

Adaptation is cost-effective to offset rising river flood risk in Europe

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1 **Adaptation is cost-effective to offset rising river flood risk in Europe**

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8 **Abstract**

9 River flooding in Europe could rise to unprecedented levels due to global warming and continued
10 development in flood-prone areas. Here we appraise the potential of four key adaptation strategies to
11 mitigate flood risk across Europe based on detailed flood risk modelling and cost-benefit analysis. We find
12 that reducing flood peaks using retention areas is economically the most attractive option. In a scenario
13 without climate mitigation, they can lower projected flood losses in Europe by the end of the century from
14 42 to 7.5 €billion/year and population exposed by 81%, or achieve a risk level comparable to today. This
15 would require an investment of 2.9 €billion/year over 2020-2100, with a return of 4€ for each 1€ invested.
16 The risk-reduction potential of economically-optimised strengthening of dykes is somewhat lower with
17 71% for a comparable annual investment. These measures avoid floods to happen and their cost-
18 effectiveness increases with the level of global warming. Implementing building-based flood proofing
19 measures and relocating people and assets are less cost-effective but can reduce impacts in localized
20 areas.

21 **Main Text**

22 **Introduction**

23 River floods are a major cause of damage in Europe (Wallemacq and House, 2018). Absolute losses have
24 generally increased over time mainly due to human encroachment and economic development on flood-
25 prone land that resulted in a strong rise in exposure and loss of natural storage capacity (Bouwer, 2011;
26 Merz et al., 2012). The number of fatalities and damage expressed relative to the exposed value and size
27 of the economy, however, have dropped (Paprotny et al., 2018) in Europe as well as other regions of the
28 world (Jongman et al., 2012a; Tanoue et al., 2016). Hence, improved protection against floods has
29 counter-balanced the effects of increasing exposure on risk and resulted in a strong reduction in
30 vulnerability (Formetta and Feyen, 2019).

31 It is less clear if and how climate change has affected the trend in flood risk. There is no consistent
32 continental-scale climatic-change signal in flood discharge observations in Europe (Hall et al., 2014),
33 despite a trend towards increasing floods in northwestern Europe and decreasing flood hazard in southern
34 and eastern Europe (Blöschl et al., 2019). However, there is growing consensus that climate change will
35 intensify the hydrological cycle (Wu et al., 2013; Madakumbura et al., 2019) and amplify the intensity and
36 probability of floods in most parts of Europe (e.g., Alfieri et al., 2015; Mentaschi et al., 2020). Fuelled with
37 continued development and urbanisation in floodplains this could give rise to an unprecedented increase
38 in flood risk (Feyen et al., 2009; Alfieri et al., 2018).

39 European societies will therefore need to implement effective flood adaption strategies. There exist a
40 broad range of measures that can be targeted at reducing the hazard, vulnerability or exposure, or at
41 managing the consequences. Despite the wide literature on flood risk mitigation, there are relatively few
42 studies that have quantified costs and benefits of different measures (Kreibich et al., 2015; Aerts, 2018;
43 Yamamoto et al., 2021). The limited continental and global scale studies considered a generic vulnerability

44 reduction (Kinoshita et al., 2018), performed a sensitivity analysis (Alfieri et al., 2016) or only focused on
45 increasing dyke height (Ward et al., 2017). However, to find the most effective strategy, limit potential
46 negative environmental effects and avoid maladaptation, it is essential to consider a range of measures
47 (Jongman, 2018). Among those, nature-based solutions have recently gained attention as more
48 environmentally sustainable ways to reduce flood risk (Faivre et al., 2017). To our best knowledge, their
49 effectiveness has not been appraised in large-scale studies.

50 Here, we present the first quantitative assessment of the costs and benefits across Europe of four key
51 flood adaptation options: raising dykes, retention areas, flood proofing and relocation. First, flood hazard
52 and risk were projected up to end of the century for different Global Warming Levels (GWL = 1.5, 2 and
53 3°C) assuming present flood protection based on a large ensemble of climate projections and long-term
54 socioeconomic projections for the EU. Each adaptation strategy was appraised using a cost-benefit
55 analysis that optimises the Net Present Value (NPV), which integrates the discounted overall costs of
56 implementation and avoided economic damages over the life time of the measure. The costs were
57 calculated as the sum of capital investments and maintenance costs, taken from a database of risk
58 reduction measures based on literature review (see Methods). The benefits are the economic damages
59 avoided by implementing the measure, calculated as the difference between future direct damages with
60 and without adaptation respectively. Flood losses, costs and benefits are presented undiscounted in
61 general, so that present and future scenarios with and without adaptation can be compared while giving
62 equal weight to each of them (more details in Methods, with an evaluation of the skill of the modelling
63 components in Supplementary information).

64 *Future flood risk scenarios*

65 We estimate that at present river flooding causes an annual damage of 7.6 €billion/year and exposes to
66 inundation around 170,000 people in the EU+UK (Figure 1). In the absence of further climate mitigation

67 (3°C in 2100) and adaptation (assuming present vulnerability), flood damage would rise to 42 €billion/year
68 by the end of the century, while annually nearly half a million Europeans would be exposed to river
69 flooding. Stringent climate mitigation would roughly halve the risk, yet the likelihood that global warming
70 be stabilized at well below 2°C above pre-industrial levels is low (Raftery et al., 2017). This means that
71 adaptation will be needed to offset the projected rise in flood risk.

72 *Adaptation through river dykes*

73 Structural defences such as river dikes have a very old tradition to protect people and economic activities
74 in floodplains (Kundzewicz, 2002). Strengthening protection through dyke systems consists of elevating
75 the river banks, through permanent or temporary barriers, to increase the maximum streamflow that the
76 watercourse can fully contain and convey downstream without causing damage. Investments in dykes
77 shows to be economically convenient to reduce the projected flood impacts on economy and society
78 (Figure 1, Table 1). The implementation of the optimal design in the whole study area (EU + UK) for the
79 3°C warming scenario would require an average annual investment of 3 €billion/year (average of
80 undiscounted costs over 2020-2100). The corresponding increased level of protection would lower annual
81 flood damages by 30 €billion/year by the end of the century, or a reduction of 71%. Also 350,000 (68%)
82 less people would be exposed to flooding (Figure 1 and Table S3). The overall discounted benefit-to-cost
83 ratio (BCR) of the dyke investments is 2.9. The BCR and impact reduction capacity grow with GWL (Tables
84 S5 and S7), showing that adaptation measures targeted at reducing flood hazard become more relevant
85 as global temperature rises.

86 Strengthening existing dyke systems is cost-effective in most countries of Europe, but with considerable
87 regional variation in risk reduction potential and BCR (Table 1 and Figure 2). The damage reduction
88 potential for single countries ranges from 7% and 10% in Portugal and Spain to 90% and 89% in Belgium
89 and the United Kingdom. The large variability in risk reduction rate is due to the variability of optimal

90 design option (i.e. the degree of implementation locally providing the highest NPV), with no change in
91 design and consequently no damage-reduction effect in regions with $BCR < 1$. In several areas of the
92 Mediterranean Region (Iberian Peninsula, southern Italy and Greece) and in some regions in Eastern
93 Europe and in the Alps, projections indicate limited increases in flood impacts that are mostly driven by
94 increasing exposure value (Figure 3 and Table S4). Hence, the costs of adaptation are not compensated
95 by avoided impacts. In other regions in Central and Western Europe (e.g. the Netherlands), protection
96 standards are already high (Figure S1), and the reduction in residual risk is not enough to make additional
97 dyke heightening economically efficient. Reduction rates in population exposed are broadly similar (Tables
98 S5 and S7).

99 Despite the favourable outcome of the cost-benefit analysis in most regions of Europe and the fact that
100 limited implementation space is required, it has to be considered that extensive reliance on dyke systems
101 can have socioeconomic and environmental drawbacks. Heightening river dykes generally increases the
102 magnitude of peak flows and hence flood risk downstream (Di Baldassarre et al., 2009). The reduction in
103 flood frequency and sense of safety favours the loss of flood memory and further development in flood-
104 prone areas. This levee effect can in the unlikely case of flood defence failure lead to catastrophic
105 consequences (Lane et al., 2011; Di Baldassarre et al., 2015). Dykes further distort the natural functioning
106 of wetlands and riparian floodplains due to the lost hydraulic connectivity with the river channel (Gumiero
107 et al., 2013).

108

109 *Adaptation through retention areas*

110 We consider natural retention in floodplains only, because structural (dams) flood control reservoirs have
111 high costs and negatively impact ecosystems (Barbarossa et al., 2020). This is achieved by setting up areas
112 within or aside the river network that can be flooded in a controlled manner to store excess water when

113 the river stage reaches critical levels (Arrighi et al. 2018). The cost-benefit analysis shows that the use of
114 natural retention areas can be highly effective to reduce flood risk in Europe. At the EU level,
115 implementing the optimal design for the 3°C warming scenario would require an average annual
116 investment of 2.9 €billion/year over the period 2020-2100 (undiscounted values), with a BCR of 4 (Figure
117 1). The resulting storage capacity would reduce flood economic damages by up to 82% (from 42 to 7.5
118 €billion/year) and population annually exposed by 81% (from 507,000 to 97,000) by the end of the
119 century, hence to a risk level that is comparable in absolute terms to that of today. Because retention
120 areas reduce the hazard by attenuating the flood hydrograph they become increasingly effective when
121 flood hazard rises with global warming.

122 Natural retention areas are economically convenient practically everywhere except for some NUTS2
123 regions (Figure 2) where projected increases in flood impacts are small (e.g. in Portugal, Spain and
124 Greece), where protection standards are high (e.g. the Netherlands), and where floodplains are too
125 narrow to accommodate retention areas (e.g. Portugal). In general, retention areas show higher BCR
126 values compared with river dykes and their optimal implementation results in a stronger reduction in
127 socioeconomic impacts (Table 1). The higher efficacy in reducing flood hazard at regional level occurs
128 because retention areas reduce flood peaks for all downstream river reaches, rather than locally
129 containing it with river dykes. Furthermore, retention areas offer additional benefits not considered in
130 the economic optimisation. Reconnecting with rivers restores the natural functioning of floodplains,
131 which improves aquatic and riparian ecosystem quality and provides a range of additional services, such
132 as the reduction of pollutants, regulation of sediment fluxes and recreational opportunities (Schindler et
133 al., 2016; Nilsson et al., 2018). The monetary evaluation and inclusion of environmental services would
134 further increase the cost-effectiveness of retention areas.

135 On the other hand, retention areas require the occupation of large portions of land (according to our
136 calculations, the largest retention areas can exceed 10 km²), which would no longer be available for

137 intensive use (e.g. agriculture, urbanisation). Moreover, narrow floodplains may not be suited for this
138 measure. For instance, the low cost-efficiency in Portugal depends on the limited increase of future flood
139 risk, but it also relates to the limited width of many floodplains and scarce availability of non-urbanised
140 land, which increases costs for implementing retention areas.

141 It is worth noting that the spatial attribution of costs and benefits of retention areas is more complex than
142 for other measures, because the design has to be carried out considering the entire river basin. This has
143 to be considered when evaluating BCR results at NUTS2 level, because areas located downstream benefit
144 from upstream retention areas, and hence reduce local implementation costs. Ideally, implementation
145 costs should be shared among all regions within a river basin. Planning in transboundary rivers, such as
146 the Danube, may be complex, although some projects have already been successfully carried out (EEA,
147 2017).

148

149 *Adaptation through flood proofing of buildings*

150 A large share of flood losses relates to damage to buildings and their content (Jongman et al., 2012b).
151 Structural and non-structural modifications can prevent water to enter the building (dry proofing) or
152 reduce damages by means of flood-adapted use and equipment of buildings (wet proofing), thus reducing
153 building and content vulnerability. Given the wide range of measures and the large variability in their costs
154 and damage reduction potential (Table 2), we based our cost-benefit analysis on central cost and damage
155 reduction estimates derived from literature (See Methods).

156 Results show that flood proofing of buildings has generally BCR values above 1 across Europe (Figure 2).
157 They can be applied at smaller scales and at lower cost compared to hazard-reduction measures such as
158 dykes and retention areas, and hence targeted where they are cost-effective (see Methods). However,
159 prioritizing only areas with a positive benefit-cost balance may leave several other areas unprotected.

160 Damage reduction measures applied at building scale have a low environmental impact, are relatively
161 easy to implement and can be adapted to changing conditions (Akadiri et al. 2012). However, they cannot
162 prevent other types of flood damage, such as to transport infrastructure (Bubeck et al., 2019) or
163 agriculture (Tapia-Silva et al., 2011). As a result the average reduction in damage attainable in Europe is
164 12%, which is considerably lower than hazard-reduction measures. More importantly, because floods are
165 not avoided population exposure is not reduced (Figure 1c), even though the degree to which people are
166 affected is lowered.

167 Flood proofing is more effective (reduction in damage above 10%) in Eastern Europe countries such as
168 Estonia, Hungary and Czech Republic, because of lower protection standards and the presence of hotspots
169 of exposed assets (Table 1). Damage reduction ratios above European average are also attainable in
170 Sweden, Belgium and United Kingdom, due to the high economic exposure. Conversely, BCR values are
171 below one in the Netherlands, where country-scale high protection standards make flood proofing less
172 likely to be used (Table 1). Overall, findings suggest that flood proofing of buildings is effective for
173 protecting areas frequently exposed to low or moderate floods and with high concentration of exposed
174 value and assets (Richert et al., 2019), whereas they are not suited for an efficient protection of large
175 areas.

176

177 *Adaptation through relocation*

178 Relocation aims at reducing the exposure of people and assets at risk of flooding by moving them to areas
179 with negligible risk (King et al. 2014). Managed relocation of individuals, businesses, and infrastructure is
180 largely ignored as a possible strategy in the EU national flood risk management policies (Mayr et al., 2020).
181 The cost-benefit analysis shows that it is the least cost-effective measure among all the adaptation
182 measures considered here (Figure 1). The implementation across the areas with a positive benefit-cost

183 balance would lead to an overall reduction in flood damages of just 2%, corresponding to a BCR of 2.3,
184 while each year only 4700 (1%) less people would be exposed to floods. This is because relocation is
185 economically convenient in a minority of NUTS2 regions concentrated in UK, Spain and around the Baltic
186 region (Figure 2). Costs of relocation are high as they include the demolishing of existing buildings, the
187 acquisition of new land and the construction of new infrastructures. Indeed, relocated people are
188 generally offered a partial compensation for their properties by the local government (Kick et al., 2011),
189 thus suggesting that financial incentives are necessary to promote relocation measures. In regions with
190 $BCR > 1$, long term flood economic damage may become comparable to and greater than the value of new
191 land and buildings, because of either low protection standards or concentration of high value assets.
192 These findings suggest that relocation can be cost-effective in localized areas, as well as for sensitive or
193 critical buildings and infrastructures frequently exposed to floods.

194 Past flood events suggest that flood relocation primarily occurs after catastrophic events for which the
195 reconstruction costs are of the same magnitude as buying a new property (López-Carr and Marter-Kenyon
196 2015). There is also a low social acceptance of relocation measures as people feel uncomfortable with
197 losing ancestral lands and properties as well as breaking long-standing ties with their communities and
198 other networks. On the other hand, relocation is the most robust long-term solution as flood risk is
199 avoided through a removal of exposure, and the land that has become available after relocation can be
200 used for the retention of flood peaks.

201

202 **Discussion**

203 We focused our analyses on adaptation scenarios based on the application of a single type of measure.
204 The outcomes suggest that 'hybrid' strategies, with different measures working in synergy and optimised
205 at the level of river basins are likely to be the best strategies to maximise local benefits and minimise

206 drawbacks of each measure, in line with recent findings (Du et al., 2020). For instance, it is advisable to
207 use dykes to protect against frequent low-magnitude events, and retention systems to mitigate extreme
208 flood peaks. Foreseeing backup risk-reduction measures, such as flood proofing of buildings, helps
209 minimising impacts when hazard-protection measures fail or are not sufficient to prevent flooding.
210 Integrating physical risk reduction measures with financial instruments such as insurance would further
211 reduce overall impacts on the economy and society (Kron et al., 2019).

212 The adoption of adaptation strategies should not be alternative to risk-informed land use planning. In past
213 decades, urban areas expanded considerably in flood-prone areas under increasing population pressure
214 and due to the benefits associated with settling close to river courses (Kummu et al. 2011), a trend that
215 has not slowed down even in recent years (Mård, et al. 2018). Our projections show that socioeconomic
216 growth and urban expansion will increase economic losses by more than 70% across Europe in 2100 (Table
217 S4). As such, taking into account flood risk in planning could be an effective way to reduce future flood
218 impacts.

219 The cost-benefit analysis does not include social, environmental and cultural aspects, which would require
220 more complex multi-criteria analyses (e.g. using the concept of social vulnerability as in Kind et al., 2020).
221 The inclusion of these aspects would likely improve the cost-effectiveness of nature-based solutions such
222 as retention areas, as highlighted in previous studies (EEA, 2017).

223 Local cost-effectiveness of measures can deviate strongly from those presented herein due to site-specific
224 characteristics. The present analysis is therefore not meant to replace detailed analyses at local and
225 regional scale, which are necessary for an effective and reliable design and implementation of adaptation
226 measures. Similarly, optimal adaptation measures should interact and require engagement of local
227 population, governments and actors. On the other hand, several large European rivers are transnational,

228 therefore our analysis can provide a consistent, pan-European framework to evaluate and compare the
229 costs and effectiveness of river flood adaptation measures under future scenarios.

230

231 **Methods**

232 We appraise costs and benefits of river flood adaptation using the IPCC risk framework (IPCC, 2014). The
233 different modelling steps and data used in the hazard, vulnerability, exposure, risk and adaptation analysis
234 are described in the following sections. We note that our analysis does not cover coastal, pluvial and flash
235 flooding. The geographical coverage of our analysis is the EU and UK, with the exception of Malta where
236 flooding is caused by pluvial and flash flood events and water courses are too small to be represented in
237 the river flood modelling framework here applied.

238 *Climate projections*

239 Projections of river streamflow with global warming are based on an ensemble of 11 bias-corrected
240 regional climate projections from EURO-CORDEX (Table S1) for Representative Concentration Pathways
241 RCP4.5 and RCP8.5 from 1981 up to 2100 (Jacob et al., 2014). The period 1981-2010, hereinafter referred
242 to as “base”, was used a reference. We consider future climate scenarios corresponding to an increase in
243 global average temperature of 1.5, 2 and 3°C above preindustrial temperature. The 1.5°C and 2°C warming
244 scenarios are explicitly considered in the Paris Agreement, while a 3°C global warming is a more realistic
245 scenario to expect by the end of the 21st century if adequate mitigation strategies are not taken. We
246 evaluate each warming scenario assuming stabilized climate from the time indicated in Supplementary
247 Table S1, i.e. there is no further warming and climate conditions remain constant after the year of reaching
248 a warming level. Climate at global warming levels derived from transient climate projections may differ
249 from stabilized climate at those warming levels. However, no high-resolution stabilized climate
250 projections are available for Europe. Moreover, studies (e.g., Maule et al., 2017; Mentaschi et al., 2020)

251 suggest that the effect of pathway to global warming levels is small compared to the models' variability,
252 except for strongly not time-invariant variables such as sea level rise.

253 *Flood hazard and risk projections*

254 We used the climate projections to generate daily streamflow simulations with Lisflood, a distributed,
255 physically based hydrological model, run at 5km grid resolution (Burek et al., 2013; van der Knijff et al.,
256 2010). The extremes of river discharge were analysed by means of the non-stationary approach proposed
257 by Mentaschi et al. (2016). This methodology allows using the whole time horizon of the simulations
258 (1981-2100) to fit the extremes, providing more reliable estimations for high return periods, compared
259 with stationary techniques that typically use 30-year windows. For more information on the
260 implementation of Lisflood, and on the fit of the extremes, the reader is referred to Mentaschi et al.
261 (2020).

262 We represent floodplain inundation processes following the approach described in Alfieri et al. (2015).
263 Specifically, flood hazard maps for a range of return periods from 10 to 500 years were derived from two-
264 dimensional hydraulic simulations with the Lisflood-FP model (Bates et al., 2010). The flood hazard maps
265 characterize the flood extent according to flood magnitude simulated along the river network.

266 We derive exposure information from the European population density map by Batista e Silva et al. (2018)
267 and the refined version of the CORINE Land Cover proposed by Rosina et al. (2018). Both maps are
268 available at the same resolution of the flood hazard maps (100m).

269 Vulnerability to floods is included in the form of damage functions and through a flood protection map.
270 We use country specific depth-damage functions from Huizinga et al. (2017) to link flood depth with the
271 corresponding direct economic damage, considering land use classes and gross domestic product (GDP)
272 per capita at local administrative level. Spatial distribution of flood protection levels in Europe is obtained

273 by combining available information on protection design levels with modelled protection standards
274 calculated by Jongman et al. (2014) and Scussolini et al. (2016) (see Supplementary information).

275 Socioeconomic Projections are based on the ECFIN 2015 Ageing Report (EC 2015). This scenario acts as a
276 benchmark of current policy, market and demographic trends in the EU. High-resolution population
277 projections are derived by the LUISA modelling platform (Jacobs-Crisioni et al., 2017). These maps capture
278 the fine-scale processes of population dynamics (e.g., urban expansion, stagnation or de-growth), and
279 concentration that represent key drivers of the future exposure of populations. The Ageing report
280 projections are available only until 2060. After that, land use was assumed static. The relative distribution
281 of people in a country in 2060 was scaled according to country projections of population up to 2100, while
282 the damage functions were corrected for the projected changes in GDP. Regarding the GDP projections,
283 the Ageing Report assumes that two out of the three determinants of economic growth, technical
284 progress and capital accumulation, would reach a steady state (with constant growth rates) by the year
285 2060. That was assumed as well for the following decades. The third contributor to growth (the labour
286 input) was assumed to evolve in a proportional way with respect to population (i.e. same growth rate).
287 That means ignoring possible changes in the labour market conditions, such as the employment rate.
288 Population projections for 2061-2100 are taken from the latest United Nations demographic report
289 (medium variant), and they are explicitly considered in the computation of the economic growth figures
290 (more details can be found in Ciscar et al., 2017).

291 We represent river flood risk as expected annual economic damage (EAD) and expected annual population
292 exposed (EAPE), following the approach described by Rojas et al. (2013). For the baseline scenario, EAD
293 and EAPE are calculated by constructing impact-probability curves based on the six return periods
294 considered by flood hazard maps and taking into account local protection levels. Changes in future flood
295 impacts are derived considering the flood frequency shift for the six reference events (i.e. magnitudes

296 corresponding to a return period of 10, 20, 50, 100, 200 and 500 years under the baseline scenario) and
297 for protection levels. All economic risk estimates in this work are expressed in €2015 values.

298 We evaluated the overall reliability of the data and models composing the risk modelling framework
299 (Supplementary Information). Most of the models and datasets have been validated to some extent
300 against observed or higher resolution data in past research studies. We also compared modelled annual
301 average economic losses against reported losses retrieved from numerous sources. We find that in a
302 number of countries (such as Czech Republic, Germany, Italy, and United Kingdom) the difference
303 between modelled and reported losses is within 50%. These countries account respectively for more than
304 50% and 70% of overall modelled and reported losses. Losses are overestimated by more than 100% in
305 France, in Scandinavian countries and in a number of medium-small countries. A detailed analysis is
306 reported in the Supplementary Information.

307

308 *Data collection for adaptation modelling*

309 For the adaptation analysis, we constructed a database of flood risk reduction investments based on a
310 review of scientific, grey and technical literature. The database provides an overview of the main types of
311 investments applied in case studies, mainly in Europe (Kuik et al., 2016; EEA, 2017; Aerts, 2018; GFDRR et
312 al., 2019). The database is available as electronic spreadsheet in the supplement material. We used
313 information on size and cost of past applications in literature to derive unit costs of adaptation measures
314 suitable for application within a pan-European framework (e.g. the cost to increase the height of one
315 linear kilometre of dyke by one meter). We also compiled information to clarify the link between
316 implementation costs and impact reduction (e.g. damage reduction factors reported for specific flood-
317 proofing measures). We decided to include in the adaptation analysis only measures for which we found
318 sufficient information on quantitative costs (especially unit costs) and performance estimates. Table S2

319 provides a description of the four adaptation measures considered in this study, while Table 2 summarizes
320 the unit costs derived from the database of adaptation measures.

321 *Modelling of the adaptation measures*

322 *Strengthening of dyke systems*

323 We model the increasing of dyke height along the river network following the approach proposed by Ward
324 et al. (2017). We first estimate the present-day height of dykes along the river network based on river
325 discharge and the level of flood protection. For instance, the height of dykes designed to contain the 1-in-
326 100-year flood event is given by the water level corresponding to the 1-in-100-year discharge. To this end,
327 we use height-discharge curves calculated by the hydrological model Lisflood. Then, for each future
328 scenario we calculate spatial maps of increases in dyke heights required to raise protection standards up
329 to the new design return levels. Implementation costs are calculated considering the overall length of
330 dikes and the additional height required. Costs are derived from literature values on dyke construction
331 and elevation costs (Table 2).

332 *Retention areas*

333 The design and modelling of retention areas requires the development of an algorithm to allocate storage
334 areas within each river basin, based on the available storage capacity and the required level of protection.
335 We first calculate the maximum storage capacity in floodplains along the river considering agricultural
336 (excluding permanent crops, e.g. orchards, vineyards) and semi-natural (e.g. permanent grassland,
337 wetlands, excluding forests) areas within the 1-in-500-year floodplain, derived from the refined CORINE
338 Land Cover (Rosina et al., 2018). Then, we calculate flood volumes that can be accommodated by present-
339 day protection standards and the flood volumes that need to be stored in each future scenario along the
340 river network. Flood volumes are estimated for each point of the river network using synthetic
341 hydrographs calculated with the Lisflood hydrological model, following the approach by Alfieri et al.

342 (2014). Finally, the required storage volumes are calculated iteratively along the river network (i.e. design
343 minus present volumes) starting from the most upstream reaches. The iterative procedure is designed to
344 calculate the reduction of flood volumes along the river network given by upstream storage. In other
345 words, part of the flood volumes stored in a section of a river basin is subtracted from the flood volumes
346 in all downstream branches. The iterative procedure is executed separately for each design level of
347 protection, and assuming a constant return period of flood peaks in the entire river network (e.g. assuming
348 to protect the entire river basin against 1-in-100-year discharge). Implementation costs are calculated
349 based on the overall flood volume to store (Table 2).

350 *Flood proofing measures*

351 There is a wide range of flood proofing measures applicable at the building scale, depending on local flood
352 and exposure characteristics (e.g. expected range of flood water depths, type and structure of the building
353 to be protected). Most research works on these measures provide an overview of costs and benefits for
354 specific case studies (Kreibich et al., 2015; Aerts, 2018), and few studies report analytical analyses of
355 different measures on real cases (Du et al., 2020) or standardized buildings (Richert et al., 2020).

356 In this work, we assume that the implementation of flood proofing measures can reduce overall damage
357 to exposed buildings by a specific fraction (e.g. 10%, 30% etc.), which is taken as design criterion. Using
358 the available database of adaptation measures, we relate damage reduction ratios with implementation
359 costs, by averaging data from all case studies in which flood proofing measures were applied. In other
360 words, “the analysis considers a standard/average flood-proofing implementation, based on available
361 literature information. Given the scale of application, we assume that damage reduction and costs can be
362 linearly correlated, because the measures can be applied over an increasing number of buildings. Note
363 that we excluded building elevation measures from the analysis because they are often not feasible for
364 existing buildings, and because their cost is comparable to relocation measures. We further assume that

365 infrastructural and agricultural damages cannot be reduced through flood-proofing measures of the built-
366 up area, meaning that potentially, on average, 90% of the expected annual damage can be reduced.

367 Cost of flood-proofing measures are usually available at building scale. These were translated in unit costs
368 related to building surface (€/m²) using building area (if available), or assuming a standard building area
369 of 100m² where no information is available (Table 2). We assume that the same costs apply to all building
370 types, even though literature studies usually focus on residential buildings). We calculate implementation
371 costs as a function of the total built-up area located within the 1-in-500-year flood extent, and the damage
372 reduction ratio required. The built-up area is derived from the Global Human Settlement maps for Europe
373 (Florczyk et al., 2019). Note that we assume that population exposed is not reduced by this adaptation
374 strategy, as building-based measures do not prevent floods from occurring

375 *Relocation*

376 Relocation measures are designed assuming that a fraction of the exposed buildings and population
377 located in flood-prone areas are moved to a flood-safe area. We consider for relocation all built-up areas
378 located within the 1-in-500-year flood extent, for consistency with the approach adopted for all the other
379 measures. Additional tests run considering only built-up areas more frequently exposed (e.g. located
380 within the 1-in-50-year flood extent) did not show significant changes at European and country scale in
381 terms of cost-benefit analysis. We do not make any assumption about the place of destination of relocated
382 assets and people, as such decision would be highly subjective, nor do we consider possible costs for
383 resettlement (e.g. realization of infrastructure networks). We assume that implementation costs increase
384 linearly with exposure reduction, and that the exposure reduction for buildings can be used to determine
385 the reduction in population exposed (e.g. relocating 20% of buildings implies the relocation of 20% of local
386 population).

387 *Analysis of adaptation strategies*

388 The evaluation of each adaptation strategy is performed using a cost-benefit analysis that optimises the
389 benefits (avoided economic damages) and the costs of implementation and maintenance over the lifetime
390 of the measures, where the lifetime was considered from 2020 up to end of this century. The calculation
391 of costs and benefits follows the framework proposed by Ward et al. (2017). For all measures except flood-
392 proofing, investment costs were calculated considering construction costs distributed between 2020 and
393 2050, while maintenance costs are considered from 2050 to 2100. Flood proofing measures have a limited
394 life span compared to the other measures (Kreibich et al., 2015), therefore during the period 2060-2090
395 we consider additional construction costs for replacement.

396 In accordance to the literature, we assume that maintenance costs amount to 1% of total construction
397 costs (Ward et al., 2017; Aerts, 2018). Similar as the implementation cost, we assume that the effect of
398 the measures applied (protection level for dykes strengthening and retention areas, or damage reduction
399 rate for flood proofing and relocation) increases linearly from 2020 (no effect) to the design value in 2050,
400 and then remains constant. Implementation costs are calculated differently for each adaptation measure
401 as described in the Section “Modelling approach for adaptation measures”.

402 For each adaptation measure we simulate different design options (e.g. raising dykes over a river stretch
403 by different height increases corresponding to a range of design return periods). For dikes strengthening
404 and building of retention areas, the optimal design level for each strategy was considered to be the one
405 providing the maximum net present value (NPV) at NUTS2 level, defined as the sum of investment costs
406 (that are negative) and economic benefits (avoided economic losses, positive) over the lifetime of the
407 project. For relocation and flood-proofing of buildings, NPV is calculated by aggregating costs and benefits
408 at 5km resolution, which corresponds to the grid used to aggregate flood impacts and derive future river
409 flow projections (Alfieri et al., 2015; Mentaschi et al., 2020).

410 Future costs and benefits are discounted to present-day values using a 5% discount rate for EU countries
411 eligible for the EU Cohesion Fund and 3% for other Member States and the UK, following the European
412 Commission’s guidelines on infrastructure investments (EC, 2014). The cost-benefit analysis is applied for
413 the three warming scenarios in order to understand the performance of the adaptation options for
414 different levels of global warming. As an indication of the performance we also present the Benefit-to-
415 Cost Ratio (BCR), which is the ratio of the total discounted benefits to costs. We calculate BCR values for
416 NUTS2 administrative regions, as well as countries and the EU + UK. For relocation and flood-proofing of
417 buildings, aggregation of results at NUTS2 and country level is done taking into account only 5km areas
418 with positive NPV. We further present benefits of adaptation in terms of the reduction in population
419 exposed to flooding.

420 Note that we could not quantify the environmental costs and benefits of the available adaptation
421 measures. However, we provide a qualitative assessment of these factors in the discussion of results.
422 Moreover, the reduction in population exposed was not monetized in the cost/benefit analysis, due to
423 the lack of accurate information on impacts (both physical and social) and sensitivity issues in attributing
424 economic value to human lives.

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432 **Data availability:**

433 The maps of river flow extremes for the global warming scenarios considered in this work are available
434 at the JRC Data Catalogue: [https://data.jrc.ec.europa.eu/dataset/20247f06-469c-4607-8af1-](https://data.jrc.ec.europa.eu/dataset/20247f06-469c-4607-8af1-a5a670082471)
435 [a5a670082471](https://data.jrc.ec.europa.eu/dataset/20247f06-469c-4607-8af1-a5a670082471) . The full dataset of river flood discharges can be provided upon reasonable request to
436 the authors.

437 The flood hazard maps are available at the JRC Data Catalogue:
438 <https://data.jrc.ec.europa.eu/dataset/1d128b6c-a4ee-4858-9e34-6210707f3c81>

439 The European population density map used to represent present-day population is available as
440 supplementary information of the work by Rosina et al. (2018) at the following address:
441 <https://doi.org/10.6084/m9.figshare.6210392>

442 The land cover map and all the spatial projections of population and land cover are available at the JRC
443 Data Catalogue: <https://data.jrc.ec.europa.eu/collection/luisa>

444 The flood damage functions and related data are available as supplementary information of the report
445 by Huizinga et al. (2017) at the JRC Publications Repository:
446 <https://publications.jrc.ec.europa.eu/repository/handle/JRC105688>

447 The flood protection dataset and the dataset of adaptation measures developed for this work are
448 available as supplementary material of the present manuscript.

449 All the other datasets generated during and/or analysed during the current study are available from the
450 corresponding author on reasonable request.

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453 **Code availability:**

454 The open-source LISFLOOD hydrological model used in this work is available in Github at [https://ec-](https://ec-jrc.github.io/lisflood/)
455 [jrc.github.io/lisflood/](https://ec-jrc.github.io/lisflood/)

456 The source code of LISFLOOD-FP8.0 is available in Zenodo under a GNU General Public License v3.0 for
457 any non-commercial use and can be downloaded at <https://doi.org/10.5281/zenodo.4073011> .

458 All the other computer codes or algorithms used to generate the results reported in the paper are
459 available from the corresponding author on reasonable request.

460

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465 **Author contributions** FD: conceptualization, formal analysis, investigation, data curation, writing
466 (original draft, review and editing); LM: methodology, formal analysis, investigation, data curation,
467 writing (review and editing); AB: data curation, validation, visualization; LA: methodology, investigation,
468 writing (review and editing); LF: conceptualization, project administration, writing (original draft, review
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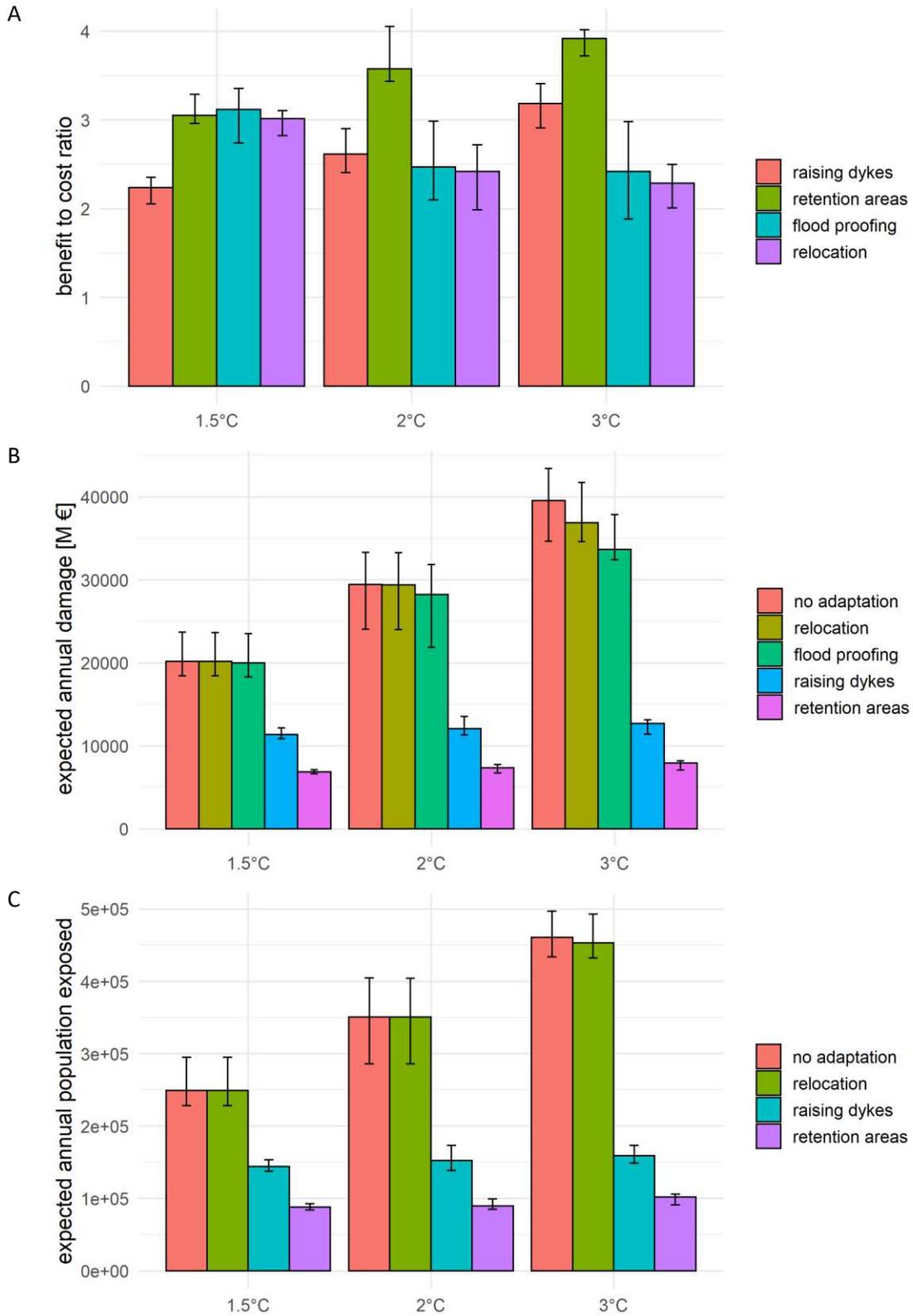
647 Table 1. Overview of four adaptation strategies at country level for the 3°C warming scenario in 2100.
648 Benefit-Cost Ratio (BCR) calculated as ratio of discounted benefits and costs over period 2020-2100.
649 Reduction (in %) in Expected Annual Damage (EAD) calculated as difference in undiscounted damage in
650 2100 with and without adaptation. Cost of implementation (in €million/year) reflect average of
651 undiscounted costs over period 2020-2100. All values refer to the ensemble average.

Country	Retention areas			Dikes strenghtening			Flood proofing			Relocation		
	BCR	EAD red.	Costs €M/y	BCR	EAD red.	Costs €M/y	BCR	EAD red.	Costs €M/y	BCR	EAD red.	Costs €M/y
Austria	2.8	83%	108.8	2.4	71%	108.1	3.4	0.3%	0.8	0.2	0%	0.0
Belgium	4.2	91%	93.1	3.6	90%	107.4	10.2	12%	37.4	5.9	0.0%	0.0
Bulgaria	3.1	78%	18.7	2.5	42%	10.8	2.0	6%	3.1	0.9	0.1%	0.0
Croatia	3.1	96%	84.1	2.2	74%	79.1	2.1	0.6%	1.2	0.5	0.0%	0.0
Cyprus	0.0	0%	0.0	0.0	0%	0.0	0.0	0%	0.0	0.0	0%	0.0
Czechia	4.4	88%	97.9	2.7	73%	129.9	1.8	15%	55.7	3.5	0.1%	0.1
Denmark	3.6	91%	6.8	1.7	59%	7.8	0.6	6%	1.5	0.0	0%	0.0
Estonia	1.1	74%	17.6	2.3	74%	11.0	18.1	27%	3.9	237.1	1.3%	0.4
Finland	3.5	80%	91.5	3.3	70%	76.5	4.4	17%	31.2	8.8	0.4%	0.8
France	3.3	87%	651.2	2.8	74%	629.6	2.3	2%	36.3	4.7	0.0%	0.2
Germany	3.8	79%	395.2	3.2	74%	439.3	2.8	1%	16.1	5.0	0.0%	0.2
Greece	2.7	73%	12.9	0.8	41%	12.4	3.8	1%	0.6	2.6	0.2%	0.0
Hungary	3.2	87%	119.5	2.8	71%	120.1	12.2	8%	35.4	14.6	0.0%	0.0
Ireland	3.2	86%	35.5	2.5	75%	39.1	1.6	5%	5.1	15.0	0.0%	0.0
Italy	4.6	88%	264.7	2.7	76%	386.3	1.9	6%	75.2	33.2	0.1%	0.3
Latvia	2.5	67%	40.0	2.1	72%	49.7	2.5	18%	27.5	5.3	0.1%	0.0
Lithuania	2.3	48%	29.0	1.8	40%	19.2	5.7	1%	1.0	52.5	0.1%	0.0
Luxembourg	4.0	91%	7.9	3.5	91%	8.8	1.0	15%	3.7	0.0	0%	0.0
Netherlands	3.9	58%	22.9	4.1	71%	26.5	0.1	0%	0.0	0.0	0%	0.0
Poland	2.3	71%	205.1	1.7	41%	137.0	3.3	3%	17.5	0.0	0%	0.0
Portugal	2.6	9%	0.9	1.8	7%	0.8	6.8	0.3%	0.0	4.6	0.1%	0.1
Romania	2.5	62%	94.1	2.7	46%	59.7	3.1	15%	35.6	5.6	0.2%	0.0
Slovakia	3.6	87%	47.6	2.5	66%	50.5	3.4	0.1%	0.3	2.1	0.7%	1.1
Slovenia	2.5	82%	23.0	1.7	59%	22.5	4.1	0.5%	0.3	0.3	0.0%	0.0
Spain	1.8	32%	64.7	1.8	10%	22.1	3.1	3%	4.1	1.5	0.1%	0.0
Sweden	5.9	87%	118.2	8.3	81%	89.3	3.7	65%	144.2	3.2	0.5%	0.6
United Kingdom	7.4	94%	181.2	3.8	89%	329.3	2.2	27%	250.1	2.7	32.5%	110.5
EU+UK	4.0	82%	2869	3.2	71%	3020	2.6	12%	788	2.2	0.2%	1.0

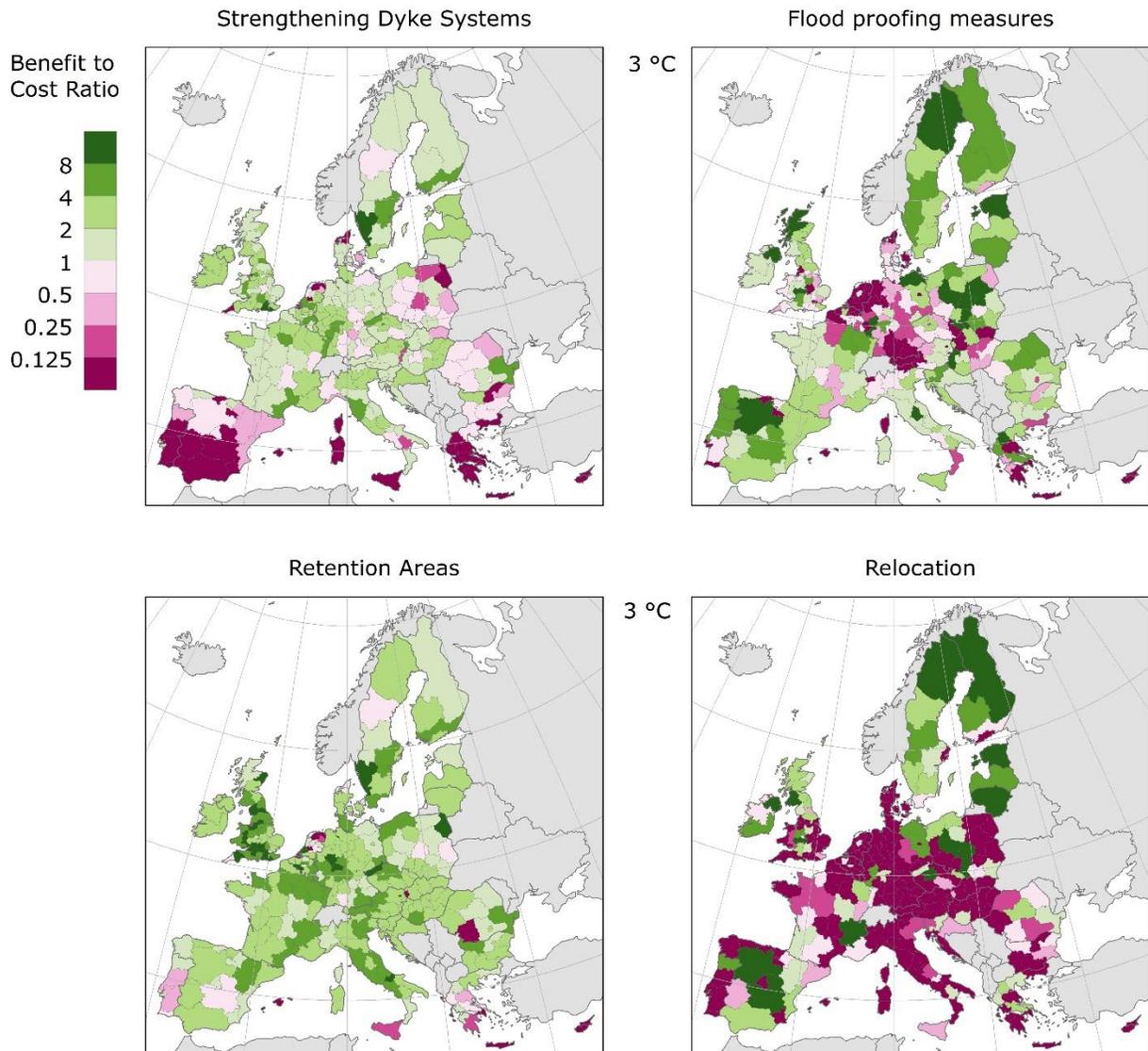
652 **Table 2.** Summary of unit costs derived from the database of adaptation measures. The table reports
 653 also the damage reduction ratios for Flood-proofing measures for buildings. The complete database is
 654 available as supplementary material.

		Normalized unit cost (2015)		
		Average	25% quantile	75% quantile
Dike systems reinforcement	€/m/m	6405	1829	9514
Retention areas	€/m3	3.73	1.05	5.00
Flood-proofing measures	€/m2	376	493	156
Relocation	€/m2	1373	906	1826
		Damage reduction ratio (average)		
		Average	25% quantile	75% quantile
Flood-proofing measures	(-)	41%	10%	80%

655



657 Figure 1. Summary of the outcomes for the four adaptation strategies considered under 1.5°C, 2°C and
658 3°C warming scenario. Panel (A) shows the benefit to cost ratio (A) calculated from total discounted
659 benefits and costs over the period 2020-2100. Panels (B) (C) describe future undiscounted economic
660 damages (B) and population exposed (C) for the year 2100, calculated under a no-adaptation scenario
661 and optimizing each of the four adaptation strategies. Coloured bars and error bars indicate respectively
662 the average and the 75th-25th quantiles of the model ensemble. All results are based on averaged at EU +
663 UK level.



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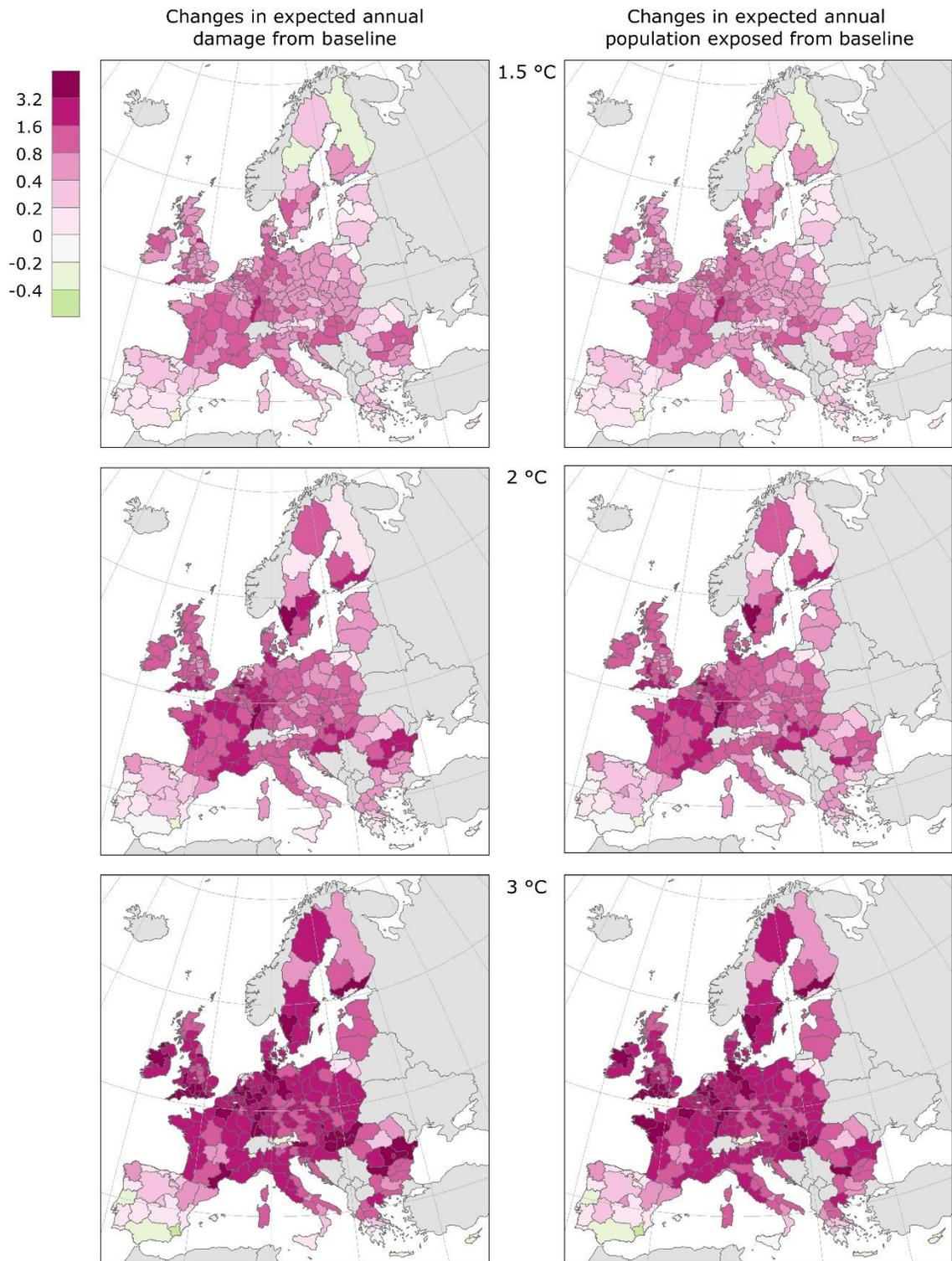
Figure 2. Benefit-cost ratio (BCR) at NUTS2 level for the adaptation strategies 'river dikes', 'retention

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areas', 'flood proofing measures' and 'relocation' for the 3°C warming scenario. The values refer to the

668

ensemble average.



670 Figure 3. Relative change in expected annual economic damage (left) and population exposed (right) for
671 1.5, 2 and 3°C warming scenarios in 2100 with respect to the baseline. The values refer to the ensemble
672 average.

Supplementary Files

This is a list of supplementary files associated with this preprint. Click to download.

- [Databaseadaptationcostspaper.xlsx](#)
- [DottorietalAdaptationsupplementv5.pdf](#)
- [floodProtectionv2019paper3.zip](#)
- [DottorietalAdaptationsupplementv6.pdf](#)