

Assessment of Alluvial Aquifer Intrinsic Vulnerability by a Generic Drastic Model; A Discussion on Data Adequacy and Pragmatic Results

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3 Assessment of alluvial aquifer intrinsic vulnerability by
4 a generic DRASTIC model; a discussion on data
5 adequacy and pragmatic results

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13

14

15 **Abstract**

16 DRASTIC is a model that is commonly used to assess vulnerability to groundwater
17 contamination at the landscape scale. When sparse data are available to populate the layers of
18 the model, it can be difficult to ascertain the true usefulness of the model produced map. In
19 this research an alluvial aquifer, the Sahneh aquifer in Kermanshah province of western Iran,
20 was mapped using the generic DRASTIC model. The data available for populating the model
21 layers were generally sparse. The model was validated using a nitrate concentration map
22 constructed from well water measurements within the DRASTIC map area. A Receiver
23 Operating Curve (ROC) analysis was conducted by placing 500 random points in the
24 DRASTIC generated map compared to the nitrate concentration map. The area under the curve
25 was compared and yielded a value of 0.72 or 72% concordance, which means it has good
26 validity. This investigation demonstrates that a generic DRASTIC model can yield acceptable
27 results without modification or increasing its complexity. If the ROC analysis had yielded a
28 value <50%, then the DRASTIC would have been considered to not be useful. A common
29 mistake in the use of DRASTIC is to modify the method to greatly increase its complexity,
30 which may actually decrease, not increase the resultant model usefulness.

31

32 **Keywords** DRASTIC · Groundwater modelling · Groundwater contamination · Receiver
33 Operating Curve (ROC) · Groundwater modelling verification

34

35 **Introduction**

36 The DRASTIC model has been evaluated to assess both its advantages and disadvantages for
37 use in groundwater pollution management (Aller et al. 1987). Despite the criticisms of the
38 model by various researchers, it is still one of the most popular and widely used methods for
39 assessing aquifer contamination potential (Rundquist et al. 1991; Al-Adamat et al. 2003;
40 Babiker et al. 2005; Baalousha 2006; Moghaddam and Fijani 2008; Hamza and Added 2009;
41 Kura et al. 2015; Vaezihir and Tabarmayeh 2015; Muhammetoglu et al. 2019; Ghosh et al.
42 2020; Torkashvand et al. 2020; Mallik et al. 2021; Nasri et al. 2021). The simplicity and
43 usability of the DRASTIC model allows it to assess vulnerability of an aquifer to pollution
44 even where sparse data are available on aquifer hydrogeological characteristics. From a
45 terminological point of view, the choice of the word DRASTIC for the seven acronyms that
46 assess intrinsic vulnerability is interesting and thought-provoking, because the word itself has
47 a strong meaning and indicates a situation that should be considered (Dictionary, C 2008).

48 Despite the apparent simplicity of the seven factors in the DRASTIC model, development
49 and presentation of thematic and practical maps with accurate data (or at least sufficient) is
50 not so easy. While DRASTIC is a simple and restrained method that gives results in any case,
51 the adequacy of the data and the realism of the results can make the output of this model
52 sufficiently reliable for use by local planners and decision makers. However, the use of the
53 model results must be constrained within the famous statement that “All models are wrong,
54 but some are useful; the practical question is how wrong they have to be to not be useful.”
55 (Box and Draper 1987). The statement by Box and Draper (1987) applies not only to the
56 accuracy of the model, which is commonly controlled by the input parameters (data), but also
57 applies to how the model is applied with consideration of the error framework.

58 The simplicity of the DRASTIC model is the evaluation methodology and use of specific
59 weights and rates as suggested by Aller et al. (1987). Advancements in software development
60 now make it is possible to easily overlap the seven layers within a GIS framework and obtain
61 a digital output that can be categorized in different ways. In this regard, there are numerous
62 articles that have provided various classifications for the final model-generated map, from
63 non-vulnerability (Metni et al. 2004; Herlinger and Viero 2007; Prasad et al. 2011; Chitsazan
64 and Akhtari 2009; Yin et al. 2012) to extreme vulnerability (Rozkowski 2010; Herlinger and
65 Viero 2007). Different input parameters used to create the final map can be varied depending
66 on the needs of the different users. The sum of the seven DRASTIC parameters yields a

67 generic DRASTIC index with the highest numbers corresponding to the highest vulnerability
68 to contamination (Corniello et al. 1997; Hua et al. 2011).

69 Hamza et al. (2015) concluded, based on a compilation of 30 studies that computed the
70 DRASTIC index and vulnerability classifications, that the generic index values range between
71 23 and 230. They suggested that to calculate the percentages within the vulnerability classes,
72 a simple formula could be used (Equation 1).

$$73 \qquad \qquad \qquad (i/230) \times 100 \qquad \qquad \qquad (1)$$

74
75 where i is the index range. The classification of vulnerability reported in the various
76 investigations included no risk, very low risk, low risk, medium risk, high risk, very high risk,
77 and extreme risk. Only 7 of the 30 studies showed any area within the no risk classification
78 and only 2 of the 30 studies showed any area within the extreme risk classification (Hamza et
79 al 2015).

80 Some researchers believe that "no risk" as one of the final map categories is basically
81 incorrect because any aquifer can be inherently vulnerable to pollution, even if the
82 vulnerability was very low (Hamza et al. 2015). The main flaw within the DRASTIC method
83 lies in the adequacy of the data, i.e., how and from what sources are these data obtained or
84 extracted? Can lack of data in one or more of the primary DRASTIC parameters be ignored
85 in the absence of better quality data in some other parameters? To what extent can
86 modification of the model increase the public acceptance of the model results? Is it useful to
87 add additional layers to this model? How can data differentiation increase the realism of the
88 results? What can be done in the absence of data to validate the model? This goes back to the
89 fundamental question posed by Box and Draper (1987) "All models are wrong, but some are
90 useful; *the practical question is how wrong do they have to be to not be useful.*"

91 The answer to all of these questions lies in the ability of the researcher(s) to extract or
92 provide adequate data and to perform sufficient analyses on the model output results to provide
93 a reasonable degree of assurance that the model is useful. To produce an accurate vulnerability
94 assessment, the researcher must be familiar with the basic principles of hydrogeology. It is
95 very important to understand the nature of various units constituting and affecting the aquifer
96 from the unsaturated zone to the base of the aquifer. In addition, there are fundamental
97 relationships between some of the seven DRASTIC parameters that need to be considered,
98 such as the soil media and the aquifer media may be the same when the water table occurs at

99 or near land surface. There is also a close relationship between impacts occurring within the
100 vadose zone and the soil media. For example, it is possible to provide a single layer based a
101 few water drilling logs for the impact of the aquifer media and the vadose zone. If the nature
102 of the upper part of aquifer system is over-simplified by ignoring the occurrence of a clay or
103 gravel unit, the results of the DRASTIC assessment could be inaccurate. This issue is quite
104 extreme in karst environments, where flow conduits could provide a bypass through which
105 contaminants could directly enter deep into an aquifer (Taheri, et al. 2017).

106 Application of the DRASTIC model could be used in basic educational for students of
107 hydrogeology, groundwater and related disciplines (Rich and Onasch 1997). Because
108 preparation of the database for the seven factors involves all the principles and methods of
109 theory in the field of groundwater. Presenting an acceptable result and preparing seven
110 parameters with sufficiently accurate data can indicate the adequacy of knowledge on the
111 aquifer being investigated and the perspective for its optimal management. In operational and
112 executive terms, if an acceptable output can be obtained with a DRASTIC model in an aquifer,
113 it can be hoped that at least the groundwater management infrastructure is available in that
114 area. Because a first step in managing an aquifer may be to have a map of the vulnerability of
115 that aquifer (Merchant 1994). Considerable research has been conducted on different technical
116 aspects and applications of the DRASTIC model (Merchant 1995; Shirazi et al. 2012; Hamza
117 et al. 2014; Khosravi et al. 2018; Kumar et al. 2019; Barbulescu 2020).

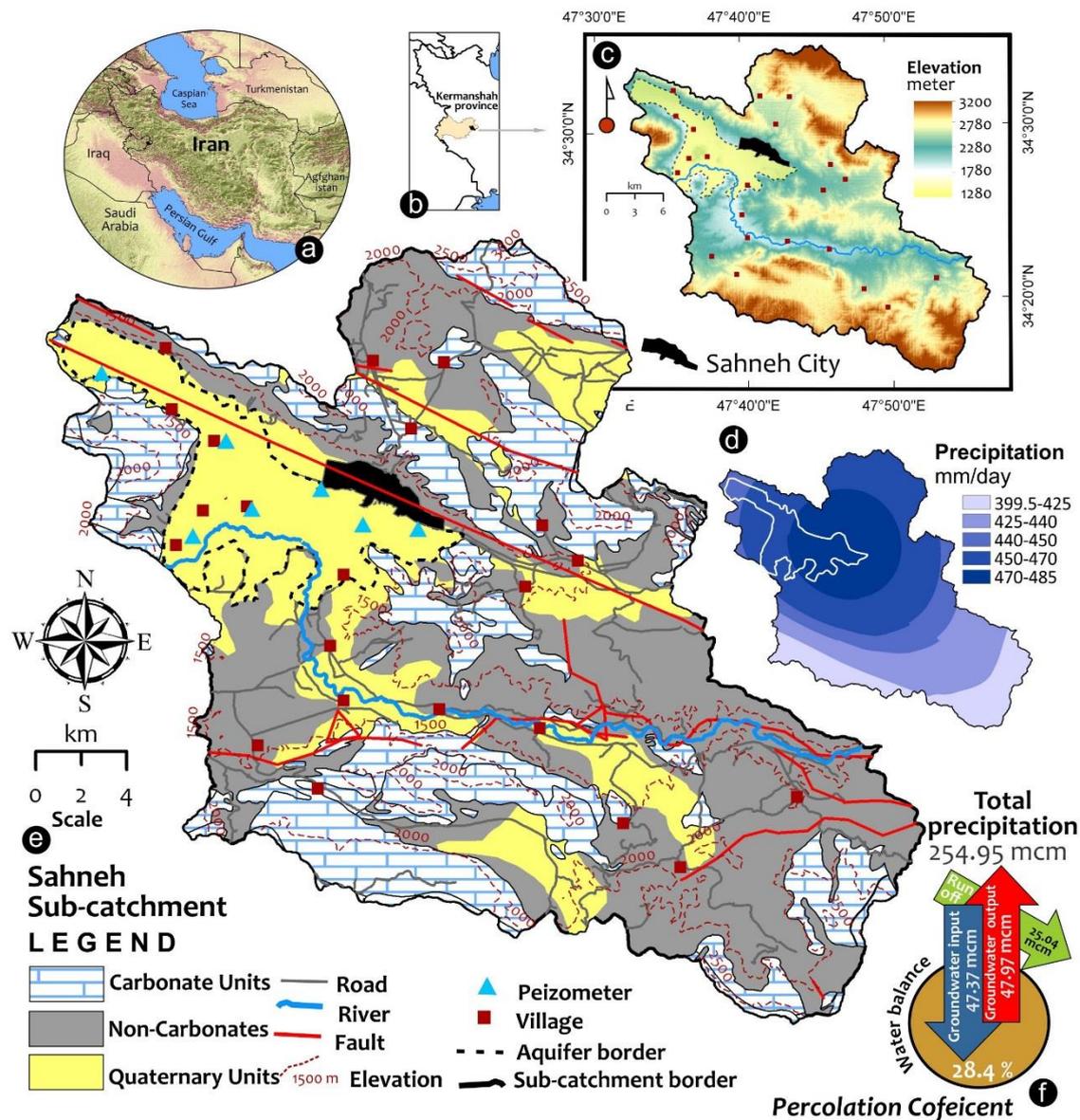
118 In this research, the Sahneh aquifer in the Kermanshah province of Iran was evaluated
119 using the method of Aller et al. (1987) based on their proposed weights of the DRASTIC
120 parameters. The results were evaluated by developing a critical assessment of the advantages
121 and disadvantages of the output and methodology. This evaluation was achieved by assessing
122 the accuracy of the available data for each DRASTIC parameter. The results of the DRASTIC
123 model output were then compared with an expert-centered scoring method applied to the same
124 land area. This comparison was used to evaluate the real variability in the DRASTIC model
125 outputs to illustrate the potential errors with the assessment and how the output should be used
126 within the context of groundwater management. A test of the model validity was used to
127 ascertain its usefulness by using nitrate data with the context of a Receiver Operating Curve
128 analysis (ROC). This research provides a method to review the generic DRASTIC method and
129 validation of its outputs to create a practical guide to water managers on the usefulness of the
130 model and what restrictions should be placed on the output use.

131 **Materials and methods**

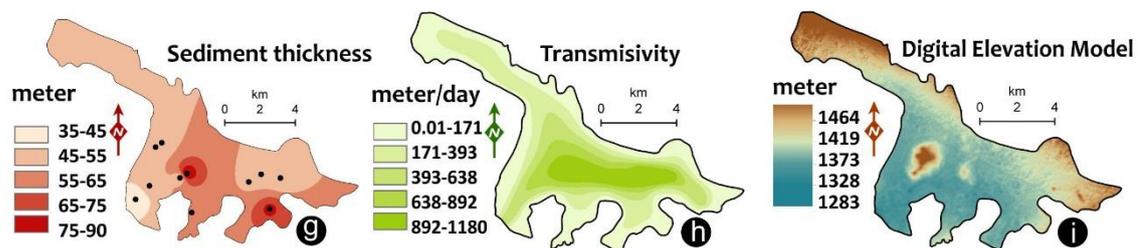
132 **Study area**

133 The study was conducted on the Sahneh aquifer that is located mainly in the Sahneh Sub-
134 catchment in the Kermanshah province of western Iran (Fig. 1a). The Sahneh aquifer covers
135 a relatively small area that is located in a cold and dry region (Fig. 1b,c). This aquifer is located
136 in the structural zone of Sanandaj-Sirjan and thrust fault zone of Zagros. The behavior of the
137 aquifer geological units is a function of the structural conditions of these two zones (Fig.1e).
138 The geological units of the region can be categorized into karst and non-karst units. Karstic
139 units generally control the altitude of the region. In addition, the carbonate units have the most
140 impact on the aquifer system recharge and discharge, although no detailed investigations of
141 those factors has been made. Non- karstic units have little effect on hydraulic exchanges with
142 the aquifer system based on aquifer hydraulic properties and the area of their occurrence. In
143 this study, the alluvial unit is considered separately from the karst units and its hydraulic
144 exchanges are ignored. This neglect of karst exchanges is common even in the study of the
145 water balance by water management organizations in the region. To eliminate the karst effect
146 in the water balance, the rate of precipitation recharge in the mountains has been added to the
147 water balance equations. Fig. 1f summarizes the groundwater aquifer balance of the scene
148 aquifer.

149



Sahneh Aquifer



150

151 **Fig. 1.** The location of studied area in Iran (a), Kermanshah province (b), and Sahneh Sub-
 152 catchment (c), isohyetal map (d), simplified geological map and distribution of carbonate, non-
 153 carbonate formations and porous media (plain) and alluvial aquifer boundary (e), water budget
 154 summary(f), sediment thickness map (g), transmissivity map (h) and digital elevation model
 155 (i).

156 Hydrogeologic assessment of the Sahneh aquifer was performed using data extracted from
 157 the seven monitoring wells. Although this number is insufficient, however, no action has been
 158 taken to correct it. Few exploratory studies of the aquifer have been conducted. The primary
 159 source of knowledge on the aquifer and its subsurface media comes from the logs of private
 160 wells used for water supply. Water levels compiled into hydrographs indicate that the
 161 groundwater balance is negative, meaning that the discharge rate is higher than the aquifer
 162 recharge rate (declining water levels) (Fig. 1f). According to the results of the water balance
 163 analysis, the recharge rates in the alluvium and higher altitudes are 3.36 million cubic
 164 meters/year (mcm) and 69.05 mcm equivalent to 4.64% and 95.36% of the total recharge from
 165 precipitation to the sub-catchment area (water budget area), respectively. With a infiltration
 166 rate of 28.4%, this means that from the total volume of rain (Fig. 1d) that falls on the alluvium,
 167 only 28.4% of this amount reaches the water table and is practically equivalent to the volume
 168 of the aquifer recharge annually. It is known that equating the infiltration rate to the recharge
 169 rate in an arid region likely overestimates recharge or underestimates infiltration. In many
 170 alluvial aquifers of Iran, the balance is done in this way, except in limited cases where detailed
 171 or case studies have been made, the infiltration rate and net recharge of the aquifer are quite
 172 approximate. Fig. 2 presents the conceptual model of the aquifer and the effective components
 173 in evaluating the DRASTIC model.

174 **DRASTIC method**

175 This research examines the DRASTIC method for the intrinsic evaluation of in an alluvial
 176 aquifer to assess groundwater pollution potential. The DRASTIC index is obtained based on
 177 the method of Aller et al. (1987) using the sum of the seven factors after multiplication of the
 178 weighting factors (Equations 2 and 3).

$$179 \quad I_{DRASTIC} = \sum_{i=1}^7 W_i \times R_i \quad (2)$$

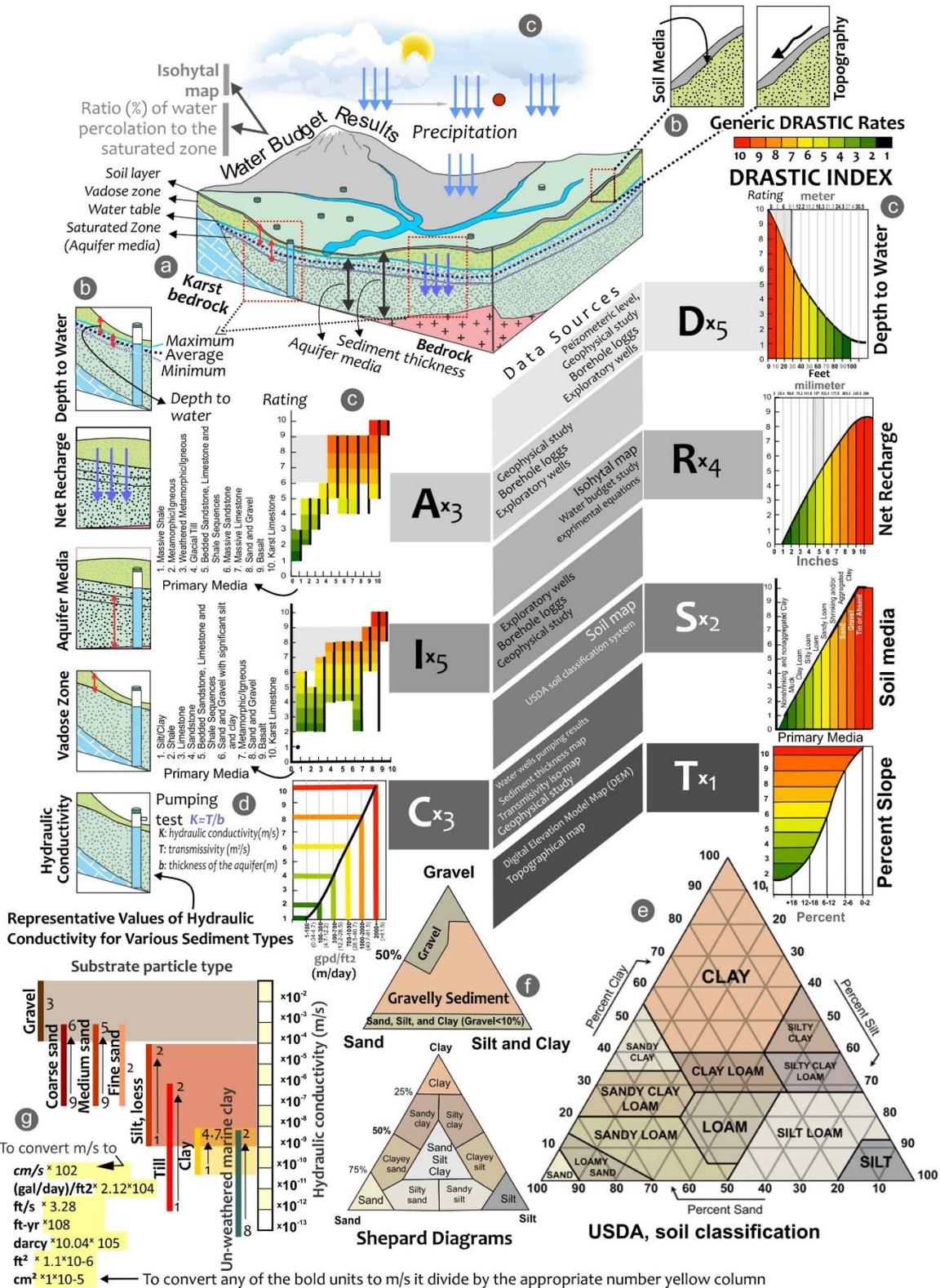
180 where W_j is the weighting of each parameter and R_j is the corresponding rate.

$$181 \quad I_{DRASTIC} = (D_w \times D_i) + (R_w \times R_i) + (A_w \times A_i) + (S_w \times S_i) + (T_w \times T_i) + (I_w \times I_i) +$$

$$182 \quad (C_w \times C_i) \quad (3)$$

183 The variables are: Depth to water (D), Net recharge (R), Aquifer media (A), Soil media (S),
 184 Topography (T), Impact of vadose zone (I), and Hydraulic conductivity (C). Each factor was
 185 calculated by multiplying the weight by the rate of that factor.

186 The importance and explanation of the seven factors and its conceptual model based on
187 Aller et al. (1987) are presented in Fig. 2a,b,c. It should be noted that a complete understanding
188 of the study area is the first step in the feasibility of using a DRASTIC model. As more
189 accurate information and data are provided, the DRASTIC model outputs will be more
190 accurate, functional, and better suited for use in water management.



191

192 **Fig. 2.** a) Conceptual model of the Sahneh aquifer and important components in the evaluation

193 of the DRASTIC model, b) seven parameters of the DRASTIC index, c) generic DRASTIC

194 model rates for seven factors, d) pumping test components, e) USDA Soil classification, f)

195 Separd diagrams, g) , Representative values of hydraulic conductivity for various sediment
196 types, and h) Unit conversions (Domenico and Schwartz 1990).

197

198 **Depth to Water Table (D)**

199 It is defined as the vertical distance from land surface to the water table (Fig. 2a). In porous
200 media, a groundwater depth map can be constructed by using the groundwater levels measured
201 in piezometers in relationship to land surface altitudes (creation of a potentiometric surface
202 maps). To obtain groundwater depth, the multi-year average of water levels in piezometers at
203 the highest and lowest positions was calculated. From the average water level in the
204 piezometers, a groundwater depth map was prepared. Since this study was performed on an
205 unconfined aquifer, most of the analysis and methods are focused on this type of aquifer. In
206 some areas where geophysical surveys were conducted, a more accurate map was obtained by
207 combining the results of the measured water levels and geophysical surveys. In the DRASTIC
208 method, the theoretical weight of this factor is 5. Figs. 3a and 2c show the groundwater depth
209 of the Sahneh aquifer in relation to those proposed by Aller et al. (1987). The most important
210 issue in preparing this factor is the lack of geophysical data and high degree of approximation
211 between points of actual measurement in the piezometers.

212 **Net recharge (R)**

213 It is the quantity of water from infiltration and through vertical percolation penetration that
214 reaches the zone of saturation (passes through the water table). Different methods have been
215 proposed by different researchers to evaluate net recharge (Williams and Kissel 1991; Piscopo
216 2001; Hamza et al. 2007; Karan et al. 2018). The use of these methods requires minimal data
217 related to actual infiltration. Since measurements of the relationship between infiltration and
218 recharge do not exist, it was assumed that the recharge estimated from the water balance and
219 the infiltration rate are approximately equal (Taheri et al. 2017). In this method, by using the
220 multiplication curves of the rainfall percentage (from the isoheytal map) (Fig. 1d), the
221 infiltration rate into the aquifer can be obtained with a good approximation. Fig. 3a shows the
222 net recharge map and Fig. 2c illustrates the related rates.

223 **Aquifer media (A)**

224 Aquifer media is one of the most important parts of the study by the DRASTIC method.
225 Preparing a map with good accuracy can give the final DRASTIC greater credibility and

226 usefulness. The best way to map the aquifer media is to use the results of geophysical surveys
227 and match it with the results of exploratory geologic well logs. In the absence of these two
228 data sources, information was used from public water well drilling logs with a high degree of
229 distribution in the aquifer area.

230 To prepare the map of the aquifer media layer, the type of sediment above and below the
231 water table was determined using statistical analysis and sedimentological diagrams. By
232 assigning numerical codes to each of these types of sediments, it is possible to interpolate and
233 prepare a sediment distribution map. In this study, the sediment diagram of Shepard (1954)
234 was used to classify sediments and prepare a map of the aquifer and vadose zone (Fig. 2f, 3b).
235 Because it is practically difficult to separate clay and silt in the desert sediments, instead of
236 separating these two factors, clay and silt were aggregated into a single class and gravel was
237 added as the third factor. This change in the naming triangle is more applicable to aquifers
238 and unsaturated sediments, although it may not be accurate in terms of sedimentologic
239 laboratory studies.

240 **Soil media (S)**

241 The soil media occurs at the top of the system and controls the infiltration of water into the
242 unsaturated zone and ultimately the recharge of the aquifer to some degree (Fig. 3a).
243 Permeable soils facilitate higher rates of water infiltration and fine-grained soils produce more
244 runoff. In many places, soil maps are available that were produced by organizations related to
245 agriculture and soil science. In this study, the soil map of Iran was used within the context the
246 U. S. Department of Agriculture (USDA) sediment classification (Fig. 2e). This classification
247 was applied, because it is the standard used by many organizations that are involved in soil
248 mapping. It is known to provide an accurate set of soil properties.

249 **Topography (T)**

250 The topography is defined in the generic DRASTIC documentation as the slope and variability
251 in the slope of land surface. The most important effect of this factor is that the slope can cause
252 runoff due to rainfall or cause water to remain on the ground. Pondered water on relatively flat
253 soil surfaces can enhance infiltration and recharge (Maliva and Missimer 2012). Steep-sloped
254 areas have enhanced runoff that lessens the potential for water infiltration. In the generic
255 DRASTIC documentation, slopes >18% have virtually no infiltration. This layer can be
256 considered the most accessible layer in Iran because many researchers use the available digital

257 elevation models and the GIS environment to convert the DEM (digital elevation model) layer
258 into a slope layer (Fig. 1i, 2b, 3a).

259 **Impact of vadose zone media (I)**

260 Aller et al. (1987) consider the impact of the vadose zone important and have assigned the
261 highest weight (i.e., weight of 5) to this layer and the groundwater depth layer. The vadose
262 zone can be very influential in the end result of the model. To prepare this layer, similar to the
263 depth layer of the aquifer, the geologic logs of private wells were used with the interpretation
264 that the sediments above the water table were considered. (Fig. 3b). Some researchers use soil
265 maps in the absence of data for the aquifer vadose zone. However, soil maps contain rather
266 generic descriptions of the lithologic soil properties. The soil profiles described in the maps
267 can be considered up to 2 m in depth based on the DRASTIC model. In aquifers where the
268 water table depth is >2 m below land surface, the use of soil descriptions associated with the
269 soil maps is not accurate enough to populate the model layer.

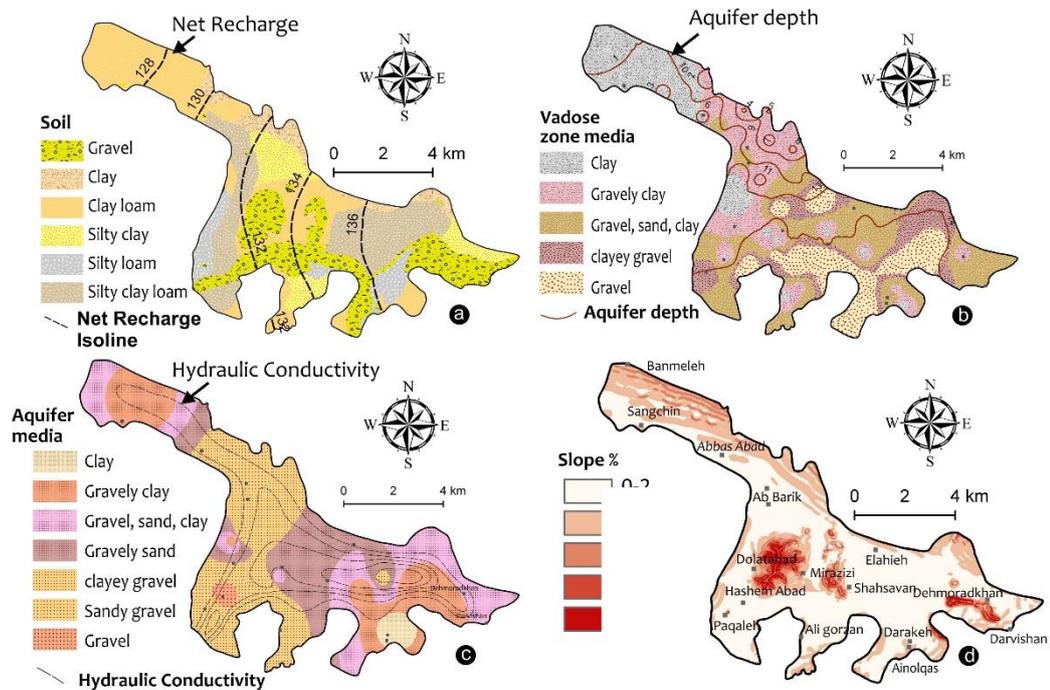
270 **Hydraulic conductivity (C)**

271 It is the last layer in the seven layers of the DRASTIC model. Hydraulic conductivity refers
272 to the ability of aquifer materials to transfer water through porous media or is the
273 proportionality coefficient in Darcy's Law. Higher hydraulic conductivity indicates the
274 potential to allow greater water movement in the aquifer. Hydraulic conductivity data were
275 obtained from the results of pumping tests, slug or withdrawal tests, and some borehole
276 geophysical methods. The best way to prepare this layer is to use the data obtained from the
277 results of aquifer pumping tests by the organizations in charge of water and reports published
278 by consulting engineers or geologists associated with a variety of water projects. These data
279 have been published in the form of thematic maps for water balance studies and aquifer
280 prohibition (Fig. 1d, 3c). In the absence of these data, Aller et al. (1987) have presented a
281 proposed table on the basis of which hydraulic conductivity values can be estimated (Freeze
282 and Cherry, 1979) (Fig. 1g, h). However, using geophysical test data and the portability and
283 thickness of the aquifer, a hydraulic conductivity layer can be obtained. In this study, average
284 hydraulic conductivity was obtained by dividing the transmissivity layer by the aquifer
285 thickness (Fig. 1g, 1h). The hydraulic conductivity of the aquifer was computed using
286 equation 4:

$$287 \quad k = T/b \quad (4)$$

288 where k is the hydraulic conductivity of the aquifer (m/s), T is the transmissivity (m^2/s), and
289 b is the thickness of the aquifer (m).

290 Groundwater experts in Iranian regional water companies and experienced drilling supervisors
291 have provided useful information on hydraulic conductivity.



292

293 **Fig. 3.** Seven layers of DRASTIC model; a) soil media, and net recharge, b) vadose zone media, and
294 aquifer depth, c) aquifer media, and hydraulic conductivity, and d) slope map.

295 Validation and sensitivity analysis

296 Various methods have been developed and used by researchers in order to validate DRASTIC
297 model outputs, and to assess model sensitivity (Benesty et al. 2009; Mogaji and San Lim 2017;
298 Kumar and Pramod Krishna 2020; Lodwick et al. 1990; Gogu and Dassargues 2000;
299 Napolitano and Fabbri 1996). The purpose of these evaluation methods is to test the adequacy
300 of the data and the final results of the model to show whether the output of the work is reliable
301 for later uses or not. Among all the published evaluation methods, the use of nitrate to assess
302 the model validity is very common (Javadi et al. 2011; Shirazi et al. 2012; Fijani et al. 2013).
303 Generally, nitrate is absent in subsurface waters and its presence indicates anthropogenic
304 influence. Therefore, groundwater nitrate concentrations were used in this study to validate
305 the vulnerability models (Malik et al. 2021). The importance of model validation with nitrate

306 stems from the fact that nitrogen is essential for plant growth and nitrate chemical fertilizers
307 are used to produce more crops on agricultural land (Schröder et al. 2004). In the absence of
308 nitrate data, some researchers have used other chemical parameters, such as chloride and
309 electrical conductivity (salinity) (Sahoo et al. 2016; Salek et al. 2018).

310 Various organizations in Iran charged with the collection of groundwater quality data may
311 also have available nitrate data, but the availability of these data is limited, because they are
312 considered to be confidential. However, the nitrate concentrations have been measured in
313 water pumped from public water wells in the area. Thereby, it was used to evaluate the model
314 output. In this study, nitrate data was used to calibrate the model based on 50 samples collected
315 and analyzed from rural drinking water wells. Some researchers have suggested the use of
316 Emergent Organic Compounds for calibration (Hamza et al. 2014). No reports or published
317 papers have been found to document this calibration method.

318 Among the various criticisms leveled at the DRASTIC method, three have received much
319 attention. The first issue is whether it is necessary to use all seven proposed layers for an
320 aquifer evaluation. Another issue is whether the weights and rates suggested by generic
321 DRASTIC are appropriate for the study area or should be they changed to meet local
322 conditions. Third, is whether it is necessary abandon the generic DRASTIC method and do
323 the evaluation with a modified DRASTIC method or another model.

324 Conducting a sensitivity analysis with a different approach is one method that can be used
325 in response to the first and second issues. Sensitivity analysis was done to evaluate the
326 influence of the assigned rates and weights of each class and thematic layer on the output of
327 the aquifer vulnerability potential map since weights and ratings used in the model are
328 theoretical, and there is a general absence of supporting experimental evidence (Napolitano
329 and Fabbri 1996). Modified DRASTIC model approaches have received much attention in
330 response to the third question (Baghapour et al. 2016; Soyaslan 2020; Rajput et al. 2020; Berhe
331 Zenebe et al. 2020). Involvement of regional factors such as land use or special features such
332 as sinkholes support changing the nature of the use of the DRASTIC model (Taheri et al.
333 2017; Warren 2019). For example, Mimi and Madi (2012) developed a modified method that
334 applies to karst areas. In using the modified methods, it is important for the researcher to reach
335 the conclusion that acceptable results cannot be achieved by using the generic DRASTIC
336 model and that the final map contradicts the nitrate content or another evaluation parameter.
337 If there is no contradiction, modifying the model is just an extra, perhaps unnecessary, task.

338 Two common types of sensitivity analysis have been used by various researchers to
339 evaluate the DRASTIC modifications and the impacts of needing all seven parameters to
340 provide a useful evaluation, including map removal sensitivity and single parameter removal.
341 Map removal sensitivity is used to examine the impacts of removing any one of the thematic
342 layers used for the computation of the groundwater vulnerability map (Lodwick et al. 1990)
343 by using equation 5

$$344 \quad S = \left(\left| \frac{\frac{V-v}{N-n}}{V} \right| \right) \times 100 \quad (5)$$

345 where S is the sensitivity measure expressed in terms of a variation index, V and v are the
346 unperturbed and the perturbed vulnerability indices, respectively, and N and n are the number
347 of data layers used to compute V and v, respectively. This method has been used for other
348 studies that are also related to groundwater or natural hazards evaluation (Taheri et al. 2020).

349 Single parameter removal sensitivity analysis examines the impact of each parameter on
350 the final vulnerability map. The effective weight in this method can be estimated using the
351 equation 6

$$352 \quad W = \left(\frac{Pr \times Pw}{V} \right) \times 100 \quad (6)$$

353 where W refers to the ‘effective’ weight of each parameter, Pr and Pw are the rating values
354 and weight of each parameter, respectively, and V is the overall vulnerability index
355 (Napolitano and Fabbri 1996).

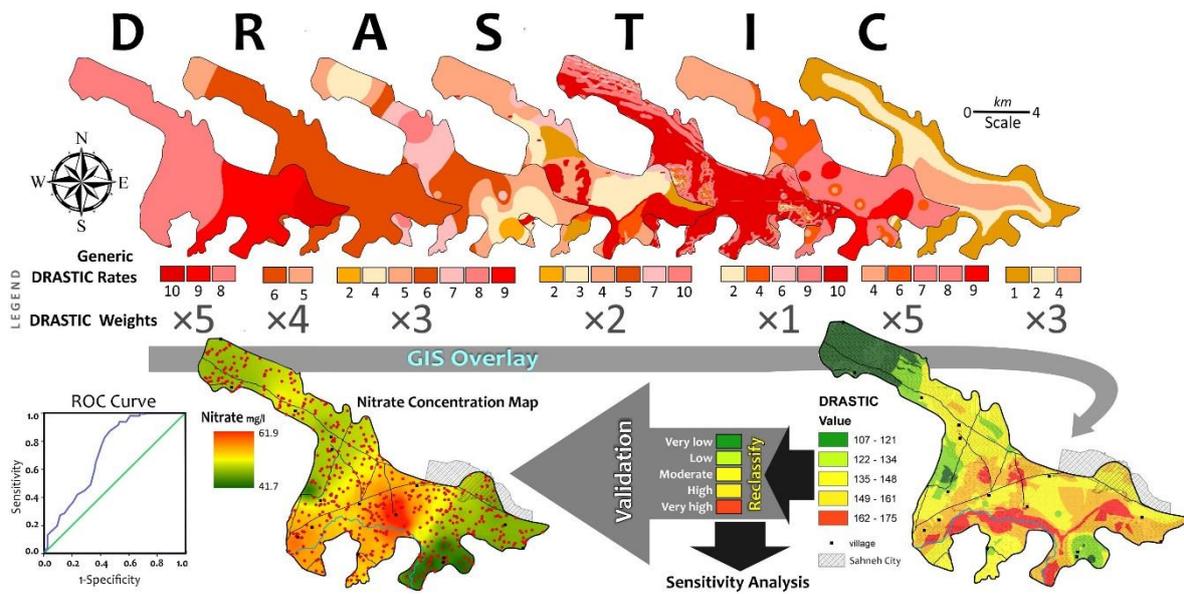
356 If the generic DRASTIC model has applicability and produces logical results after
357 validation, there is no need for sensitivity analysis. In some studies, some parameters do not
358 have significant variation in the region (for example, the net recharge layer in this study, which
359 includes only two categories). Therefore, such layers are well-defined in the sensitivity
360 analysis with a minimal error range. During validation and sensitivity the model
361 designer/researcher is able to determine which layer(s) are not really effective due to lack of
362 data or lack of significant impact on the final result of the model. However, many reviewers
363 of articles and journals consider sensitivity analysis to be a research necessity.

364 **Results**

365 Based on the data available in Kermanshah Regional Water Authority (KRWA) and its
366 affiliated departments as well as general soil maps of Agriculture Organization and
367 Climatology office of the Meteorological Organization, seven maps required for the
368 DRASTIC model were prepared (Figures 3a,b,c,d and 4). These maps provide the most
369 accurate output possible based on the available data. However, these maps are interpolated
370 with minimal data and require much more basic studies, such as geophysics, pedological
371 studies, and field measurements to improve their accuracy. Not only is it not possible to
372 provide more data at this time, but it is also unlikely for the next 10 years, based on the trend
373 of routine studies being conducted in the region. Accordingly, one of the goals of this research
374 was to check the adequacy of the results based on the available data. Data processing, data
375 sources and evaluation methods are provided for each parameter as summarized in Fig. 2.

376 By overlying the seven weighted layers, the final aquifer vulnerability map was prepared
377 (Figs. 4, 5a). The range of the calculated values was from 107 to 175, and they were divided
378 into 5 categories based on the equal interval classifier method in GIS. By applying the equal
379 interval classifier on the final map, it was divided into 5 classes, which are very low (107-
380 121), low (121-134), moderate (134-148), high (148-161), and very high (161-175) (Fig. 5b).

381 The results show that only 8.6% and 6.7% of the study area fall under very low and low
382 vulnerability class, respectively. The moderate and high vulnerability classes cover 26%, and
383 17.2% of the total study area. Most of the area, consisting of 42%, is classified as having high
384 vulnerability (Fig. 5a,b). The southern part of the study area falls under the very high
385 vulnerability class and most parts of the flood plain environment occur in the very high
386 vulnerability class. Most of these areas are lands used for agricultural (crops) and rangeland
387 (grazing). Most industrial and residential areas are located in moderate vulnerability class.



388

389 **Fig. 4.** Flowchart of the final output model of the general DRASTIC model in the study area

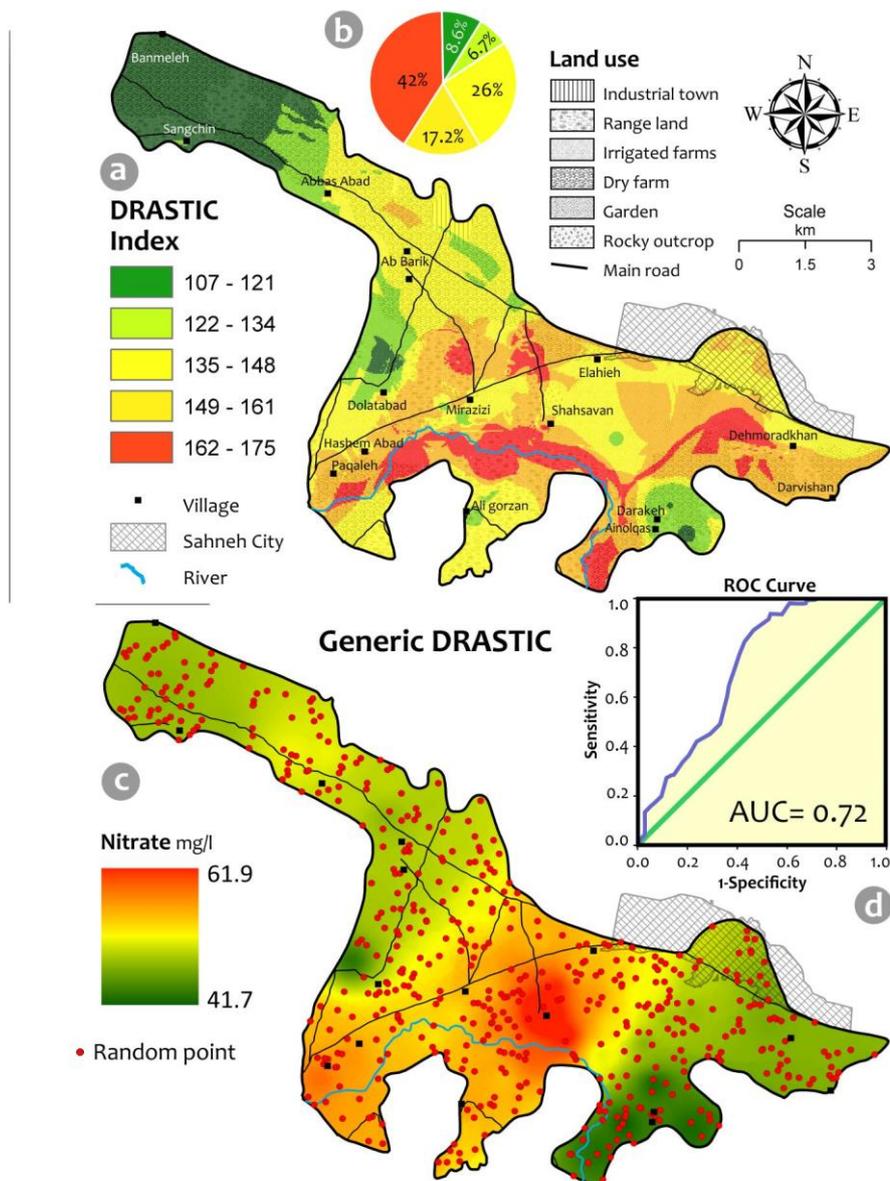
390 **Validation**

391 In this study, using the analysis of drinking water wells, the nitrate concentration in water
 392 produced from these wells was mapped in the GIS environment (Fig. 5b). There are several
 393 ways to validate the DRASTIC map using the concentration of nitrate as a proxy. The first
 394 method is to use linear regression between some or all pixel values to assess the correlation
 395 between nitrate and vulnerable zones. Another method is overlying the two maps in software
 396 environments such as Idrisi, which shows the degree of correlation (Fig. 5e). Another one of
 397 the validation methods used in many research papers in various fields including evaluation of
 398 landslides, groundwater and other sciences is the Receiver Operating Curve (ROC) curve
 399 (Chen et al., 2018; Taheri et al. 2019, 2020).

400 The ROC method is commonly used to assess validity of modeling results in water
 401 resources management (Khosravi et al. 2018). ROC is considered to be an evaluation of a
 402 binary classification system whose detection threshold is also variable. In this method, the
 403 accuracy of the prediction as well as the quality of the created model is evaluated based on the
 404 area under the curve (AUC). The ROC method was used to assess the validity and usefulness
 405 of the created DRASTIC model. To do this, 500 random points were placed within the map
 406 grid using GIS (Fig. 5c). The data corresponding to these points for both the nitrate
 407 concentration map and the final DRASTIC vulnerability map were assessed. For more nitrate

408 concentrations greater than the standard limit of 50 mg/L a value of 1 was assigned and a
 409 value 0 of was assigned to the points below the nitrate limit. Using SPSS software, the
 410 characteristic curve of the graph was created. In this method, a curve showing above 50%
 411 indicates the validity of the model, and the higher the value, the more valid the model.
 412 Yesilnacar and Topal (2005) classified AUC values with respect to prediction accuracy into 5
 413 quantitative ranges, which are 0.5–0.6 for low/poor, 0.6–0.7 for moderate/average, 0.7–0.8 for
 414 good, 0.8–0.9 for very good, and 0.9–1 for excellent. Based on this approach in this study
 415 AUC value obtained was 0.72 or 72%, which is considered to be good validity (Fig. 5d).

416



417

418 **Fig. 5.** Nitrate analysis results (a) and its adaptation to general DRASTIC model (b) and
419 DRASTIC with expert weights (b) linear regression (c) ROC curve with 500 random points
420 (d) and Idrisi output (e).

421 **Discussion**

422 **Validation and usefulness of the DRASTIC model map**

423 Given the widespread use of the DRASTIC model in the world and the importance of this
424 model in evaluating vulnerability of aquifers, methods of evaluating model accuracy in
425 differing hydrogeologic settings is an important topic. It is doubtful that the results of many
426 DRASTIC studies have been evaluated by researchers with nitrate results using the ROC
427 method. Attempts to change the weights and rates for model validation with nitrate analysis
428 results can be very misleading. The results of this study showed that even where there is a
429 minimum of sufficient data, the use of the generic DRASTIC method can provide acceptable
430 results and there is no need to make extensive changes in the structure of the model. Therefore,
431 there is a method available to determine the usefulness of the model output and answer the
432 fundamental question posed by Box and Draper (1987) “All models are wrong, but some are
433 useful; *the practical question is how wrong do they have to be to not be useful.*” The DRASTIC
434 model created to assess the groundwater vulnerability of the aquifer studied is useful.

435 In some studies, using the DRASTIC model, due to the lack of data related to the aquifer,
436 vadose zone, and soil environment have been combined to generalize subsurface conditions
437 (Iqbal et al. 2015; Amiri et al. 2020). This amount of generalization and over-simplification
438 of geological conditions can have a negative impact on the model results. Conversely, in many
439 groundwater vulnerability investigations, validation of the model results using nitrate content
440 is an integral part of the model. While some aquifers do not have much variability of nitrate
441 concentration data, the aquifer geological materials may still susceptible to contamination.
442 Use of other chemical and physical parameters such as EC or sulfate can be used, but these
443 chemical parameters may not be valid proxies for anthropogenic pollution. Sometimes the
444 geological nature of the formations in contact with water determines the ionic and elemental
445 content and has nothing to do with the nature of the aquifer susceptibility to pollution.

446 Based on Hamza et al. (2014), it is concluded that a high degree of subjectivity exists in
447 assigning the index value ranges in a mapped area, and consequently, the type of quintile
448 classification adopted in a particular study. It seems that an expert-centered perspective can

449 be better at determining the most appropriate classification model, rather than a purely
450 statistical perspective. Field studies of soil conditions and geological environments that are
451 prone to pollution show some of the issues can be real or not so.

452 **Use of Fuzzy logic to refine zonal boundaries**

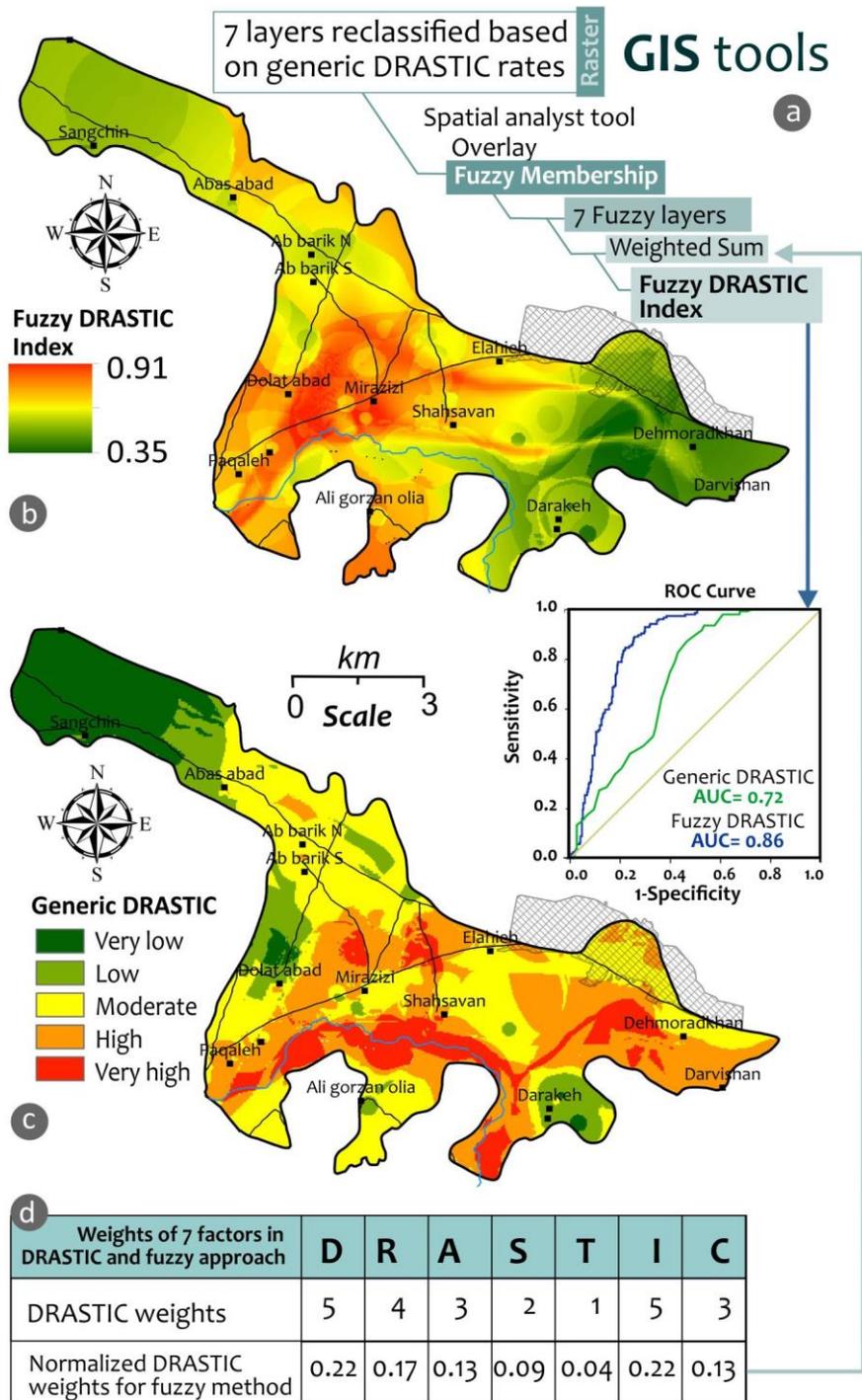
453 In recent years, there have been many criticisms of the sharpness of the boundaries between
454 different classes on the vulnerability map. This criticism may be inherently correct, but with
455 the available data, there may be no definitive solution. Some researchers believe that the fuzzy
456 method (Zadeh 1983) is the best way to better refine the boundaries. The concept of fuzzy
457 logic involves assessment of defined classes and how to define membership within the defined
458 classes. Zadeh (1996) used fuzzy logic to deal with language uncertainty. It is a superset of
459 Boolean logic, which can be used to differentiate between the concepts of partial truth, and
460 degrees of truth from “completely true” and “completely false”. The function of membership
461 with strict boundaries or classical (crisp sets) membership would assign only two classes
462 defining non-membership and 1 being membership. In the realm of fuzzy logic, membership
463 values can be assigned from 0 to 1 with the end members being non-membership and
464 membership. Mohammadi et al. (2009) states that no point lies in a specific rating, but in all
465 ratings.

466 When applying fuzzy-logic to DRASTIC modeling, a set of parameters are developed that
467 are associated with linguist terms and variables. The terms are then linked to differing aspects
468 of the hydrogeologic setting that are indicative of specific vulnerabilities to contamination.
469 The fuzzy system then adjusts itself to the variation of the input indices. The Fuzzy
470 Interferences System (FIS) generates a continuous contamination vulnerability function that
471 provides clear boundaries between different boundaries (Afshar et al. 2007).

472 Numerous articles have been published using fuzzy methods (Huicheng et al. 1999; Afshar
473 et al. 2007; Nobre et al. 2007; Rezaei et al. 2013; Jafari and Nikoo 2019; Pathak and Bhandary
474 2020; Saranya and Saravanan 2021). Some of the fuzzy logic applications have also been
475 applied to investigations using multi-criteria decision-making methods such as AHP (Sener
476 and Davraz 2013; Yang et al. 2017; Mallik et al. 2021). However, applied fuzzy methods still
477 cannot eliminate the definiteness of boundaries in layers, such as occur in soil, aquifer media
478 boundaries, and the vadose zone. Fuzzy methods applied to quantitative layers such as water
479 depth, slope, hydraulic conductivity and net recharge can provide a mental picture of the
480 integration of the change process. Since many data maps are extracted from equivalent curves,

481 there is practically a definite structure in the nature of the data and only the problem form is
482 fuzzy. However, in navigating the water wells to which a number is assigned, the lowest
483 number can be selected as the representative of the finest sediments and the highest as the
484 representative of the coarsest grains, and the final map can be multiplied by the selected weight
485 of the generic DRASTIC model applied. This method is a great idea in aquifers with gradual
486 changes of sedimentary facies, but it is a clear fact that the aquifer layers and the vadose zone
487 and soil media are not practically homogeneous and are dependent on sedimentary paleo-
488 environments conditions and post-sedimentation events.

489 Related software has recently introduced options for fuzzy maps that can provide fuzzy
490 outputs between zero and 1. A fuzzy map for the DRASTIC vulnerability was applied to the
491 Sahneh aquifer for comparison to the generic DRASTIC map (Fig. 6a,b). They are moderately
492 similar in nature, however, the value of AUC on ROC curve in this method is quite more than
493 sharp boundaries in generic DRASTIC (AUC=0.86). The various steps and operating
494 instructions for using GIS software to fuzzy 7 different layers are shown in Fig. 6a. Due to the
495 fact that in the generic DRASTIC method, the algebraic multiplication of weights in rates has
496 been used to overlap the seven layers in the fuzzy DRASTIC. This method cannot be used in
497 a straight-forward manner for a multiplication function. In this case, Normalized DRASTIC
498 weights for fuzzy pattern recognition model were applied (Fig. 6d). The most important reason
499 for increasing the validity of this method compared to the generic DRASTIC approach is the
500 gradual boundaries between different layers is more compatible with the measured nitrate
501 concentration changes. These results showed that although fuzzification of the seven layers
502 increases the accuracy of the model. However, because of the non-continuous nature of the A,
503 S, I layers and practically the effects of sharp boundaries in the final map, this problem cannot
504 be overcome by applying fuzzification.



505

506 **Fig. 6.** The fuzzy-DRASTIC vulnerability of the Sahneh aquifer (a) flowchart of GIS tools
 507 application and build fuzzy DRASTIC map (b) fuzzy Drastic compared to the generic
 508 DRASTIC map (c), normalized DRASTIC weights for fuzzy method (d).

509

510 **Conclusions**

511 Despite the controversy concerning the DRASTIC method, it is still the most common and
512 popular model for assessing the vulnerability for contamination of aquifers, especially alluvial
513 aquifers. The global expansion of this model and all its modified variants has shown that it is
514 a good measure of the ability to manage groundwater resources in terms of managing risk for
515 contamination. The results of this research showed that in an aquifer with relatively limited
516 data, very good results were obtained and verified. These results indicate the acceptable
517 performance of this model without changing weights and rates. In different validation
518 methods, results above 70% indicate the efficacy of the model.

519 Using the DRASTIC model as a minimum requirement for aquifer vulnerability
520 assessments is justified. Use of DRASTIC generated maps can be used by governmental
521 organizations in charge of land use and industrial development to reduce the serious
522 consequences of aquifer pollution. This research with a flowchart can provide a step-by-step
523 and educational approach to land use planners and operations managers to understand the
524 methods, the usefulness of the produced map, and its weaknesses. While certain land uses
525 could be banned in specific very high vulnerability areas, where high quality data support this
526 map accuracy, a lesser degree of regulation could be instituted where map accuracy is in
527 question. Methods such as full containment of hazardous materials (e.g., underground refined
528 petroleum storage tanks) or mandatory groundwater monitoring to assess potential impacts
529 (e.g., agricultural use) could be instituted. It must be clear, that the regulatory water
530 management structure must be linked to the accuracy of the DRASTIC maps, particularly
531 where sparse data are used using the method.

532 It is noteworthy that the more efforts that are made to complicate the model and use
533 multiple relationships to optimize the model, the more ambiguity occurs in the understanding
534 and use of the final map(s) by local water-managers and decision-makers. The incompatibility
535 of some case studies with nitrate can be considered as the inadequacy of the data used to
536 produce some layers, because if the data are sufficient, the concordance of nitrate and other
537 statistical methods with the generic DRASTIC model will be >50%. When the verification
538 applied shows the DRASTIC model has a <50% agreement with the nitrate test, then the model
539 may not be useful as a planning tool. Because of the nature and behavior of porous media are
540 the same based on the properties of the sediments, hydrogeological properties, and hydraulic
541 gradients, any deviations found these fundamental properties within the context of a model

542 could create a poor water-resources regulatory climate with public distrust. DRASTIC is also
543 a universal tool, if properly and ethically applied, can support scientifically supported
544 groundwater quality management. This should be well articulated and enforced at the
545 beginning of the training for DRASTIC users.

546

547 **Ethical Approval**

548 Not applicable, because this article does not contain any studies with human or animal
549 subjects.

550 **Consent to Participate**

551 I understand that I am free to contact any of the people involved in the research to seek
552 further clarification and information.

553 **Consent to Publish**

554 I, Amjad Maleki hereby declare that I participated in the study and in the development of the
555 manuscript title Assessment of alluvial aquifer intrinsic vulnerability by a generic DRASTIC
556 model; a discussion on data adequacy and pragmatic results. I have read the final version and
557 give my consent for the article to be published in Environmental Science and Pollution
558 Research .

559 **Authors Contributions**

560 All authors contributed to the study conception and design. Material preparation, data
561 collection and analysis were performed by Kamal Taheri, Amjad Maleki and Reza Omidipour.
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569

570 **Competing Interests**

571 The authors have declared that no competing interests exist.

572 **Availability of data and materials**

573 The authors confirm that the data supporting the findings of this study are available within
574 the article. The detailed data analyzed during the current study are available from the
575 corresponding author on reasonable request .

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Figures

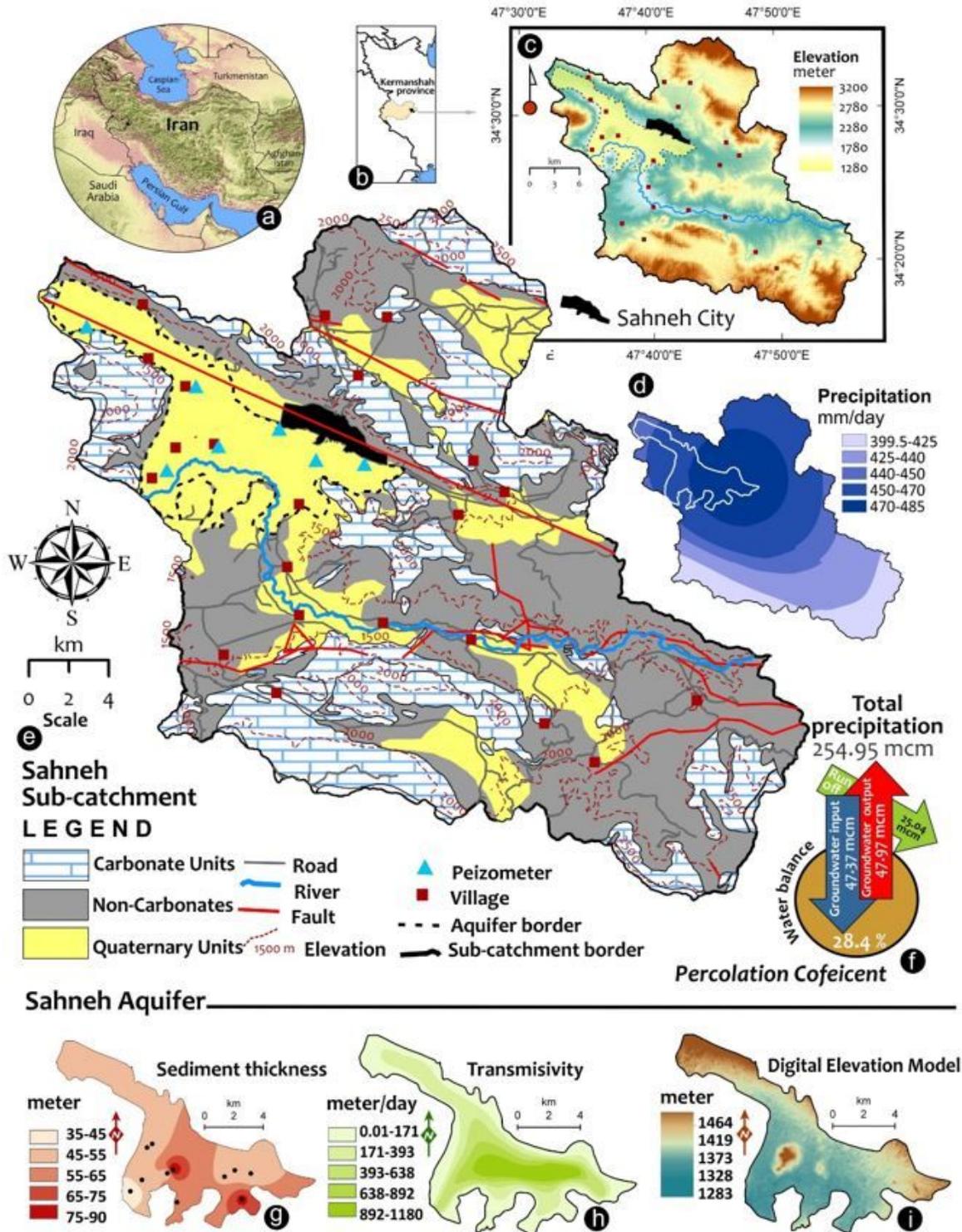


Figure 1

The location of studied area in Iran (a), Kermanshah province (b), and Sahneh Sub-catchment (c), isohyetal map (d), simplified geological map and distribution of carbonate, non-carbonate formations and porous media (plain) and alluvial aquifer boundary (e), water budget summary (f), sediment thickness

map (g), transmissivity map (h) and digital elevation model (i). Note: The designations employed and the presentation of the material on this map do not imply the expression of any opinion whatsoever on the part of Research Square concerning the legal status of any country, territory, city or area or of its authorities, or concerning the delimitation of its frontiers or boundaries. This map has been provided by the authors.

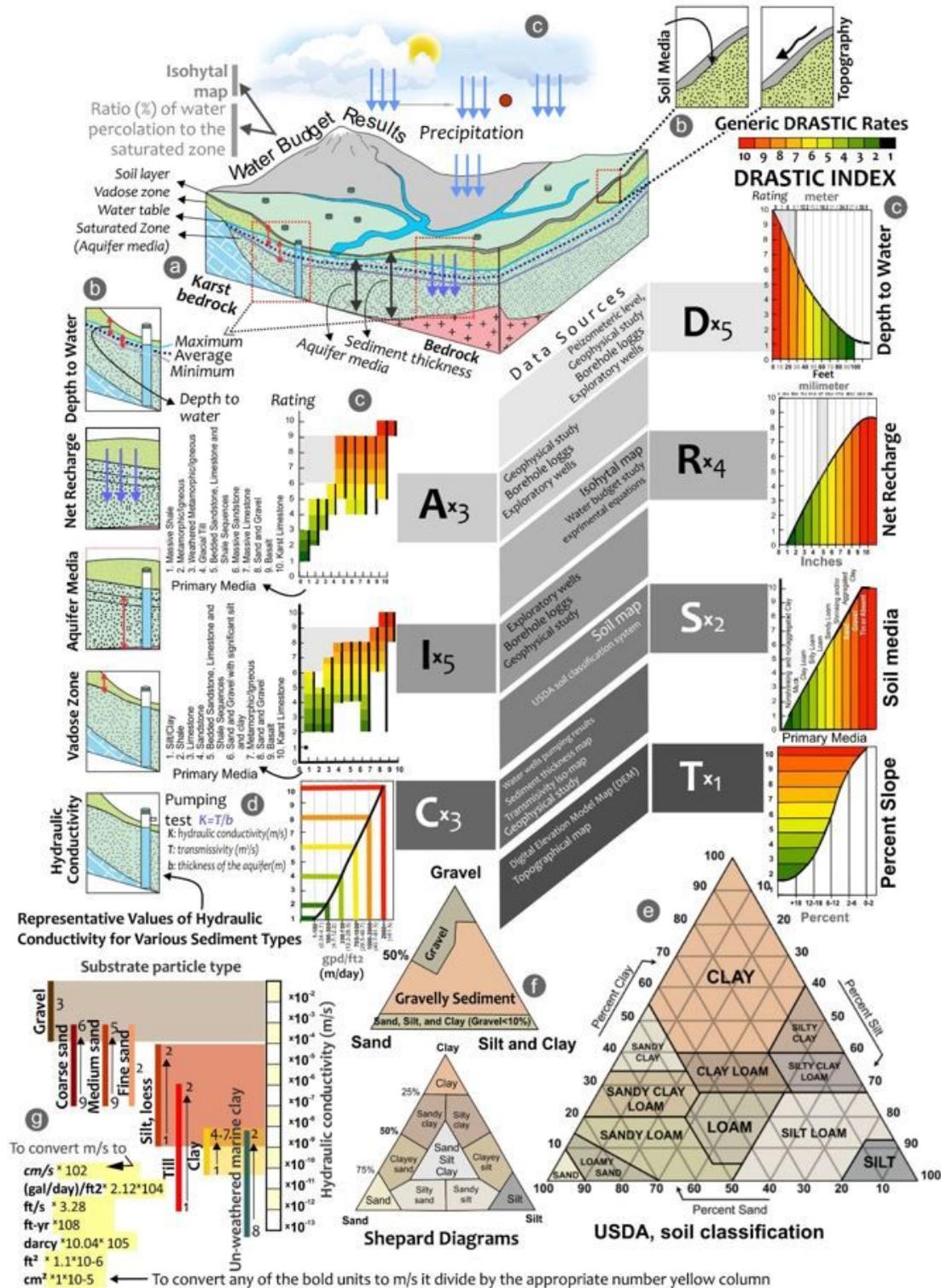


Figure 2

a) Conceptual model of the Sahneh aquifer and important components in the evaluation of the DRASTIC model, b) seven parameters of the DRASTIC index, c) generic DRASTIC model rates for seven factors, d) pumping test components, e) USDA Soil classification, f) Separd diagrams, g) , Representative values of hydraulic conductivity for various sediment types, and h) Unit conversions (Domenico and Schwartz 1990).

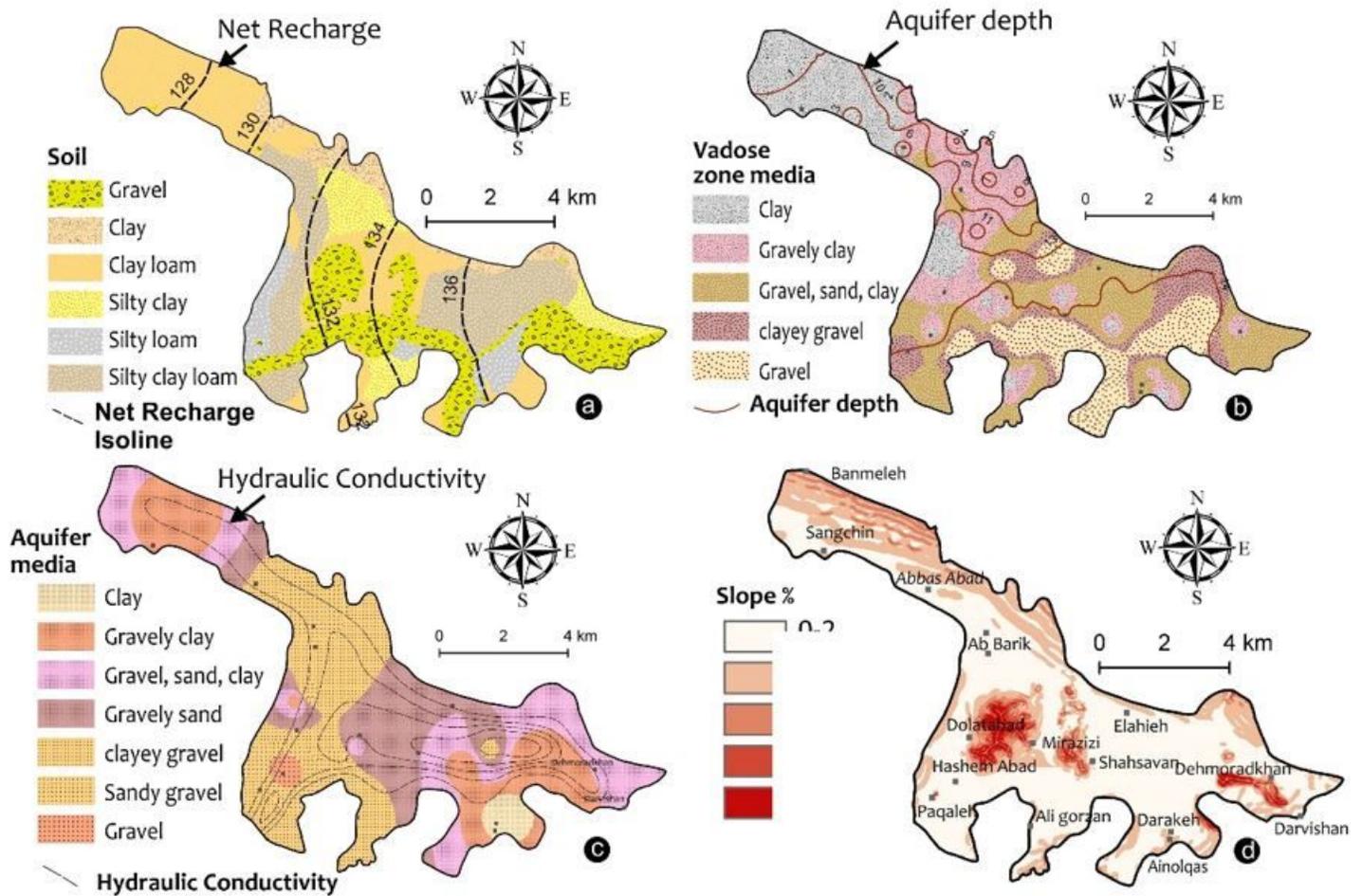


Figure 3

Seven layers of DRASTIC model; a) soil media, and net recharge, b) vadose zone media, and aquifer depth, c) aquifer media, and hydraulic conductivity, and d) slope map. Note: The designations employed and the presentation of the material on this map do not imply the expression of any opinion whatsoever on the part of Research Square concerning the legal status of any country, territory, city or area or of its authorities, or concerning the delimitation of its frontiers or boundaries. This map has been provided by the authors.

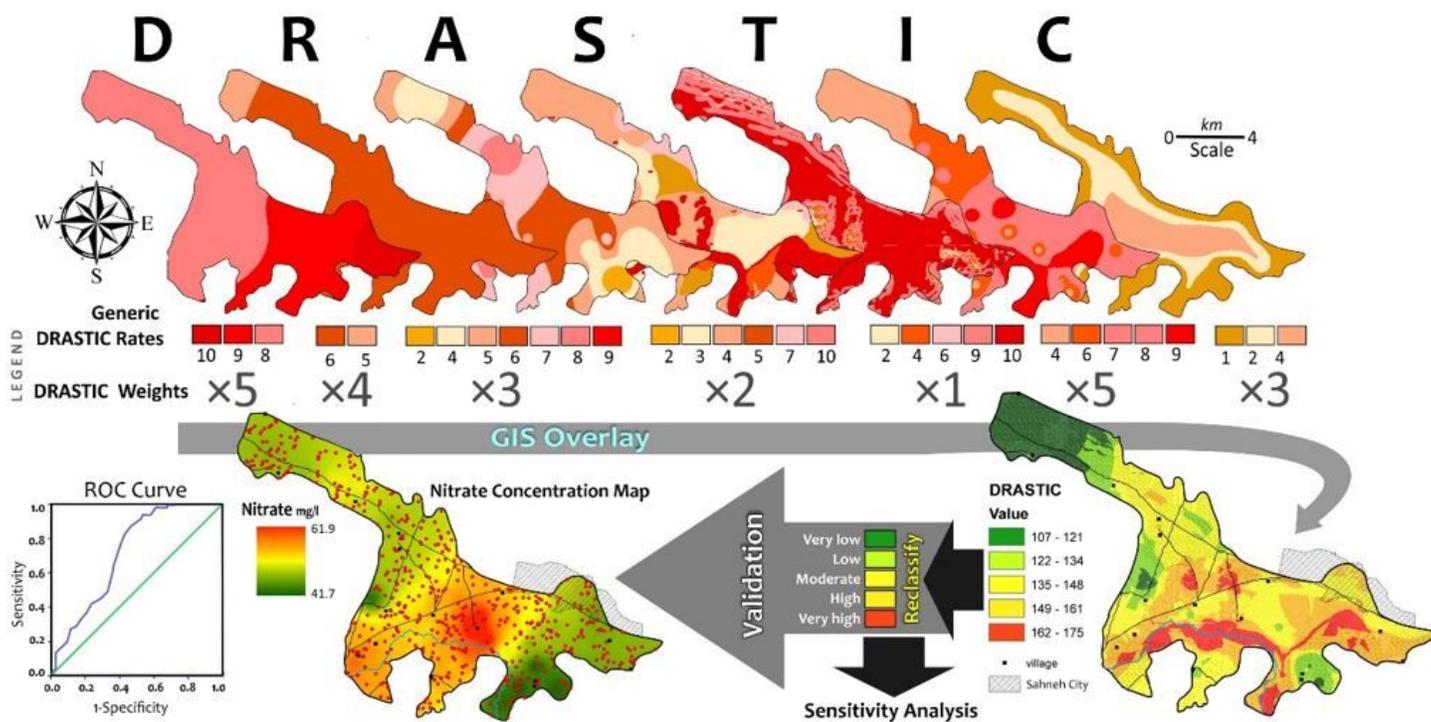


Figure 4

Flowchart of the final output model of the general DRASTIC model in the study area. Note: The designations employed and the presentation of the material on this map do not imply the expression of any opinion whatsoever on the part of Research Square concerning the legal status of any country, territory, city or area or of its authorities, or concerning the delimitation of its frontiers or boundaries. This map has been provided by the authors.

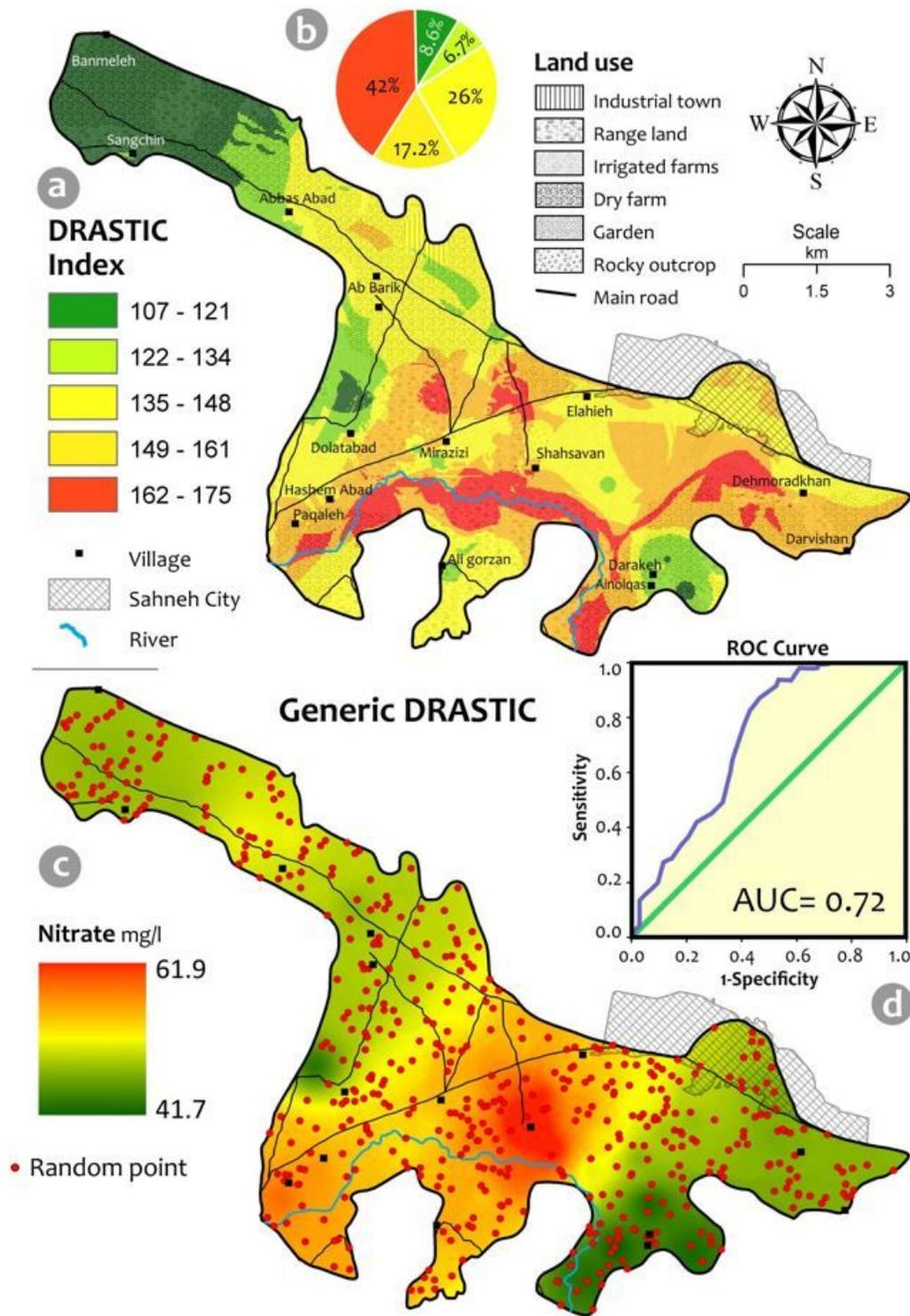


Figure 5

Nitrate analysis results (a) and its adaptation to general DRATIC model (b) and DRATIC with expert weights (b) linear regression (c) ROC curve with 500 random points (d) and Idrisi output (e). Note: The designations employed and the presentation of the material on this map do not imply the expression of any opinion whatsoever on the part of Research Square concerning the legal status of any country,

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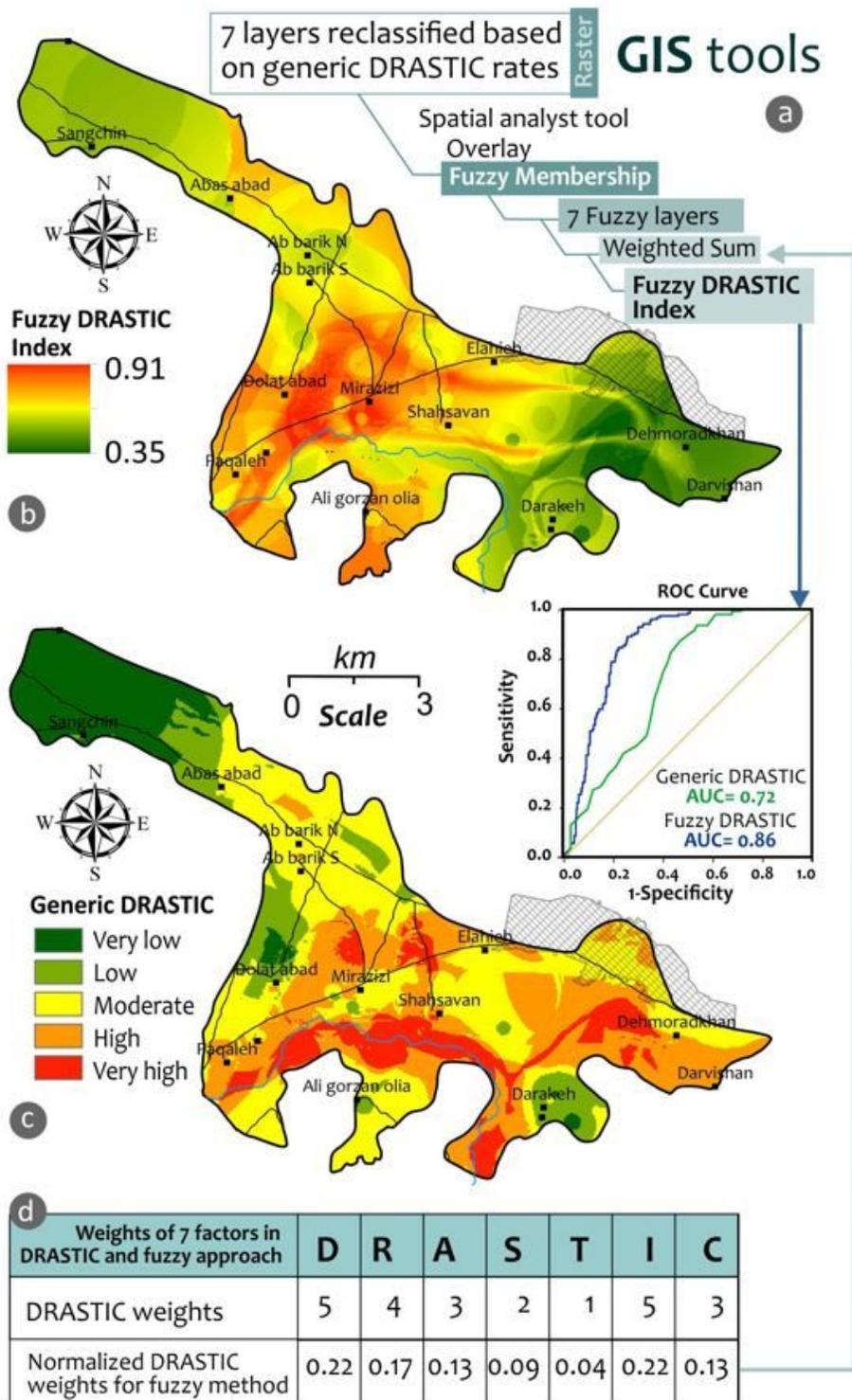


Figure 6

The fuzzy-DRASTIC vulnerability of the Sahneh aquifer (a) flowchart of GIS tools application and build fuzzy DRASTIC map (b) fuzzy Drastic compared to the generic DRASTIC map (c), normalized DRASTIC weights for fuzzy method (d). Note: The designations employed and the presentation of the material on

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