.epa.gov/iris>. The Rfd value of arsenic was based on the value of inorganic arsenic.  
473

474 **Table 3** Mean concentrations ( $\mu\text{g/g}$  wet weight) of arsenic (As), cadmium (Cd), copper (Cu), mercury (Hg), lead  
 475 (Pb) and zinc (Zn) in oysters from the northern South China Sea in 1989-2015. Oyster heavy metal levels in  
 476 Guangxi and Hainan were not available until 2006 due to logistic constraints

Zone	Year	Cu	Zn	Pb	Cd	As	Hg	MPI
Northern South China Sea	1989-1990	429.45	166.98	0.72	1.05	1.44	0.04	<b>3.82</b>
	1991-1995	85.58	198.34	0.67	0.62	0.95	0.03	<b>2.42</b>
	1996-2000	43.33	127.94	0.32	1.58	0.89	NaN	<b>3.68</b>
	2001-2005	47.34	140.47	0.21	0.78	0.34	0.02	1.4
	2006-2010	28.76	153.72	0.2	0.84	0.1	0.01	0.95
	2011-2015	80.02	135.86	0.34	0.86	0.52	0.01	1.6
	Overall	53.85	149.23	0.31	0.85	0.33	0.01	1.38
East Guangdong	1989-1990	139.45	71.35	0.8	0.57	1.71	0.05	<b>2.7</b>
	1991-1995	63.29	217.03	0.55	0.43	1.15	0.03	<b>2.2</b>
	1996-2000	32.33	117.40	0.37	1.22	0.92	NA	<b>3.41</b>
	2001-2005	32.02	127.44	0.25	0.46	0.38	0.02	1.24
	2006-2010	24.29	151.96	0.19	0.75	0.12	0.01	0.93
	2011-2015	74.59	141.49	0.38	0.81	0.51	0.01	1.6
	Overall	42.25	146.19	0.32	0.65	0.4	0.01	1.31
Pearl River Estuary	1989-1990	927.44	293.82	0.71	1.79	1.26	0.03	<b>4.85</b>
	1991-1995	162.7	294.5	0.77	0.95	0.89	0.03	<b>3.13</b>
	1996-2000	117.62	203.55	0.52	2.59	0.88	NaN	<b>5.52</b>
	2001-2005	191.92	232.65	0.32	2.09	0.27	0.02	<b>2.33</b>
	2006-2010	59.99	185.18	0.27	1.39	0.07	0.01	1.2
	2011-2015	84.75	158.37	0.32	1.17	0.4	0.01	1.65
	Overall	122.35	210.49	0.42	1.49	0.32	0.01	1.93
West Guangdong	1989-1990	512.83	192.58	0.69	1.1	1.41	0.05	<b>4.17</b>
	1991-1995	92.95	162.11	0.77	0.78	0.8	0.03	<b>2.45</b>
	1996-2000	39.7	116.46	0.23	1.71	0.85	NaN	<b>3.4</b>
	2001-2005	48.7	134.24	0.15	1.04	0.32	0.02	<b>2.62</b>
	2006-2010	33.9	148.96	0.18	0.94	0.1	0.01	0.97
	2011-2015	88.54	133.75	0.34	0.91	0.6	0.01	1.67
	Overall	59.64	141.85	0.29	0.98	0.35	0.02	1.6
Guangxi	2006-2010	30.9	224.6	0.18	1.29	0.11	0.01	1.1
	2011-2015	79.23	155.4	0.34	0.79	0.52	0.01	1.61
	Overall	49.48	186.61	0.25	1.01	0.24	0.01	1.33
Hainan	2006-2010	18.79	125.7	0.24	0.53	0.09	0.01	0.8
	2011-2015	72.77	113.98	0.27	0.8	0.47	0.02	1.6
	Overall	36.98	119.7	0.26	0.65	0.21	0.02	1.21

477 NaN: not available

478

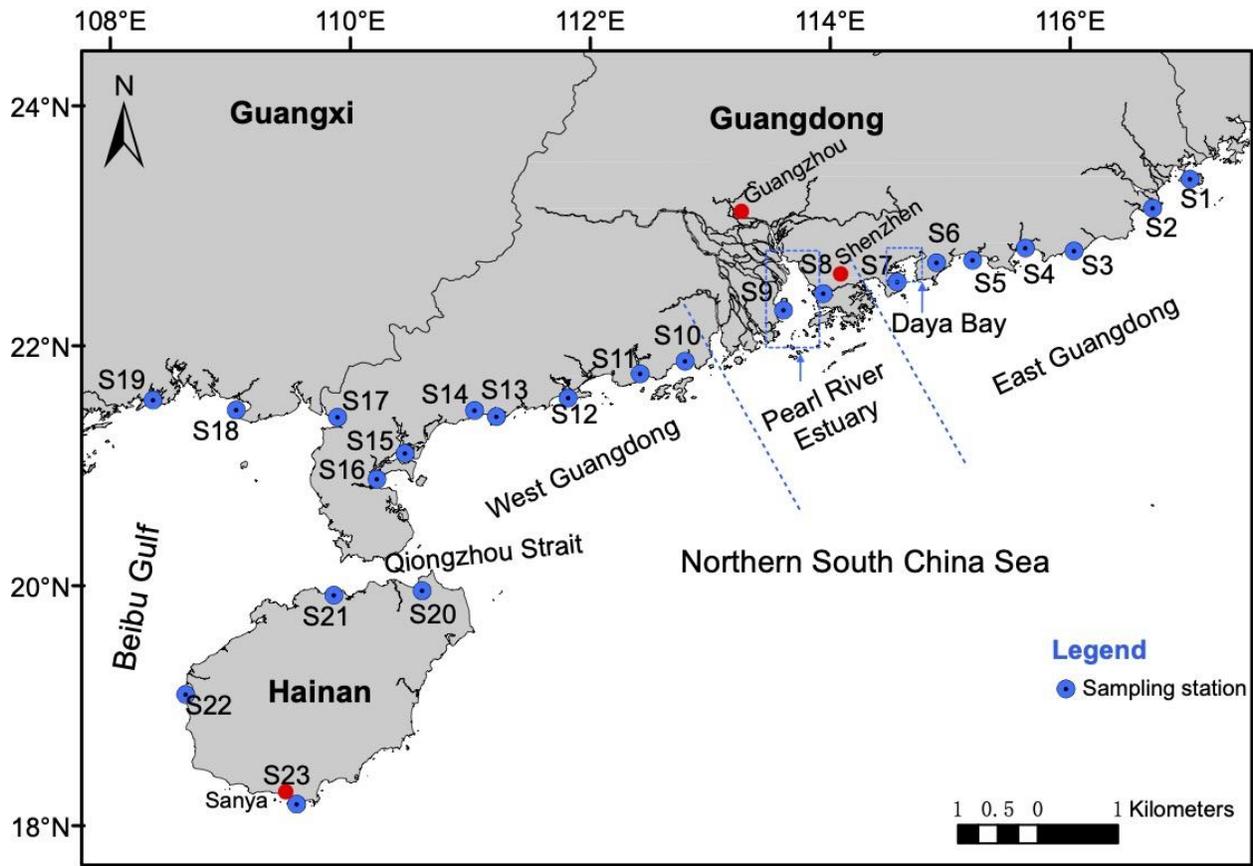
479 **Table 4** Target hazard quotient (THQ) and hazard index (HI) of arsenic (As), cadmium (Cd), copper (Cu), mercury  
 480 (Hg), lead (Pb) and zinc (Zn) exposure through oyster consumption in the coastal areas of the northern South China  
 481 Sea in 1989-2015

Zone	Year	THQ <sub>Cu</sub>	THQ <sub>Zn</sub>	THQ <sub>Pb</sub>	THQ <sub>Cd</sub>	THQ <sub>As</sub>	THQ <sub>Hg</sub>	HI
Northern South China Sea	1989-1990	2.055	0.107	0.039	0.201	0.919	0.015	<b>3.336</b>
	1991-1995	0.410	0.127	0.037	0.119	0.606	0.011	<b>1.309</b>
	1996-2000	0.207	0.082	0.018	0.302	0.568		<b>1.177</b>
	2001-2005	0.227	0.090	0.011	0.149	0.217	0.008	0.702
	2006-2010	0.138	0.098	0.011	0.161	0.064	0.004	0.475
	2011-2015	0.383	0.087	0.019	0.165	0.332	0.004	0.989
	Overall	0.258	0.095	0.017	0.163	0.211	0.004	0.747
East Guangdong	1989-1990	0.667	0.046	0.044	0.109	1.091	0.019	<b>1.976</b>
	1991-1995	0.303	0.138	0.030	0.082	0.734	0.011	<b>1.299</b>
	1996-2000	0.155	0.075	0.020	0.234	0.587		<b>1.070</b>
	2001-2005	0.153	0.081	0.014	0.088	0.242	0.008	0.586
	2006-2010	0.116	0.097	0.010	0.144	0.077	0.004	0.448
	2011-2015	0.357	0.090	0.021	0.155	0.325	0.004	0.952
	Overall	0.202	0.093	0.018	0.124	0.255	0.004	0.696
Pearl River Estuary	1989-1990	4.438	0.187	0.039	0.343	0.804	0.011	<b>5.823</b>
	1991-1995	0.779	0.188	0.042	0.182	0.568	0.011	<b>1.770</b>
	1996-2000	0.563	0.130	0.028	0.496	0.562		<b>1.779</b>
	2001-2005	0.918	0.148	0.018	0.400	0.172	0.008	<b>1.664</b>
	2006-2010	0.287	0.118	0.015	0.266	0.045	0.004	0.735
	2011-2015	0.406	0.101	0.018	0.224	0.255	0.004	<b>1.007</b>
	Overall	0.586	0.134	0.023	0.285	0.204	0.004	<b>1.236</b>
West Guangdong	1989-1990	2.454	0.123	0.038	0.211	0.900	0.019	<b>3.744</b>
	1991-1995	0.445	0.103	0.042	0.149	0.510	0.011	<b>1.262</b>
	1996-2000	0.190	0.074	0.013	0.327	0.542		<b>1.147</b>
	2001-2005	0.233	0.086	0.008	0.199	0.204		0.730
	2006-2010	0.162	0.095	0.010	0.180	0.064	0.004	0.515
	2011-2015	0.424	0.085	0.019	0.174	0.383	0.004	<b>1.089</b>
	Overall	0.285	0.091	0.016	0.188	0.223	0.008	0.810
Guangxi	2006-2010	0.148	0.143	0.010	0.247	0.070	0.004	0.622
	2011-2015	0.379	0.099	0.019	0.151	0.332	0.004	0.984
	Overall	0.237	0.119	0.014	0.193	0.153	0.004	0.720
Hainan	2006-2010	0.090	0.080	0.013	0.101	0.057	0.004	0.346
	2011-2015	0.348	0.073	0.015	0.153	0.300	0.008	0.896
	Overall	0.177	0.076	0.014	0.124	0.134	0.008	0.534

482 **Table 5** Contributions of arsenic (As), cadmium (Cd), copper (Cu), mercury (Hg), lead (Pb) and zinc (Zn) to the  
483 principal components (PC) in principal component analysis (PCA) for oyster heavy metal levels in the northern  
484 South China Sea in 2006-2015

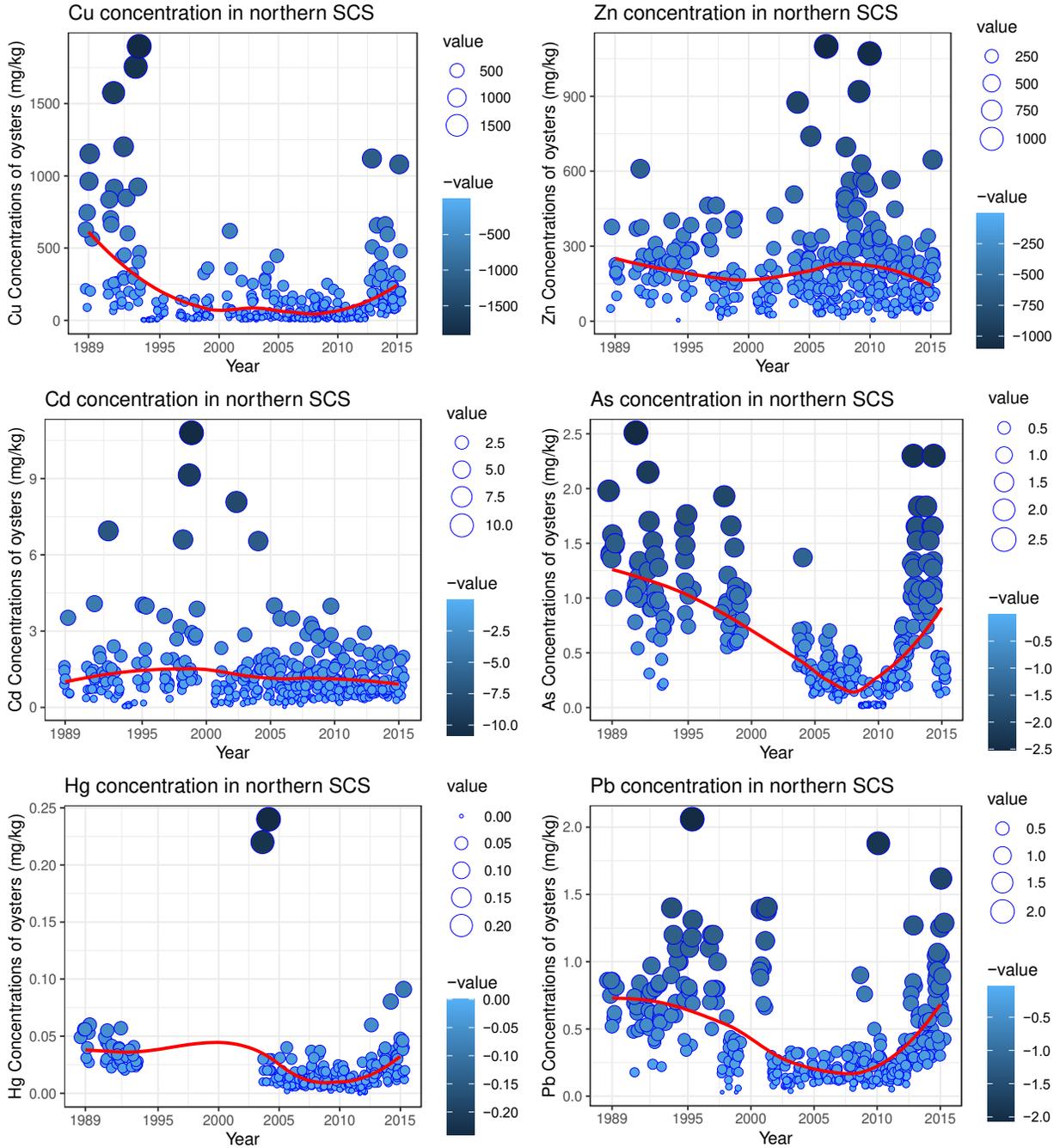
	PC1	PC2	PC3	PC4	PC5	PC6
Cu	29.78	0.92	5.8	22.87	0.24	40.39
Zn	1.02	46.6	1.59	0.01	47.26	3.52
Pb	23.76	0.31	36.71	13.82	5.9	19.5
Cd	3.8	39.02	8.61	6.53	40.3	1.74
As	18.87	10.77	37.74	0.6	5.42	26.6
Hg	22.77	2.38	9.54	56.16	0.88	8.26

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486

487 **Fig. 1** Map of 23 sampling sites for cultured oysters collected from the northern South China Sea in 1989-2015.  
 488 Sampling sites S1 to S23 were Zhelin Bay, Guangao Bay, Jiazi Harbor, Jieshi Bay, Changsha Bay, Kaoyang  
 489 Estuary, Yaling Bay, Shenzhen Bay, Tangjia Bay, Guanghai Bay, Zhenhai Bay, Mawei Bay, Bohe Harbor,  
 490 Shuidong Harbor, Zhanjiang Harbor, Leizhou Bay, Anpu Harbor, Beihai Harbor, Fangcheng Harbor, Dongzhai  
 491 Harbor, Maniao Harbor, Basuo Harbor, and Yulin Harbor. S1-S7 were in east Guangdong, S8 and S9 were in the  
 492 Pearl River Estuary, S10-S17 were in west Guangdong, S18 and S19 were in Guangxi, and S20-S23 were in Hainan  
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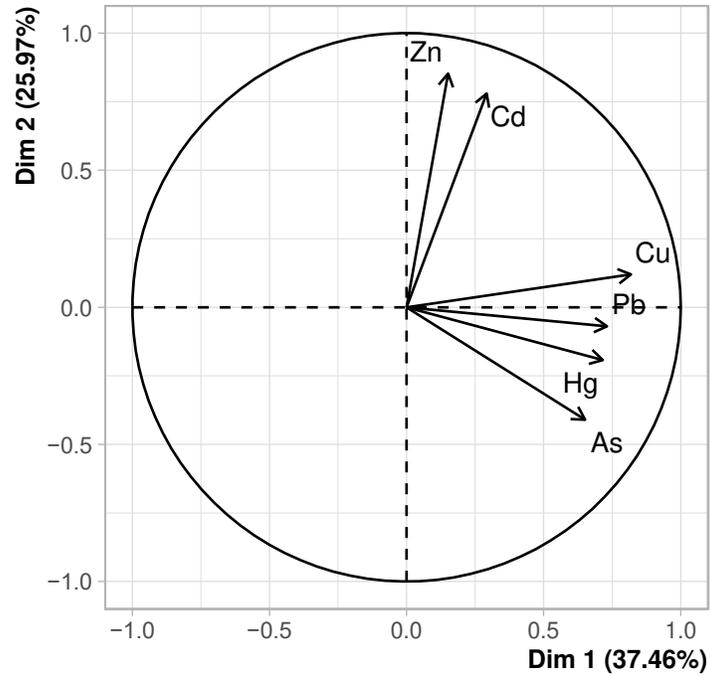
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**Fig. 2** Temporal variation of arsenic (As), cadmium (Cd), copper (Cu), mercury (Hg), lead (Pb) and zinc (Zn) in oysters from the northern South China Sea (SCS) in 1989-2015, according to locally weighted least squares regression scatterplot smoothing (LOWESS)



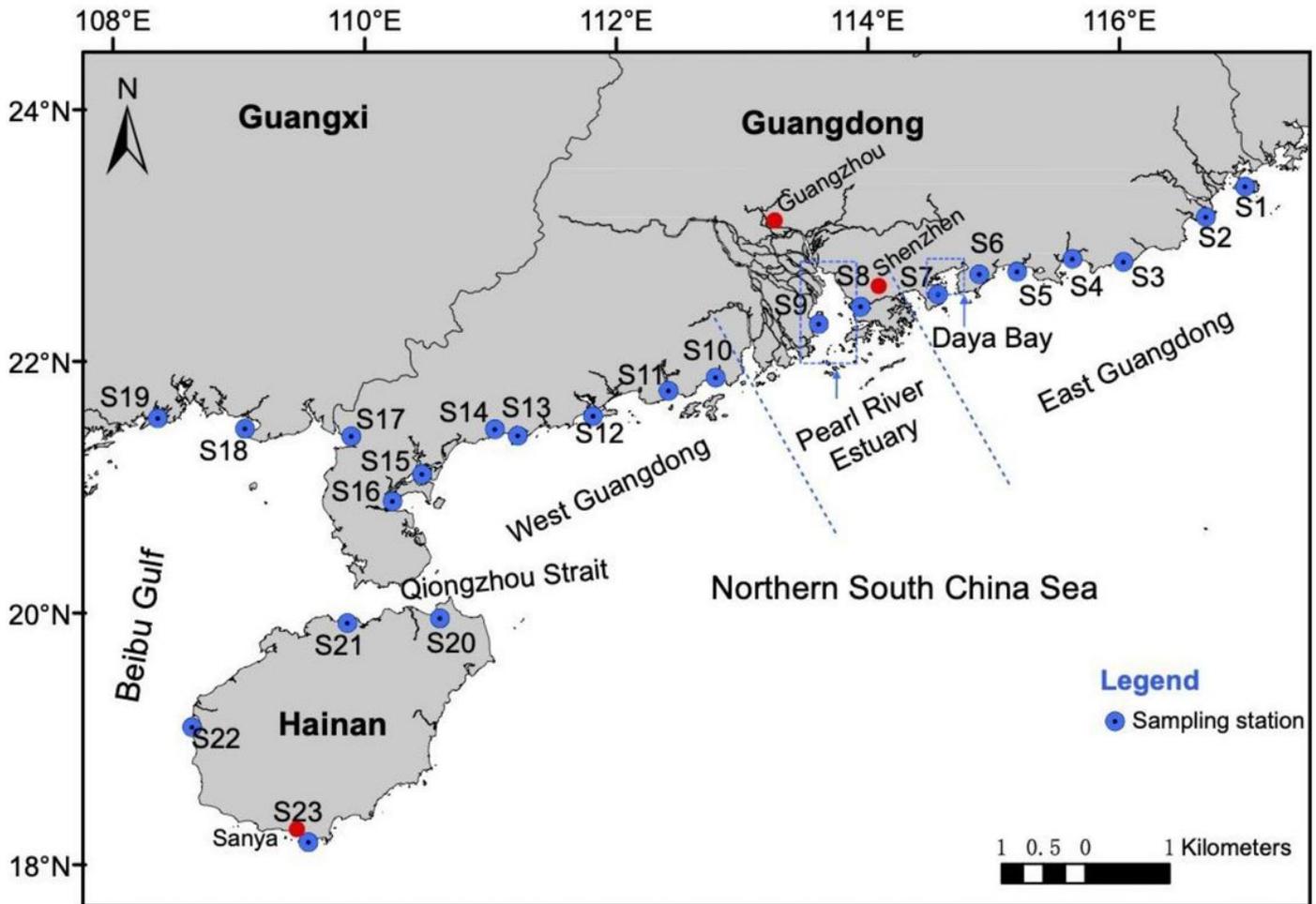
501

502 **Fig. 3** Relationships between arsenic (As), cadmium (Cd), copper (Cu), mercury (Hg), lead (Pb) and zinc (Zn)

503 based on the first two principal components (PC) in principal component analysis (PCA) for oyster heavy metal

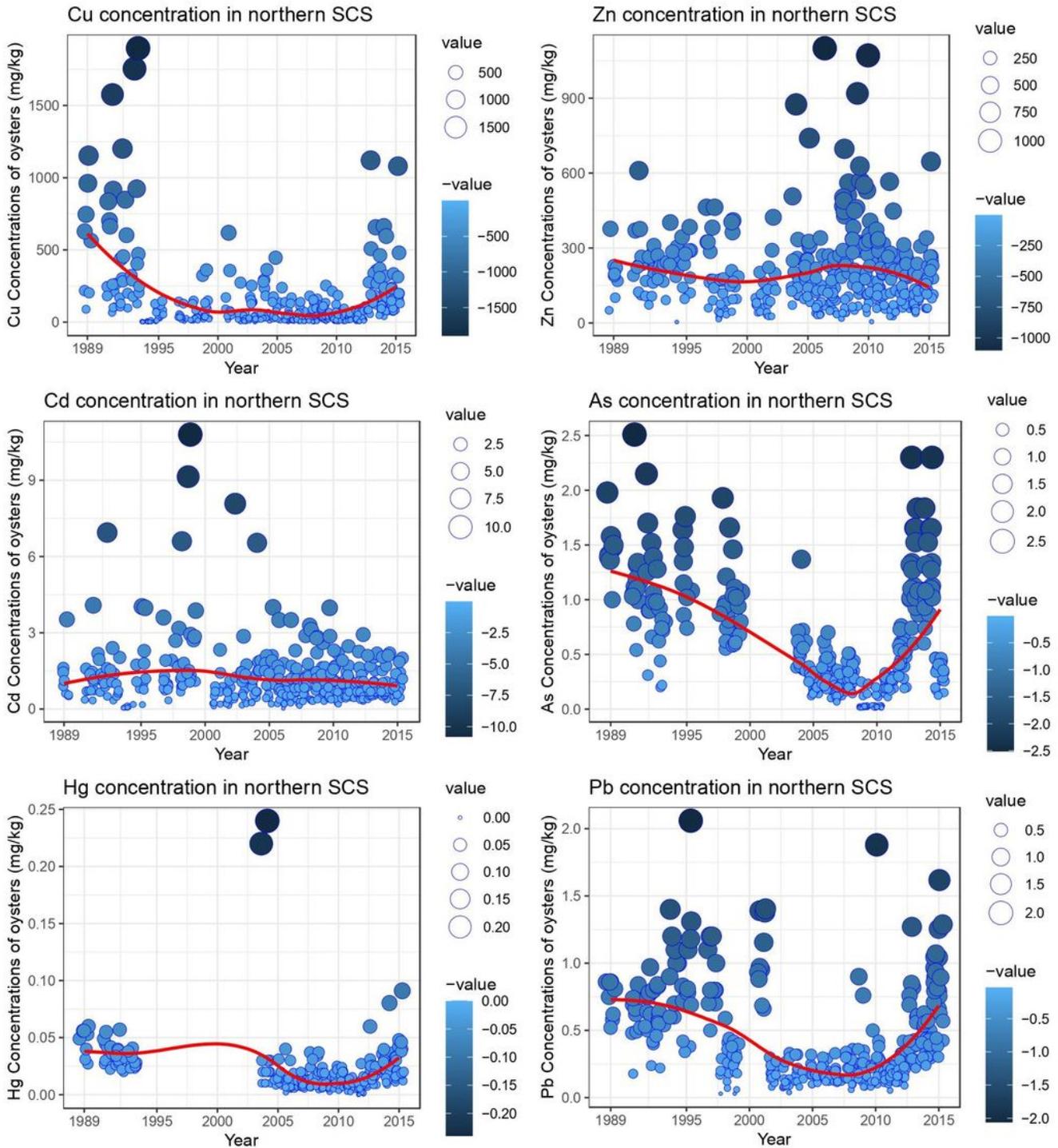
504 levels in the northern South China Sea in 2006-2015

# Figures



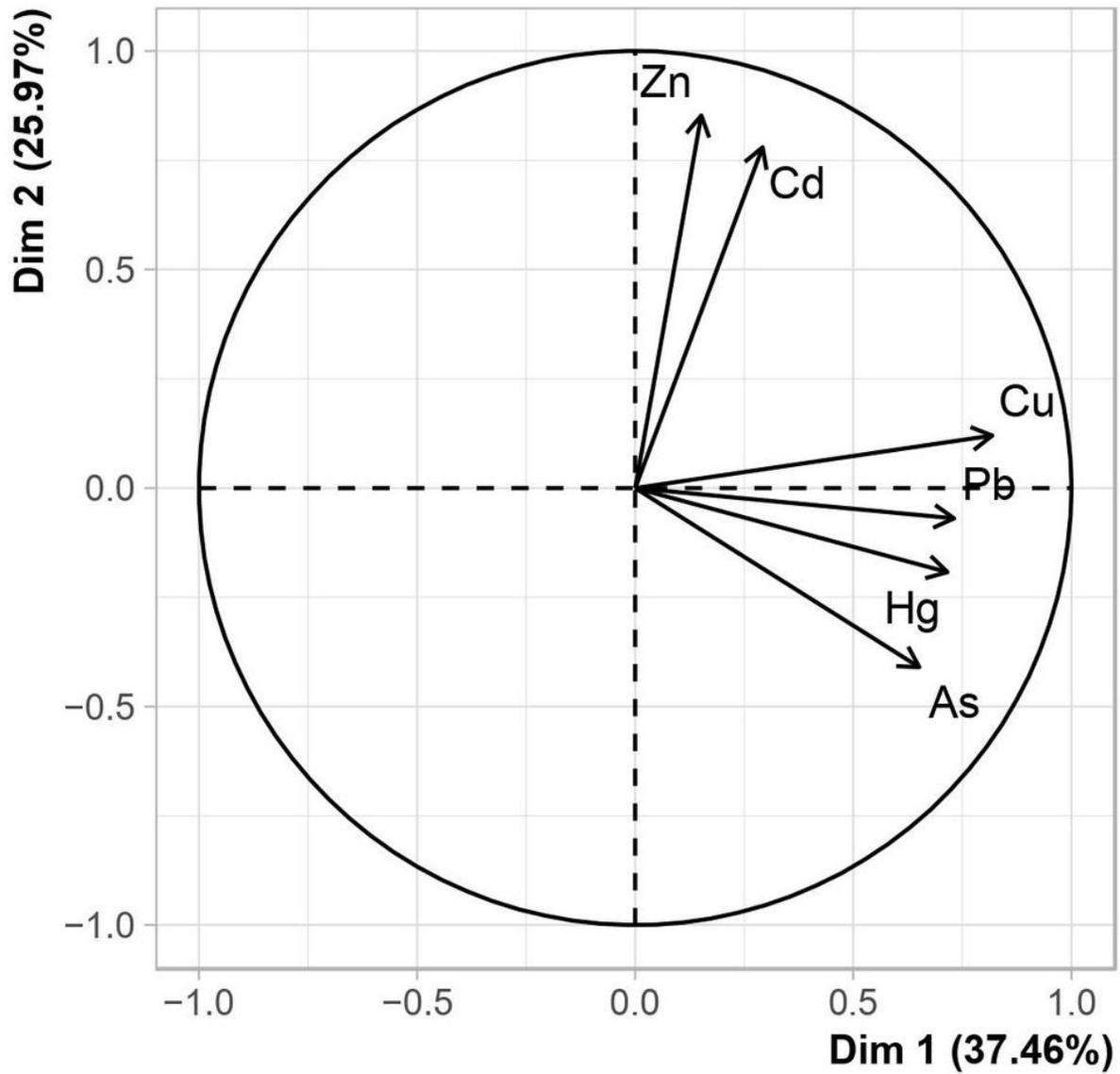
**Figure 1**

Map of 23 sampling sites for cultured oysters collected from the northern South China Sea in 1989-2015. Sampling sites S1 to S23 were Zhelin Bay, Guangao Bay, Jiazi Harbor, Jieshi Bay, Changsha Bay, Kaoyang Estuary, Yaling Bay, Shenzhen Bay, Tangjia Bay, Guanghai Bay, Zhenhai Bay, Mawei Bay, Bohe Harbor, Shuidong Harbor, Zhanjiang Harbor, Leizhou Bay, Anpu Harbor, Beihai Harbor, Fangcheng Harbor, Dongzhai Harbor, Maniao Harbor, Basuo Harbor, and Yulin Harbor. S1-S7 were in east Guangdong, S8 and S9 were in the Pearl River Estuary, S10-S17 were in west Guangdong, S18 and S19 were in Guangxi, and S20-S23 were in Hainan Note: The designations employed and the presentation of the material on this map do not imply the expression of any opinion whatsoever on the part of Research Square concerning the legal status of any country, territory, city or area or of its authorities, or concerning the delimitation of its frontiers or boundaries. This map has been provided by the authors.



**Figure 2**

Temporal variation of arsenic (As), cadmium (Cd), copper (Cu), mercury (Hg), lead (Pb) and zinc (Zn) in oysters from the northern South China Sea (SCS) in 1989-2015, according to locally weighted least squares regression scatterplot smoothing (LOWESS)



**Figure 3**

Relationships between arsenic (As), cadmium (Cd), copper (Cu), mercury (Hg), lead (Pb) and zinc (Zn) based on the first two principal components (PC) in principal component analysis (PCA) for oyster heavy metal levels in the northern South China Sea in 2006-2015