

# Estimating Combined Health Risks of Nanomaterials and Antibiotics From Natural Water: a Proposed Framework

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## Research Article

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1 **Estimating combined health risks of nanoparticles and antibiotics from natural water:**

2 **A proposed framework**

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10

11

12 **Abbreviations**

13	ABs	antibiotics
14	ADD	average daily dose
15	ATSDR	Agency for Toxic Substances and Disease Registry
16	BAF	bio-accessibility fraction
17	CDI	chronic daily intake
18	CuO	copper oxide
19	Fe <sub>3</sub> O <sub>4</sub>	iron oxide
20	HQ	hazard quotient
21	IRIS	Integrated Risk Information System
22	NPs	nanoparticles
23	PNEC	Predicted No-Effect Concentration
24	RfD	reference dose
25	TiO <sub>2</sub>	titanium dioxide

26	USEPA	United States Environmental Protection Agency
27	WoE	Weight of Evidence
28	ZnO	zinc oxide

29  
30

### 31 **Abstract**

32 Nanoparticles are the major class of emerging contaminant detected at relatively high  
33 concentrations in aquatic environment. They are likely to co-exist with other chemical  
34 pollutants such as antibiotics in natural water systems. There are chances that if they are  
35 taken up orally, might pose adverse effects to human health. To address this issue, a risk  
36 framework is developed to study the combined exposure of nanoparticles and antibiotics in  
37 natural waters for the first time. The framework was applied to a hypothetical exposure of  
38 nanoparticles (CuO, ZnO, Fe<sub>3</sub>O<sub>4</sub> and TiO<sub>2</sub>) and antibiotics (ciprofloxacin, CIP; ofloxacin,  
39 OFX; norfloxacin, NOR; levofloxacin, LEVO) to estimate human risks in a six-step approach  
40 for two different exposure scenarios i.e. availability adsorption isotherm data and vice versa.  
41 Risk was also estimated for the released fragments of antibiotics, nanoparticles and metal  
42 ions in the human digestive system. Mixture toxicity risk assessment was conducted for pairs  
43 (i) antibiotics and metal ions, (ii) antibiotics and nanoparticles, and (iii) nanoparticles and  
44 metals ions. Though the estimated risk values were observed to be less than 1 (both hazard  
45 quotients and hazard interactions less than 1) for all the conditions and assumptions made but  
46 it requires through monitoring of the studied contaminants in water to protect humans from  
47 their adverse effects, if any. Maximum allowable concentrations at which no risk occurs to  
48 humans was found to be (maximum values): antibiotics (233.8 µg/L, NOR); metal ions (1.02  
49 × 10<sup>9</sup> mg/L, Ti<sup>2+</sup> ions), and nanoparticles (6.68 × 10<sup>5</sup> mg/L, TiO<sub>2</sub>), respectively.

50 **Keywords: Antibiotics; nanoparticles; interaction; water; health risk; oral exposure**

51

## 52 **1 Introduction**

53 In the past few years, increased concerns have been raised due to occurrence of a wide  
54 variety of emerging contaminants including nanomaterials, pharmaceutical drugs etc., in  
55 natural water systems. Studies suggest that the concentration of these contaminants in water  
56 ranged from  $\mu\text{g/L}$  to  $\text{ng/L}$  (Chen et al., 2016; Ebele et al., 2017) and even high concentration  
57 values has also been reported. Nanoparticles usually display unique physical and chemical  
58 properties, and because of their inherent reactivity with other pollutants, nanoparticles may  
59 act as a carrier and co-occur with other pollutants producing long-term environmental and  
60 health risks (Azizi et al., 2016; Wang et al., 2016).

61 Upon release and emission, contaminants like nanoparticles may interact with  
62 chemicals (antibiotics) in the environment, potentially leading to a co-exposure of organisms  
63 and the occurrence of mixture effects (Naasz et al., 2018). That co-exposure to  
64 nanoparticles and antibiotics may occur is a valid assumption. Both substances are likely to  
65 be present and co-exist in the aquatic environment (Lammel et al., 2019). Although studies  
66 have been conducted to identify the occurrence of these contaminants but very little or  
67 restricted information is available about their environmental exposure (Coll et al., 2016;  
68 Holden et al., 2014). In real life scenarios, they might differ in their toxicity or can undergo  
69 transformation to produce products which under certain circumstances might show harmful  
70 effects and thus needs to be addressed.

71 Organisms are usually exposed to multiple mixtures of contaminants instead of single  
72 compounds (Uwizeyimana et al., 2017). During the process of passage, it is possible that  
73 nanoparticles or pharmaceutical drugs can form nanoparticle-toxin complexes or antibiotics

74 complexes or even produce inter-category complexes of two contaminants due to  
75 nanoparticles high surface area and large aggregates (Zhu et al., 2011) thus, there are ongoing  
76 concerns on evaluating the environmental risk for the mixtures containing NPs. However,  
77 studies on the interaction of two different types of emerging contaminants have so far not  
78 been reported. In recent years the scientific community has undertaken enormous efforts to  
79 assess the eco-toxic potential of nanomaterial (Menard et al., 2011), but there are  
80 comparatively few studies that have investigated their interaction and combined toxicological  
81 effects with co-existing “traditional” environmental pollutants (Canesi et al., 2015; Hartmann  
82 and Baun, 2010; Naasz et al., 2018).

83         Looking into the potential adverse effects of these contaminants on human health, risk  
84 assessment studies have been conducted for some classes, for example, nanoparticles (Parsai  
85 and Kumar, 2020), pharmaceutical drugs (antibiotics) (Kumari and Kumar, 2020) etc.,  
86 (Supplementary Table S1) but none of the reported studies (as per authors best knowledge)  
87 have tried to capture the interaction aspect linked with nanoparticles and antibiotics in natural  
88 water systems. Lack of available guidelines and regulations adds to the ongoing problem and  
89 makes it even more difficult, if not properly taken care of. Therefore, it becomes imperative  
90 to study and identify the interaction of these contaminants in water so that guideline values  
91 can be formulated for exercising appropriate control measures.

92         This study aimed at proposing a framework to determine the risk exposure effects of  
93 nanoparticles and antibiotics to the human digestive media or GI-tract followed by oral  
94 ingestion. Widely detected nanoparticles, NPs (ZnO, CuO, Fe<sub>3</sub>O<sub>4</sub>, and TiO<sub>2</sub>) and  
95 pharmaceutical drugs, antibiotics (Ciprofloxacin, Ofloxacin, Norfloxacin, and Levofloxacin)  
96 in natural water were selected.

97

98 **2 Methodology**

99 Figure 1 shows the flow diagram of the proposed framework for determining health risk  
100 estimates due to the interaction of nanoparticles and antibiotics in natural water followed by  
101 oral ingestion. The study used a six-step risk assessment approach to determine risk exposure  
102 effects of nanoparticles and antibiotics to human health (Kumari and Kumar, 2020a; Kumar  
103 et al., 2014). Briefly, this framework assumes that when these contaminants enters into the  
104 human body via oral route they might get disintegrate into respective antibiotics and  
105 nanoparticles with due course of time in the human digestive system. The released  
106 nanoparticles might get further dissolved into their respective ions, releasing free  
107 nanoparticles. These released metals can cause harmful health effects in spite of the fact that  
108 zinc, copper, and iron are essential elements for all living organisms if present in excessive  
109 quantities (Evangelou et al., 2007; Twining et al., 2005). Therefore, this approach assumes  
110 the probability of risk exposure effects due to released antibiotics and released nanoparticles  
111 in the human digestive system. Risk assessment study was also conducted for free  
112 nanoparticles and free metals ions after the dissolution of nanoparticles in human digestive  
113 liquid to check whether they pose any possible health risks or not. This study does not  
114 consider the size, shape, charge, and surface area of nanoparticles as they might show large  
115 variation after their dissolution in human digestive system as it is reasonably tough to capture  
116 these aspects in real life scenarios.

117

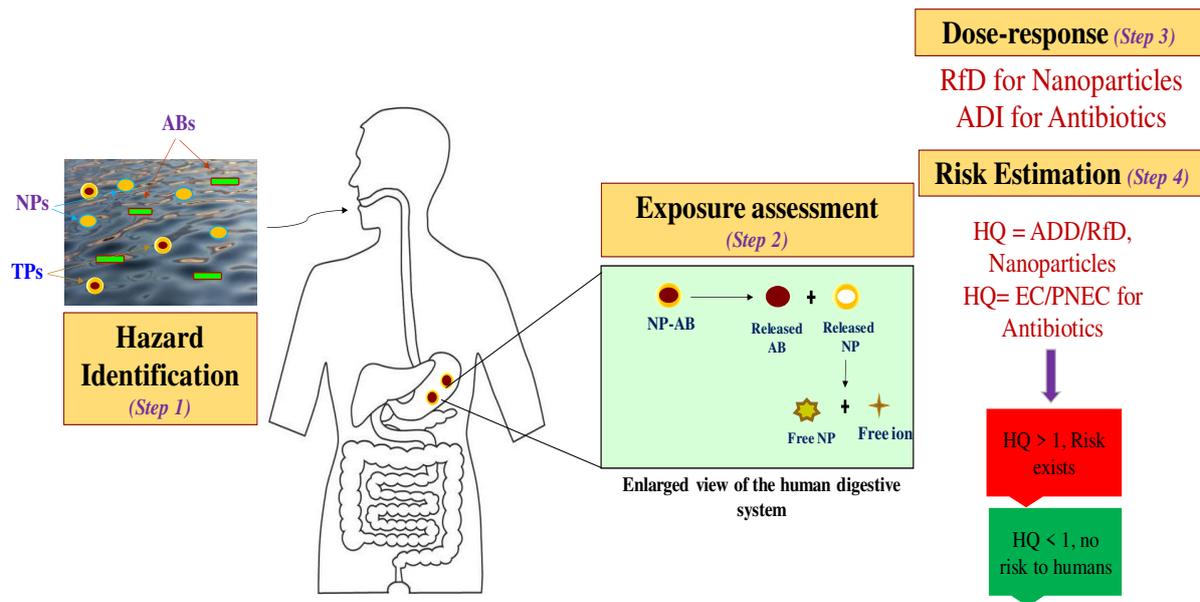
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123

124 **Fig. 1** Hypothetical schematic diagram showing the uptake of nanoparticles-antibiotics

125 transformed products and their dissolution in the human digestive system (NPs =

126 nanoparticles, ABs = antibiotics, TPs = transformed products, HQ = hazard quotient, EC =

127 Environmental concentration, PNEC = Predicted no-effect concentration; RfD = Reference

128 dose; ADI = Acceptable daily intake)

129

### 130 2.1 Hazard Identification

131 Hazard identification involves identifying the material of interest and collecting information

132 for which risk evaluation needs to be done. To determine the suitability of the suggested

133 framework, this study considered the hypothetical exposures of nanoparticles and antibiotics

134 for illustrative purpose. Nanoparticles such as CuO, ZnO, Fe<sub>3</sub>O<sub>4</sub> and TiO<sub>2</sub> were considered as

135 they are used as antimicrobial agents and additives in consumer and health-care products. The

136 risk assessment of selected nanoparticles is essential as several research investigations have

137 shown their adverse effects to human health (Croteau et al., 2014; Ye et al., 2018). Amongst

138 the nanoparticles selected iron oxide nanoparticles (Fe<sub>3</sub>O<sub>4</sub>, γ-Fe<sub>2</sub>O<sub>3</sub> and superparamagnetic

139 IONPs) have been extensively used for pharmaceutical applications (Ding & Guo, 2013;  
140 Namvar et al., 2014). CuO NPs might induce oxidative stress resulting in destruction of  
141 human liver cell (Shukla et al., 2013). ZnO is one of the most frequently detected NPs in  
142 surface water (Kurlanda-Witek et al., 2014) and has been reported be cytotoxic and harmful  
143 compared to other metallic nanoparticles (Li et al., 2020).

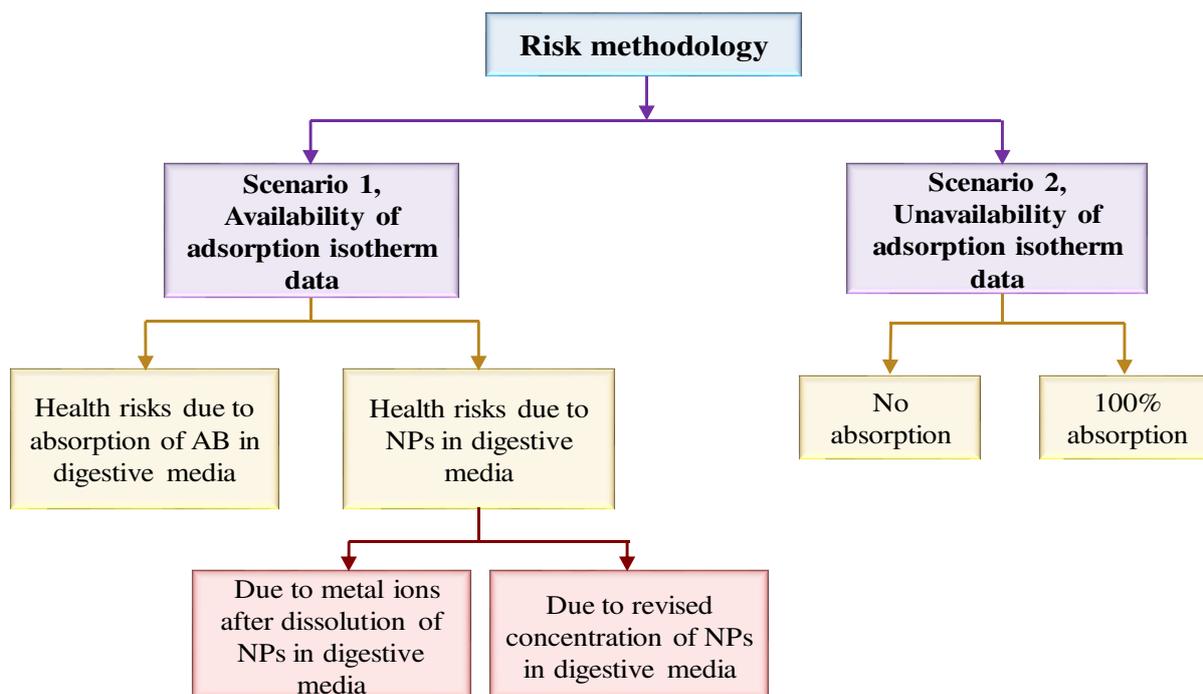
144 To study the interaction of nanoparticles with antibiotics, the most widely detected  
145 fluoroquinolones (FQs) antibiotics [ciprofloxacin (CIP), ofloxacin (OFX), norfloxacin  
146 (NOR), and levofloxacin (LEVO)] in the environmental media were selected owing to their  
147 use in the treating pulmonary, urinary, and digestive infections. The selected antibiotics are  
148 also effective in treating wide range of pathogenic bacteria (gram-negative and gram-  
149 positive) and mycoplasmas (Hooper and Wolfson, 1993; MacGowan and Andersson, 2003).  
150 CIP, NOR, and OFX common antibiotics administered to humans (Alder et al. 2001; Choma  
151 2003). Amongst these, CIP is one of the most extensively used drug in the world (Sukul and  
152 Spitteller, 2007) and LEVO has been listed as one of the most essential medicine of human  
153 use by the World Health Organization (Hooper and Rubinstein, 2003). The environmentally  
154 occurring concentration (EC) of antibiotics and nanoparticles is taken from published  
155 scientific literature. The information is provided as supplementary text Table S2 and S3.

156

## 157 2.2. Exposure assessment

158 This research study evaluated health risks due to the exposure of inter category emerging  
159 contaminants (nanoparticles and antibiotics) to children as they have been recognised to be  
160 the most sensitive sub-population compared to adults (Preston, 2004). To determine exposure  
161 effects, health risks was estimated considering two different scenarios as mentioned above in  
162 Fig. 1a. Detailed information is provided below.

163



164

165 **Fig. 1a** Scenarios considered for determining risk exposure to humans

166

167 Under scenario 1, the study considered nanoparticles for which the adsorption  
 168 isotherm data was available in literature. Scenario 2 indicates the condition where the  
 169 adsorption isotherm data for the selected nanoparticles was not available in published  
 170 literature. For scenario 1, risk was estimated for two different exposure routes (1) Risks due  
 171 to absorption of antibiotics in the digestive media only, and (2) Risk due to absorption of  
 172 nanoparticles in digestive media. Under exposure route 2, two sub-routes were considered:  
 173 (a) Route 2a: risk due to exposure to metal ions after the dissolution of nanoparticles in  
 174 digestive liquid, (b) Route 2b: risk due to the revised concentration of nanoparticles in  
 175 digestive media. For scenario 2, two cases (i) no absorption, and (ii) 100% absorption were  
 176 considered. The detailed information is provided in the sections to follow.

177

178

179 2.2.1 Risk estimation of NPs in the digestive media followed by GI-tract absorption when the  
180 isotherm data is available

181 It is believed that oral ingestion serves as the main and primary route for the probable uptake  
182 of NPs to target sites or structures followed by their dissolution in the human digestive  
183 system (Fig. 1). In risk assessment, dissolution of NPs in the human digestive system gives a  
184 realistic and accurate quantity of nanoparticle loading to the digestive system. During risk  
185 estimation process, bio accessibility fraction of nanoparticles ( $BAF_{NP}$ ) was considered to  
186 show the influence of parameters governing the interaction of NPs with digestive liquid, and  
187 also to calculate the realistic exposure dose.  $BAF_{NP}$  serves an important parameter as it  
188 considers the dissolution of nanoparticle to metals ions in the human digestive system. This  
189 study used the BAF values of nanoparticles available in literature for risk assessment purpose  
190 (Zhong et al., 2017). It is observed that the dissolution of nanoparticles led to the release of  
191 metal ions in the digestive system. Therefore, it is also important to determine exposure doses  
192 for the residual concentrations of released metal ions and nanoparticles in the digestive  
193 system to get the overall risk exposures of NPs absorption in the digestive media.

194 Risk estimation of metal ions released after bio-assimilation of NPs in the digestive  
195 media was carried out in terms of the average daily dose of metal ions ( $ADD_{Mi, DM}$ ) to the  
196 human digestive system using the  $BAF_{NP}$  as mentioned above.  $ADD_{Mi, DM}$  (mg/kg/day) for  
197 individual metal ions in the digestive system was estimated using Eq. (1). Similarly, risk of  
198 revised concentration of NPs in the digestive media followed by GI-Tract absorption was  
199 calculated as  $ADD_{R\_NP, DM}$ .  $ADD_{R\_NP, DM}$  of individual NP was determined using Eq. (2).

200

$$201 \quad ADD_{Mi, DM} = \frac{C_{Mi} \times BAF_{NP} \times IR_w \times EF \times ED}{BW \times AT} \quad (1)$$

202

$$203 \quad ADD_{R_{NP,DM}} = \frac{C_{NP,DM} \times BAF_{NP} \times IR_w \times EF \times ED}{BW \times AT} \quad (2)$$

204

205           Where,  $C_{Mi}$  is the concentration of released metal ions in the digestive media;  $BAF_{NP}$   
 206 is the bio-assimilation potential of NPs;  $C_{NP, DM}$  is the revised concentration of NPs in the  
 207 human digestive system.

208

209 *2.2.2 Risk estimation of ABs in the digestive media followed by oral administration when the*  
 210 *isotherm data is available*

211 In this case, concentration of nanoparticles was assumed to be 1 mg/L. Concentration of ABs  
 212 after adsorption by nanoparticles (mg/g) is taken from published literature, provided as  
 213 supplementary information, Table S2. To determine the concentration of antibiotics in  
 214 adsorbed form on nanoparticles in mg/L, the concentration of nanoparticles (mg/L) was  
 215 multiplied by concentration of antibiotics after absorption on NPs (mg/g) and divided by  
 216 1000. 1000 is the conversion factor. To calculate the concentration of antibiotics in the  
 217 human digestive media followed by GI-tract absorption ( $Conc_{AB\_DM}$ ) after oral ingestion,  
 218 dissolution rate of antibiotics in the digestive media and the rate at which antibiotics is  
 219 getting absorbed by the GI-tract is taken into account. Dissolution rate and absorption rate of  
 220 data of antibiotics is taken from those reported in literature and is provided as supplementary  
 221 Table S4.  $Conc_{AB\_DM}$  is calculated using Eq. (3).

222

$$223 \quad Conc_{AB\_DM} = Conc. \text{ of AB in adsorbed form of NPs} \times Dissolution \text{ rate of AB} \\ \times Absorption \text{ rate} \quad (3)$$

224

225 To determine the risk exposure effects, chronic daily intake (CDI) values (mg/kg-bw/day)  
226 were estimated using Eq. (4).

227

$$228 \quad CDI_{AB\_DM} = \frac{C_{AB\_DM} \times IR \times ET \times ED}{BW \times AT} \quad (4)$$

229 Where,  $C_{AB\_DM}$  is the concentration of antibiotics in the digestive media followed by GI-tract  
230 absorption in GI-tract in mg/L, BW is the body weight (Kg); IR is the intake rate of water, EF  
231 is exposure frequency (365 days/year), and ED is the exposure duration (70 Yrs.).

232

### 233 *2.2.3 Risks estimation when the adsorption isotherm data were not available*

234 This study also tried to determine risk exposure effects for human health for a hypothetical  
235 situation where the adsorption isotherm data of ABs on adsorption by nanoparticles is not  
236 available. Under this scenario, two different cases (i) no absorption, and (ii) 100% absorption  
237 were considered to determine risk estimates.

238 (i) No absorption: Under this case, it is assumed that no absorption of NPs takes place in the  
239 GI-tract, and the ABs ingested remains as free ABs in the digestive media. Risk exposure  
240 of ABs was estimated using the surface water concentration of ABs as mentioned in the  
241 hazard identification section. Acceptable daily intake (ADI) values of individual  
242 antibiotics were used to calculate the Predicted no-effect concentration (PNEC) values.  
243 ADI values specifies the level of daily intake that should not result in any damaging  
244 effects to human health from direct exposure (Cunningham et al., 2009) whereas PNECs  
245 represent the lowest concentration values at which no harmful effects are anticipated.  
246 Input parameters used to estimate PNEC values and ADI values of the antibiotics is given  
247 in Table 1. PNEC values were estimated using Eq. (5) as given below.

248

249 
$$PNEC = \frac{ADI \times BW}{IR_w \times GI_{AF}} \quad (5)$$

250 Where, ADI is Acceptable Daily Intake ( $\mu\text{g kg-day}^{-1}$ ); BW is body weight of children in  
251 Kg;  $IR_w$  is the intake rate of water in  $\text{L day}^{-1}$ ; and  $GI_{AF}$  is the gastrointestinal absorption  
252 factor of antibiotics, assumed to be 1 in this study.

253 (ii) 100% absorption: Under this case, it is assumed that the amount of NPs and ABs is  
254 getting fully absorbed in the GI-tract. Risk was estimated similar to the approach  
255 mentioned in section 2.2.1.

256

### 257 2.3 Dose-response assessment

258 Limited available data on the reference doses (RfDs) of nanoparticles, for instance,  $\text{TiO}_2$  and  
259  $\text{Fe}_3\text{O}_4$  in the human digestive system makes it difficult to determine possible risks to human  
260 health. To fill this knowledge gap, this study used the recommended reference dose values of  
261 ions to denote the reference dose values of nanoparticles, where the RfD values of  
262 nanoparticles were considered equal to their corresponding ions. As mentioned earlier, since  
263 the RfD values for a majority of nanoparticles considered in this study is not available in the  
264 published literature hence, toxicity values of metal ions were used to represent the toxicity of  
265 NPs. A reference dose value for iron is neither available in the Integrated Risk Information  
266 System (IRIS) (U.S. EPA, 2006a) nor the Drinking Water Standards and Health Advisories  
267 list (U.S. EPA, 2005). Therefore, in this case, the provisional RfD (p-RfD) values for iron,  
268 derived using the lowest-observed-adverse-effect level (LOAEL) and uncertainty factor (UF)  
269 was considered (U.S. EPA, 2006b). RfD values of other nanoparticles i.e., ZnO NPs and CuO  
270 NPs is taken from recently published work (Parsai and Kumar, 2020). ADI values of  
271 individual antibiotics used to estimate the PNEC values is taken from published literature.  
272 Table 1 lists information about the parameters used for risk estimation.

273

274 **Table 1** Information about parameters used to determine risk estimates

Parameters	Units	Values	References
Body weight, BW	Kg	16.7	Argall et al., 2003
Average time, AT	Year	365 × 70	ATSDR, 2005
Exposure Duration, ED	Year	70	
Exposure Frequency, ED	Days	365	
Reference dose values, RfD			
ZnO NPs	mg kg <sup>-1</sup> day <sup>-1</sup>	0.0315	Parsai and Kumar, 2020
CuO NPs	mg kg <sup>-1</sup> day <sup>-1</sup>	0.0262	
TiO <sub>2</sub> NPs as Ti <sup>2+</sup> ion	mg kg-bw <sup>-1</sup> day <sup>-1</sup>	3	Ramoju et al., 2020
Fe <sub>3</sub> O <sub>4</sub> NPs as Fe <sup>3+</sup> ion	mg kg-day	0.7	U.S. EPA, 2006b
Zn <sup>2+</sup>	mg kg <sup>-1</sup> day <sup>-1</sup>	0.3	IRIS, 2005
Cu <sup>2+</sup>	mg kg <sup>-1</sup> day <sup>-1</sup>	0.04	

275

276

277 **2.4 Risk estimation and characterization**

278 Hazard quotient (HQ) values were calculated to determine risk to children. HQ is the ratio of  
 279 the possible exposure to a contaminant and the extent to which no adverse effects are  
 280 expected to occur. If the HQ is observed to be less than 1, then no adverse health  
 281 consequence is anticipated as a result of exposure (Kumari et al., 2015). HQ values were  
 282 calculated using the CDI and ADD values estimated in the exposure assessment section and  
 283 RfD values taken from table 1 as given in Eq. (6-9). Estimated HQ values were used the  
 284 calculate the hazard index (HI) values for interactions of (i) antibiotics with metal ions, (ii)  
 285 antibiotics with nanoparticles, and (iii) metal ions with nanoparticles in the human digestive

286 system as mentioned in Eq. (10-12). Limited information is available in published literature  
287 on how antibiotics interact with metal ions or nanoparticles. Urbaniak et al. (2007) used 'k'  
288 values to provide information on the strength of interaction between fluoroquinolones and  
289 metals and observed strong interaction between them. Few studies reported synergistic effects  
290 for the interactions of antibiotics with nanoparticles (Abo-Shama et al., 2020) or metal ions  
291 (Nazari et al., 2012) however antagonistic effects do occur as well. Turel (2002) in his study  
292 showed that the activity of fluoroquinolones reduces in the presence of metal ions.  
293 Complexation with metal ions is one of the primary reasons for the reduced activity of  
294 fluoroquinolones (Seedher and Agarwal, 2010), and also modifies their solubility and binding  
295 capacity (Djurdjevic et al., 2007). This study assumed that synergistic effects occur for the  
296 interaction of antibiotics with metal ions or nanoparticles. For case (i) and (ii), HI values  
297 were estimated using the modified USEPA Weight of Evidence (WoE) approach adopted  
298 from Kumari and Kumar (2020a) study. The detailed information and formula is provided as  
299 supplementary Text information, T1. Under WoE approach,  $M_{ij}$  which indicates the  
300 magnitude of interaction i.e. the influence of  $j^{\text{th}}$  compound on the toxicity of  $i^{\text{th}}$  compound  
301 was taken as 5, the default value as per USEPA recommendations, and  $B_{ij}$  denotes the  
302 strength of evidence for which scores were assigned as per the USEPA classification scheme  
303 (USEPA 2009a, b). For case (iii) i.e., HI values for the interaction of metal ions with NPs, it  
304 was assumed that they do not interfere with each other and therefore does not pose any toxic  
305 effects to human health. Dose-addition method was used to estimate HI values as mentioned  
306 in Eq. (12).

307

$$308 \quad HQ_{AB\_DM} = \frac{CDI_{AB\_DM}}{ADI} \quad (6)$$

$$309 \quad HQ_{Mi\_DM} = \frac{ADD_{Mi\_DM}}{RfD_{Mi}} \quad (7)$$

310 
$$HQ_{RNP\_DM} = \frac{ADD_{RNP\_DM}}{RfD} \quad (8)$$

311 
$$HQ = \frac{EC}{PNEC} \quad (9)$$

312 
$$HI = HQ_{antibiotics} + HQ_{metal\ ions} \quad (10)$$

313 
$$HI = HQ_{antibiotics} + HQ_{NPs} \quad (11)$$

314 
$$HI = HQ_{metal\ ions} + HQ_{NPs} \quad (12)$$

315

## 316 2.5 Risk management

317 Maximum allowable concentration ( $C_{max}$ ) can be defined as the concentration beyond which  
318 no adverse effects or risk exposure can occur.  $C_{max}$  specifies the upper limit values of  
319 substance under study and can provide a helping hand to regulatory agencies for managing  
320 the risk. To calculate the  $C_{max}$  values, the HQ values in Eq. (6-9) was set as 1, and the  
321 concentration was calculated.

322

## 323 3 Result and discussion

### 324 3.1 Estimation of nanoparticles and antibiotics loading in the digestive media

325 The results revealed that  $Fe_3O_4$  NPs showed high bio-accumulation in the digestive system  
326 with a value of  $2.58 \times 10^{-3}$  mg/L. The comparative analysis indicated that  $Fe_3O_4$  NPs has the  
327 highest accumulation in the digestive media which was followed by ZnO NPs ( $5.77 \times 10^{-8}$   
328 mg/L), CuO NPs ( $2.40 \times 10^{-9}$  mg/L) and  $TiO_2$  NPs ( $7.19 \times 10^{-10}$  mg/L). Overall,  
329 concentration of  $Fe_3O_4$  NPs was found to be highest amongst all the nanoparticles considered  
330 for this study. High accumulation of nanoparticles like ZnO and CuO NPs in the digestive  
331 system can be related to their size. Larger the size of nanoparticles higher is the accumulation

332 potential in the human digestive system (Bergin and Witzmann, 2013). The observed  
333 sequence of nanoparticles in the human digestive system is similar to those observed by  
334 Parsai and Kumar (2020).

335 Similar to nanoparticles, the concentration of antibiotics in the human digestive media  
336 ( $C_{AB\_DM}$ ) after nanoparticles dissolution was also calculated to determine the accumulation  
337 potential of antibiotics in the human digestive system. The concentration of levofloxacin  
338 cannot be determined due to the lack of adsorption isotherm data. Hence, as a result further  
339 risk assessment studies were not carried out although, it can be done only if the data is  
340 available. Amongst the antibiotics considered, ofloxacin ( $1.8 \times 10^{-3}$  mg/L) showed highest  
341 accumulation potential in the human digestive system next to ciprofloxacin ( $3.87 \times 10^{-7}$   
342 mg/L) and norfloxacin ( $4.44 \times 10^{-7}$  mg/L). Wingender et al. (1985) in his study reported rapid  
343 absorption of ciprofloxacin in the upper GI-tract however comprehensive information on the  
344 absorption of ciprofloxacin or any other FQ in different parts of the human GI-tract does not  
345 exist (Harder et al., 1990). Furneri et al. (2000) suggested that accumulation of FQs  
346 antimicrobial agents is reduced by lowered pH and, under some conditions, by divalent  
347 cations as well.

348

## 349 3.2. Risk estimation

350 *3.2.1 Risk estimation of NPs in the digestive media followed by GI-tract absorption when the*  
351 *isotherm data is available*

352 3.2.1.1 Risk estimation of metal ions released in the digestive media after dissolution of  
353 nanoparticles

354 The results revealed that the HQ values for all types of metals ions released in the human  
355 digestive system after nanoparticle dissolution were observed to be less than 1 under the  
356 conditions assumed in this study. This indicated that the metal ions released from the  
357 nanoparticles in the human digestive system does not show any significant health risks to  
358 human health [HQ values ranged from  $1.44 \times 10^{-11}$  (for  $\text{Ti}^{2+}$  ions) to  $2.21 \times 10^{-4}$  (for  $\text{Fe}^{3+}$   
359 ions)]. HQ values more than 1 shows possible health concerns.

360

361 3.2.1.2 Estimation of risks due to the revised concentration of nanoparticles in the digestive  
362 media after dissolution of nanoparticles

363 HQ values calculated for the revised concentration of nanoparticles in the human digestive  
364 system were in the sequence of (low to high): ZnO NPs,  $3.43 \times 10^{-8}$ ;  $\text{TiO}_2$  NPs,  $3.19 \times 10^{-7}$ ;  
365 CuO NPs,  $2.28 \times 10^{-5}$ ;  $\text{Fe}_3\text{O}_4$  NPs,  $2.85 \times 10^{-3}$  (Table 2). The results showed that the  
366 estimated HQ values for the revised concentration of nanoparticles were smaller than 1, the  
367 acceptable risk level, indicating no significant risks to human health. Amongst the  
368 nanoparticles considered,  $\text{Fe}_3\text{O}_4$  NPs showed comparatively high HQ values than other  
369 nanoparticles. It is important here to mention that till the time of writing no such studies (as  
370 per the author's best knowledge) has been reported in literature to determine risks for the  
371 released concentration of metal ions as well as nanoparticles in the human digestive system.  
372 It is also not known how the released metal ions behave after their dissolution from  
373 nanoparticles in the digestive media. Moreover, no guidelines or recommendations is  
374 available on the use of nanomaterials as a coating agent for drug delivery. In spite of being  
375 used in numerous biomedical and other industrial uses, the safety, toxicity, and its interaction  
376 with and within the biological systems are still unclear (Snyder-Talkington et al. 2012; Zhong  
377 et al., 2017).

379 **Table 2** Summary of the estimated risk of nanoparticles released in the digestive media

<b>Metal ions released in the digestive media</b>	<b>Risk of metal ions in the digestive media (<math>HQ_{Mi}</math>)</b>	<b>Nanoparticles</b>	<b>Risk of revised concentration of nanoparticles in the digestive media (<math>HQ_{R\_NP}</math>)</b>
$Zn^{2+}$	$1.15 \times 10^{-8}$	ZnO	$3.43 \times 10^{-8}$
$Ti^{2+}$	$1.44 \times 10^{-11}$	TiO <sub>2</sub>	$3.19 \times 10^{-7}$
$Fe^{3+}$	$2.21 \times 10^{-4}$	Fe <sub>3</sub> O <sub>4</sub>	$2.85 \times 10^{-3}$
$Cu^{2+}$	$3.59 \times 10^{-9}$	CuO	$2.28 \times 10^{-5}$

380

381 3.2.1.3. Risk of revised concentration of antibiotics in the digestive system after release from  
382 nanoparticles and absorption in GI-tract

383 HQ values estimated using the revised concentration of antibiotics in the human digestive  
384 system ranged from  $1.90 \times 10^{-12}$  (NOR) to  $3.37 \times 10^{-8}$  (OFX). The observed values indicate  
385 that no possible risks exist to human health as the obtained HQ values were less than the  
386 acceptable risk level. Amongst all, it was observed that CIP concentration may pose risks to  
387 human health due to low values. Overall, the results of risk evaluations disclosed that there  
388 exist no risks to human health for the consumption of nanoparticles coated antibiotics though  
389 oral ingestion of water under the conditions assumed in this study and for the scenario's  
390 considered. Although the estimated risk values were below the acceptable risk level but still

391 the amount of nanoparticles to be used as a coating material for antibiotics needs to be  
392 regulated prior to its use for human.

393

### 394 *3.3. Risk estimation when the adsorption isotherm data is available*

#### 395 3.3.1 No absorption

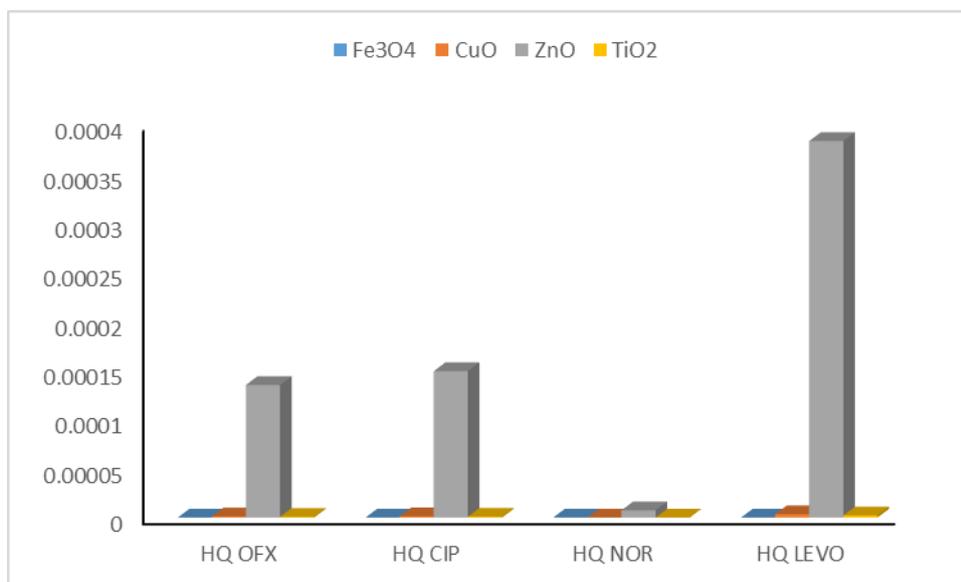
396 The term "No absorption" implies that under the hypothetical scenario considered no  
397 adsorption of antibiotics on nanoparticles takes place. As mentioned in the exposure  
398 assessment section, in this case, risk was estimated only for the oral ingestion of surface  
399 water contaminated with antibiotics. PNEC values were estimated to determine risk exposure  
400 effects. PNECs provides more accurate estimates and can be refined using detailed  
401 information considering different assessment factors (Bopp et al., 2019). The calculated  
402 PNEC values were observed to be (high to low): norfloxacin (233.80 µg/L) > ofloxacin  
403 (53.44 µg/L) > ciprofloxacin (26.72 µg/L) > levofloxacin (2.505 µg/L). HQ values ranged  
404 from  $1.68 \times 10^{-3}$  (norfloxacin) to  $4.3 \times 10^{-2}$  (levofloxacin) and are observed to less than the  
405 acceptable risk level ( $HQ < 1$ ), therefore does not pose any significant risks to human health.  
406 An inverse relation was observed between PNEC and HQ values. Lower the PNEC values,  
407 higher the risk will be. Moreover, the PNEC values is directly dependent on the concentration  
408 of antibiotics in the environmental media. The results presented in this study is somehow  
409 similar to those reported by the authors (Kumari and Kumar, 2020a) in their previous work  
410 on sulfamethoxazole, ampicillin, and amoxicillin. Even though the presence of antibiotics in  
411 the water environment pose insignificant risks to human health, the risk management of these  
412 substances is required so as to protect human beings from their detrimental effects, if any. A  
413 thorough monitoring of these antibiotics in water bodies is required to protect the human  
414 health.

415

416 3.3.2 100% absorption

417 The term "100% absorption" implies that under the hypothetical scenario, it was assumed that  
418 total adsorption of antibiotics by nanoparticles occurs. Figure 2 shows the HQ values of  
419 individual antibiotics for different nanoparticles. As can be seen, the HQ values for all the  
420 antibiotics were observed to be less than 1, the acceptable risk, indicating no significant risks  
421 to human health. Amongst the NPs studied, high risk values were observed for ZnO NPs  
422 compared to others. Similar to the other cases, here also, absorption of ZnO NPs in the  
423 human digestive system was found to be the highest. Parsai and Kumar (2020) found high  
424 HQ values for ZnO NPs thorough fish consumption exposure. The toxicity of ZnO  
425 nanoparticles might be due to their solubility. It is reported that dissolution of ZnO  
426 nanoparticles takes place in the extracellular region, which in turn increase the level of  
427 intracellular Zn<sup>2+</sup>. However, the mechanism behind the increased level of intracellular Zn<sup>2+</sup>  
428 ions and dissolution of ZnO nanoparticles in the medium is still unclear (Pandurangan and  
429 Kim, 2015).

430



431

432 **Fig. 2** HQ values of nanoparticles for nanoparticles

433

434 This study estimated risks due to the exposure of single type of nanoparticle at-a-time.  
435 The chance assembly of more than one type of nanoparticles or antibiotics was not  
436 considered due to lack of information on the BAF values of mixture of NPs, dissolution and  
437 absorption rate of mixture of antibiotics in the GI-tract, dissolution of nanoparticles to their  
438 corresponding ions in the digestive system and RfD values. The present study showed an  
439 example of determining risk estimates in digestive system after including all the important  
440 parameters and highlighted its significance for human health. This type of risk assessment  
441 studies has not been conducted and this is the first time an attempt has been made to predict  
442 risk of contaminants within the human digestive system. The developed framework will help  
443 in assessing risk effects of the exposure of NP coated ABs to human in the digestive system.  
444 Though, few studies have estimated risks due the exposure of nanoparticles (Pizzol et al.,  
445 2019; Yang et al., 2017) and antibiotics (Kumari and Kumar, 2020a; 2020b) but none of them  
446 have considered the effect of human digestive media on the fate of nanoparticles and  
447 antibiotics including the reference dose values of NPs. The proposed framework included the  
448 effects of dissolution of nanoparticles to their corresponding ions in digestive system and  
449 estimated the revised concentration of antibiotics in the digestive media so as to provide more  
450 realistic values of both nanoparticles and antibiotics in the human digestive system. The  
451 suggested framework provides a step-wise approach to determine the risks exposure effects  
452 of NPs and ABs-complexes in the human digestive system.

453

454 3.4 HI interactions

455 Under the conditions assumed in this study, the NPs coated ABs after oral ingestion  
456 undergoes absorption in the GI-tract and dissolved to produce different fractions as  
457 mentioned in the exposure assessment section, therefore, it is essential to calculate the risk of  
458 these released components in the human digestive system to provide more realistic risk  
459 estimates. For case (i) HI interaction of antibiotics with metal ions, the results revealed that  
460 the interactions of antibiotics with metal ions in the human digestive system does not pose  
461 any risks to human health as the estimated HI values were less than 1 ( $HI_{int}$  for antibiotics-  
462 metal ion pairs  $< 1$ ) for all the conditions assumed in this study. In case (ii) HI interaction of  
463 antibiotics with nanoparticles, the calculated  $HI_{int}$  values were also observed to be less than 1,  
464 the acceptable risk level and therefore, does not pose any risk to human health, if present  
465 together. For case (iii) HI interaction of antibiotics with metal ions, similar to the results  
466 obtained for the above two cases, here also the interaction results were smaller than 1,  
467 indicating no significant health risks to human health. Overall, it was observed that no  
468 significant health was observed for the three mixture pairs as mentioned above. Although the  
469 HI interaction values were observed to be less than the acceptable risk level, still guideline  
470 values needs to be developed so as to regulate the amount and use of nanoparticles in targeted  
471 drug delivery systems as even a nano-gram increase in their concentrations might show  
472 adverse effects and can be detrimental to human health. Supplementary Table S5 shows  
473 information about the results obtained for all the cases and combinations mentioned above.

474

475 3.5 Maximum allowable concentration of metal ions, nanoparticles, and antibiotics in water  
476 Table 3 provides the calculated maximum allowable concentration for antibiotics, metal ions,  
477 and nanoparticles.  $C_{max}$  values of metal ions at which no risks occur ranged from (high to  
478 low)  $6.68 \times 10^5$  mg/L (for  $Ti^{2+}$  ions) to 6.27 mg/L (for  $Zn^{2+}$  ions). The estimated  $C_{max}$  values

479 of nanoparticles at which no health risks effects were observed ranged from 0.658 mg/L  
480 (ZnO NPs) to  $6.68 \times 10^5$  mg/L (TiO<sub>2</sub> NPs). Amongst the NPs studied, ZnO NPs showed  
481 highest risk, and therefore, the C<sub>max</sub> values for ZnO NPs was found to be lowest followed by  
482 Fe<sub>3</sub>O<sub>4</sub>, CuO, and TiO<sub>2</sub> NPs, respectively. If we compare the C<sub>max</sub> values of metal ions with  
483 that of nanoparticles, it can be seen that the Zn<sup>2+</sup> ions (0.658 mg/L) showed comparatively  
484 high C<sub>max</sub> values than ZnO NPs (6.27 mg/L). Previous studies also reported high C<sub>max</sub> values  
485 for ZnO NPs in water bodies for the inadvertent ingestion of NPs through fish consumption  
486 exposure (Parsai and Kumar, 2020). Similar C<sub>max</sub> values for Ti<sup>2+</sup> ions and TiO<sub>2</sub> NPs, and Fe<sup>3+</sup>  
487 ions and Fe<sub>3</sub>O<sub>4</sub> NPs as can be seen from Table 3 is due to the use of similar RfD values for  
488 metal ions and NPs (due to unavailability of RfD values for these NPs) during risk  
489 estimation. Therefore, there is a need for conducting *in vivo* and *in vitro* studies for  
490 determining RfD of nanoparticles to get accurate risk estimates. The observed results  
491 demonstrated that stern actions and control measures must be taken to reduce the risk  
492 exposure effects of metal ions and nanoparticles.

493 The C<sub>max</sub> values of antibiotics were also calculated to determine their allowable  
494 concentration in water systems. The study observed that C<sub>max</sub> of antibiotics beyond which no  
495 risk effects can occur was observed to be 2.5 µg/L (for levofloxacin), 53.44 µg/L (for  
496 ofloxacin), 26.72 µg/L (for ciprofloxacin), and 233.80 µg/L (for norfloxacin). On the basis of  
497 C<sub>max</sub> values, it can be said that levofloxacin pose maximum risk to children whereas  
498 norfloxacin shows minimum risk. Different C<sub>max</sub> values of antibiotics have been reported by  
499 researchers (Lubasch et al., 2000; Owen et al., 1997) which might be related to the  
500 administered dose of antibiotics and age of the population considered. It was observed that an  
501 inverse relationship exists between the maximum allowable concentration and hazard  
502 quotient values i.e., the lower the C<sub>max</sub> values the higher the risk will be and vice versa. The  
503 results obtained in this study can be used by the regulatory bodies like USEPA, OECED and

504 WHO for setting up the guidelines values for metal ions, nanoparticles, and antibiotics in  
 505 water.

506 **Table 3** Maximum allowable values ( $C_{\max}$ ) of metal ions, nanoparticles and antibiotics  
 507 assuming HI values as 1; lowest values are shown in bold text and are italicised

<b>Metal ions (mg/L)</b>				
Maximum allowable concentrations, $C_{\max}$	<b><i>Zn<sup>2+</sup>, 6.27</i></b>	Cu <sup>2+</sup> , $2.23 \times 10^3$	Fe <sup>3+</sup> , $4.79 \times 10^{1*}$	Ti <sup>2+</sup> , $6.68 \times 10^{5\#}$
<b>Nanoparticles (mg/L)</b>				
Maximum allowable concentrations, $C_{\max}$	<b><i>ZnO, 0.658</i></b>	CuO, $1.46 \times 10^3$	Fe <sub>3</sub> O <sub>4</sub> , $4.79 \times 10^{1*}$	TiO <sub>2</sub> , $6.68 \times 10^{5\#}$
<b>Antibiotics (µg/L)</b>				
Maximum allowable concentrations, $C_{\max}$	<b><i>Levofloxacin, 2.51</i></b>	Ofloxacin, 53.44	Ciprofloxacin, 26.72	Norfloxacin, 233.80

508

509 **4 Effect of assumptions used on risk values**

510 Due to lack of information on parameters used to determine risk estimates due to exposure of  
511 nanoparticles and antibiotics from natural waters, a lot of assumptions were made to fill the  
512 data gaps which includes (i) concentration of NPs: to calculate risk estimates of due to the  
513 interaction of nanoparticles with antibiotics, this study assumed the nanoparticle  
514 concentration to be 1 mg (NPs) per mg of antibiotics (data taken from literature) to determine  
515 the concentration of antibiotics adsorbed on nanoparticles (mg/L). Under this assumption, it  
516 was observed that the estimated risk values for all the conditions studied does not pose any  
517 concerns to human health. However, risk estimates might vary as it directly depends on the  
518 concentration of substances used for the study (Kumari and Gupta, 2018) (ii) RfD values:  
519 Due to unavailability of reference dose values for nanoparticles considered in this study (as  
520 mentioned in the exposure assessment section), the RfD values of metal ions were taken  
521 (assuming that the values are equivalent and similar to that of nanoparticles) to estimate the  
522 risk values. At the assumed RfD values, no exists to human health for the conditions studied.  
523 The observed risk estimates might show different results, if estimated using accurate RfD  
524 values of nanoparticles (iii) BAF values: it determines the behavior of nanoparticles in  
525 solution. Transformation and actions of nanoparticles within the human digestive system is  
526 hard to anticipate (iv) ADI values of antibiotics: the values are taken from literature.  
527 Different values of the selected antibiotics have been reported by Wang et al. (2018) and  
528 Hanna et al. (2018) which creates a dilemma on which values to take for determining risk  
529 estimates, and thereby creating uncertainty in the overall process (v)  $B_{ij}$  values: The study  
530 assumed the  $B_{ij}$  values as 1 due to lack of information on the interaction of antibiotics with  
531 nanoparticles or metal ions. The risk estimates presented in this study indicate the point  
532 estimate values and can vary depending on variability of these parameters, creating  
533 uncertainty in risk estimates. Uncertainty analysis using Monte Carlo simulations needs to  
534 performed to overcome these issues in risk estimates. Moreover, it is essential to recognize

535 those parameters which adds high variability in HQ estimation so that efforts could be taken  
536 for reducing their variability in risk estimation process.

537

## 538 **5 Implications of the proposed framework**

539 This study proposed a framework using six-step risk assessment approach to determine the  
540 risk exposure effects of nanoparticles and antibiotics followed by oral ingestion and their  
541 possible interaction within the human digestive system. This study provides a systematic  
542 information on risk assessment involving the fate of nanoparticles and antibiotics within the  
543 human digestive system. Agencies like USEPA and FDA have suggested use of alternative  
544 testing strategies and data requirement for nanoparticles (Aschberger et al., 2016), however a  
545 systematic approach in dealing with nanoparticles and antibiotics does not exist. The  
546 proposed framework can be used by regulatory bodies such as USEPA, EU agencies and  
547 OECD for monitoring of nanoparticles and antibiotics in water. The outcome of the study  
548 will help in identifying the possible concentration of released compounds (antibiotics,  
549 nanoparticles, metal ions) in the human digestive system and the data generated can be used for  
550 formulating guideline limits of both antibiotics and nanoparticles. Studies by Parsai and  
551 Kumar (2021) on nanoparticles and Kumari and Kumar (2020) has developed risk assessment  
552 framework for determining risk exposure effects alone and in mixture combination however  
553 none of the reported study analysed the interaction between these two contaminants. In this  
554 regard, this study will be helpful in understanding the interaction as well as the fate and  
555 behaviour two contaminants. Besides, the maximum allowable concentration values of  
556 contaminants derived in the study can be used by the regulatory bodies to regulate the  
557 concentration of antibiotics, nanoparticles and metal ions in natural water systems.

558

## 559 7 Summary and Conclusions

560 The major findings of the study are presented below:

- 561 • The present study proposed a framework for assessing health risk exposure effects caused  
562 due to the interaction of nanoparticles and antibiotics followed by oral administration.  
563 The developed framework was applied to a hypothetical scenario where environmentally  
564 occurring concentration of nanoparticles ( $\text{Fe}_3\text{O}_4$ ,  $\text{ZnO}$ ,  $\text{CuO}$ , and  $\text{TiO}_2$ ) and antibiotics  
565 (levofloxacin, ofloxacin, ciprofloxacin, norfloxacin) were taken for illustrative purpose.
- 566 • The study estimated the loading of antibiotics and nanoparticles in the human digestive  
567 system after their release. Amongst the nanoparticles,  $\text{Fe}_3\text{O}_4$  NPs ( $2.58 \times 10^{-3}$  mg/L)  
568 presented maximum accumulation in the digestive media whereas  $\text{TiO}_2$  NPs the minimum  
569 ( $7.19 \times 10^{-10}$  mg/L). Similarly, for antibiotics, ofloxacin has the highest accumulation rate  
570 in the human digestive system ( $1.8 \times 10^{-3}$  mg/L).
- 571 • The risk estimated for two different scenarios showed hazard quotient values less than 1  
572 under the conditions and assumptions made in this study. Therefore, on the basis of  
573 results obtained it can be said that the interaction of two contaminants does not pose any  
574 risks to human health followed by their release and dissolution in the human digestive  
575 system.
- 576 • Mixture toxicity (HI interactions) studies was conducted for three different binary  
577 combinations (i) antibiotics with metal ions, (ii) antibiotics with nanoparticles, and (iii)  
578 metal ions with nanoparticles. The estimated HI values for all the mixture combinations  
579 was observed to be less than 1, the acceptable limit, and therefore indicated no significant  
580 risks to human health. However, more detailed studies on the interactions antibiotics with  
581 metal ions and nanoparticles is required ( $B_{ij}$  and  $M_{ij}$  values in the WoE approach) for  
582 accurate risk predictions.

583 • Overall, this work significantly increases our understanding on the fate of nanoparticles  
584 as well as antibiotics in the human digestive system and provides the knowledge base for  
585 better assessment of risk estimates of the studied contaminants in natural water systems.  
586 Efforts are required for conducting proper *in vitro* and *in vivo* eco-toxicity studies so that  
587 better understanding can be made on the fate and behaviour of released fragments of  
588 nanoparticles and antibiotics in the human digestive system.

589

#### 590 **Data availability**

591 All the data supporting the results reported in the article are included in the manuscript and  
592 can be found in supplementary file. Data sharing is not applicable to this article as no datasets  
593 were generated or analyzed during the current study.

594

#### 595 **Declarations**

596 Ethics approval and consent to participate: Not applicable

597 Consent for Publication: Both the authors have their consent for publishing the manuscript

598 Availability of data and material: Not applicable

599 Competing interests: Not applicable

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607

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# Figures

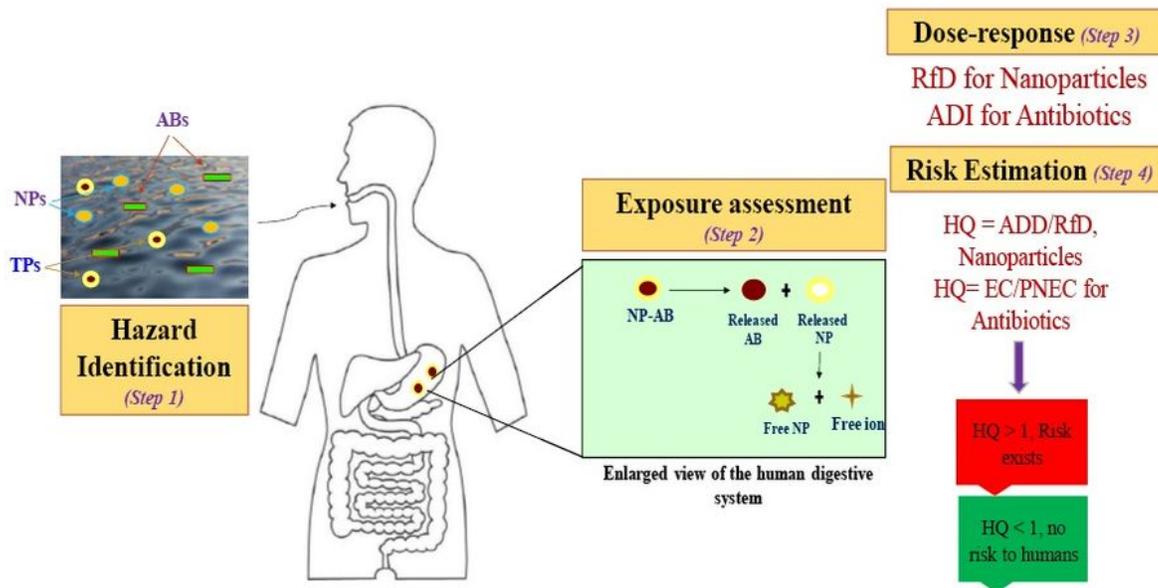


Fig. 1

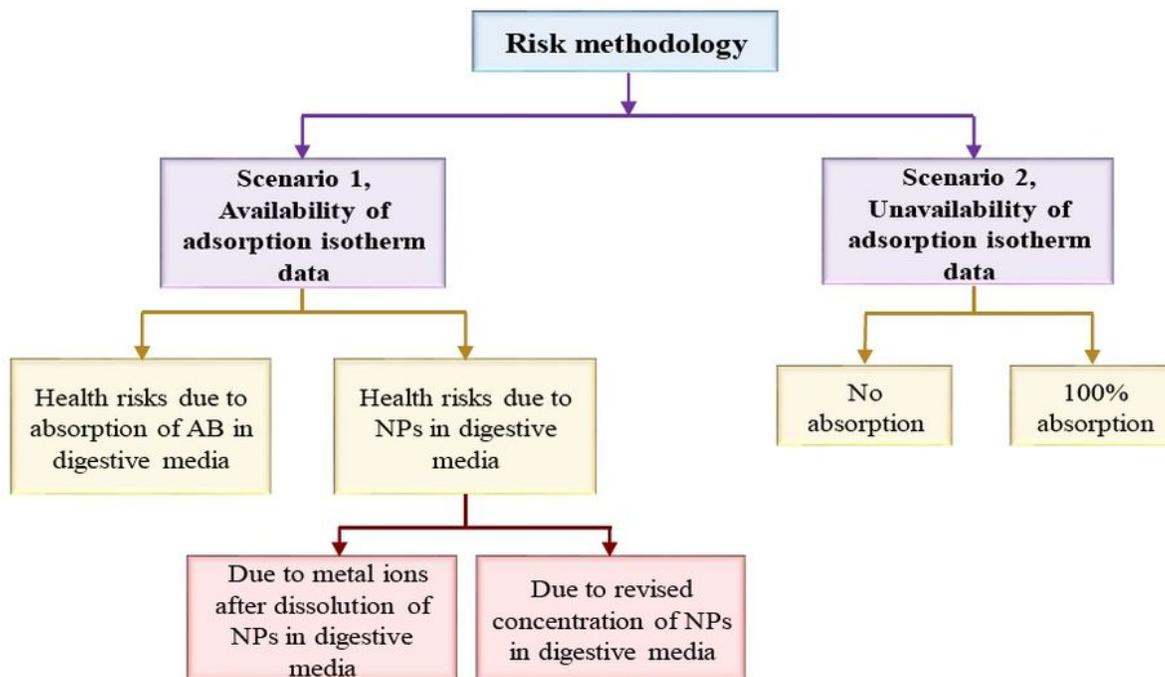
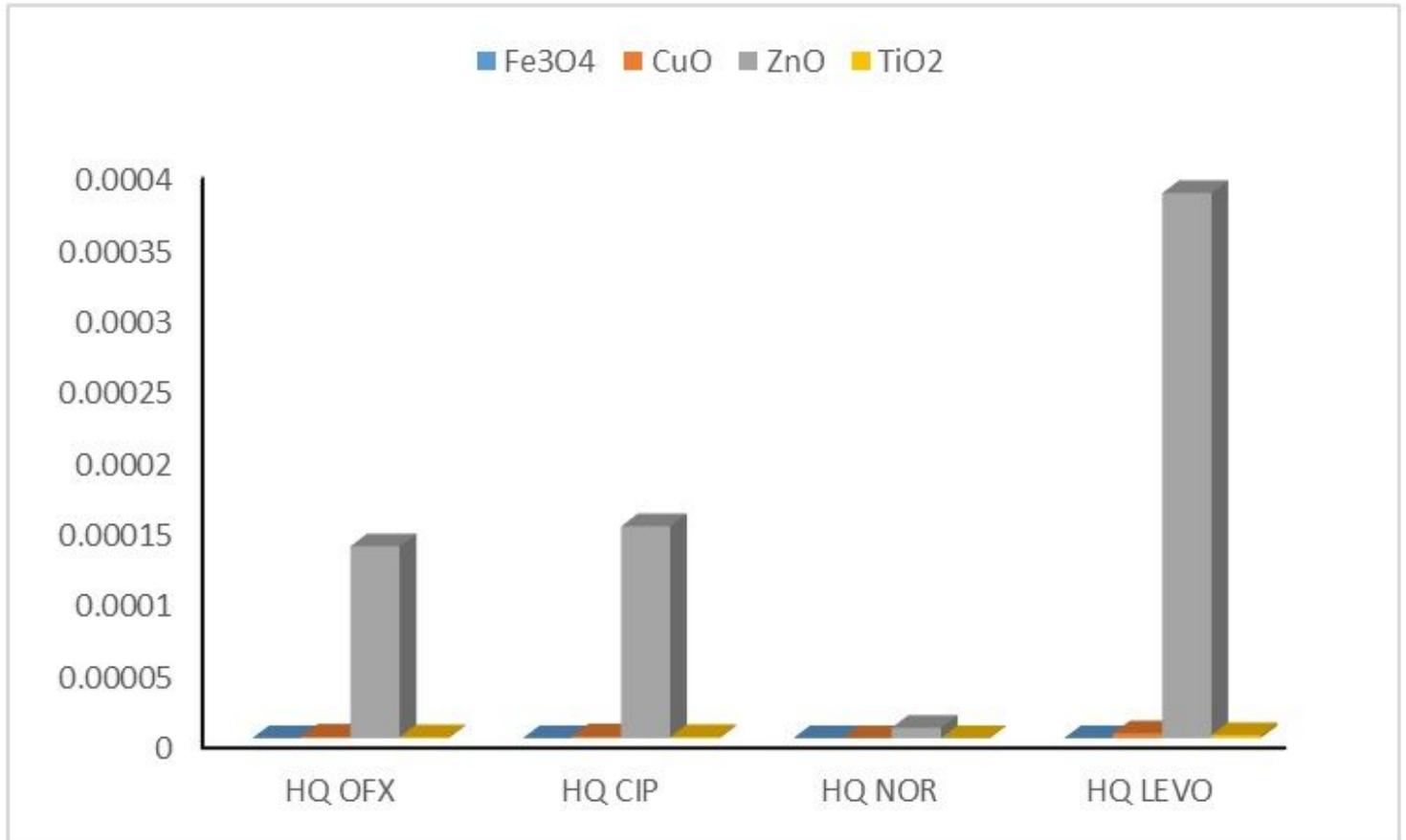


Fig. 1a

Figure 1

Hypothetical schematic diagram showing the uptake of nanoparticles-antibiotics transformed products and their dissolution in the human digestive system (NPs = nanoparticles, ABs = antibiotics, TPs = transformed products, HQ = hazard quotient, EC = Environmental concentration, PNEC = Predicted no-

effect concentration; RfD = Reference dose; ADI = Acceptable daily intake). 1a Scenarios considered for determining risk exposure to humans



**Figure 2**

HQ values of nanoparticles for nanoparticles

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