

Deposition of potentially toxic metals in the soil from surrounding cement plants in a karst area of Southeastern Brazil

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14 **Abstract**

15

16 Cement factories are the main sources of environmental pollutants among the different industrial activities,
17 including soil contamination by potentially toxic metals and the Karst region of Southeastern Brazil is
18 known for the implementation of large cement producing facilities. This study aims to evaluate whether
19 there is an increase in the concentration of PTM in the soil surrounding the cement plants and to estimate
20 their harmfulness to both local human population and environment. In total, 18 soil samples were collected
21 from the surroundings of three cement plants as well as four soil samples from areas outside the influence
22 of cement plants and concentration of the following potentially toxic metals (PTM) were estimated: Cd,
23 Pb, Co, Cu, Cr, Mn, Ni, and Zn. The results revealed that all PTM concentrations from cement plant
24 surroundings were significantly higher than PTM concentrations from control areas and no PTM
25 concentrations from CPS or CA soil samples exceeded national and global contamination thresholds.
26 However, Igeo Index indicated low level soil contamination by Pb, Cu and Cr and high levels for Co. We
27 could not verify significant Non-carcinogenic risk to health for any soil sample, but carcinogenic risk
28 analysis revealed different levels of carcinogenic risk among the sampled locations, for both adults and
29 children. Our results indicate that exclusively evaluating the concentration of potentially toxic metals is not
30 enough to verify the potential harmful effects of cement production for the surrounding population. Here
31 we evidence that additional indices, based on both contamination indices and health risk assessments,
32 should be considered for better evaluation of the impacts of cement production activity.

33

34 **Keywords:** heavy metals, soil, cement, environmental risk, pollution

35

36 **Introduction**

37

38 Soils are remarkable systems, constituted of an extraordinarily diverse range of mineral and
39 organic components, organized and interactive in particular ways that result in the delivery of the range of
40 ecosystem services, contributing to basic human needs such as food, water, and air supply (Keesstra et al.
41 2016, Wall et al. 2012). Soil ecosystem services depend on soil properties and their interactions, and are
42 mostly influenced by its use and management (Adhikari and Hartemink 2016). The intensification and
43 expansion of human activities has increased the pressure on land resources and led to soil degradation
44 (Levin et al., 2017). One of the main concerns regarding the impact of human activities in soil systems are
45 soil contamination events by potentially toxic metals (PTM), since they poses alarming threats to
46 agricultural productivity, food safety, and human health (Naila et al. 2019). Potentially toxic metals (PTM)
47 – the most widespread term for heavy metals – are those with high density ($> 5\text{g cm}^{-3}$), compared to other
48 elements and atomic number greater than 20. These metals are resistant to degradation and can accumulate
49 in the components in which they manifest their toxicity (Yadav et al. 2019).

50 Unlike organic pollutants, PTM cannot be biodegraded and, mainly due to their cumulative power
51 in living organisms, posing risks to health and to the environment (Karimi et al. 2020, Tsezos 2009).
52 Moreover, PTM can remain in the soil for long periods, depending on soil retention capacity and physical-
53 chemical properties (El Sherbiny et al. 2019, Kabata-Pendias and Mukherjee 2007). Soil contamination by
54 PTM is usually the result of anthropogenic activities related to urbanization, including automotive exhaust
55 pipe, and industrialization (Qasemi et al. 2018, Li and Feng 2012). Cement plants are one of the most
56 common sources of PTM pollution by gas emissions and production of cement powder (Jafari et al. 2019).

57 The cement industry has a high polluting potential at all stages of production (Yadegarnia et al.
58 2019). The levels and characteristics of pollutant emissions depend on the technological and operational
59 characteristics of the industrial process, especially rotary kiln furnaces, the chemical and mineralogical
60 composition of raw materials, the chemical composition of the fuels used; the operational speed of kiln
61 furnaces, and the efficiency of the emission control systems of installed pollutants (Chen et al. 2015). The
62 primary pollutants emitted in the cement manufacturing process are particulate matter, carbon dioxide,
63 sulfur oxides, and nitrogen oxides (Bermudez et al. 2010). According to the U.S. environmental agency,
64 cement manufacturing plants are among the largest sources of hazardous pollutant emissions, including
65 PTM (Ogunkunle and Fatoba 2014, USEPA 2013, Yahaya et al. 2013). Significant parts of these pollutants

66 are deposited in the soil, including the burning of fuels for the kiln furnaces, generating fine dust particles
67 that are deposited in the soil (Mandal and Voutchkov 2011). This dust contains PTM that are deposited in
68 the soil and cause serious environmental and health impacts since it is not biodegraded, accumulating
69 (Ogunkunle and Fatoba 2013).

70 About 2-3% of the Brazilian territory is formed by karst environments, structured in 19 karst
71 regions (Travassos, 2019). These regions are characterized by vast areas of sedimentary carbonate rocks,
72 with limestone, dolomite rock, and marbles being the most common components. Due to calcareous
73 composition, these regions in Brazil are commonly used for limestone exploitation for use in agriculture,
74 but also as an input in the manufacturing process of Portland cement, in which it is common to install this
75 type of facility in the karstic regions of Brazil (Travassos, 2019). This study was carried out in an area
76 located at the Southern portion of one of the largest Brazilian karst regions (called Grupo Bambuí), at
77 Southeastern Brazil. This area represents a mosaic of environments characterized by areas used for
78 agricultural production, limestone mining, small urban settlements, and presence of cement plants (IBGE
79 2010).

80 Here, we hypothesized that cement production in the study area leads to the deposition of PTM in
81 its surroundings and that these events of increasing in PTM concentrations can be potentially harmful to
82 vicinal human population of the cement plants. Thus, our main goal is to evaluate whether there is an
83 increase in the concentration of PTM in the soil surrounding the cement plants and to estimate their
84 harmfulness to both local human population and environment, applying worldwide used indices. This
85 investigation also aims to provide evidence that conventional strategies for estimating of soil contamination
86 by cement production plants adopted by local regulatory agencies are not sufficient to estimate the risks of
87 PTM soil contamination to human health.

88 **Materials and methods**

89

90 *Sampling and quantification of potentially toxic metals in the soil*

91

92 The study area comprises the Southern portion of the widest karstic region in Brazil, called Grupo
93 Bambuí, located in the Southeast region of Brazil (Figure 1). In this area, soil samples were collected from
94 the surroundings of three Portland Cement plants (named CPS1, CPS2 and CPS3) and also soil samples in
95 two agricultural production areas located outside the influence of cement production plants (10 km far from

96 CPSs) taken as control areas (CA). Six sampling points were selected around each cement production plant
97 (CPS), in which each point presented a plant distance of less than 500 meters (a circular scheme). In each
98 control area, two sampling points were selected. The samples were collected in a single round, performed
99 in May 2020 (dry season). In each sampling point, approximately 2000 g of surface soil was collected with
100 depth between 0 and 15 cm, following procedures described in Pacchioni et al. (2014).

101 Concentration of the following potentially toxic metals (PTM) were estimated in soil samples: Cd,
102 Pb, Co, Cu, Cr, Mn, Ni, and Zn. The quantification of heavy metals was performed following the
103 methodology of air-acetylene flame atomic absorption spectrometry proposed by USEPA 3050B (USEPA,
104 1998a) and USEPA 3051A (USEPA, 1998b). All quantifications were performed by a laboratory facility
105 certified by local regulatory agencies (Oceanus-Hidroquímica Laboratory, certification code:
106 UN015590/55.11.10).

107

108 *Analysis of potentially toxic metals concentrations in the soil*

109

110 Soil PTM contamination analyses were performed using the following internationally recognized
111 indices: Enrichment Factor (EF) and Geoaccumulation Index (Igeo). In addition to the indexes, raw data
112 about PTM concentration were also included to compare them with previous studies and also to compare
113 with soil PTM contamination limits established in both Brazilian (CONAMA No. 420/09, BRASIL 2009)
114 and local (COPAM No. 166/11, MINAS GERAIS 2011) laws.

115 The Enrichment Factor (EF) was used to evaluate the degree of enrichment of PTM in soil. EF is
116 used to assess the level of human impact, by differentiation anthropogenic source from natural sources of
117 PTM (Barbieri 2016). The EF value near to unity indicates natural origin, those less than unity suggests
118 possible mobilization/depletion of metal, while EF value greater than unity recommends that the element
119 is of anthropogenic origin (Yadav et al. 2019). EF index was calculated by equation as follows, where C_i
120 is the concentration of individual elements in soil (mg/kg); C_{Ref} is the concentration of reference element
121 for normalization (mg/kg). In this study, used Fe as reference element, so that Fe concentration was
122 determined for all soil samples to perform EF calculations.

123

$$EF = (C_i/C_{Ref})_{soil} / (C_i/C_{Ref})_{background}$$

124

125 The Geoaccumulation Index (Igeo) was calculated to estimate the degree of soil pollution by a
given PTM, comparing observed PTM concentration with its reference concentration, proposed by Muller

126 (1969). Also, Igeo index was ranked in seven classes of soil pollution, described in Mohammadi et al.
127 (2020). Igeo was calculated by equation as follows, where C_i is the PTM concentration in soil samples, and
128 B_i is the background concentration of these elements. Background values (B_i) used in this study were based
129 on reference values proposed for carbonate sedimentary rocks by Turekian and Wedepohl (1961) and they
130 were presented in Table S1.

$$131 \quad I_{geo} = \log_2 \left(\frac{C_i}{1.5 \cdot B_i} \right)$$

132 We performed non-paired Wilcoxon tests to compare PTM concentration as well as adopted
133 contamination indices (EF and Igeo) between the CPS and CA soil samples, using R statistical program (R
134 version 3.3.2).

135

136 *Health risk assessment*

137

138 In this study, non-carcinogenic health risks assessments (NCR) for CPS vicinal human populations
139 by soil PTM contamination was estimated using USEPA health risk methods for oral, dermal, and contact
140 pathways (USEPA 2013). First, the Average Daily Dose (ADD) in milligrams per kilogram per day values
141 of PTM through oral, dermal, and inhalation pathways was calculated using the following equations.
142 Reference parameters for both adult and children exposure were presented in Table 1.

$$143 \quad ADD_{ing} = C \times \frac{IngR \times EF \times ED \times CF}{BW \times AT}$$

$$144 \quad ADD_{inh} = C \times \frac{InhR \times EF \times ED}{PEF \times BW \times AT}$$

$$145 \quad ADD_{derm} = C \times \frac{SA \times SL \times ABS \times EF \times ED \times EF}{BW \times AT}$$

146 As a second step, ADDs were used to calculate the Hazard Quotient (HQ) by dividing the daily
147 exposure doses (ADD_{ing} , ADD_{inh} , and ADD_{derm}) for each PTM with its corresponding reference doses (RfD)
148 ($\text{mg kg}^{-1} \text{ day}^{-1}$). Adopted RfDs and calculated ADDs values for the eight PTM were presented in Table S2
149 and Table S3, respectively. Then, to estimate the non-carcinogenic risks of PTM for humans (NCR), we
150 calculated the Hazard Index (HI) which corresponds to the sum of HQ indices: $NCR = HI = \sum HQ_i$. $HI < 1$
151 indicates that there are no significant risks of non-cancerous effect, and $HI > 1$ indicates a chance that non-
152 carcinogenic effects may occur, with a probability that it tends to increase as HI value increases (Wang et
153 al. 2020).

154 The carcinogenic risks to humans (CR) was estimated as described by Yadav et al. (2019), which
155 CR corresponds to the average daily doses (Σ ADD) multiplied by respective slope factor (SF). A slope factor
156 is an upper bound probability of an individual developing cancer as a result of a lifetime exposure to an
157 agent by ingestion or inhalation. Adopted Slope Factors for the eight PTM were presented in Table S2.
158 When $CR < 10^{-6}$, no cancer risk exists. When $10^{-6} < CR < 10^{-4}$, the risk is within the acceptable range. When
159 $CR > 10^{-4}$, human tolerance is exceeded.

160 **Results**

161

162 Our first approach in this study was to compare PTM concentrations estimated for CPS and CA
163 soil samples. The results of these comparative analyses revealed that all PTM concentrations from CPS
164 were significantly higher than PTM concentrations from CA areas (Figure 2). Moreover, when we
165 compared PTM concentrations between the three CPS (CPS1, CPS2 and CPS3), we could not verify
166 significant differences in PTM concentrations between the three CPS, except for Cr element, in which soil
167 samples from CPS3 presented values higher than CPS1 and CPS2 (Table S4).

168 Here, we also compared PTM concentrations from CPS soil samples with soil contamination
169 thresholds established by local (COPAM No. 16/11, MINAS GERAIS, 2011), national (CONAMA No.
170 420/09, BRAZIL, 2009) and global (WHO, 1993) recommendations. The results revealed that several CPS
171 soil samples exceeded local PTM contamination threshold for Co and Pb. However, no PTM concentrations
172 from CPS or CA soil samples exceeded national and global contamination thresholds (Table S1). In
173 addition, when we compared PTM concentrations estimated from CPS or CA soil samples with PTM
174 concentration values reported for PTM soil contamination events worldwide by previous studies, we could
175 observe that the levels of PTM concentration in CPS or CA soil samples were lower than those described
176 in reported PTM soil contamination events (Table 2).

177 The results of our PTM enrichment analyses, estimated by the EF Index, revealed that the soil
178 enrichment by PTM was superior in CPS areas when compared with control areas (CA) for all studied PTM
179 (Table 3). However, we could verify high enrichment level only for Co element in CPS soil samples ($EF >$
180 5), with a single sampling point showing a very high level of enrichment ($EF > 20$). Similarly, our Igeo
181 analysis of soil contamination by PTM revealed that the levels of soil contamination in CPS soil samples
182 were higher when compared with CA samples, for all PTM (Figure 3). Considering the mean values of Igeo
183 Index, we did not verify significant contamination by Mn, Ni and Zn in both CPS and CA soil samples.

184 Low level soil contamination were verified for Pb, Cu and Cr elements. However, Igeo Index for Co
185 element showed varying degrees of contamination between soil samples, ranging from moderate to extreme
186 level of soil contamination (Figure 3).

187 Our non-carcinogenic health risk (NCR) analyzes, estimated by Hazard Quotient (HQs) and
188 Hazard Index (HI) calculations, revealed that for all sampled locations (CPS1, CPS2, CPS3 and CA), we
189 could not verify significant risk to health ($HI > 1$), for all studied PTM. Still, similar NCR results were
190 verified for both adults and children estimates. On the other hand, carcinogenic risk analysis (CR) revealed
191 different levels of carcinogenic risk (CR) among the sampled locations (CPS1, CPS2, CPS3 and CA), for
192 both adults and children (Table 5). In addition, the five PTMs, which was possible to perform the CR
193 analysis (Cd, Pb, Co, Cr and Ni), also presented distinct levels of CR among the sampled locations where
194 soil were accessed (Table 5).

195 **Discussion**

196

197 Several studies reporting events of soil enrichment contamination by PTM in Brazilian territory
198 have been published recently: Brito et al. (2020) also estimated PTM concentration in soils from areas
199 under agricultural exploration and natural Cerrado vegetation, but in Northeast region; Silva et al. (2016)
200 verified and differentiated the sources of emission of PTMs in soils used for sugarcane cultivation in
201 Northeast region; Smidt et al. (2011) identified potential sources of pollution by PTMs in soils highly
202 influenced by a phosphorus fertilizer factory. However, this is the first study performed in Brazil focused
203 on estimating potential soil PTM contamination caused by cement plants operation and their potential health
204 risks to local human population. Some studies have already been published relating concerning events of
205 soil contamination by PTM and the anthropic activity of cement production in some regions of the world,
206 reporting events of PTM soil contamination more severe than those observed in this study (Table 2),
207 probably related to the fact that these previous studies have been performed in countries whose laws about
208 gas emissions as well as soil contamination by PTM are less restrictive than in Brazil (El-Sherbiny et al.
209 2019; Kolo et al. 2018, Ogunkunle and Fatoba, 2014, Mandal and Voutchkov, 2011; Al-Khashman and
210 Shawabkeh, 2006). On the other hand, when considering studies on PTM contamination events in general,
211 China is the country where most of these studies have been reported, indicating that soils from areas of
212 intense urban and industrial activity are severely contaminated by PTM, creating alarming risks to the

213 environment and the health of local populations (Han et al. 2021, Ma et al. 2020, Han et al. 2020, Jin et al.
214 2019, Yadav et al. 2019).

215 In this study we hypothesized that PTM concentration around the cement plants would be higher
216 than in control areas, which was in accordance with our findings, even at moderate levels when compared
217 with other similar studies (El-Sherbiny et al. 2019; Kolo et al. 2018, Ogunkunle and Fatoba, 2014, Mandal
218 and Voutchkov, 2011; Al-Khashman and Shawabkeh, 2006) and also with the concentration thresholds of
219 PTM recommended by local (COPAM No. 16/11, MINAS GERAIS, 2011) and national (CONAMA No.
220 420/09, BRAZIL, 2009) regulatory laws as well WHO recommendations (WHO, 1993). On the other hand,
221 when we accessed the status of PTM contamination not by PTM concentration solely, but by applying
222 contamination (Igeo) and enrichment (EF) indices, these parameters could indicate that some PTM
223 contamination presented above the “negligible status”, attributing environmental concerns to soil
224 contamination in cement plant surroundings which are operating in the study area. Thus, we consider that
225 our results constitute strong evidences of the need to incorporate the contamination and enrichment indices
226 (Igeo and EF) not only in instruments of risk assessment to the environmental contamination by local
227 environmental agencies, but also as a tool for monitoring programs of PTM pollution to be incorporated in
228 Environmental Management Systems of cement industries.

229 PTMs can be introduced in soil environment by natural routes, such as volcanic activities and
230 weathering of rocks, or by anthropogenic disturbances (Antoniadis et al. 2019, Weissmannová and
231 Pavlovský, 2017). The entry of PTMs into the ecosystem via the food chain can collapse the ecosystem
232 since they: affect the biodegradability of organic pollutants and magnifying their toxic effects; change soil
233 properties such as pH and porosity; and can disrupt the taxonomic and functional structure of biological
234 communities that maintain part of the ecosystem services provided by the soil (Okereafor et al. 2020, Abdu
235 et al. 2017, Giller et al. 2009). In addition to interfering in the ecosystem, PTMs also directly influence
236 people's health, causing inhibition of enzymatic activities, altering protein synthesis, nucleic acid functions
237 and changes in membrane permeability, cause lipid peroxidation, dehydration of sulfhydryl proteins, along
238 with other effects (Fu and Xi 2020, Briffa et al. 2020).

239 In our Igeo results, we found moderate to extreme soil contamination by Cobalt, since observed
240 concentrations of this metal were higher the reference values defined for sedimentary carbonatic rocks
241 which form the geological context of the study area (Turekian and Wedepohl 1961). Natural occurrence of
242 Cobalt is highly concentrated in mafic rocks as well in black shales, commonly forming minerals with S,

243 As and Se, such as cobaltite, smaltite, linneite and arsenosulfide. In soils, high level of Co were found in
244 loamy (Cambisols) and in organic (Histosols) soils (Kabata-Pendias, 2010). Events of soil contamination
245 by Cobalt has been associated with mining and smelting activity, fertilizer use, and sewage sludge spreading
246 (Hamilton 2000) and this study represents the first record of soil contamination by Cobalt in studies focused
247 on soils from cement plant surroundings. Although Cobalt is considered a plant micronutrient (component
248 of several enzymes and co-enzymes), it has been shown to affect growth and metabolism of plants,
249 depending on the concentration and status of cobalt in rhizosphere and soil (Palit et al. 1994). In addition
250 to harmful effects on plants, soil contamination events by Cobalt are also associated with disturbances in
251 the soil microbiota (Shaheen et al. 2016, Zaborowska et al. 2016, Gál et al. 2008). Cobalt contamination
252 has direct interactions with human health, with skin and respiratory issues (Leyssens et al. 2017). Cobalt
253 has been noted to have a high affinity to the sulphhydryl group, thus causing inhibition of crucial enzymes
254 (Simonsen et al. 2012). Moreover, has also been linked to carcinogenic effects, possibly related to the
255 evidence that Cobalt interfere in process of DNA repair and can cause direct induction of DNA-protein
256 crosslinking DNA damage and sister-chromatid, as well evidences of cobalt-mediated free radicals
257 formation (Valko et al. 2005). Here, we also verified low levels of soil contamination by Lead, Copper and
258 Chromium in soil samples from cement plant surroundings. Small particles of inorganic Lead can be
259 absorbed through the respiratory and gastrointestinal tract and chronic exposure to lead contributes to
260 alterations in hormonal and neuronal systems, toxicity in renal cells and affect the hematopoietic system
261 (Briffa et al. 2020). Chromium which is not absorbed through the lungs, may then enter the gastrointestinal
262 tract, primarily absorbed in the jejunum. Toxicity mechanism and carcinogenicity of chromium is a
263 complex process, including higher redox potential, free radicals production and DNA lesions (Dayan and
264 Paine 2001). Oral intake of Copper may cause hepatic and kidney disease, including Wilson's disease,
265 related to copper accumulation in organs instead of being excreted by bile (Harris and Gitlin 1996). In
266 addition to the harmful effects of these PTMs on human health, there are strong evidences that events of
267 soil contamination by these three MTPs are related to disturbances of the soil ecosystem which is related
268 to deterioration of soil quality (Gong et al. 2021, Song et al. 2021, Tang et al. 2019, Lin et al. 2019, Fajardo
269 et al. 2019, Giller et al. 1998, Kandeler et al. 1996).

270 Our Carcinogenetic Risks (CR) estimates showed significant carcinogenic risks for children
271 related to soil contamination by Chromium and Cobalt in all sampled locations, with the exception of two
272 sampling points in CPS2 and CPS3. Considering the estimated CR for adults, risks for Chromium were

273 also detected, the other PTMs (Co, Pb, Ni and Cd) presented low or nonsignificant risks (Qing et al., 2015).
274 Similar results were also observed in previous studies performed in China Wan et al. (2016) and Mao et al.
275 (2019). Other studies related significant carcinogenic risks related to soil contamination by Cd, Ni and Zn
276 (Mohammadi et al. 2020, Karimi et al. 2020, Yadav et al 2019). Our results raised the need for constant
277 monitoring of the levels of PTMs in the soil from cement plant surroundings, using both NCR and CR
278 indices of health risk for local human populations. These indices proved to be more sensitive to
279 conventional metrics recommended by local and national environmental laws (CONAMA, 420/09,
280 COPAM, 166/11). Thus, we consider that our results constitute strong evidences of the need to incorporate
281 the contamination and enrichment indices (Igeo and EF) as well as Health Risk Assessments (NCR and CR
282 indices) not only in instruments of risk assessment to the environmental contamination by local
283 environmental agencies, but also as a tool for monitoring programs of PTM pollution to be incorporated in
284 Environmental Management Systems of cement industries.

285 **Conclusion**

286

287 In this study, we hypothesized that cement production at karstic areas from Southeast Brazil could
288 be related to soil PTM contamination at cement plant surroundings and that these events could be potentially
289 harmful to vicinal human population. Our findings revealed that all PTM concentrations from Cement Plant
290 Surroundings were significantly higher than PTM concentrations from Control areas. Moreover, although
291 the soil contamination events seen in this study can be considered mild when compared to other similar
292 studies, health risk assessments revealed there are concerning carcinogenic risks by the local human
293 population due to long-term exposure to PTM found in soil of the surrounding the cement plants.

294

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562 **Figure captions**

563

564 **Figure 1:** Google Map of study area (on the right) comprising the Southern part of the karst region, located
565 in the Southeast region of Brazil (on the left). Blue and Yellow circles represents control and cement plant
566 surroundings sampling sites, respectively.

567

568 **Figure 2:** Boxplot of PTM concentrations by sampled environments: Cement Plant Surroundings (CPS)
569 and Control (CA). Wilcoxon tests p-values (CPS vs. CA) were described on the on the lower right corner.

570

571 **Figure 3:** Boxplot of Geoaccumulation Index (Igeo Index) for each accessed PTM, by sampled
572 environments: Cement Plant Surroundings (CPS) and Control (CA). Wilcoxon tests p-values (CPS vs. CA)
573 were described at the bottom.

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Figures

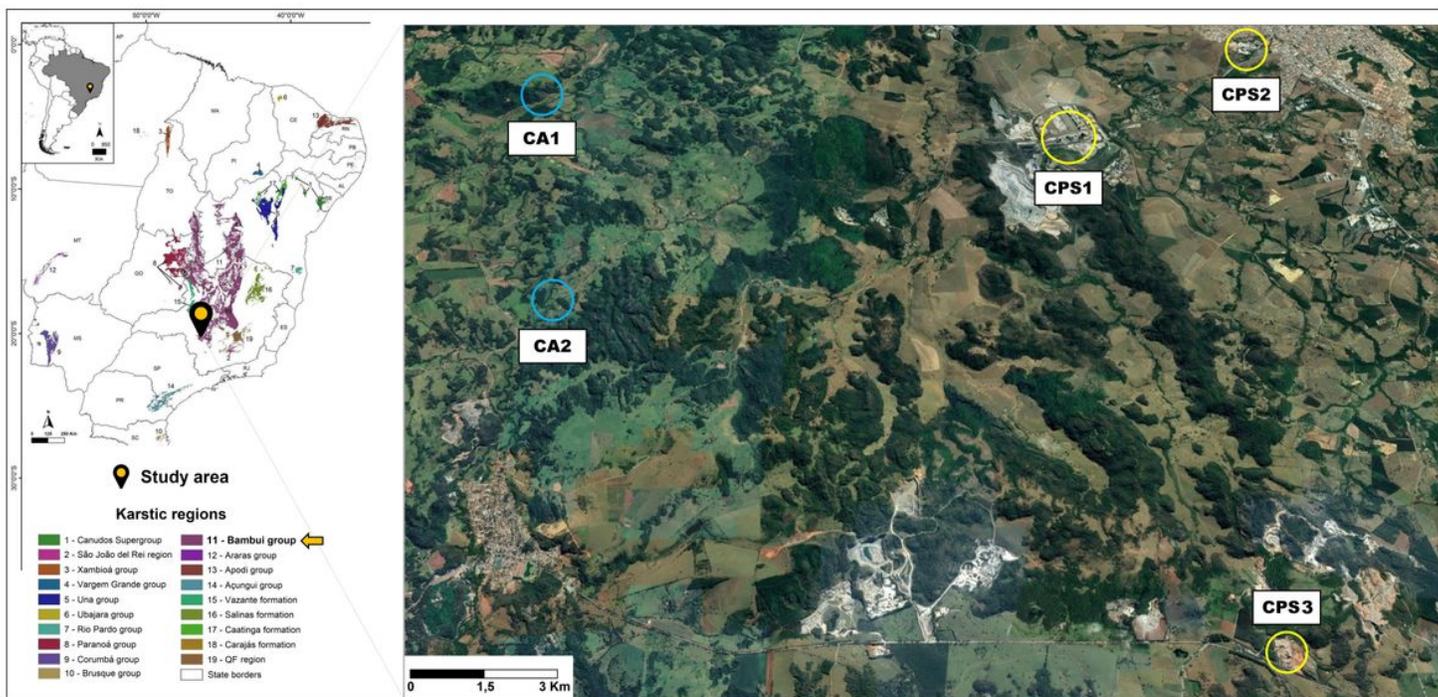


Figure 1

Google Map of study area (on the right) comprising the Southern part of the karst region, located in the Southeast region of Brazil (on the left). Blue and Yellow circles represents control and cement plant surroundings sampling sites, respectively.

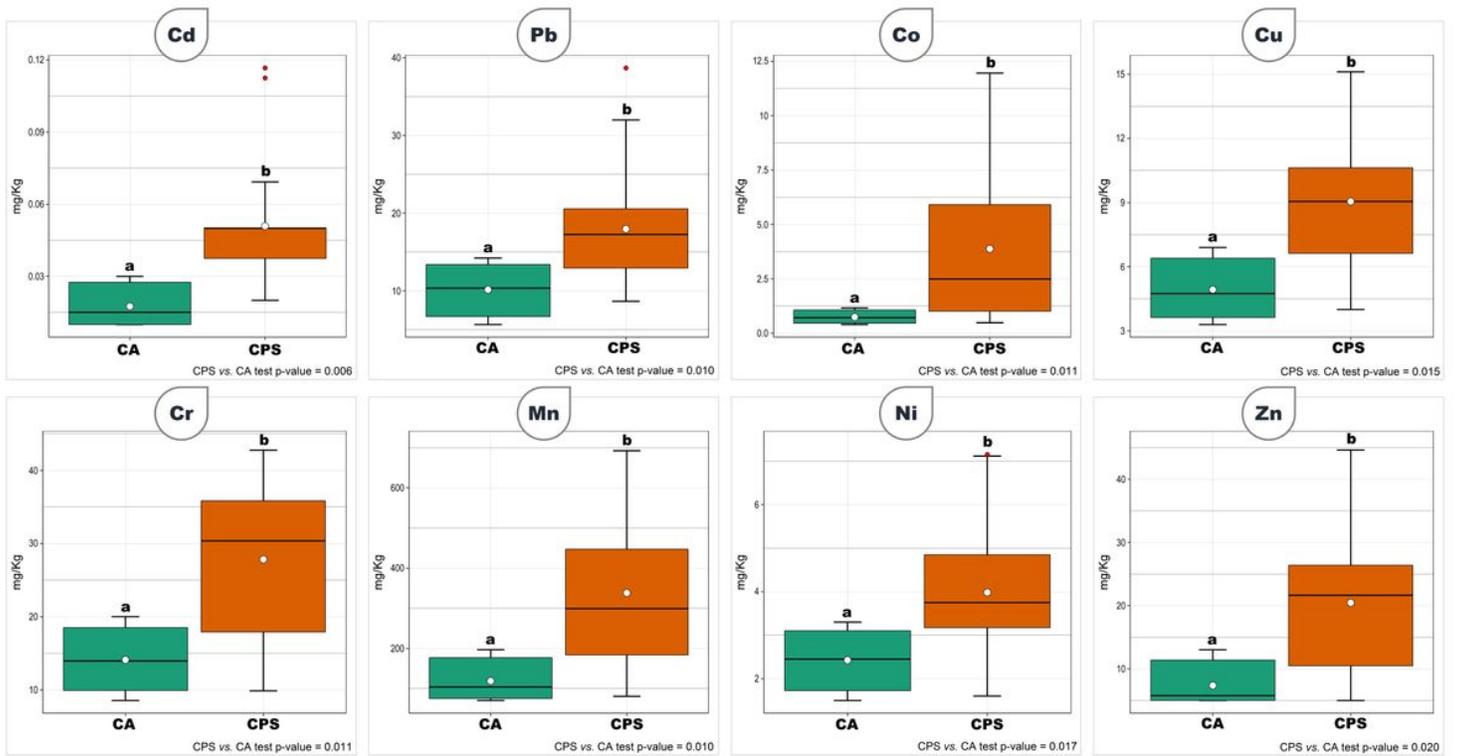


Figure 2

Boxplot of PTM concentrations by sampled environments: Cement Plant Surroundings (CPS) and Control (CA). Wilcoxon tests p-values (CPS vs. CA) were described on the on the lower right corner.

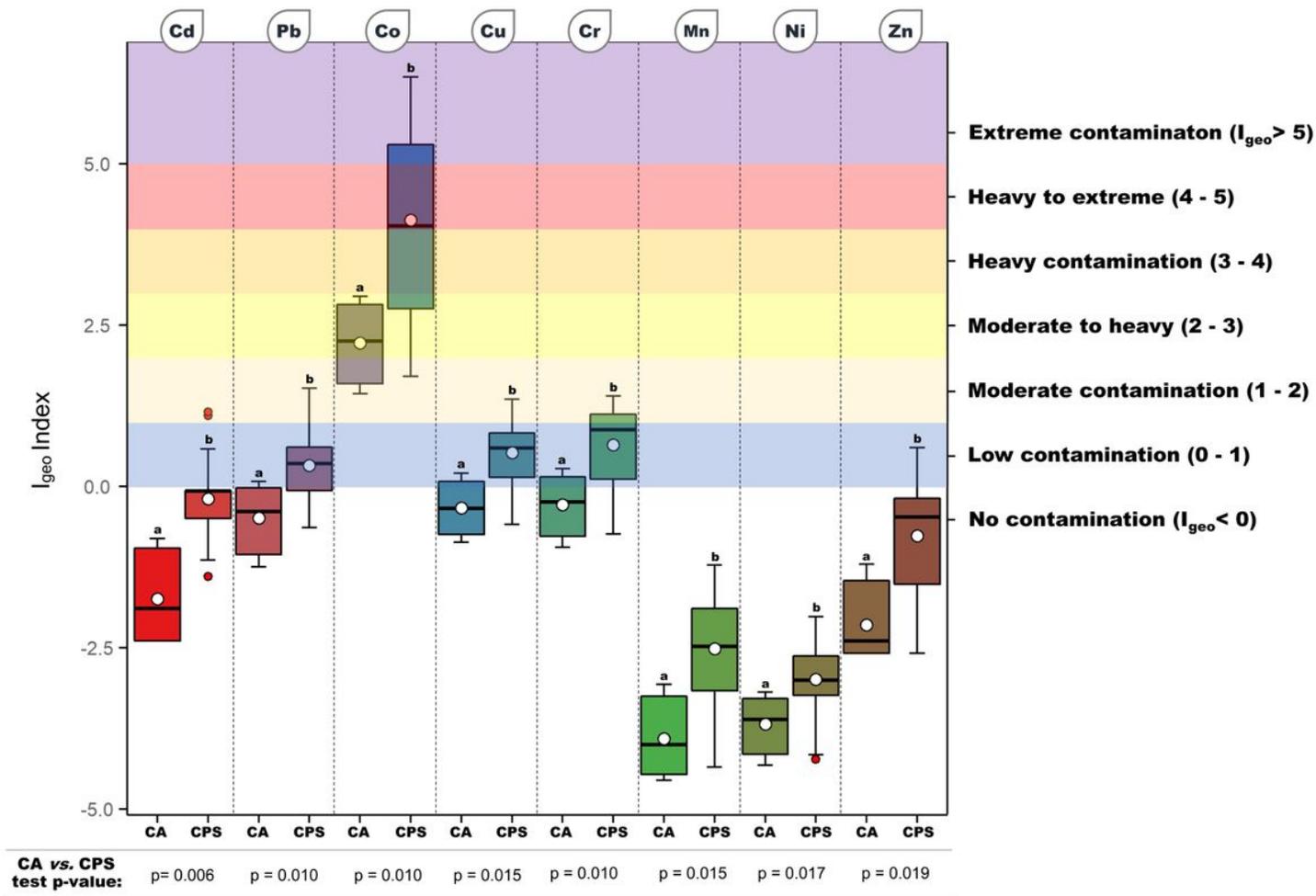


Figure 3

Boxplot of Geoaccumulation Index (I_{geo} Index) for each accessed PTM, by sampled environments: Cement Plant Surroundings (CPS) and Control (CA). Wilcoxon tests p-values (CPS vs. CA) were described at the bottom.

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