

# Simulating Grazing Effects on Soil Organic Carbon Dynamics in Semi-arid Rangelands (Southern Iran)

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## Research Article

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1 **Simulating grazing effects on soil organic carbon dynamics in semi-arid**  
2 **rangelands (Southern Iran)**

3 RUNNING TITLE: Simulating soil organic carbon

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19 **Simulating grazing effects on soil organic carbon dynamics**  
20 **in semi-arid rangelands (Southern Iran)**

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22 **Abstract**

23 Grazing is one of the main causes of rangeland degradation worldwide, due to the  
24 effects of overgrazing on vegetation cover and biodiversity. But few data are available on  
25 the effect of grazing intensity on the dynamics of soil organic carbon (SOC) and soil labile  
26 organic carbon (SLOC). So far, very few studies have addressed the modeling of SOC  
27 dynamics under different grazing intensities, and SLOC dynamics has not been modeled  
28 yet. In this study, we used the CENTURY model to select the most effective grazing  
29 management in terms of carbon sequestration (SOC and SLOC stocks) in semi-arid  
30 rangelands of Southern Iran. The effect of four different scenarios of grazing intensity was  
31 simulated: no grazing, light grazing (LG), moderate grazing (MG), and heavy grazing  
32 (HG). The results of long-term model simulations (2015-2100), indicated that SOC stocks  
33 will change by 2.7, 1.7, -23.4, and -24.6% in the scenarios of exclusion, LG, MG, and HG  
34 respectively compared to 2014. With increasing grazing intensities, SLOC stocks in LG,  
35 MG, and HG scenarios significantly decreased compared to the no grazing scenario by  
36 26.1, 59.6, and 70%, respectively. Thus, this study suggests recommending light grazing  
37 management for semi-arid rangelands of Iran and also SLOC as a suitable index for  
38 studying the effect of grazing on soil carbon.

39 Key words: CENTURY model; SLOC; SOC; grazing intensity; semi-arid rangelands.

40

41 **Introduction**

42 Soils have a prominent role in maintaining the balance of the global carbon cycle (La,  
43 2008), and any small change in soil organic carbon (SOC) has high impacts on the

44 concentration of CO<sub>2</sub> in the atmosphere (Smith *et al.*, 2008; Muñoz-Rojas *et al.*, 2015).  
45 Rangelands have a high potential to sequester the atmospheric CO<sub>2</sub> in the soil due to their  
46 prevalence in about 50% of land area worldwide, and globally can store up to 30% of SOC  
47 (Derner and Schuman, 2007). In particular, one of the most effective factors in SOC storage  
48 is grazing management (Mcsherry and Ritchie, 2013; Waters *et al.*, 2016). Grazing is one  
49 of the main causes of rangeland degradation particularly in arid and semi-arid  
50 environments, due to the effects of overgrazing on vegetation cover, biodiversity of plant  
51 species, unpalatable species, and livestock trampling that enhance soil loss by erosion,  
52 reduce the production potential of rangelands, and negatively affect SOC, SOC pools and  
53 soil biological activity (Derner and Schuman, 2007; Cao *et al.*, 2013; Al-Rowaily *et al.*,  
54 2015; Sepe *et al.*, 2015).

55 Recent research has shown that grazing intensity affects SOC stocks, net primary  
56 production, root growth, plant shoot/root allocation, soil C/N ratio, and organic matter  
57 decomposition (Derner *et al.*, 2006; Derner and Schuman, 2007; Pineiro *et al.*, 2010;  
58 Ritchie, 2014; Papanastasis *et al.*, 2015; Orgill *et al.*, 2016). Furthermore, other studies  
59 have shown that short-term periods are not adequate for the evaluation of grazing effects  
60 on SOC stocks, so long-term investigations have been recommended (Medina-Roldán *et*  
61 *al.*, 2012).

62 To investigate the effects of grazing on soil carbon, two pools of soil decomposable  
63 carbon including soil labile organic carbon (SLOC), and light fraction organic carbon  
64 (LFOC) have been considered to be suitable indicators (Chen *et al.*, 2012; Sheng *et al.*,  
65 2015). SLOC refers to carbon with high solubility, quick movement and easy  
66 mineralization influenced by plants and soil microorganisms (Cao *et al.*, 2013). SLOC  
67 includes particulate organic carbon (POC), readily oxidized carbon (ROC), soil microbial  
68 biomass carbon (SMBC), dissolved organic carbon (DOC), and light fraction organic  
69 carbon (LFOC) (Geng *et al.*, 2009). SLOC has a relatively short turnover time and has  
70 shown higher sensitivity to management practices when compared to the total SOC stock.  
71 Thus, this pool has been suggested as suitable and sensitive indicator to study the effect of  
72 grazing on SOC (Soon *et al.*, 2007; Cao *et al.*, 2013).

73 In recent years, simulation models have been recognized as effective tools for decision-  
74 making and ecosystems management in relation to soil carbon sequestration. The  
75 CENTURY model is a process-based ecological model to simulate SOC dynamics,  
76 integrating the effects of climate, soil driving variables and management in different  
77 ecosystems (croplands, grasslands, forests and savannas) on soil fertility parameters and  
78 water dynamics (Parton *et al.*, 1987). The model includes specific options to simulate the  
79 effect of grazing on plant production and soil carbon.

80 About 70% of Iran's rangelands are located in arid and semi-arid regions, and are  
81 generally used as grazing pastures. At present, no specific research has been conducted to  
82 simulate the effect of grazing on SOC stocks in semiarid environments, especially in terms  
83 of SLOC pools. Bajgah rangeland is one of the semi-arid rangelands of southern Iran that  
84 after the Iranian Revolution and the establishment of an army garrison around this  
85 rangeland and the placement of another part of the rangelands at the Faculty of Agriculture,  
86 Shiraz University, these pastures were not grazed after the revolution. Our hypothesis was  
87 that no grazing and light grazing would not have a negative effect on soil carbon stock, and  
88 with increasing grazing intensity in the long term, soil organic carbon stock and the soil  
89 carbon sequestration rate would decrease in these rangelands. If the results of the studies  
90 showed us that a type of grazing intensity management cannot have a negative effect on  
91 soil carbon, we would be able to propose this type of grazing management to the school  
92 authorities, which will not reduce the soil organic carbon stock (soil health), and it can also  
93 provide animal husbandry and livestock products. Therefore, the aims of this study are: i)  
94 to assess SOC stocks under different scenarios of grazing intensities in semi-arid  
95 rangelands of Southern Iran, ii) to simulate the long-term variations of SOC and SLOC  
96 pool stocks with the CENTURY model, and iii) to indicate the more effective grazing  
97 management.

## 98 **Materials and methods**

### 99 *Study area*

100 The Bajgah rangelands are located in the northwest of the School of Agriculture Shiraz  
101 University (latitude: 29°36' N, longitude: 52°32' E, elevation: 1820 m) in Southern Iran  
102 (Figure 1). Based on the 43-year (1972-2014) statistics of the local meteorological station,  
103 the average annual precipitation and temperature are 388.4 mm and 13.4°C, respectively.  
104 The rangelands of Bajgah are classified as semi-arid, and grasses are mainly C<sub>3</sub> type.  
105 *Bromus sp.*, *Agropyron sp.*, *Onobrychis sp.*, *Medicago sp.*, *Hordeum sp.*, and *Poa sp.* have  
106 been observed predominantly within the studied region. Soils are mainly Fluvisols and  
107 moderately deep. The historical management of rangelands belong to three periods: the  
108 period before nationalization of forests and rangelands (before 1963), the period of  
109 nationalization of rangelands (1963-1979), and the contemporary (Iranian Revolution)  
110 period (1979 to present), and includes light grazing, moderate grazing, and no grazing  
111 management respectively (Vanaee *et al.*, 2017). After the Iranian Revolution and the  
112 change of government and also occurrence of a war between Iran and Iraq, the grazing  
113 intensity on rangelands in Iran was increased to produce more food and the supply of meat.  
114 But the Bajgah rangelands because they were within the garrison of the army and the lack  
115 of permission to enter the ranchers in the area caused the rangelands not to be grazed. The  
116 management of Bajgah rangelands has been similar to the rest of Iran until the occurrence  
117 of the revolution (that's mean the management of this region has been the same as Vanaee  
118 *et al.* (2017)) and since that time, the grazing of livestock has been eliminated in the Bajgah  
119 rangeland. For the present study, four scenarios were selected light grazing (LG) with 25%  
120 of live shoots removed by grazing, moderate grazing (MG) with 50% of live shoots  
121 removed by grazing, heavy grazing (HG) with 75% of live shoots removed by grazing, and  
122 no grazing management. According to surveys with native people, the duration of grazing  
123 in the rangelands of Bajgah is four months from the beginning of December to the end of  
124 March.

125

126 *CENTURY model*

127 The CENTURY model simulates the long-term dynamics of C, N, P and S for different  
128 ecosystems. This model represents SOC in three conceptual pools: active, passive and slow  
129 (Parton *et al.*, 1988) and the sum of carbon in these pools represents the SOC stock  
130 (Tornquist *et al.*, 2009). The passive pool contains a high level of lignin, chemically  
131 resistant to decomposition and with a long (800-1200 years) turnover time (Parton *et al.*,  
132 1988). Slow pool mainly contains cellulose, hemi-cellulose, and organic matter physically  
133 protected inside soil aggregates, with a turnover time of 20-50 years. Active pool contains  
134 microorganisms and their products (proteins, amino acids, sugars, and starches) and  
135 represents the soil labile organic carbon pool (SLOC), with a turnover time between 2 and  
136 4 years. In the Century model, the total organic carbon of soil is the sum of soil organic  
137 carbon contained in the three pools of slow, passive and active. In this study, we considered  
138 the soil organic carbon stock in the active pool equivalent to SLOC (however, it appears  
139 that the active pool is the same as SLOC). In CENTURY, SLOC is included in a variable  
140 called *som1c (1)* and is highly influenced by management practices such as grazing (Bot  
141 and Benites, 2005). The model has three options (GRZEFF = 0, 1, 2) to consider the effect  
142 of grazing on grass production (aboveground and below ground), root/shoot ratio and soil  
143 carbon. For option 1 (GRZEFF=0) there are no direct impacts of grazing on plant  
144 production except for the removal of vegetation and return of nutrients by the animals.  
145 Option 2 (GRZEFF=1) is referred to as the lightly grazed effect (Holland *et al.*, 1992) and  
146 includes a constant root/shoot ratio (not changing with grazing) and a linear decrease in  
147 potential plant production with increasing grazing intensity. Option 3 (GRZEFF=2) is  
148 referred to as the heavy grazed (Holland *et al.*, 1992) option, and includes a complex  
149 grazing optimization curve for aboveground plant production. These options are adjusted  
150 by some parameters available in the *CROP.100* file (Parton *et al.*, 1987). The input data to  
151 the CENTURY model are included in 12 files, each containing a specific set of variables  
152 (Figure 2).

153 *Sampling and laboratory measurements*

154 Previous soil organic carbon measurements were available for 1987, 1995, 1996, 2010  
155 and 2012 (30, 30, 45, 40 and 30 samples, respectively) under no grazing management. In  
156 the present research, an additional soil sampling on the same area (4000 ha) was carried  
157 out from September to December 2014. A simple random sampling was adopted to collect  
158 independent and unbiased samples as well as remain within the limits of time, money, and  
159 staff available for sampling. After removing the litter layer, 90 soil samples were taken  
160 from the top soil (0-20 cm). Soil samples were air dried, visible plant materials were  
161 removed, and the analyses were made on the < 2 mm dried soil fraction after sieving.  
162 Particle-size distribution, soil reaction, soil organic carbon, and total soil nitrogen were  
163 determined by the Hydrometry method (Bouyoucos, 1962), pH meter, Walkley and Black  
164 method (Walkley and Black, 1934), and Kjeldahl method (Bremner *et al.*, 1982)  
165 respectively. To determine the soil bulk density, two soil core samples (in addition to  
166 routine soil samples) were collected in each sampling point using the Core method (Blake  
167 and Hartge, 1986). Finally, SOC stock was calculated in the 0-20 cm layer using Eq. (1):

$$\text{SOC Stock} = \text{OC (\%)} \times \text{layer thickness (cm)} \times \text{bulk density (g cm}^{-3}\text{)} \quad (1)$$

169  
170 For vegetation measurements, first the proper number of plots was determined to  
171 collect representative data. Thereafter, 17 quadrats (plots) with dimensions of 1×1 m  
172 were randomly established throughout the rangelands. Then vegetation sampling was  
173 carried out during the main growing season from May to July 2014. In each plot, litter and  
174 plant biomass (above and below ground) samples were dried at 60°C for 48 h and weighed.  
175 Thereafter, 2 g samples were oven ashed at 550°C for 6 h, and plant carbon and the total  
176 plant nitrogen were measured using the ash weight, primary weight, and ratio of organic  
177 carbon to organic material relationship (Birdsey *et al.*, 2000) and Kjeldahl method (Kirk,  
178 1950), respectively.

179

180 *Model initialization*

181 In this study, the CENTURY model version 4.0 (Parton *et al.*, 1987) was used to  
182 simulate SOC stock and SLOC stock. Values of monthly mean maximum and minimum  
183 temperatures and precipitation were obtained from the 43-year (1972-2014) statistics of the  
184 meteorological station of the College of Agriculture in Shiraz University. Data on climatic  
185 parameters for future simulation periods were stochastically generated by CENTURY  
186 model based on the skewed distribution of climate data (Wang *et al.*, 2008; Tornquist *et*  
187 *al.*, 2009) from 1972 to 2014. The site-specific parameters and the soil and plant cover  
188 parameters are shown in Table 1. The physiological and ecological parameters for C<sub>3</sub> plants  
189 were specified in the *CROP* file of the model.

#### 190 *Model calibration*

191 CENTURY model is able to simulate total soil organic carbon (SOC) and carbon pools  
192 under the equilibrium state (Tornquist *et al.*, 2009), that is a method to run the model for  
193 an initial period (7000-10000 years) based on the conditions of soil and vegetation before  
194 any anthropic disturbance (Kamoni *et al.*, 2007). The equilibrium condition represents the  
195 baseline for the evaluation of the management effects on soil organic carbon in the  
196 ecosystem. Therefore, the user should first determine the conditions of “equilibrium” and  
197 then import them to the CENTURY model. In this research, the CENTURY model was run  
198 for 10,000 years considering (Tornquist *et al.*, 2009; Wilson *et al.*, 2009) that the Bajgah  
199 rangelands have been under coverage of C<sub>3</sub> grasses and light grazing before 1963. Then,  
200 the CENTURY model was run for the periods after 1963. Finally calibrated using the data  
201 obtained from previous samplings (following the initial sampling in 1987). The calibration  
202 process included repetitive running of the model and reviewing the model outputs, then  
203 adjusting the model’s default parameters until measured carbon (Table 1) was equal to the  
204 simulated SOC stock (Tornquist *et al.*, 2009; Wilson *et al.*, 2009).

#### 205 *Model validation*

206 SOC data measured in September and December 2014, as well as the data of the  
207 previous sampling (1995, 1996, 2010 and 2012) were used for the model validation (265

208 data). To validate the model, we compared the model output to a set of data independent  
209 from the calibration stage. Certainly, the data that was measured in this study was in a no  
210 grazing management.

211 The SOC stock data are inserted into Century model on an average basis and this is one of  
212 the defects of the Century model that does not consider spatial distribution and works based  
213 on the central indices (the Century model based on mean). Finally, we will have 6 pairs (or  
214 6 points) of data due to the use of 6 sampling times to model validation. Some statistical  
215 comparisons between the simulated and measured values including determination factor  
216 ( $R^2$ ), correlation coefficient ( $r$ ), root mean square error (RMSE) (Eq. 2),  $RMSE_{0.05}$  (Eq. 3)  
217 and modelling efficiency (EF) (Eq. 4) were used for model validation where  $O_i$ ,  $P_i$ ,  $\bar{O}$ ,  $SE_i$ ,  
218  $t_{m0.95}$ , and  $n$  are the observed value, simulated value, the mean of observed values, the  
219 standard deviation of the observed values, t-student value (95% probability level), and the  
220 number of observations, respectively.

221

$$222 \quad RMSE = \frac{100}{\bar{O}} \times \sqrt{\frac{\sum_{i=1}^n (O_i - P_i)^2}{n}} \quad (2)$$

$$223 \quad RMSE_{0.05} = \frac{100}{\bar{O}} \sqrt{\frac{\sum_{i=1}^n (SE_i \times t_{m0.95})^2}{n}} \quad (3)$$

$$224 \quad EF = 1 - \frac{\sum_{i=1}^n (P_i - O_i)^2}{\sum_{i=1}^n (O_i - \bar{O})^2} \quad (4)$$

225

### 226 *Simulation of grazing scenarios*

227 To investigate the effect of different scenarios of grazing intensities on SOC and SLOC  
228 stocks, the model was run in the four different grazing scenarios from 2015 to 2100 in the  
229 rangelands of Bajgah: no grazing, light grazing, moderate grazing and heavy grazing. To  
230 test the effect of different scenarios of grazing intensity on the simulated parameters, LSD  
231 test was used. All statistical analyses were carried out by SPSS software.

## 232 **Results**

### 233 *Model calibration and validation*

234 The model default parameters were adjusted during the calibration until the simulated  
235 SOC stock with a value of 3251.6 (g m<sup>-2</sup>) was close to the measured value of 3228.5 (g m<sup>-2</sup>)  
236 in 1987 representing the original levels of SOC stock. Thus, the CENTURY model was  
237 considered suitable for simulating the dynamics of SOC of the study area. The comparison  
238 of variations in the carbon pools curves during the equilibrium period (Figure 3) indicated  
239 that the slow and active carbon pools increased rapidly during the first 300 and 100 years,  
240 respectively, while the passive pool decreased slowly throughout the entire equilibrium  
241 period. At the equilibrium state, the fractions of slow, active, and passive organic carbon  
242 pools were 50%, 3.8%, and 46.2% of the total SOC respectively.

243 There was a significant linear relationship ( $R^2=0.86$ ) between measured and simulated  
244 SOC stocks (Figure 4), and the correlation coefficient ( $r$ ) was 0.93 (Table 2). The root mean  
245 square error (RMSE), that indicates the total difference between the measured and  
246 simulated values, was lower than  $RMSE_{0.05}$ , also EF (modelling efficiency) was 83%  
247 (Table 2), thus the observed and simulated data are not significantly different. Overall,  
248 results indicate that the CENTURY model accurately simulates the dynamics of SOC stock  
249 in the Bajgah rangelands (Figure 4).

### 250 *Changes of SOC stocks under grazing*

251 Long-term changes of SOC stocks in response to grazing until 2100 showed significant  
252 differences ( $p<0.01$ ), and the minimum (2883.4 g m<sup>-2</sup>) and maximum (3436.2 g m<sup>-2</sup>) SOC  
253 stocks were observed in the heavy and no grazing scenarios (Table 3). Figure 5 shows the  
254 trend of changes in the SOC stocks from 1963 to 2014, followed by the variations of SOC  
255 stock from 2015 to 2100 under the four different grazing scenarios.

256 The simulations showed that the SOC stocks during the years from 2015 to 2100  
257 increased by 2.7% under no grazing and by 1.7% under the light grazing scenario, but there  
258 was no significant difference ( $p<0.01$ ) between them (Table 3 and Figure 5). However,  
259 despite the higher increase of SOC stock in the no grazing scenario compared to the light

260 grazing scenario, both changes were not significant in relation with the baseline year (2014)  
261 (Table 3).

262 At higher grazing intensities, SOC stocks decreased significantly in the two scenarios  
263 of moderate and heavy grazing in comparison with the no grazing and light grazing  
264 scenarios (Table 3 and Figure 5). The simulations indicated that SOC stocks decreased by  
265 23.4 and 24.6% in 2100 compared to 2014 under the scenarios of moderate and heavy  
266 grazing, respectively (Table 3). Despite the considerable reduction of SOC stocks under  
267 the moderate and heavy grazing scenarios, no significant difference was observed between  
268 them (Table 3).

#### 269 *Changes of SLOC stock under grazing*

270 The simulation of soil labile organic carbon (SLOC) stocks across the different grazing  
271 scenarios (Table 4) indicated significant differences among all scenarios ( $p < 0.01$ ). Figure  
272 6 shows the annual variations of SLOC stock from 1963 to 2014, and then from 2015 until  
273 2100 under the different grazing scenarios. Simulations also indicated that SLOC stocks  
274 will be decreased by 26.1, 59.6, and 70% in the light, moderate, and heavy grazing  
275 scenarios respectively, compared to the no grazing scenario (Table 4).

## 276 **Discussion**

277 The results of validation indicated that the CENTURY model was able to simulate the  
278 changes of SOC (Table 2), and can be applied for the long-term simulation of the SOC  
279 changes in semi-arid rangelands of Bajgah. Wang *et al.* (2008) by using a validated Century  
280 model simulated the effects of different grazing intensities on SOC changes in Northeast  
281 China, and results showed that SOC will keep constant at lower grazing intensities. So far,  
282 different studies indicated the capability of the Century model for SOC simulation in  
283 rangelands for example Zhang *et al.* (2007) in grasslands on the Qinghai-Tibetan Plateau  
284 under alpine climatic conditions, Brown *et al.* (2010) in rangelands of southwestern United  
285 States, Vanaee *et al.* (2017) in different meadows of Kurdistan province with continental  
286 climate and high rainfall.

287 The simulation results showed increased SOC stocks by adopting the no grazing  
288 management in the Bajgah rangelands (Figure 5). No grazing has an enhanced canopy  
289 cover and a high density of litter (Shifang *et al.*, 2008), aboveground and below-ground  
290 litter accumulation (Chen *et al.*, 2012), and a positive effect on improvement of vegetation  
291 diversity (Al-Rowaily *et al.*, 2015) that finally lead to greater SOC stocks (Derner and  
292 Schuman, 2007; Mekuria *et al.*, 2007; Chen *et al.*, 2012). Most studies have reported  
293 increased soil carbon levels by applying no grazing management, in agreement with the  
294 results obtained from this study (Mekuria *et al.*, 2007; Chen *et al.*, 2012). Nevertheless,  
295 other researches have shown that soil organic carbon may increase under grazing compared  
296 to no grazing management due to the C immobilization in the excessive plant litter material,  
297 and the development of annual grasses with a different rooting system favoring organic  
298 matter accumulation (Reeder *et al.*, 2002; Orgill *et al.*, 2016). Waters *et al.* (2016) argued  
299 that the soil inherent properties, the climate, the landscape morphology and position, and  
300 also the composition of the vegetation community are involved in the difference of the  
301 ecosystem's carbon response to the no grazing management.

302 In the Bajgah rangelands, long-term simulations indicated that the light grazing  
303 intensity presents increased SOC stocks compared to moderate and heavy grazing (Figure  
304 5), with not significant difference compared with the no grazing scenario (Table 3). Wang  
305 *et al.* (2001) stated that appropriate grazing decreases the amount of the mature and old  
306 tissues of the plant, and therefore improves the photosynthetic rate of the plant leaves  
307 remained after grazing, as well as the cycle of nutrients and water in the plant. Wright *et al.*  
308 (2004) reported that long-term grazing at low grazing intensity of Bermuda-grass pastures  
309 can increase SOC and SON concentrations and could have strong potential for C and N  
310 sequestration. This is mainly due to enhanced turnover of plant material and excreta under  
311 low grazing intensity. Mcsherry and Ritchie (2013) also demonstrated that in the sites  
312 where C<sub>3</sub> grasses are predominant, light grazing leads to increased SOC stocks, thus  
313 supporting the results obtained from this study. Some studies reported that light grazing  
314 intensity was a useful management for enhancing C sequestration (Da Silva *et al.*, 2014).

315 Cecagno *et al.* (2018) showed a higher potential of the soil for C sequestration with a low  
316 grazing intensity.

317 The long-term simulation of different scenarios of grazing from 2015 to 2100,  
318 demonstrated also that as the grazing intensity increases, SOC stocks would decrease  
319 (Figure 5). Other recent studies stated that the change of the SOC stocks of rangelands in  
320 response to grazing is dependent on the grazing intensity (Mcsherry and Ritchie, 2013;  
321 Ritchie, 2014; Papanastasis *et al.*, 2015). In this study, the simulations indicated a SOC  
322 stock reduction from 2015 until 2100 in the two scenarios of moderate (with 50% of live  
323 shoots removed by grazing) and heavy grazing (with 75% of live shoots removed by  
324 grazing) (Figure 5); and percent reduction of SOC stocks was 23.4 and 24.6, respectively  
325 (Table 3). Wang *et al.* (2008), by using Century model showed that high grazing intensities  
326 would have higher probability to release more carbon into atmosphere, thus grassland  
327 ecosystems would act as a carbon source. They also showed that when 40% live shoots  
328 were removed by grazing event per month, about 20% soil organic carbon was lost in good  
329 agreement with our results. Vanaee *et al.* (2017) in Dehgolan meadows in Kurdistan  
330 province (in west of Iran) by using Century model predicted that SOC stock under  
331 moderate grazing and high grazing in the period from 2014 to 2100 will be reduced by 23  
332 and 25 percent, respectively. Their results were very close to our study results.

333 Increased grazing intensity leads to a decreased carbon input from litterfall due to  
334 vegetation destruction of plant cover and consumption of litter by herbivores, a reduction  
335 of leaves, an increase of soil temperature in the areas where soil is bare, and the  
336 development of proper conditions for microbial decomposition (Abril and Bucher, 2001).  
337 All of these factors lead to lowered SOC stocks (Derner *et al.*, 2006; Wang *et al.*, 2008).

338 Some studies have reported a negative effect of grazing on SOC in regions with  
339 precipitation lower than 600 mm per year (Golluscio *et al.*, 2009). Pineiro *et al.* (2010)  
340 stated that the levels of SOC under semi-arid climate conditions decrease under grazing.  
341 According to Mcsherry and Ritchie (2013) the intensity of grazing and the type of grass  
342 are the most important factors regulating SOC in rangelands after environmental variables.

343 Furthermore, decreased SOC levels with increased grazing intensity in rangelands where  
344 C<sub>3</sub> grasses are predominant have been reported by other researches (Potter *et al.*, 2001).  
345 Mcsherry and Ritchie (2013) also suggest that grazing in the rangelands where C<sub>3</sub> grasses  
346 are predominant, under moderate and heavy grazing conditions, has a negative effect on  
347 the soil organic carbon. Considering that in the semi-arid rangelands of Bajgah annual  
348 precipitation is lower than 600 mm per year, and the dominant grasses are of C<sub>3</sub> type, the  
349 decreased levels of SOC stock under moderate and heavy grazing conditions in these  
350 rangelands is reasonable and in agreement with the existing literature.

351 The results of the simulation indicated decreased levels of soil labile organic carbon  
352 (SLOC) with increased grazing intensity in the Bajgah rangelands (Figure 6). Cao *et al.*  
353 (2013) also stated that with increased grazing intensity, the SLOC stocks would reduce.  
354 The SLOC stocks decreased by 59.6 and 70% under moderate and heavy grazing scenarios  
355 respectively, in comparison with the no grazing scenario (Table 4). Although only a few  
356 studies are available on the effect of grazing on SLOC, Chen *et al.* (2012) indicated that  
357 SLOC stocks in grazed sites is 73.3% less than in no grazing sites, and this finding is close  
358 and in agreement with our results.

359 Although the amount of SOC stocks had no significant difference between the light  
360 grazing and no grazing scenarios, as well as between moderate and heavy grazing scenarios  
361 (Table 3), there was a significant difference ( $p < 0.01$ ) among the levels of SLOC stocks in  
362 all grazing scenarios (Table 4). In addition, the annual changes of the SLOC pools in  
363 comparison with the annual variations of SOC stocks showed that SLOC is more sensitive  
364 to grazing and it responds much faster to grazing than SOC (Figure 5 and 6). Accordingly,  
365 the SLOC stock can be proposed as a proper and sensitive parameter to be preferred to  
366 SOC stock for researches on grazing intensity. Due to the slow reaction of the SOC stock  
367 (Medina-Roldán *et al.*, 2012), the majority studies have proposed the SLOC as a proper  
368 parameter to study the grazing effects on soil carbon (Soon *et al.*, 2007; Cao *et al.*, 2013).

## 369 **Conclusions**

370 In this research, a reduction of SOC and SLOC stocks with the increase of grazing  
371 intensity was observed in semi-arid rangelands of Bajgah (Southern Iran). The long-term  
372 simulation of light grazing had not a significant decreasing effect on the SOC stock.  
373 Therefore, light grazing management is recommended in Bajgah rangelands. Furthermore,  
374 this research suggests and confirms the effectiveness of SLOC stock as a sensitive and  
375 proper parameter for the investigation of different grazing scenarios effects in the future  
376 research.

377 **Author contribution** Contributions of all authors to this work were as  
378 follows:

379 Sayed Fakhreddin AFZALI: Conceptualization, writing—original draft and editing

380 Bijan AZAD: Conceptualization, writing—original draft

381 Rosa FRANCAVIGLIA : writing , Literature reviewing

382

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386 **Declarations**

387 **Ethics approval and consent to participate** Not applicable.

388 **Consent for publication** Not applicable.

389 **Competing interests** The authors declare that they have no competing  
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391

392 **References**

393 Abril, A.; Bucher, E.H. Overgrazing and soil carbon dynamics in the western Chaco of Argentina. *Appl.*  
394 *Soil Ecol.* **2001**, 16 (3), 243-249.

395

396 Al-Rowaily, S.L.; El-Bana, M.I.; Al-Bakre, D.A.; Assaeed, A.M.; Hegazy, A.K.; Ali, M.B. Effects of  
397 open grazing and livestock exclusion on floristic composition and diversity in natural ecosystem of  
398 Western Saudi Arabia. *Saudi J. biol. Sci.* **2015**, 22(4), 430-437.

399 Birdsey, R.; Heath, I.S.; Williams, D. Estimation of carbon budget model of the United States forest  
400 sector. *Advances in Terrestrial Ecosystem carbon Inventory, Measurements, and Monitoring*  
401 *Conference in Raleigh, North Carolina, October 3-5, 2000*, 51-59.

402

403 Blake, G.R.; Hartge, K.H. Bulk density. *Methods of Soil Analysis. Part 1. Physical and mineralogical*  
404 *methods. Soil Sci. Soc. Am. Pub. 1986*; pp. 363-376.

405

406 Bot, A.; Benites, J.; 2005. The importance of soil organic matter: Key to drought-resistant soil and  
407 sustained food production. Food Agriculture Organization of the United Nations, Rome, **2005**, p. 95.

408 Bouyoucos, G.J. Hydrometer method improved for making particle size analyses of soils. *Agron. J.*  
409 **1962**, 54 (5), 464-465.

410

411 Bremner, G.J.; Mulvaney, C.S. Nitrogen total. In: Page, A.L.; Miller, R.H.; Keenry, R.R.  
412 (Eds.), *Methods of soil analysis, part 2. Seconded. American Society of Agronomy, Madison,*  
413 *WI, 1982*, pp.595-624.

414

415 Brown, J.; Angerer, J.; Salley, S.W.; Blaisdell, R.; Stuth, J.W. Improving estimates of rangeland carbon  
416 sequestration potential in the US Southwest. *Rangeland Ecol. Manage.* **2010**, 63 (1), 147-154.

417

418 Cecagno1, D.; Veloso Gomes, M.; Andrade Costa, S.E.V.G.; Martins, A.P.; Oliveira Denardin, L.G.;  
419 Bayer, C.; Anghinoni, I.; Faccio Carvalho, P.C. Soil organic carbon in an integrated crop-livestock  
420 system under different grazing intensities. *Rev. Bras. Cienc. Agrar.* **2018**, 13, 1-7.

421

422 Cao, J.; Wang, X.; Sun, X.; Zhang, L.; Tian, Y. Effects of grazing intensity on soil labile organic carbon  
423 fractions in a desert steppe area in Inner Mongolia. *SpringerPlus* **2013**, 2 (1), 1-8.

424

425 Chen, Y.; Li, Y.; Zhao, X.; Awada, T.; Shang, W.; Han, J. Effects of Grazing Exclusion on Soil  
426 Properties and on Ecosystem Carbon and Nitrogen Storage in a Sandy Rangeland of Inner Mongolia,  
427 Northern China. *Environ. Manage.* **2012**, 50, 622-632.

428

429 Da Silva, F.D.; Amado, T.J.C.; Ferreira, A.O.; Assmann, J.M.; Anghinoni, I.; De Faccio Carvalho, P.C.  
430 Soil carbon indices as affected by 10 years of integrated crop-livestock production with different  
431 pasture grazing intensities in Southern Brazil. *Agric. Ecosyst. Environ.* **2014**, 190, 60–69.

432

433 Derner, J.D.; Boutton, T.W.; Briske, D.D. 2006. Grazing and ecosystem carbon storage in the North  
434 American Great Plains. *Plant Soil*, **2006**, 280 (1-2), 77-90.

435

436 Derner, J.; Schuman, G. Carbon sequestration and rangelands: a synthesis of land management and  
437 precipitation effects. *J. Soil Water Conserv.* **2007**, 62 (2), 77-85.

438

439 Geng, Y.Q.; Yu, X.X.; Yue, Y.J.; Li, J.H.; Zhang, G.Z. Active organic carbon pool of coniferous and  
440 broad-leaved forest soils in the mountainous areas of Beijing. *For. Stud. China* **2009**, 11 (4), 1-6.

441 Golluscio, R.; Austin, A.; Martinez, G.; Gonzalez, P.M.; Sala, O.; Jackson, R. Sheep grazing decreases  
442 carbon and nitrogen pools in the Patagonian Steppe: combination of direct and indirect effects. *Ecosyst.*  
443 **2009**, 12, 686-697.

444 Holland, E.A.; Parton, W.J.; Detling, J.K.; Coppock, D.L. Physiological responses of plant populations  
445 to herbivory and their consequences for ecosystem nutrient flow. *Am. Nat.* **1992**, 140 (4), 685-706.

446

447 Kamoni, P.; Gicheru, P.; Wokabi, S.; Easter, M.; Milne, E.; Coleman, K.; Falloon, P.; Paustian, K.;  
448 Killian, K.; Kihanda, F. Evaluation of two soil carbon models using two Kenyan long term experimental  
449 datasets. *Agric. Ecosyst. Environ.* **2007**, 122, 95-104.

450

451 Kirk, P.L. Kjeldahl method for total nitrogen. *Anal. Chem.* **1950**, 22 (2), 354-358.

452

453 Lal, R.. Carbon sequestration. *Philos. Trans. Roy. Soc.* **2008**, 363, 815-830.

454

455 Mcsherry, M.E.; Ritchie, M.E. Effects of grazing on grassland soil carbon: a global review. *Global*  
456 *change biol.* **2013**, 19 (5), 1347-1357.

457

458 Medina-Roldán, E., Paz-Ferreiro, J.; Bardgett, R.D. Grazing exclusion affects soil and plant  
459 communities, but has no impact on soil carbon storage in an upland grassland. *Agric. Ecosyst. Environ.*  
460 **2012**, 149, 118-123.

461

462 Mekuria, W.; Veldkamp, E.; Haile, M.; Nyssen, J.; Muys, B.; Gebrehiwot, K. Effectiveness of  
463 exclusions to restore degraded soils as a result of overgrazing in Tigray, Ethiopia. *J. Arid Environ.*  
464 **2007**, 69 (2), 270-284.

465

466 Muñoz-Rojas, M.; Doro, L.; Ledda, L.; Francaviglia, R. Application of CarboSOIL model to predict  
467 the effects of climate change on soil organic carbon stocks in agro-silvo-pastoral Mediterranean  
468 management systems. *Agric. Ecosyst. Environ.* **2015**, 202, 8-16.

469

470 Orgill, S.E.; Condon, J.R.; Conyers, M.K.; Morris, S.G.; Alcock, D.J.; Murphy, B.W.; Greene, R.S.B.  
471 Removing grazing pressure from a native pasture decreases soil organic carbon in southern new south  
472 wales, australia. *Land Degrad Dev.* **2016**, 29 (2), 274-283.

473

474 Papanastasis, V.P.; Bautista, S.; Chouvardas, D.; Mantzanas, K.; Papadimitriou, M.; Mayor, A.G.;  
475 Koukioumi, P.; Papaioannou, A.; Vallejo, R.V. Comparative assessment of goods and services  
476 provided by grazing regulation and reforestation in degraded mediterranean rangelands. *Land Degrad.*  
477 *Dev.* **2015**, 28 (4), 1178-1187.

478

479 Parton, W.J.; Schimel, D.S.; Cole, C.; Ojima, D. Analysis of factors controlling soil organic matter  
480 levels in Great Plains grasslands. *Soil Sci. Soc. Am. J.* **1987**, 51 (5), 1173-1179.

481

482 Parton, W.J.; Stewart, J.W.; Cole, C.V. Dynamics of C, N, P and S in grassland soils: a model.  
483 *Biogeochem.* **1988**, 5 (1), 109-131.

484 Pineiro, G., Paruelo, J.M.; Oesterheld, M.; Jobbágy, E.G. Pathways of grazing effects on soil organic  
485 carbon and nitrogen. *Rangeland Ecol. Manage.* **2010**, 63 (1), 109-119.

486

487 Potter, K.; Daniel, J.; Altom, W.; Torbert, H. Stocking rate effect on soil carbon and nitrogen in  
488 degraded soils. *J. soil water conserv.* **2001**, 56, 233-236.

489

490 Reeder, J.; Schuman, G. Influence of livestock grazing on C sequestration in semi-arid mixed-grass and  
491 short-grass rangelands. *Environ. Pollut.* **2002**, 116 (3), 457-463.

492

493 Ritchie, M.E. Plant compensation to grazing and soil carbon dynamics in a tropical grassland. *PeerJ*  
494 **2014**, 2:e233, 1-27.

495

496 Sepe, L.; Salis, M.; Francaviglia, R.; Fedrizzi, M.; Carroni, A.M.; Sabia, E.; Bruno, A.; Rufrano, D.;  
497 Ruda, P.; Dell'Abate, M.T.; Alianello, A.; Velocchia, M.; Masetti, O.; Renzi, G.; Fanigliulo, R.; Pagano,  
498 M.; Sperandio, G.; Guerrieri, M.; Puri, D.; Claps, S. Environmental effectiveness of the cross  
499 compliance Standard 4.6 'Minimum livestock stocking rates and/or appropriate regimens'. *Ital. J.*  
500 *Agron.* **2015**, 10 (1), 1-9.

501

502 Sheng, H.; Zhou, P.; Zhang, Y.; Kuzyakov, Y.; Zhou, Q.; Ge, T.; Wang, C. Loss of labile organic  
503 carbon from subsoil due to land-use changes in subtropical China. *Soil Biol. Biochem.* **2015**, 88, 148-  
504 157.

505

506 Shifang, P.; Hua, F.; Changgui, W. Changes in properties and vegetation following exclusion and  
507 grazing in degraded Alxa desert steppe of Inner Mongolia, China. *Agric. Ecosyst. Environ.* **2008**, 124,  
508 33-39.

509

510 Smith, P.; Fang, C.; Dawson, J.; Moncreiff, J. Impact of global warming on soil organic carbon. *Adv*  
511 *Agron.* **2008**, 97, 1-43.

512

513 Soon, Y.; Arshad, M.; Haq, A.; Lupwayi, N. The influence of 12 years of tillage and crop rotation on  
514 total and labile organic carbon in a sandy loam soil. *Soil Tillage Res.* **2007**, 95 (1), 38-46.

515

516 Tornquist, C.G.; Mielniczuk, J.; Cerri, C.E.P. Modeling soil organic carbon dynamics in Oxisols of  
517 Ibirubá (Brazil) with the Century Model. *Soil Tillage Res.* **2009**, 105 (1), 33-43.

518

519 Vanaee, F.; Karami, P.; Joneydi Jafari, H.; Nabialahi, K. Simulation of soil organic carbon dynamic in  
520 meadow ecosystems under different management practices using CENTURY model. *J. Rangeland*  
521 **2017**, 10 (4), 439-449.

522

523 Walkley, A.; Black, I.A. An examination of the Degtjareff method for determining soil organic matter,  
524 and a proposed modification of the chromic acid titration method. *Soil sci.* **1934**, 37, 29-38.

525

526 Wang, M.J.; Wan, X.R.; Zhong, W.Q. The interaction between the vegetarian and plant. *Chin. J. Ecol.*  
527 **2001**, 20 (5), 39-43.

528 Wang, Y.; Zhou, G.; Jia, B. Modeling SOC and NPP responses of meadow steppe to different grazing  
529 intensities in Northeast China. *Ecol. Modell.* **2008**, 217, 72-78.  
530

531 Waters, C.M.; Orgill, S.E.; Melville, G.J.; Toole, I.D.; Smith, W.J. Management of grazing intensity in  
532 the semi-arid rangelands of southern australia – effects on soil and biodiversity. *Land Degrad. Dev.*  
533 **2016**, 28 (4), 1363-1375.  
534

535 Wilson, C.; Papanicolaou, A.; Abaci, O. SOM dynamics and erosion in an agricultural test field of the  
536 Clear Creek, IA watershed. *Hydrol. Earth Syst. Sci. Discuss.* **2009**, 6 (2), 1581-1619.  
537

538 Wright, I.J.; Reich, P.B.; Westoby, M.; Ackerly, D.D.; Baruch, Z.; Bongers, F.; Cavender Bares, J.;  
539 Chapin, T.; Cornelissen, J.H.C.; Diemer, M.; Flexas, J.; Garnier, E.; Groom, P.K.; Gulias, J.; Hikosaka,  
540 K.; Lamont, B.B.; Lee, T.; Lee, W.; Lusk, C.; Midgley, J.J.; Navas, M.L.; Niinemets, U.; Oleksyn, J.;  
541 Osada, N.; Poorter, H.; Poot, P.; Prior, L.; Pyankov, V.I.; Roumet, C.; Thomas, S.C.; Tjoelker, M.G.;  
542 Veneklaas, E.J.; Villar, R. The worldwide leaf economics spectrum. *Nat.* **2004**, 428, 821–827.  
543

544 Zhang, Y.; Tang, Y.; Jiang, J.; Yang, Y. Characterizing the dynamics of soil organic carbon in  
545 grasslands on the Qinghai-Tibetan Plateau. *Sci. China Ser D: Earth Sci.* **2007**, 50, 113-120.  
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570 **Table 1.** Soil and vegetation characteristics and climatic parameters used as inputs for the  
 571 CENTURY model.

Parameter	Value
Location	52°N , 29°E
Precipitation (mm)	388.44
Temperature (C°)	13.4
Sand (%)	12.72
Silt (%)	53.38
Clay (%)	33.88
Bulk density (Mg m <sup>-3</sup> )	1.265
pH	7.93
Field capacity (FC v/v%) <sup>a</sup>	0.383
Wilting point (WP v/v%) <sup>a</sup>	0.159
Initial total SOC (g m <sup>-2</sup> ) <sup>b</sup>	3228.53
Soil C/N	11.6
Litter carbon (g m <sup>-2</sup> )	29.469
Aboveground biomass carbon (g m <sup>-2</sup> )	38.947
Belowground biomass carbon (g m <sup>-2</sup> )	56.401
Aboveground biomass nitrogen (g m <sup>-2</sup> )	1.121
Belowground biomass nitrogen (g m <sup>-2</sup> )	1.646
Litter C/N	34.72

572 <sup>a</sup> data provided by the Water Engineering Department of Agricultural College; <sup>b</sup> Average of 30 samples in 1987.

573

574 **Table 2.** Quantitative statistical analysis between measured and simulated SOC stocks.

R	R <sup>2</sup>	RMSE <sub>0.05</sub>	RMSE	EF
0.93	0.86	7.07	2.35	83%

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579 **Table 3.** Long-term changes of SOC stocks under the different grazing intensity scenarios (2015-  
 580 2100).

Scenario	SOC stock in 2014 (g m <sup>-2</sup> )	SOC stock in 2100 (g m <sup>-2</sup> )	Changes compared to 2014 (%) <sup>*</sup>
No grazing	3356.1	3446.7 <sup>a</sup>	+2.7
LG	3356.1	3412.5 <sup>a</sup>	+1.7
MG	3356.1	2570.4 <sup>b</sup>	-23.4
HG	3356.1	2529.8 <sup>b</sup>	-24.6

581 Values followed by the same letter are not significantly different among grazing scenarios at p<0.01 (LSD tests);  
 582 SOC soil organic carbon; LG light grazing; MG moderate grazing; HG heavy grazing.

583 <sup>\*</sup> Changes compared to 2014 (%) calculated as [100 × (SOC stock in 2100- SOC stock in 2014)/ SOC stock in 2014]

584

585 **Table 4.** Long-term changes of SLOC stocks under the different grazing intensity scenarios  
586 (2015-2100).

Scenario	SLOC stock (g m <sup>-2</sup> )	Changes compared to No grazing (%)
No grazing	20.87a	0
LG	15.43b	-26.06
MG	8.43c	-59.6
HG	6.27d	-69.95

587 Values followed by the same letter are not significantly different among grazing scenarios at p<0.01; SLOC soil  
588 labile organic carbon; LG light grazing; MG moderate grazing; HG heavy grazing.

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# Figures

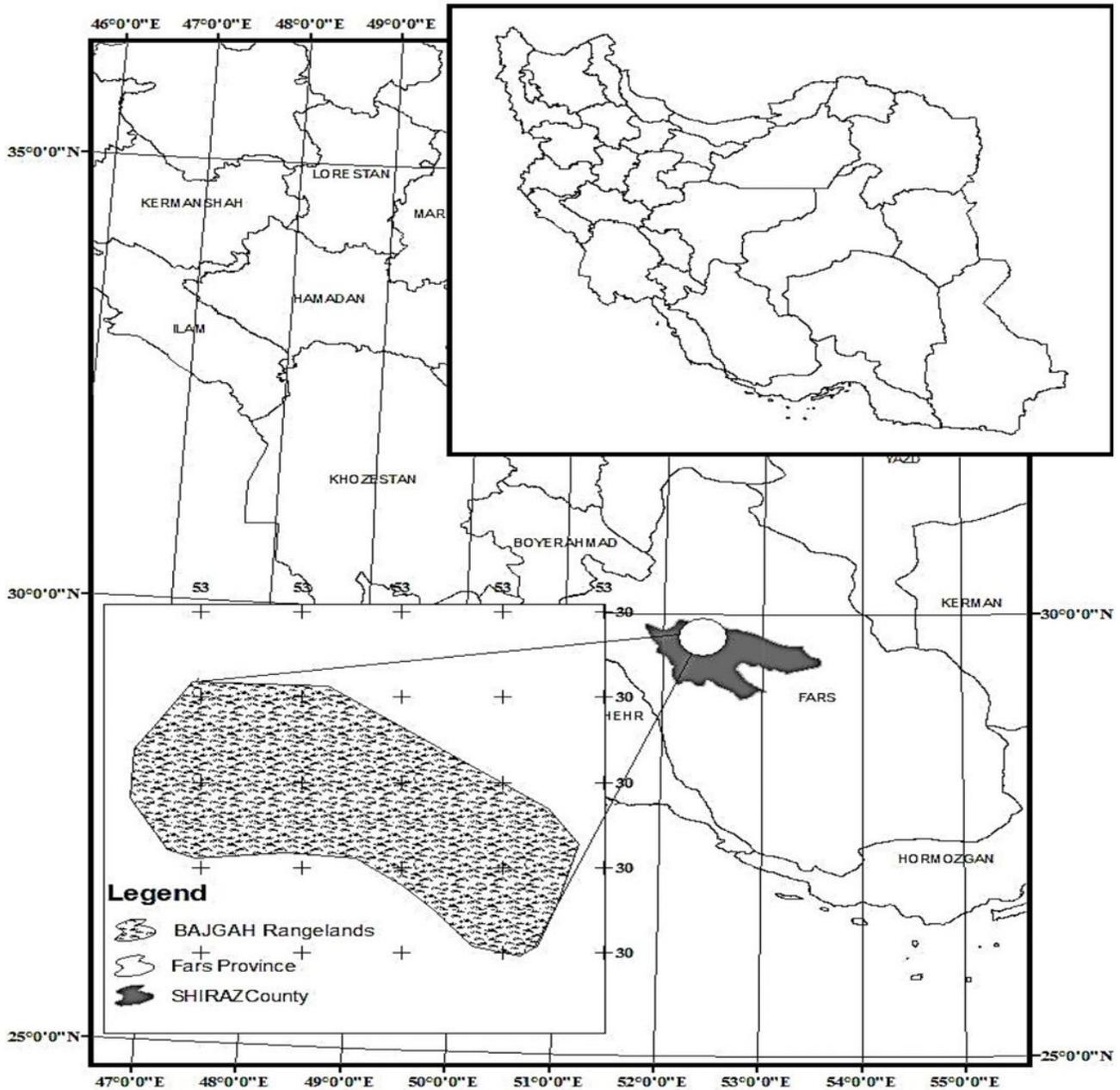


Figure 1

Bajgah region in the South of Iran, Shiraz County, province of Fars. Note: The designations employed and the presentation of the material on this map do not imply the expression of any opinion whatsoever on the part of Research Square concerning the legal status of any country, territory, city or area or of its authorities, or concerning the delimitation of its frontiers or boundaries. This map has been provided by the authors.

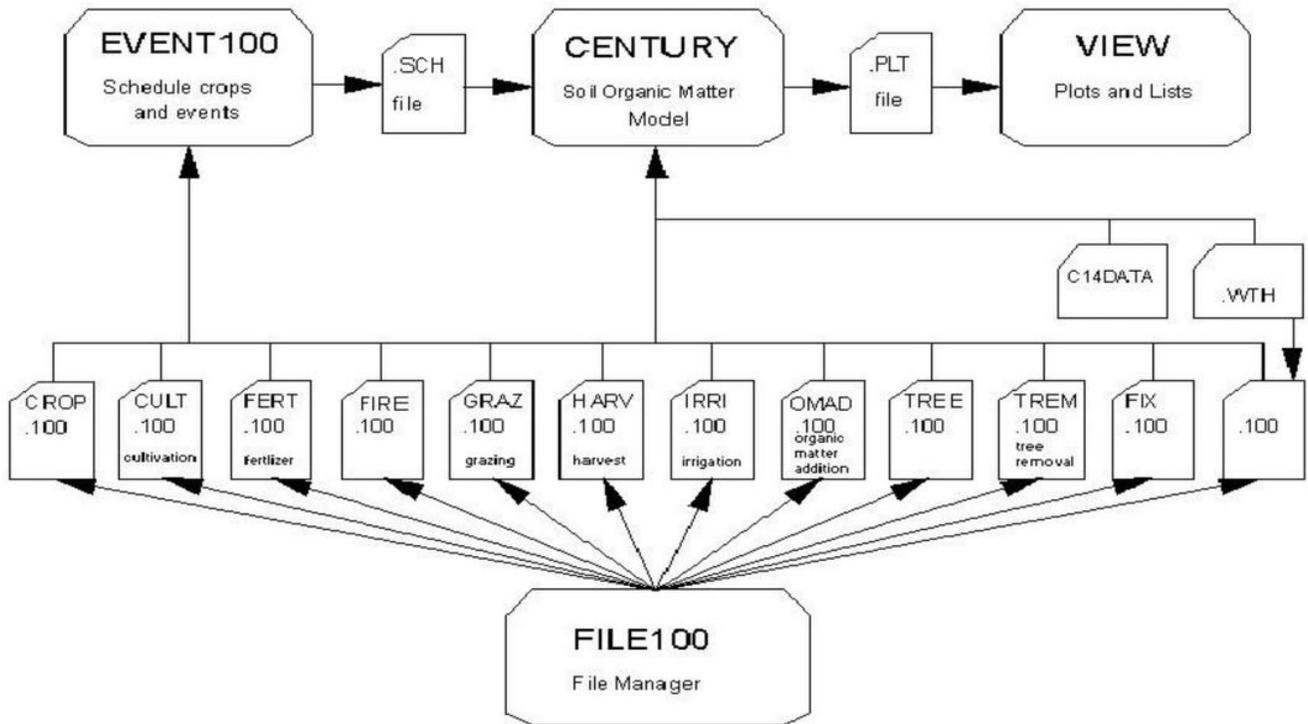


Figure 2

The Century model flowchart.

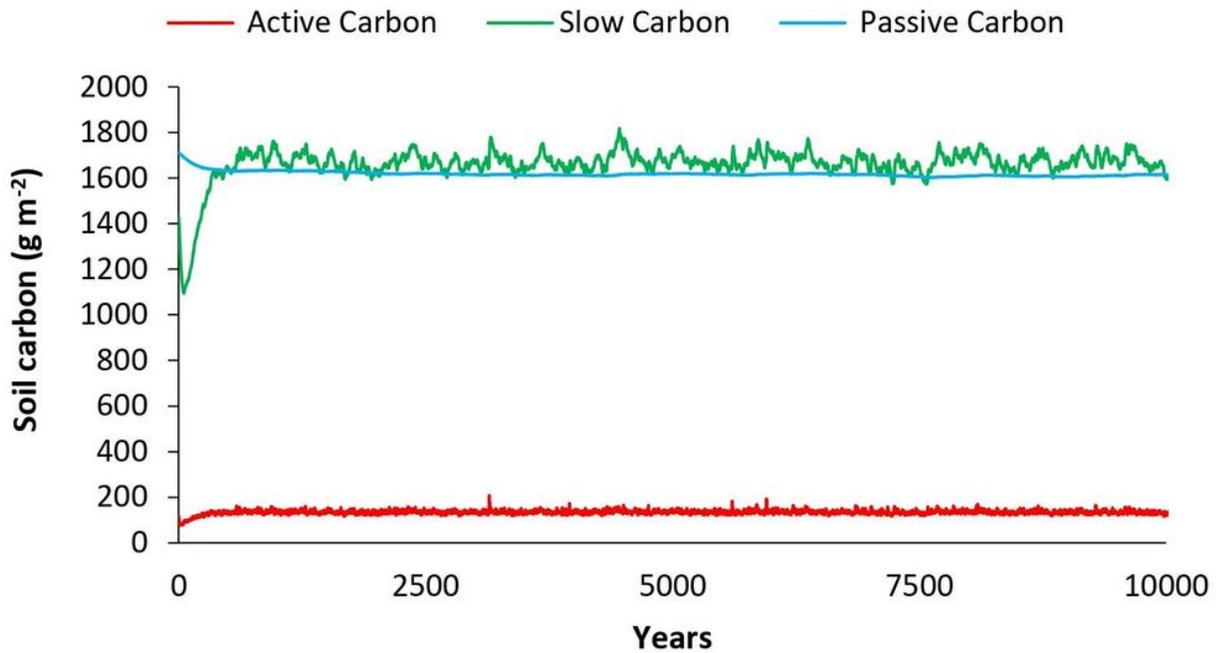
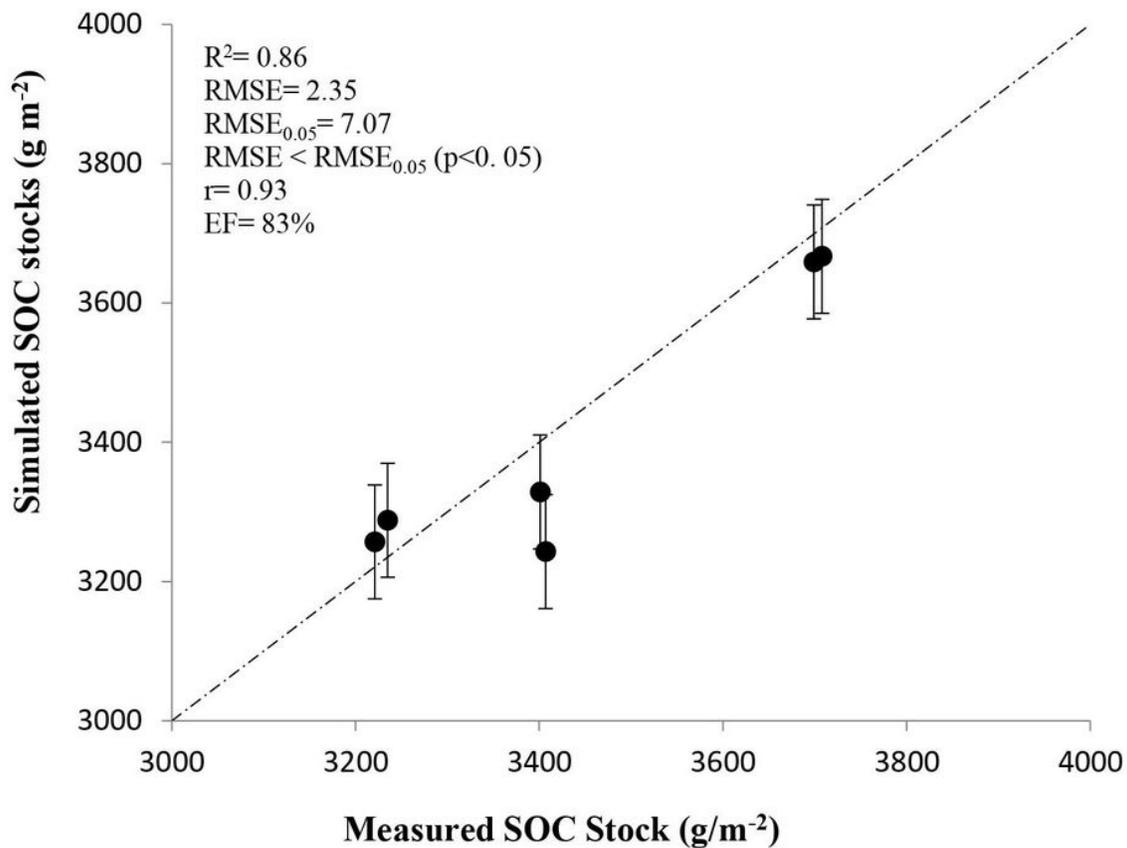


Figure 3

Change of carbon pools during the equilibrium state process.



**Figure 4**

Measured and simulated SOC Stock in compared with 1:1 line. Vertical bars indicate the difference between measured and simulated values. EF compares simulations or predictions and observations on an average level, and ranges from - 1 to 1, with best performance at EF = 1.

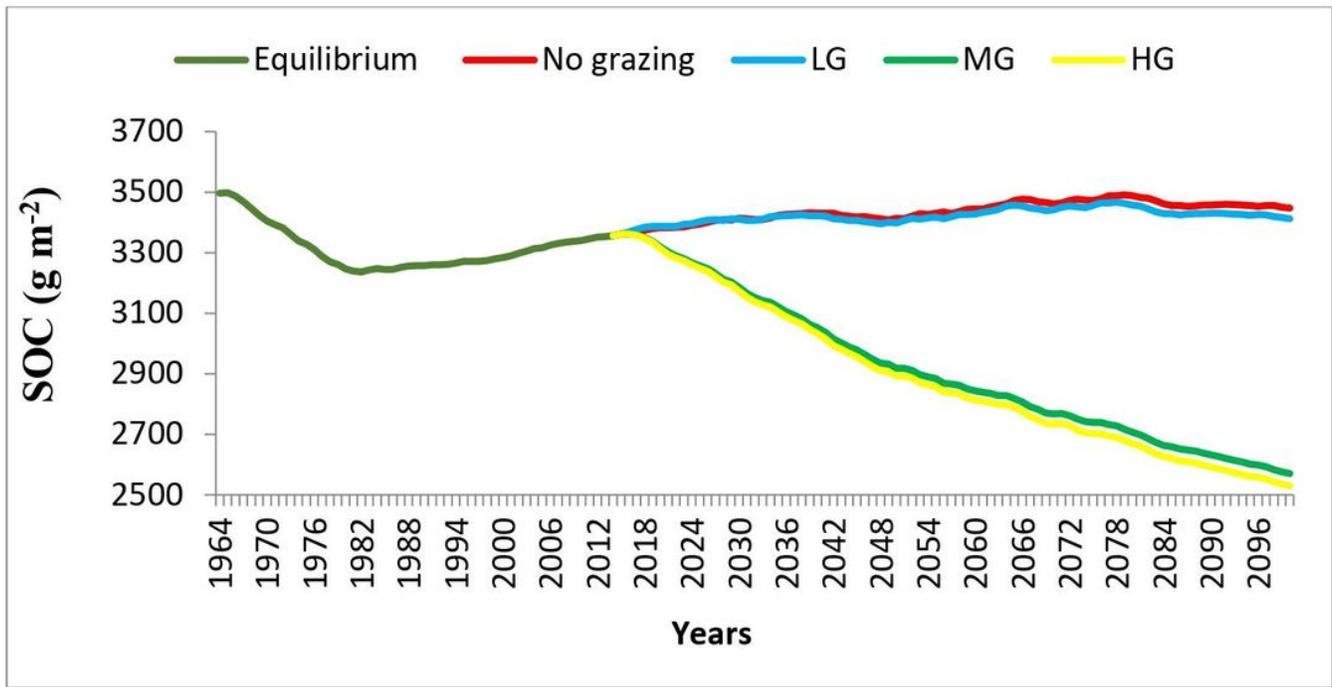
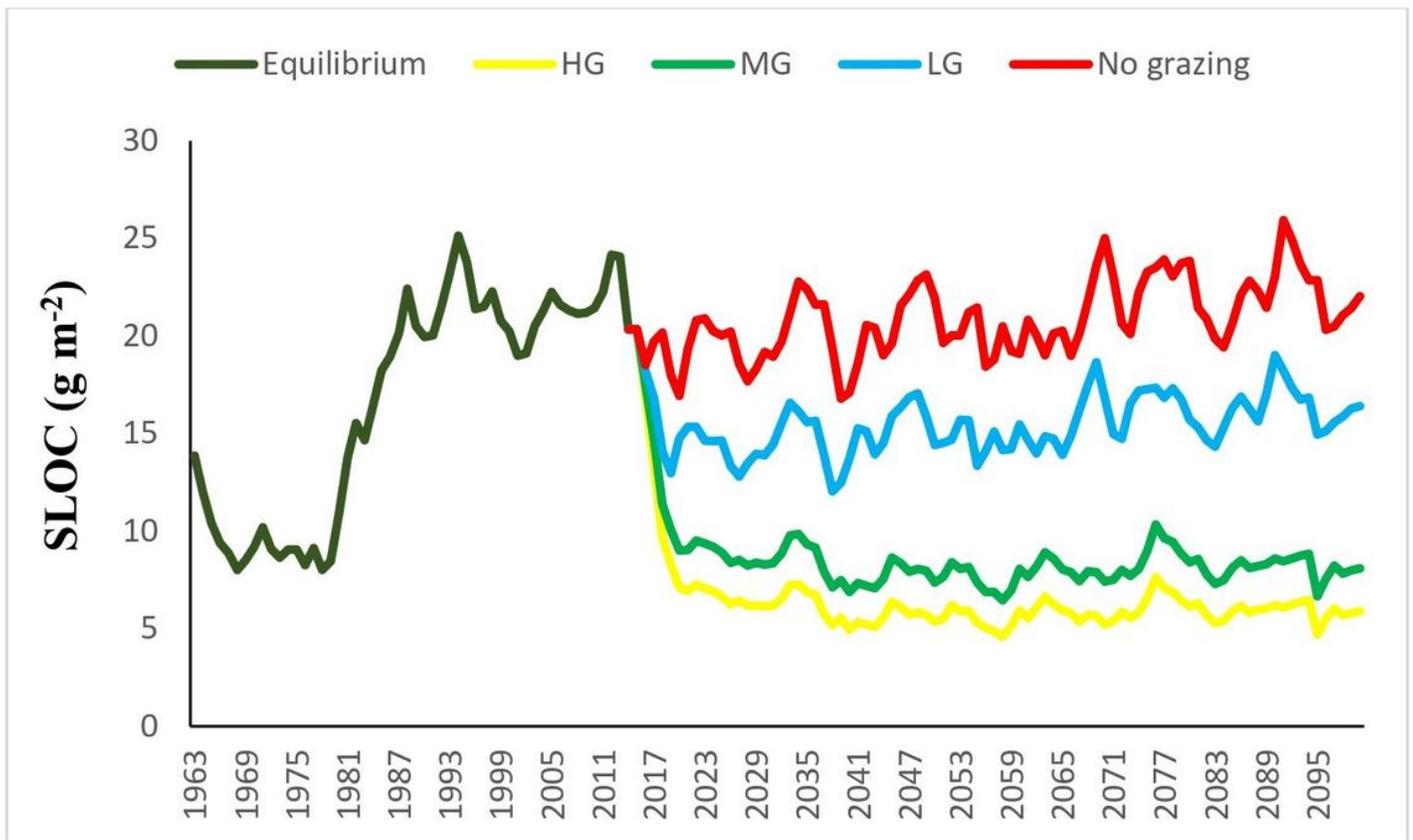


Figure 5

Simulated soil organic carbon (SOC) dynamics under different grazing intensity scenarios (No grazing, LG = Light Grazing, MG = Moderate Grazing, HG =Heavy Grazing).



## Figure 6

Simulated soil labile organic carbon (SLOC) dynamics under different grazing intensity scenarios (No grazing, LG = Light Grazing, MG = Moderate Grazing, HG =Heavy Grazing).