

Effects of Magnetically Treated *Sedum Alfredii* Seeds On The Dissolved Organic Matter Characteristics of Cd-Contaminated Soil During Phytoextraction

Youjun Tang

Yangtze University

Shuaizhi Ji

sinopec Zhongyuan oilfield

Dan Chen

Yangtze University

Jiawei Wang

Yangtze University

Min Cao

University of Leicester

Jie Luo (✉ gchero1216@hotmail.com)

Yangtze University <https://orcid.org/0000-0001-6543-9346>

Research Article

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Abstract

The effects of magnetic field treatments on the two determining factors of phytoremediation, growth status and element uptake capacity, of *Sedum alfredii* have been thoroughly studied; however, minimal studies have been performed to determine the influence of the Cd hyperaccumulator *Sedum alfredii*, grown from magnetically treated seeds, on the dissolved organic matter (DOM) characteristics in its rhizosphere. A series of pot experiments were conducted to evaluate the variations in the DOM concentration and fractionations in the rhizosphere of *S. alfredii* treated with external magnetic fields. Compared with the untreated seeds, *S. alfredii* grown from magnetically treated seeds excreted more DOM in its rhizosphere. Additionally, the hydrophilic DOM fractionation proportion, which presented a greater capacity to mobilize Cd in the soil, increased from 42.7 % in the control sample to 47.2 % in the 150 mT magnetically treated *S. alfredii* sample. The water-soluble and exchangeable forms of Cd (extracted using deionized water and NH_4NO_3 , respectively) in the rhizosphere of the magnetically treated *S. alfredii* were significantly lower than those of the control sample. Furthermore, the Cd extraction capacity of DOM from the rhizosphere of the magnetically treated *S. alfredii* was greater than that of the control sample, thereby increasing the Cd uptake ability of the magnetically treated species. These results suggest that the secretion of large amounts of DOM, especially acid and hydrophilic fractionations, is an essential mechanism of magnetically treated *S. alfredii* to mobilize Cd in the soil.

1. Introduction

Cd, a well-known trace inorganic contaminant with high bioavailability and a long biological half-life, has detrimental effects on human health (Xiao et al., 2017). Cd-polluted soil decontamination using conventional technologies such as soil washing, excavation, in-situ solidification, and electrokinetic remediation is expensive and could damage the physical structure and ecological system of the soil (Komínková et al., 2018).

Phytoextraction, the application of hyperaccumulators to remove pollutants from contaminated sites via plant uptake and migration to harvestable tissues, is an economical, ecologically friendly alternative to the traditional physical and chemical decontamination methods of contaminated sites. Various hyperaccumulator types, including *Noccaea caerulescens* for Cd and Zn (Escarre et al., 2000), *Pteris vittata* for As (Gonzaga et al., 2009), *Arabis paniculate* for Pb (Tang et al., 2009), and *Phytolacca acinosa* for Mn (Xue et al., 2004) have been identified and studied to evaluate the mobilization, extraction, transportation, hyperaccumulation, and detoxification mechanisms of contaminants in these plants. The major constraint restricting the widespread utilization of hyperaccumulators for soil remediation is the low biomass generation abilities of these plants (Li et al., 2018).

Improving the dry weight and metal uptake efficiency of hyperaccumulators is critical for phytoextraction in metal-polluted fields; therefore, a thorough understanding of hyperaccumulator mechanisms that activate and accumulate metals is necessary. Previous studies have suggested that the application of an appropriate external magnetic field can enhance the phytoremediation efficiencies of *Eucalyptus globulus*

(Luo et al., 2019a), *N. caerulea* (Luo et al., 2019b), and *Festuca arundinacea* (Luo et al., 2020). For instance, an appropriate magnetic field can improve biomass yield and metal uptake efficiency by activating membrane transporters in plant tissues, providing energy for Mg^{2+} to biosynthesize chlorophyll, and enhancing the antioxidant enzyme activities in plants to scavenge excessive free radicals as a result of balancing the endogenous biological magnetic field.

Root-soil interactions can also modulate the metal activity in the rhizosphere and consequent alterations in the metal uptake abilities of plants. Many factors, including pH, cation exchange capacity, alkalinity, organic matter, microbial activity, and moisture content, can influence metal activity in the rhizosphere. The variation in pH in hyperaccumulator-growing soils has been thoroughly studied; however, inconsistent results have been reported. For example, Loosemore et al. (2004) found that the pH in the *Nicotiana tabacum* rhizosphere significantly decreased after growth occurred; however, McGrath et al. (1997) found that the pH in the *N. caerulea* rhizosphere did not change, and Kim et al. (2010) reported that rhizosphere pH increased by 1.3 after *B. juncea* cultivation. These inconsistent results suggest that pH reduction in the rhizosphere is not the only metal activating mechanism in the rhizosphere by accumulators.

Previous studies suggest that dissolved organic matter (DOM) controls metal fractionation and bioavailability rather than pH when the latter is higher than 6.5; therefore, DOM features in the rhizosphere should be considered during the phytoremediation process (Christensen and Christensen, 2000). DOM that can pass through a 0.45- μ m filter has an affinity for bonding with metals, thus increasing metal bioavailability in the rhizosphere (Beiyuan et al., 2018; Bradney et al., 2019; El-Naggar et al., 2018). Kim et al. (2010) reported that plants could spontaneously adjust their rhizosphere to improve the uptake of beneficial elements by various mechanisms, such as acidification, lowering of oxidation-reduction potential, and organic matter excretion. Furthermore, hyperaccumulators secrete more DOM into their rhizosphere when subjected to high-metal environments, resulting in a decreased extraction rate of the corresponding metal hyperaccumulation by these plants (Guilpain et al., 2018; Tao et al., 2020). Based on previous studies on the major DOM characteristics and its impact on metal bioavailability in the rhizosphere, it has been determined that DOM can significantly reduce Cu, Pb, Ni, and Cd adsorption by soil particles, which increases the solubilities of these metals (Li et al., 2013c). However, the effects of magnetically treated plants on DOM fractionation and metal bioavailability are not thoroughly understood; therefore, further evaluation is necessary to determine their effects on the rhizosphere of these plants during the phytoextraction process.

S. alfredii, a Crassulacean acid metabolism species found in South China, is a candidate for phytoextraction in Cd-polluted fields owing to its physiological features (Li et al., 2013b). Previous studies reported that an appropriate external magnetic field stimulated the growth rates of hyperaccumulators and high-biomass producing plants and improved their metal uptake capacities (Luo et al., 2019a, 2019b, 2020). Therefore, magnetic field application could alter the DOM characteristics in the rhizosphere of *S. alfredii*. The primary goals of this study were to evaluate the effects of magnetic field treatment on DOM concentrations and fractionations in the rhizosphere of *S. alfredii*, investigate the Cd extraction capacity

of DOM from the rhizosphere of magnetically treated *S. alfredii*, and verify the feasibility of using magnetic fields to improve the phytoremediation efficiency of *S. alfredii*.

2. Materials And Methods

2.1 Plant and soil materials

S. alfredii seeds were gathered from a lead-zinc mine in Shangfang Town, Quzhou City, China, and separated into four groups. Three groups were placed in non-magnetic containers and treated with a 50, 100, and 150 mT intensity static magnetic field for 20 min each day for one week, and the fourth group was exposed to a geomagnetic field and used as the control. Each treatment had five replicates. A field generator was used to produce the static magnetic field, and intensity stability was monitored using a Gauss magnetometer. After pre-sowing treatment, all plants were cultivated for 14 days under greenhouse conditions in a nutrient solution (Lu et al., 2008).

The heavily contaminated soils used in this experiment were obtained from the top layer (20 cm) of a well-known electronic waste recycling center, Guiyu, located in south China, in which the soils have been heavily polluted owing to electronic waste dismantling activities using primitive recycling methods (Jiang et al., 2019). After air-drying, the soils were ground and sieved using a 2-mm nylon mesh. Four rounds of the wetting and drying processes were performed to equilibrate the soil, and then the mixed soil was divided into 6 kg aliquots and stored in PVC pots for *S. alfredii* transplantation. The Cd concentrations in the homogenized soil were $3.83 \pm 0.46 \text{ mg kg}^{-1}$.

Four pre-cultured seedlings were transplanted into each pot. During the growth process, soil moisture was maintained at 70 % field capacity using deionized water and measuring the weights every day. Ninety days after the start of the experiment, all plants were carefully harvested, and rhizosphere soils from each pot were obtained by shaking the roots (Luster et al., 2009). The soil pH was measured before and after transplantation using a pH meter in a 1:2.5 soil-water suspension.

The harvested plant tissues were washed with tap water to remove foreign materials and then soaked in a 10 mM EDTA- Na_2 solution to eliminate adsorbed ions. The cleaned tissues were dried in a 70°C oven until a constant weight was achieved.

2.2 DOM extraction and fractionation

The DOM in the rhizosphere of *S. alfredii* was extracted following the procedures suggested by Jones and Willett (2006) with minor modifications, wherein 2.5 g of the rhizosphere soil was mixed with 25 mL of distilled water and shaken at 200 rpm for 120 min. The extracts were centrifuged at $10,000 \times g$ for 20 min, and the recovered supernatant was sieved using 0.45- μm filters. A total organic carbon analyzer (TOC-5000, Shimadzu, Japan) was used to measure the dissolved organic carbon in the filtrate. Leenheer and Croue (2003) developed a dissolved organic carbon fractionation method that classified dissolved organics into six fractionations, including hydrophilic acid, hydrophilic base, hydrophilic neutral,

hydrophobic acid, hydrophobic base, and hydrophobic neutral based on their polarity, alkaline-acidic properties, and specific compound characteristics. The DOM fractionation was performed according to the method suggested by Leenheer and Croue (2003), and the six DOM fractions were measured using the total organic carbon analyzer (TOC-5000, Shimadzu, Japan).

2.3 Cd extraction capacity of DOM

Cations in the extracted DOM were removed using an Amberlite cation exchange H resin (Sigma-Aldrich, USA). After dilution to 100 mg L^{-1} with deionized water, 20 mL of the solution was mixed with 2 g of soil in a centrifugal tube and centrifuged at $8000 \times g$ for 30 min. After being sieved through a $0.45\text{-}\mu\text{m}$ membrane, the Cd concentration in the recovered supernatant was measured using an inductively coupled plasma mass spectrometer (ICP-MS) (Agilent 7700, USA). The soil samples (2 g) were mixed with deionized water and prepared using the same procedure as the control.

2.4 Cd analysis

The dried plant and rhizosphere soil samples were ground and sieved using a $74\text{-}\mu\text{m}$ nylon mesh. The plant samples were digested through the aqua regia dissolution method (Ok et al., 2011), wherein the prepared plant samples were digested in a solution of HNO_3 and HCl (1:3 ratio) and heated for 100 min at 130°C . After cooling, the digestion was filtered through a $0.45\text{-}\mu\text{m}$ membrane, and the Cd concentration in the recovered filtrate was analyzed via ICP-MS (Agilent 7700, USA).

The water-soluble and exchangeable fractions of Cd in the rhizosphere were determined according to the methodology described in a previous study (He et al., 2020), wherein 1.2 g of dried soil was shaken with 25 mL CaCl_2 (10 mmol L^{-1}) in a centrifugal tube at 200 rpm for 30 min and then centrifuged at $4000 \times g$ for 15 min. After filtration through a $0.45\text{-}\mu\text{m}$ filter, the supernatant was collected and the water-soluble Cd was measured using the ICP-MS (Agilent 7700, USA).

The residue recovered from the water-soluble extraction was shaken with 25 mL of NH_4NO_3 (1 mol L^{-1}) in a centrifugal tube at 200 rpm for 30 min. During shaking, the pH of the mixture was maintained at 7.0 (He et al., 2020). The centrifugation and analysis procedures in the exchangeable Cd fraction were identical to those used for the water-soluble fraction.

Carbonate, Fe-Mn oxide, organic matter, and the residual fractions of Cd in the remaining sedimentation reclaimed from the exchangeable extraction were extracted using 1 M NaOAc , a mixture of 0.04 M $\text{NH}_2\text{OH}\cdot\text{HCl}$ and 25 % HOAc , a mixture of 0.02 M HNO_3 and 30 % H_2O_2 , and aqua regia in sequence, according to Li and Thornton (2001). The sum of these six chemical fractions represented the total Cd concentration in the rhizosphere.

Quality control was performed by analyzing two reference materials, GBW07410 (soil) and GBW10012 (plant), which were obtained from the China Standard Materials Research Center (Beijing, China).

2.5 Statistical analysis

The data are represented as the average of the five replicates used in this experiment. The magnetic field treatment effects on the dry weight of *S. alfredii*, DOM concentration and fractionations in the rhizosphere, mobile Cd in the soil, and Cd extraction ability of DOM were tested via one-way analysis of variance. Means of significant difference ($p < 0.05$) were evaluated by Fisher's protected least significant difference test, and SPSS 15.0 software was used to conduct the statistical analyses in this study.

3. Results And Discussion

3.1 Soil pH

A significant pH variation among the treatments was not observed during this study. Compared with the initial soil pH (6.1 ± 0.2) at transplantation, the magnetically treated and untreated *S. alfredii* reduced the pH by 0.2 and 0.1, respectively, in its rhizosphere during the 90-day experiment. These results indicate that the soil used can buffer the plant-induced pH variation, and the pH changes in the *S. alfredii* rhizosphere are not responsible for Cd mobilization in this experiment.

Soil pH significantly influences the metal activity in the substrate, which is enhanced by reducing the pH value (Hu et al., 2018; Römkens et al., 2002; Wang et al., 2020). The pH variation in the rhizosphere soil has been reported for other plants. For instance, *N. tabacum* reduces the pH in the rhizosphere (Loosemore et al., 2004), *B. juncea* increases the soil pH significantly (Kim et al., 2010), and *N. caerulea* has no effect on the pH in the rhizosphere (McGrath et al., 1997). The pH variation in the soil is a complex process that is dependent on soil properties, plant types, and weather conditions. Generally, Cd is present as an oxyanion in soil, and its uptake by *S. alfredii* could enhance the pH in the rhizosphere because the plant releases OH^- in the soil to maintain the charge balance after extracting Cd. Decreased soil pH was caused by increased root secretions, including rhamnolipids, surfactin, and humic and fulvic acids (Davin et al., 2018). The combination of these two processes could determine the final soil pH during phytoremediation.

3.2 DOM concentration and fractionation

The final DOM concentrations in the rhizospheres of the 50, 100, and 150 mT treated *S. alfredii* were 156.9, 192.5, and 198.3 mg kg^{-1} , respectively, which are higher than the initial concentration of the experiment (129.3 mg kg^{-1}). Compared with the control that was exposed to a geomagnetic field, the 100 and 150 mT treatments significantly increased the DOM concentrations in the rhizosphere of *S. alfredii* (Fig. 1), while the 50 mT treatment had a slightly lower value ($p > 0.05$). The 100 and 150 mT treated samples showed 17.6 % and 21.1 % increases in the DOM concentrations in the soil, respectively, relative to the control; however, no significant DOM concentration differences were observed between the 100 and 150 mT treatments (Fig. 1).

The different fraction distributions in the soil of the various treatments are presented in Fig. 2. The DOM fractions in the *S. alfredii* soil exhibited different distribution trends among various treatments, and the hydrophobic fraction concentrations (hydrophobic acid, base, and neutral) did not significantly vary

among the treatments during phytoremediation. In contrast, the hydrophilic fraction concentrations (hydrophilic acid, base, and neutral) were significantly enhanced when treated by 100 and 150 mT than those treated with the geomagnetic field and 50 mT. Generally, the control had higher hydrophobic fraction concentrations than the other treatments, while the 150 mT treatment showed higher hydrophilic fraction concentrations than the other treatments. The acid fraction concentrations (hydrophilic and hydrophobic acid) were significantly higher in the 100 and 150 mT treatments than in the control and 50 mT treatments.

Furthermore, the acid fraction was the predominant DOM component in all the soil treatments, accounting for 77.8 %, 76.5 %, 77.1 %, and 74.9 % of the DOM in the soil of the control, 50, 100, and 150 mT treatments, respectively. Meanwhile, the hydrophobic fraction proportion decreased from 57.3 % to 43.0 %, and the hydrophilic fraction proportion increased from 42.7 % to 57.2 % in the control and 150 mT treatment, respectively.

During the phytoremediation process, root secretion is one of the most significant DOM sources (Martin et al., 2017); therefore, the elevated DOM concentrations in the soils at the end of the experiment might be induced by the presence of *S. alfredii*. Although the effects of external magnetic fields on root secretion have not been determined, many studies suggest that magnetic fields regulate plant metabolism and growth rate on the cellular, subcellular, and molecular levels (Ćirković et al., 2017; Shokrollahi et al., 2018; Teixeira da Silva and Dobránszki, 2016). Therefore, the increased DOM concentration in the rhizosphere of magnetically treated *S. alfredii* could be partially induced by the stimulation of the applied magnetic fields.

DOM contains many functional components, including sulfhydryl, amino, and phosphoryl groups, which have a strong affinity to metals and increase the metal bioavailability in soil (Borggaard et al., 2019; Zhou et al., 2019). The chemical activity of hydrophilic acids is greater than that of hydrophobic acids because the former have more functional groups that can activate the metals in soil (Huang et al., 2019). Therefore, the chemical components of DOM in the rhizospheres of the 100 and 150 mT treated *S. alfredii* are more conducive to increasing Cd mobility. The mechanisms in which magnetic fields regulate root exudation regarding the physiological characteristics of plants require further investigation.

3.3 Chemical fractions of Cd in the rhizosphere

S. alfredii growth did not change the total Cd concentration in the soil compared to the initial state; however, it increased the variable coefficient of the metal (Fig. 3), which concurs with the findings of a previous study, suggesting that plant growth can heterogenize the soil over time (He et al., 2020).

The chemical fractions of Cd in the soil were analyzed at the beginning and end of the experiment. The initial water-soluble and exchangeable Cd concentrations in the soil decreased significantly after the cultivation of *S. alfredii* under all treatments, and the reduction in the rhizosphere increased with increasing magnetic intensity. Compared with the control, the 50 mT treatment did not significantly change the water-soluble Cd concentration, but it significantly decreased the exchangeable fraction

concentration (Fig. 3). However, the 100 and 150 mT treatments decreased the water-soluble and exchangeable Cd concentrations, and their lowest concentrations were observed in the *S. alfredii* rhizosphere exposed to 150 mT. Li et al. (2013a) compared the water-soluble Cd concentration variations in soils after the growth of hyperaccumulating and non-hyperaccumulating ecotypes of *S. alfredii*. They found that the hyperaccumulating ecotype of *S. alfredii* decreased the water-soluble Cd in slightly contaminated soil, which agrees with the results of this study. In contrast, the non-hyperaccumulating ecotype of *S. alfredii* increased the water-soluble Cd in the rhizosphere because *S. alfredii* activates Cd in the soil through increased root excretions, and Cd solubilization is faster than the Cd uptake by the plant.

The water-soluble and exchangeable Cd concentrations in the soils decreased significantly after the *S. alfredii* growth with or without magnetic field treatment. However, the chemical fraction reductions in the rhizosphere were less than 1 % of the extraction amount in the *S. alfredii*, indicating that more than 99 % of the Cd uptake migrated from other fractions. Soils can spontaneously maintain a dynamic balance of a specific element during plant growth via a pair of opposite processes, i.e., mobilizing the element in soil and its extraction by plants. This study shows that the effective Cd extraction root system of *S. alfredii* can eliminate water-soluble and exchangeable Cd faster than the rate Cd is supplemented from other chemical fractions, and *S. alfredii*, particularly the magnetically treated species, can mobilize Cd that is not initially bioavailable. The mobilizing ability of *S. alfredii* increased with increasing magnetic field intensity.

Increased DOM in the rhizosphere of the species can regulate the Cd mobilization process. As shown in Fig. 3, the proportion of organic matter Cd increased significantly from 11.0 % in the control to 25.7 % and 25.9 % in the 100 and 150 mT treatments, respectively, while no significant changes in the carbonate, Fe-Mn oxide, and residual fraction proportions were observed. This finding is consistent with the results suggested by Krishnamurti et al. (2004) who suggested that Cd-DOM complexes can be taken up intact by *Chlorococcum* sp. and result in increased Cd toxicity. Generally, soil pH and DOM concentration are the two dominant factors controlling bioavailability and chemical activity in soil (Christensen and Christensen, 2000). In this study, soil pH did not vary after plant growth; therefore, the variation in the chemical fraction of Cd in the soil is mainly driven by the increased DOM in the rhizosphere.

3.4 Cd extraction capacity of DOM from the rhizosphere

The DOM in the magnetically treated soils of *S. alfredii* significantly enhanced the Cd solubility in the soils relative to deionized water (Fig. 4). For example, the Cd concentration extracted using 100 mg L⁻¹ DOM from the control and the 50, 100, and 150 mT treatments were 48.3 %, 58.6 %, 137.9 %, and 151.7 % higher, respectively, than that extracted using deionized water, indicating that the Cd extraction capacity of DOM improved with increasing magnetic field intensity.

Borggaard et al. (2019) reported that increased DOM concentration improved the metal solubility in the soil and attributed the metal extractability to hydrophilic fractions. Chen et al. (2018) studied the influence of DOM on Zn extractability in calcareous soil and suggested that the addition of maize straw resulted in increased amounts of applied Zn recovered by the DTPA-extractable fraction, which is

attributed to the increased DOM concentration and consequently, increased formation of Zn-fulvic acid complexes. Additionally, the DOM from the soils of the 100 and 150 mT treated *S. alfredii* had a greater ability to mobilize Cd in the rhizosphere compared with the control and 50 mT treatment. For instance, the Cd concentration extracted by the DOM in the 150 mT treatment was 0.73 mg L^{-1} , which was 0.70 and 0.59 times greater than that of the control and 50 mT treatments, respectively. This indicates that the DOM fractions under the different magnetic field treatments varied and potentially influenced Cd mobilization. Furthermore, the magnetically treated *S. alfredii* had a greater capacity to generate chelates with Cd compared with the control.

3.5 Phytoextraction efficiency

The average soil Cd concentration difference before and after transplantation could not be used to estimate the phytoextraction efficiency of *S. alfredii* because the duration of the experiment was not long enough for the *S. alfredii* to change the total Cd concentration in its rhizosphere. Therefore, the Cd uptake in the harvested tissues was calculated to evaluate the phytoremediation effect of *S. alfredii*.

The Cd uptake of *S. alfredii* was defined as the product of the plant's dry weight (belowground and aerial plant parts) and Cd concentration. As shown in Table 1, the 150 mT treated *S. alfredii* had the largest dry weight, and that of the 100 and 50 mT treatments and the control were successively lower. Except for the 50 mT treatment, magnetic field treatments significantly improved the dry weight. Similarly, Cd concentrations in the belowground and aerial plant parts increased with increasing magnetic field intensity (Table 1), consistent with the DOM concentration and fraction variation trends in the *S. alfredii* rhizosphere.

Table 1
Dry weight, Cd concentrations, and Cd decontamination capacity of *S. alfredii* in different treatments

	Dry weight (g)		Cd content		Cd accumulation (mg pot ⁻¹)
	Roots	Shoots	Roots	Shoots	
control	0.73 ± 0.12	8.21 ± 1.25	22.3 ± 2.9	71.2 ± 5.6	2.4
50 mT	0.69 ± 0.08	9.22 ± 1.03	25.6 ± 2.2	66.5 ± 6.1	2.5
100 mT	1.02 ± 0.09	10.63 ± 1.52	29.7 ± 3.1	90.5 ± 9.1	4
150 mT	1.12 ± 0.15	12.39 ± 2.56	38.1 ± 2.6	95.7 ± 7.1	4.9

Based on the dry weight and Cd content, 2.4, 2.5, 4.0, and 4.9 mg of Cd were extracted from the control, 50, 100, and 150 mT treatment pots. The excess Cd in each pot that required removal was calculated as the product of the difference between the initial soil Cd concentration (3.83 mg kg^{-1}) and its safe threshold (0.30 mg kg^{-1}) and the soil mass (6 kg). The required number of planting cycles to clean the soil was determined by the excess Cd divided by the Cd uptake capacity of *S. alfredii*; therefore, 9, 9, 6,

and 5 planting cycles are required for *S. alfredii* to decontaminate the soil in the control, 50, 100, and 150 mT treatments, respectively.

A previous study concluded that an appropriate magnetic field improves the plant dry weight by increasing antioxidant enzyme activity that scavenges oxyradicals resulting from stresses, balancing the internal magnetic field in plants, and increasing cytomembrane permeability to enhance nutrient pass through (Luo et al., 2020). The study focused on the plant improvement effects of magnetic fields rather than effects on the rhizosphere. Previous studies have shown that many plants can modulate the chemical properties of their rhizosphere to improve pollutant extraction via various pathways, including acidification, stimulating microbial proliferation, and DOM excretion (Cao et al., 2018; Gattullo et al., 2015). This study found that the increased Cd uptake capacity of magnetically treated *S. alfredii* was related to increased DOM concentrations and its acid and hydrophilic fractions. Furthermore, higher Cd extractability was observed for DOM under magnetic field treatments, particularly when exposed to 150 mT, indicating that DOM from magnetically treated soils exhibits a greater capacity to mobilize Cd in the soil by generating a complex.

The Cd extraction capacity of magnetically treated *S. alfredii*, particularly at 150 mT, was much greater than that of the control and decreased the required remediation time by half. This study demonstrated that applying an appropriate magnetic field is an effective method to enhance the phytoextraction efficiency of *S. alfredii* based on the variation in DOM concentrations and its fractions in the soil.

Conclusions

This study determined that magnetic field treatment, particularly a 150 mT treatment, improves the phytoextraction of *S. alfredii* by increasing its dry weight and Cd uptake capacity. Compared to the control, the DOM concentrations and acid fractions were significantly higher in the 100 and 150 mT treatments. Additionally, the hydrophilic fraction proportions significantly influence metal mobilization in the soil, which increases with increasing magnetic intensity. Compared with the initial state, water-soluble and exchangeable Cd in the rhizosphere decreased significantly after the *S. alfredii* growth, and the decreases were significantly higher in the magnetically treated plants than in the control. The Cd extraction capacity of DOM from the rhizosphere of magnetically treated *S. alfredii* was significantly higher than that of the control, thus increasing the Cd concentrations in the plant tissues. This study reveals that the primary mechanism to increase Cd extraction in *S. alfredii* is the DOM in magnetically treated *S. alfredii*, which increases the Cd mobilization capacity in the rhizosphere by generating DOM-Cd (organic matter Cd fraction) and consequently increases the phytoextraction efficiency of *S. alfredii* in Cd-polluted soil.

Declarations

Corresponding author's email address: gchero1216@hotmail.com

Ethical Approval: Not applicable. This work did not describe experiments with animals, human subjects, or human tissue samples.

Consent to Participate: Not applicable. This work did not describe experiments with animals, human subjects, or human tissue samples.

Consent to Publish: The manuscript entitled, “Effects of magnetically treated *Sedum alfredii* seeds on the dissolved organic matter characteristics of Cd-contaminated soil during phytoextraction” is prepared in accordance with the Guide for authors available on the journal’s website and it has not been published elsewhere in part or in its entirety. All authors attest to the validity of its contents, and agree to its submission in Environmental Science and Pollution Research.

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Figures

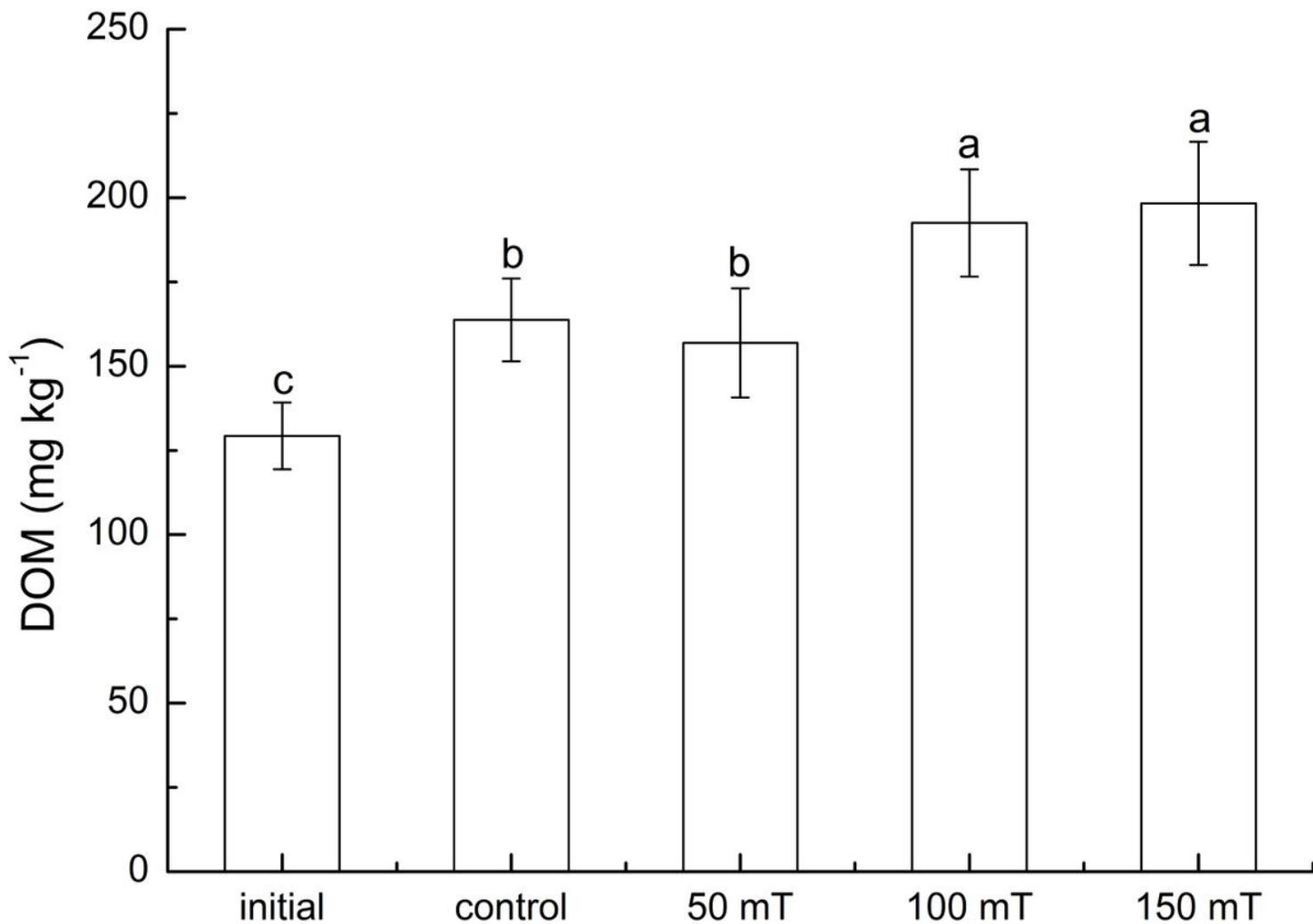


Figure 1

DOM concentrations in the rhizosphere of *S. alfredii*. Different letters represent significant differences in DOM content ($p < 0.05$) evaluated by Fisher's LSD post-hoc tests in different treatments.

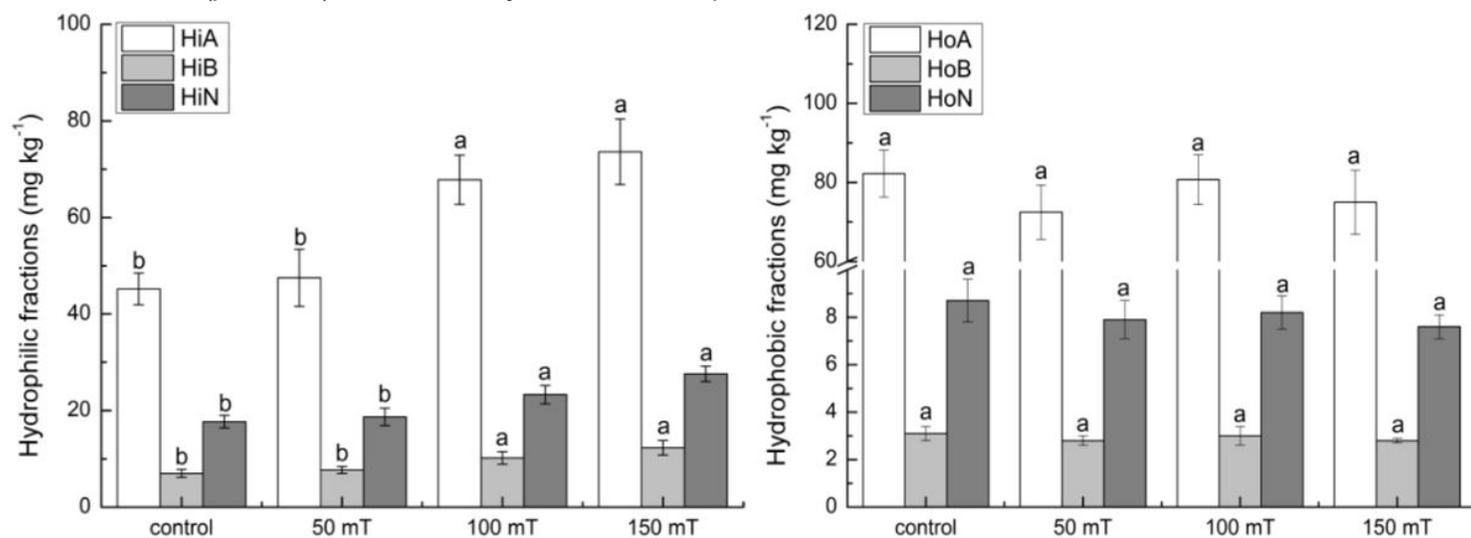


Figure 2

DOM fractions in the rhizosphere of *S. alfredii* Different letters represent significant differences in DOM fractions ($p < 0.05$) evaluated by Fisher's LSD post-hoc tests in different treatments.

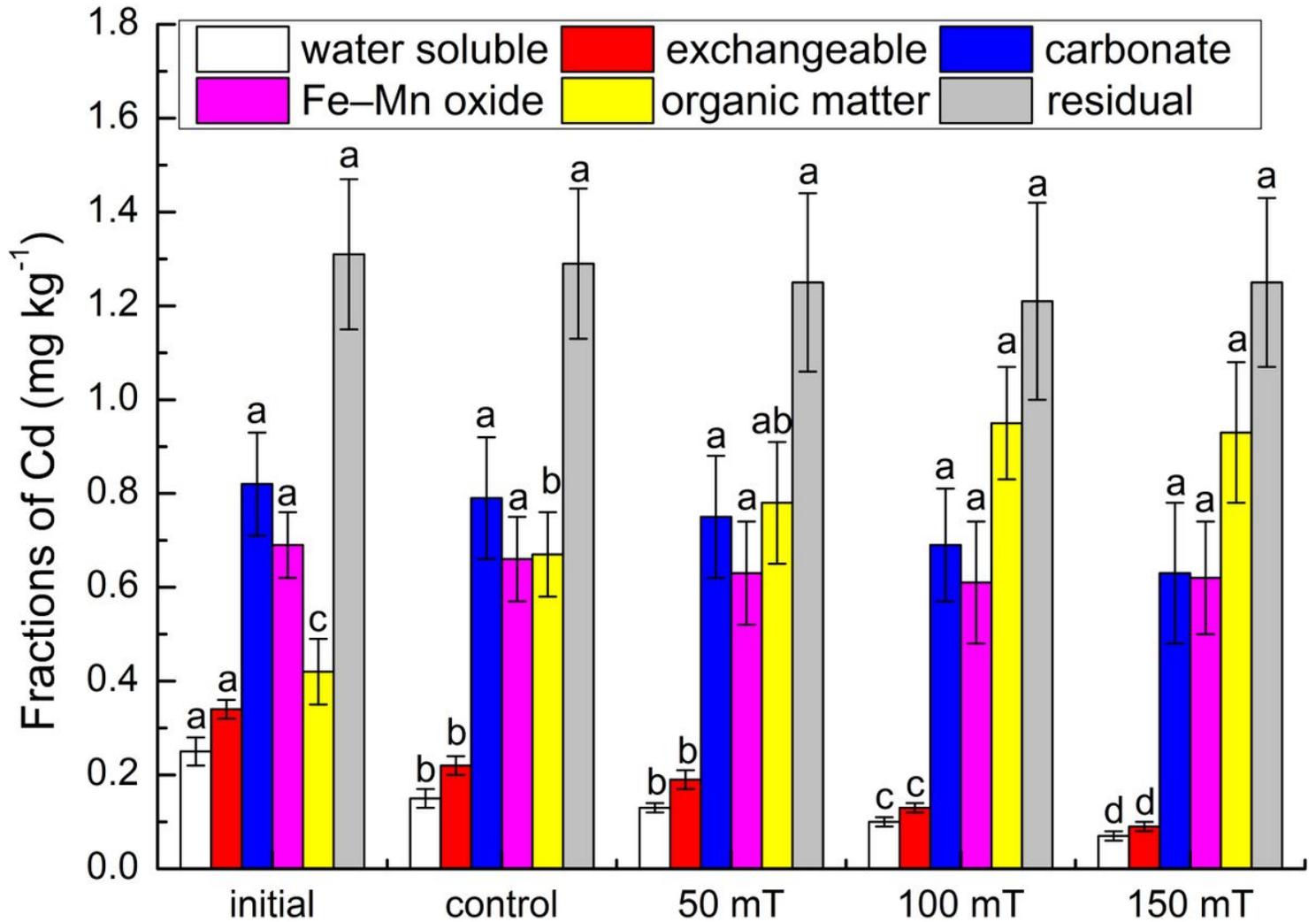


Figure 3

Chemical fractions of Cd in the rhizosphere of *S. alfredii* Different letters represent significant differences in Cd fractions ($p < 0.05$) evaluated by Fisher's LSD post-hoc tests in different treatments.

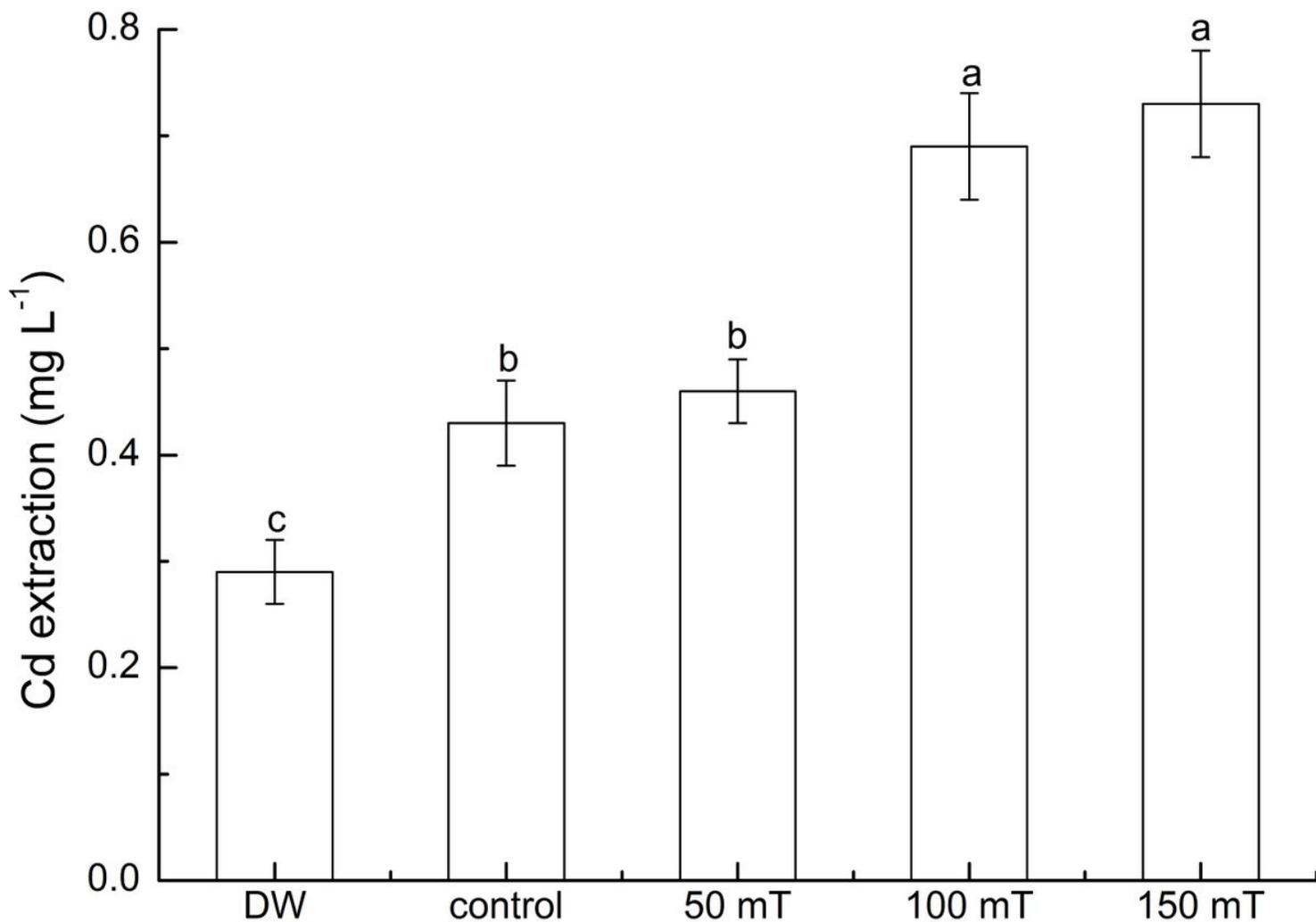


Figure 4

Cd extraction abilities of DOM in different treatments Different letters represent significant differences in Cd extraction ($p < 0.05$) evaluated by Fisher's LSD post-hoc tests in different treatments.