

# Global decline in subsistence-oriented and smallholder fire use

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## Analysis

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## 2

### 3 Abstract

4 Controlled fire use by hunter-gatherers and smallholder agriculturalists and pastoralists shapes  
5 ecologies and enhances livelihoods worldwide. Yet, at the global scale, we know little about how  
6 these practices influence human wellbeing, ecologies, and wildfire risk. As a basis for global  
7 syntheses, we collated information from the literature about fire practices in 587 case study  
8 locations spanning the globe. Here, we assess the coverage and completeness of this data. Limited  
9 quantitative data, particularly, presents a challenge for improved modelling of anthropogenic  
10 influences on fire regimes. We also analyse global trends in fire practices from these studies, finding  
11 evidence that subsistence-oriented fire practices have declined in recent decades, while market-  
12 oriented fire practices have increased. The case studies point to important drivers of these changes,  
13 especially economic pressures, and state governance. We discuss the implications of these findings  
14 for fire policy, and future research.

## 15

### 16 Introduction

17 Worldwide, controlled fire use is culturally and economically significant across human societies for  
18 subsistence-oriented or smallholder livelihoods, including hunting and gathering, pastoralism and  
19 agriculture<sup>1</sup>. These fire practices often draw on deep, place-based knowledge of the feedbacks  
20 between fire and its biophysical and social environment<sup>2</sup>. Beyond immediate livelihood benefits that  
21 structure the application of anthropogenic fire to landscapes, such fires have co-benefits. People  
22 tend to make small, frequent fires that produce a more heterogeneous landscape, supporting greater  
23 biodiversity and fragmenting fuels, which reduces wildfire risk<sup>3,4</sup>. Simultaneously, especially in  
24 changing climates and environments, or where local governance institutions and fire knowledge are  
25 being lost, anthropogenic fire can potentially lead to uncontrolled wildfires, posing risks to human  
26 health and wellbeing<sup>1</sup>. Generally, such wildfire events dominate public and policy discourse around  
27 fire<sup>5</sup>, not the social and ecological relevance of controlled fire use.

28

29 Recent global studies demonstrate that annual burned area has decreased in recent decades<sup>6</sup>, while  
30 extreme wildfire events are becoming more frequent and intense<sup>6,7,8</sup>. Drivers of these trends include  
31 fragmentation of fuel landscapes with land use change, accumulation of flammable vegetation after  
32 suppression of human fire use, and lengthening fire weather windows due to climate change<sup>9,10</sup>. In  
33 this context we need to better understand the conditions under which the loss of livelihood-oriented  
34 fire use can lead to wildfires. Since the 1990s, a growing number of case studies have examined local

35 fire knowledge and practices, but with limited synthesis of this literature at the global scale<sup>1</sup>. Existing  
36 literature reviews have drawn on few sources, have often been biased geographically towards North  
37 America and Australia, and examined narrow research questions<sup>2,4,11,12,13,14</sup>. We do not adequately  
38 understand the social conditions in which smallholder and subsistence livelihood-oriented fire use is  
39 currently changing, nor the implications for fire ecologies, human wellbeing, and wildfire events<sup>15</sup>.  
40 Most global fire models represent human fire use and suppression as a function of variables such as  
41 population density or Gross Domestic Product (GDP), failing to account for different ways in which  
42 people use fire, or how this varies across environmental and social contexts<sup>16</sup>. Similar variables are  
43 used to make inferences about the human signal in global fire remote sensing studies<sup>6,17</sup>.

44

45 To build an empirical basis for studying livelihood-oriented fire practices at the global scale, we  
46 conducted a systematic review of 592 published and unpublished sources to create an open-source  
47 database, the Livelihood Fire Database (LIFE), describing 1708 contemporary subsistence-oriented  
48 and smallholder fire use and mitigation practices in 587 case study locations. LIFE contains data on  
49 both social and biophysical aspects of fire practices. This resource can support, for example, future  
50 analyses covering the spatiotemporal patterns of fires set for diverse livelihood purposes, different  
51 forms of fire governance, and drivers of change in fire practices. Here, we assess the coverage and  
52 completeness of the LIFE database, present a typology of fire use purposes developed from the  
53 database, and assess contemporary trends in these fire practices.

54

## 55 **Results and discussion**

56 The LIFE database demonstrates that fire use and mitigation practices have remained important  
57 within subsistence-oriented and smallholder livelihoods worldwide over the past 30 years. Yet,  
58 subsistence-oriented fire practices have declined at the global scale, while market-oriented fire  
59 practices have increased. The following sections speak to these findings in turn, and discuss their  
60 implications for human wellbeing, ecologies, and wildfire risk. Finally, we reflect on gaps in our  
61 knowledge of these fire practices as highlighted by LIFE, and implications for future research and fire  
62 policies.

63

### 64 **Fire practices contribute to diverse livelihoods worldwide.**

65 LIFE contains information about 1708 fire practices, of which 92 percent were actively practiced at  
66 the time of the case study research and eight percent had been lost but were practiced within the  
67 living memory of research participants. 95 percent of practices in LIFE are examples of fire use. For  
68 23 percent of these fire use practices, we have information about associated fire control measures.

69 One percent of practices in LIFE are examples of ‘opportunistic’ use of fires or burned areas created  
70 by lightning or other people. The remaining three percent of practices are general fire mitigation  
71 measures taken independently of fire use, including fire prevention and suppression practices. Our  
72 focus here is largely on fire use, but there is need for future research addressing fire control and  
73 mitigation specifically.

74

75 These practices were recorded for case study locations spanning 84 countries across all flammable  
76 continents, with the most locations in Brazil, Indonesia, Australia, Mexico, and India (Fig. 1). The  
77 USA, Canada, Australia, and China have the highest number of practices that were lost at the time of  
78 the case study research. Research biases shape the distribution of case studies to some extent. The  
79 high number of cases in Australia and the USA, for example, is probably more reflective of a bias  
80 towards research in these countries than it is of widespread contemporary subsistence-oriented and  
81 smallholder fire practices. Conversely, the low number of cases in eastern Europe and northern,  
82 western, and central Asia probably reflects lack of research rather than lack of fire practices. Despite  
83 research biases, the high concentration of cases in the equatorial tropics likely reflects a high  
84 prevalence of subsistence-oriented and smallholder fire practices in these regions.

85

86 The fire practices in LIFE contribute to livelihoods on a spectrum between subsistence-oriented and  
87 market-oriented. 43 percent of fire practices in LIFE are subsistence-oriented (associated with a  
88 product produced, gathered, or hunted mainly for subsistence), 15 percent are market-oriented  
89 (associated with a product produced, gathered, or hunted mainly for sale), two percent are practices  
90 that are themselves marketed as a service, and 20 percent are neither associated with markets or  
91 subsistence.

92

93 Using LIFE, we developed a hierarchical fire use purpose classification system including eight higher  
94 tier categories: agriculture, pastoralism, hunting and fishing, gathering, charcoal and firewood  
95 production, movement, human health and wellbeing, and social signals. Each of these higher tier  
96 categories is associated with several of 29 lower tier categories (Table 1). This scheme aligns broadly  
97 with, but is more detailed than, those developed in previous studies<sup>4,12,18</sup>. Our lower tier categories  
98 include both direct reasons for fire use, such as driving game for hunting or clearing vegetation for  
99 swidden agriculture, as well as co-benefits, such as reducing fuel loads or maintaining cultural  
100 identity. It is likely that many anthropogenic fires, especially those set in tropical savannas, are  
101 associated with multiple purposes, including proximate reasons and co-benefits<sup>19</sup>. We listed multiple  
102 purposes against a fire practice only where a source explicitly stated that multiple reasons were

103 associated with the same ignitions (in 17 percent of cases). It is probable that a higher proportion of  
104 anthropogenic fires are associated with multiple purposes. Most sources analysed did not distinguish  
105 between direct reasons for fire use and co-benefits, so we could not make this distinction in LIFE.  
106 This is a limitation for future analyses, in that it is the more proximate reasons that are more likely to  
107 sustain fire use and consciously shape how fire is applied to the landscape<sup>20</sup>.

108

109 **Subsistence-oriented fire practices are declining globally.**

110 We recorded trends in the proportion of the population in the study area practising the practice, the  
111 overall proportion of the study area affected by the practice annually, and the frequency with which  
112 the practice takes place in each affected landscape patch. Practices associated with two thirds of our  
113 29 fire use purposes were significantly more likely to show decreasing trends in at least one of the  
114 variables in a two-tailed binomial test (Table 2). Except for pre- and post-harvest crop residue  
115 burning, for those fire purpose types for which we had insignificant results in the binomial tests, we  
116 had very small sample sizes ( $n < 15$ ).

117

118 Declining trends in fire use have a historical precedent. From the 15<sup>th</sup> century, European colonialism,  
119 development of capitalist economies and intensification of agriculture directly and indirectly  
120 suppressed fire use associated with subsistence and smallholder livelihoods worldwide<sup>21,22,23</sup>. The  
121 LIFE database sheds light on contemporary drivers of changes in fire use. Subsistence-oriented fire  
122 use practices in LIFE are significantly more likely to show decreasing trends, while market-oriented  
123 fire use practices are more likely to show increasing trends in all three variables (Fig. 4). This  
124 suggests that economic pressures are a global driver of changing fire use.

125

126 Case studies in LIFE illustrate some of the mechanisms by which economic pressures are driving  
127 decreasing subsistence-oriented fire use. Sometimes, commercial land uses limit the land available  
128 to populations with subsistence-oriented livelihoods. In the Brazilian *cerrado*, for instance,  
129 expansion of industrial-scale capitalised agriculture is driving Indigenous and traditional populations  
130 into smaller territories, limiting fire use associated with hunting and gathering<sup>24</sup>. Elsewhere, market-  
131 oriented activities are replacing subsistence-oriented activities within livelihoods, reducing  
132 subsistence-oriented fire use. For example, in Guinea-Bissau, structural adjustment policies in the  
133 late 1980s created new export markets for cashew, driving smallholders to clear land for cashew  
134 orchards in following decades<sup>25</sup>. Adoption of the cashew cash crop has driven declines in swidden  
135 and savanna fire use because farmers have less time and land available for more traditional land  
136 uses, and they are wary of the risks of wildfire damage to their cash crops. Deepening capitalist

137 relations also see the replacement of subsistence activities with wage labour and purchased goods.  
138 For example, in Australia, many Aboriginal people now live and work in large settlements ‘off-  
139 country’, decreasing their use of fire for hunting and gathering<sup>26</sup>. Notably, the substitution of  
140 market-oriented production or wage labour for subsistence livelihoods may only be partial. For  
141 Maya people in southern Belize, for instance, the insecurity and seasonality of labour and cash crop  
142 markets means that swidden agriculture remains important within diversified livelihoods<sup>27</sup>.

143

144 State governance can be another direct or indirect driver of decreasing fire use. Over recent  
145 centuries, a narrative that fire use was incompatible with ‘rational’ land use was used as justification  
146 for policies aiming to eradicate anthropogenic fire in many countries, especially in European  
147 colonies<sup>21,22,23</sup>. Though many nation states have since adopted ‘prescribed fire’ (the controlled  
148 application of fire for specific land management objectives), this is usually solely deemed the  
149 prerogative of specialised agencies, while livelihood-oriented fire use is heavily regulated. For  
150 instance, in the Borana rangelands of Ethiopia, pastoral fire use has largely been lost after  
151 enforcement of a general ban on fire use<sup>28</sup>. Fire regulations may be specific to, or only see  
152 enforcement in, protected areas such as the Biligiri Rangaswamy Temple Tiger Reserve in India,  
153 where Soliga people have stopped practising customary fire use since it was made illegal<sup>29</sup>.  
154 Regulations may also impose restrictive conditionality upon fire use. In Laos, laws that severely limit  
155 the amount of land allocated to each household and restrict fallow periods to 5 years have made it  
156 difficult to maintain swidden agriculture<sup>30</sup>. Many countries have strong anti-fire regulations, but it is  
157 important to note that these are not always enforced: it is often easy for fire users to remain  
158 anonymous; states may lack the resources for strong enforcement; and those state officials  
159 responsible for enforcement may collude with fire users<sup>22,31,32</sup>.

160

161 Besides regulation, economic governance mechanisms are driving declines in fire use. In recent  
162 decades one such mechanism has been Payments for Ecosystem Services (PES) schemes, where  
163 these have made payments conditional upon eliminating fire use. For instance, in the *páramos* of  
164 Ecuador, communities participating in the Socioparamo PES scheme are paid to cease burning and  
165 grazing in certain areas<sup>33</sup>. Elsewhere, such as in Mexico, agricultural subsidies from the Government  
166 have only been available for permanent agriculture rather than swidden, and grants have supported  
167 the adoption of agrochemicals and mechanical tillage, which replace fire use<sup>34</sup>.

168

169 Case studies in LIFE highlight potential implications of declines in or loss of subsistence-oriented fire  
170 use. Declining fire use can severely undermine the livelihoods of rural people. For example, for

171 Borana pastoralists in Ethiopia, loss of fire use in rangelands has led to bush encroachment and  
172 forage scarcity<sup>28</sup>, and prohibition of fire use has reduced household income from agriculture and  
173 forest produce for Soliga people in India<sup>29</sup>. Sometimes, as for some Pemón communities in  
174 Venezuela, declining fire use can lead to a loss of cultural identity and conflict between older  
175 generations who retain fire knowledge and younger generations who are losing this knowledge<sup>35</sup>.  
176 Notably, its association with cultural identity may also contribute to the persistence of fire use in the  
177 face of countervailing drivers, as with the swidden agriculture of Tagbanua people in the  
178 Philippines<sup>36</sup>.

179

180 Loss of human fire can reduce the biodiversity of fire-dependent ecosystems. In the Biligiri  
181 Rangaswamy Temple Tiger Reserve in India, loss of Soliga fire use has increased vegetation density  
182 to the point of suppressing the growth and recruitment of many tree and plant species<sup>29</sup>.  
183 Meanwhile, in Australia, hill kangaroos are more abundant in regions with fine-grained vegetation  
184 mosaics resulting from Martu hunting fires, than in regions dominated by lightning fire<sup>37</sup>. Loss of  
185 smallholder and subsistence-oriented fire use can also lead to increased wildfire risk where it leads  
186 to more homogenous fuel landscapes. Case studies in LIFE suggest that fire use associated with  
187 reducing fuel loads at the landscape scale, or creating firebreaks is declining. In the Brazilian *cerrado*,  
188 Indigenous Reserves where Xavante fire use for hunting and gathering is retained suffer significantly  
189 less wildfires than reserves where fire use has been lost or has declined<sup>24</sup>. When fire use reduces in  
190 importance or is made illegal, local governance institutions and knowledge that ensure that fire use  
191 is controlled can also be lost, increasing the likelihood of anthropogenic wildfires. For example, anti-  
192 fire legislation in Ghana is leading to the loss of traditional village institutions that organised  
193 collective early-dry season burning to protect trees of spiritual and economic importance<sup>38</sup>. In Belize,  
194 Maya farmers with less time for swidden agriculture due to engagement in wage labour, cash  
195 cropping and pasture development are more likely to practice agricultural burns alone than in  
196 shared labour groups, and less likely to construct firebreaks before burning or remain onsite until a  
197 fire is extinguished<sup>27</sup>.

198

#### 199 **Market-oriented fire practices are increasing globally.**

200 Economic pressures can also drive increases in fire use: while subsistence-oriented fire use is  
201 declining, market-oriented fire use practices are increasing. New markets for certain products are  
202 driving increases in fire use associated with their production. For example, in Indonesia, a high value  
203 market for turtle meat emerged in the 1990s, driving increases in fire use to hunt turtles<sup>39</sup>. New  
204 technologies or infrastructural development may also drive increasing fire use. On the Indo-Gangetic

205 Plains, 'green revolution' technologies have enabled smallholders to adopt double cropping of rice-  
206 wheat, with the use of combine harvesters<sup>40</sup>. Under this cropping system there is limited time  
207 between harvest of rice and planting of wheat, and a greater biomass of crop residue left in-field.  
208 This means that farmers are now more likely to burn crop residues than collect them for other uses.  
209

210 Among specific fire use purposes, only fire use to clear land for permanent agriculture, and protest  
211 fires (arson) show increasing trends in all three variables (proportion of population involved,  
212 proportion of landscape affected, and frequency of fire use in affected landscape patches) (Table 2).  
213 The increasing trend in protest fires is often linked to disputes over land, especially where  
214 commercial enterprises or protected areas restrict subsistence or smallholder land use. Its  
215 anonymity makes fire a well-documented 'weapon of the weak'<sup>22,41</sup>. Arson is commonly used by  
216 smallholder farmers in Indonesia, for example, to assert their rights to land where large landholders  
217 have obtained land concessions from government<sup>42</sup>.  
218

219 Transitions to cash cropping in permanent agricultural systems tend to be driven by economic  
220 pressures, as with adoption of cashew as a cash crop by farmers in Guinea-Bissau<sup>25</sup>. In many cases,  
221 government policies also support transitions to permanent agriculture, like in Mexico, where it is  
222 promoted through agricultural subsidies<sup>34</sup>. Where economic pressures and policies support market-  
223 oriented permanent agriculture this can also be an indirect driver of increasing fire use associated  
224 with crop residue burning. This has been the case with smallholder rice-wheat cropping on the Indo-  
225 Gangetic Plains<sup>40</sup>. Crop residue burning does not, however, show a significant increasing trend across  
226 all case studies. This may relate to recent Government regulations on crop residue burning in many  
227 countries. In China, for example, crop residue burning has been banned since 1997<sup>43</sup>.  
228

229 While burning to clear land for swidden agriculture is significantly more likely to be decreasing in  
230 terms of the proportion of population practising it, or the proportion of the landscape affected, it is  
231 more likely to be increasing in frequency in affected landscape patches. This accords with observed  
232 global trends of declining area and declining fallow periods in swidden agriculture<sup>44</sup>. Shorter  
233 rotations and corresponding increases in fire use frequency are often linked to declining land  
234 availability per household. This can be related to expansion of commercial land uses or protected  
235 areas, or to population growth. It may also be that Government policies or incentive programs  
236 supporting permanent agriculture drive shortening fallow periods. In Brazil, participants in the Bolsa  
237 Floresta Program receive payments for 'zero deforestation', where 'deforestation' refers to opening

238 fields in mature forest areas<sup>45</sup>. Clearing plots in secondary forests under a certain age of regrowth is  
239 permitted, incentivising farmers to practice short fallow swidden.

240

241 Generally, where economic pressures are driving increasing fire use, this fire use is less likely to be  
242 sensitive to environmental cues, which may increase wildfire risk and lead to negative ecological  
243 impacts. For example, in Indonesia, repeated burning to open areas for market-oriented fishing is  
244 transforming peatland forests into open floodplains<sup>39</sup>. Positive feedbacks between fire, removal of  
245 tree cover and loss of peat are increasing the susceptibility of these areas to wildfires.

246

247 **Our knowledge of livelihood fire practices remains limited.**

248 In each source, we looked for information about 37 variables for each fire practice (see  
249 supplementary table 1). Most sources only provided information for some of the variables of  
250 interest, so the coverage of LIFE is incomplete (Fig. 5). In part because this kind of data may be  
251 sensitive and difficult to collect, most case studies do not provide quantitative data regarding the  
252 fire return intervals or burned area of anthropogenic fires. This presents particular limitations for use  
253 of the database to improve global fire models. There is also particularly little data available on the  
254 social or demographic factors, such as age or gender, that shape participation in fire practices.  
255 Geographically, there are also data gaps, such as in eastern Europe and northern, western, and  
256 central Asia (Fig. 1). There are also far more studies discussing fire use than fire control and  
257 mitigation practices. These are areas that future case study research might address.

258

259 Despite the data gaps, there is potential to combine other forms of fire data with the quantitative  
260 and qualitative data from LIFE and future case studies, to improve our understanding of the  
261 implications of changes to smallholder and subsistence livelihood-oriented fire practices. Local or  
262 regional studies provide a precedent here. For example, in Mexico, social research on contemporary  
263 fire use has been used to help interpret the human fire signal in a dendrochronological record<sup>46</sup>, and  
264 in Indonesia, local case study data has been combined with fire remote sensing data to study how  
265 local fire use practices shape the regional fire regime<sup>42</sup>. So far, case study data on fire use has not  
266 been combined with these other types of data at the global scale, nor has it been used to inform  
267 global fire models<sup>16</sup>. Our work to collate case study data at the global scale opens future possibilities  
268 to analyse how fire regimes and ecologies are shaped by fire use directed towards different  
269 livelihood outcomes.

270

271 Since the 1990s, there have been some efforts to design fire management programs based on  
272 traditional fire knowledge, in recognition of the ways in which subsistence-oriented fire use has  
273 shaped ecosystems or contributed to wildfire risk reduction. Perhaps the most advanced example of  
274 this is in Australia, where Aboriginal fire practices have informed the management of protected  
275 areas<sup>47</sup>, and Aboriginal rangers are now being funded to conduct fire management via government-  
276 accredited methods for counting carbon credits<sup>48</sup>. Importantly, such programs often assume that  
277 standardised fire management can substitute for the contingent, livelihood-oriented practices of fire  
278 users, but this does not necessarily have the same ecological outcomes, nor lead to meaningful  
279 recognition of local fire users<sup>22,49,50</sup>. Future policies must make space for controlled fire use as it is  
280 already practised within local livelihoods. This will only be possible if the drivers of changes to these  
281 livelihoods are explicitly considered. We must better understand the social and environmental  
282 conditions under which economic pressures or state governance drive loss of fire use or less  
283 controlled fire use. We must also consider the cultural and economic factors that can drive the  
284 persistence of fire use in some places. The case studies in LIFE will be an important resource for  
285 further systematic analyses of these drivers.

286

## 287 **Methods**

288 A systematic literature review was used to find sources for inclusion in the LIFE database. Below we  
289 outline our search methodology, database structure, and methods for the analysis of trends in fire  
290 practices presented in this article.

291

## 292 **Data sources**

293 Published articles and books and ‘grey literature’ in the form of dissertations and unpublished  
294 reports, including non-English language sources, were included in LIFE. We were unable to include  
295 sources where we could not access them either online, or in libraries at the time of the research. The  
296 following criteria were used for inclusion of a source in LIFE: a) is written or published during or after  
297 the year 1995, with relevant data collection carried out during or after the year 1990; b) has a sub-  
298 national scale of analysis; c) is based on original empirical research (sources based on secondary  
299 data, or review articles were not included); d) discusses fire use or mitigation practices by  
300 smallholders or households with non-agricultural subsistence-oriented livelihoods, as practiced at  
301 the time of the research or practised within the living memory of research participants. We defined  
302 ‘smallholders’ as households for which agriculture is the principal source of livelihood, and which  
303 rely predominantly on family labour<sup>51</sup>. Such households usually produce at least a proportion of  
304 their crops and/or livestock for subsistence but may also produce a substantial amount for the  
305 market. We did not consider farm size in our definition, given that the amount of land required to

306 make agriculture a viable livelihood varies significantly between countries, and different  
307 governments vary significantly in their legal definitions of the farm area considered a  
308 'smallholding'<sup>52</sup>.

309

310 The systematic literature review was initially conducted using the Clarivate Analytics Web of Science  
311 Core Collection. We applied a search string to the titles, keywords, and abstracts of all sources in the  
312 collection (see supplementary methods 1). The search results were filtered to remove all sources  
313 published before 1995, and to only include sources classified into a limited number of the Web of  
314 Science Research Areas (see supplementary methods 1). Our search yielded 3718 sources, of which  
315 226 sources met our criteria for inclusion in LIFE. Searches were then made of several databases of  
316 fire-specific literature. We considered for inclusion all sources listed in six existing databases and  
317 literature reviews focused on human fire use<sup>5,11,12,13,14,53</sup>. Finally, we applied another search string to  
318 the titles and abstracts of all material in the catalogue of the Fire Research Institute Library<sup>54</sup> (see  
319 supplementary methods 1). These additional searches resulted in a further 87 sources for inclusion  
320 in LIFE. 279 more sources were then included based on 'snowball sampling' of the bibliography of  
321 each of the sources.

322

### 323 **Structure of the LIFE database**

324 In a preliminary version of LIFE, we mostly used free text to record information against the variables  
325 of interest. Later, based on qualitative analysis of this text, we developed categories to code the data  
326 for some of the variables. Where this was the case, the free text is maintained alongside the  
327 categorical data in the published database.

328

329 LIFE was created as a Microsoft Excel spreadsheet with three tabs, which respectively record  
330 information about individual sources, individual case study locations, and individual fire practices in  
331 those locations (see supplementary methods 1, supplementary table 1 and supplementary data 1).  
332 We grouped sources where they relate to research by the same author(s) in the same case study  
333 location(s) over a continuous period. Where a source or group of sources provided data separately  
334 for case studies in multiple locations, these were recorded as separate cases. Case studies in LIFE  
335 vary greatly in study area, from studies focused on single settlements to regional studies. For all  
336 cases, a single latitude and longitude point is provided to indicate the approximate centre of the  
337 study area, and each is coded into one of four categories to indicate the broad scale of the study  
338 area. When determining what constituted separate fire practices for inclusion in LIFE, we used the  
339 criteria by which the author(s) of the source(s) distinguished between practices. In most cases,

340 authors distinguished fire practices based on their purposes (for example, ‘burning to clear land for  
341 swidden agriculture’ or ‘burning to drive game for hunting’). Sometimes authors made it clear that  
342 multiple fire use purposes related to the same ignitions on the landscape, in which case these were  
343 listed together as a single practice. In some cases, authors distinguished practices based on the  
344 seasons or locations in which they take place (for example, ‘fires set in the early dry season’).

345

#### 346 **Analysis of trends in fire practices**

347 For each fire practice, we recorded whether a practice was increasing or decreasing in each of three  
348 variables: 1) the proportion of the population in the study area practising the practice; 2) the overall  
349 proportion of the study area affected by the practice annually; 3) the frequency with which the  
350 practice takes place in each patch of the landscape. Few sources reported on trends in practices in  
351 these terms, so we made judgements based on the information available. For example, if a source  
352 noted that expansion of intensive cropland in the study area had taken place at the expense of  
353 savanna areas in which subsistence hunting took place, we would record fire use for hunting as  
354 taking place over a decreasing proportion of the study area. Or, for example, if a source reported  
355 that younger generations were no longer learning to use fire for hunting, we would report fire use  
356 for hunting as taking place among a decreasing proportion of the population. For over 50 percent of  
357 practices, it was not possible to record the trend in any of these three variables (Fig. 3). We also  
358 noted where practices had been lost, but were remembered by research participants, or where  
359 practices were being (re)introduced based on traditional knowledge.

360

361 We then ran a two-tailed exact binomial test for each of the three variables for all subsistence-  
362 oriented practices, all market-oriented practices, and all fire use practices categorised under each of  
363 our 29 lower-tier fire use purpose types (Table 1). Here we tested whether the proportion of  
364 practices in each of these categories with a decreasing trend, or that had been lost, differed from the  
365 null hypothesis with a *P*-value of less than 0.05. In running this test, we assumed that the case  
366 studies in LIFE were an unbiased sample of all smallholder and subsistence-oriented fire practices,  
367 i.e., that the authors of the studies in LIFE were not more likely to focus on practices that were  
368 increasing rather than decreasing, or *vice versa*. We believe that this assumption holds, given that  
369 LIFE draws on studies from a wide range of academic disciplines, and with a diversity of primary  
370 research topics (many studies discussed fire practices tangentially to other phenomena).

372

#### 373 **Data availability statement**

374 The full LIFE database is included as supplementary data 1.

375

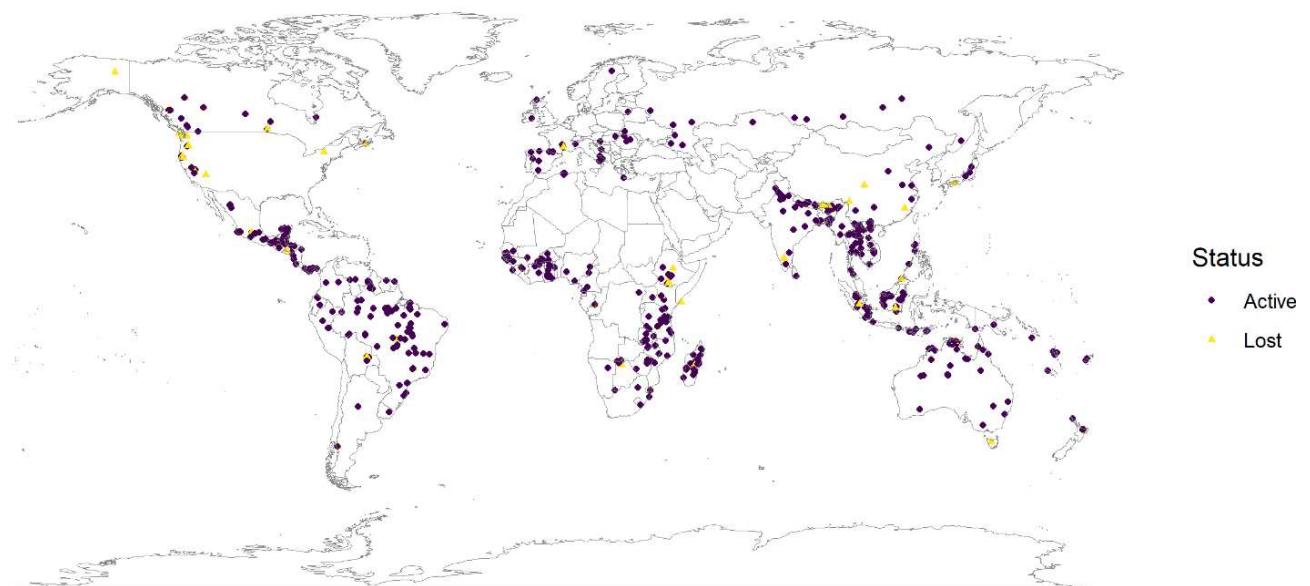
### 376 **Code availability statement**

377 The R computer code supporting the analysis presented in this study is included as supplementary  
378 methods 2.

379

380 **Fig. 1.** Map showing case study locations in the LIFE database, including those with fire practices that  
381 were actively practiced at the time of the case study research and those where practices had been  
382 lost but were practiced within the living memory of research participants.

383



384

385

386 **Table 1.** Fire use purpose categories, and the number of examples of each in the LIFE database (*n*).

Fire use purpose		Example	<i>n</i>
Higher tier category	Lower tier category		
Agriculture	A1. Clear vegetation for swidden or semi-permanent agriculture	Lacandon Maya <i>milpa</i> swidden agriculture of corn, beans, and maize in Mexico <sup>55</sup> .	325
	A2. Clear vegetation for permanent agriculture	Establishment of cashew plantations in Guinea-Bissau <sup>25</sup> .	46
	A3. Clear weeds and/or crop residues during the growing season	Kayapó people in Brazil burning weeds in cassava swidden plots <sup>56</sup> .	21
	A4. Clear weeds and/or crop residues after harvest to enable planting	Rice crop residue burning on the Trans-Gangetic Plains of India <sup>40</sup> .	133
	A5. Reduce crop pests	Kanak people in New Caledonia burning to remove cover for wild pigs to protect tuber crops <sup>57</sup> .	34
Pastoralism	P1. Clear vegetation to establish new pasture areas	Establishment of pasture by smallholder farmers in Pará, Brazil <sup>58</sup> .	18
	P2. Enhance forage for grazing livestock	Pastoral burning in the Basque region of France <sup>59</sup> .	216

	P3. Herd livestock	Use of smoke by Evenk people in Siberia to draw reindeer close to camp, because of relief from biting insects. <sup>60</sup>	17
	P4. Reduce livestock pests and predators	Rangeland burning by Oromo people in the Bale Mountains, Ethiopia <sup>61</sup> .	50
Hunting and fishing	HF1. Create or improve habitat for hunted or fished species	Burning to create pools for fishing in Indonesian peatlands <sup>39</sup> .	24
	HF2. Renew forage to draw hunted or fished species into a particular area	Warlpiri people in Australia burning for fresh growth to attract kangaroos for hunting <sup>62</sup> .	65
	HF3. Improve visibility or access specifically for hunting or fishing	Pemón people in Venezuela burning to remove tall and cutting grasses from river edges for fishing <sup>63</sup> .	67
	HF4. Drive animals when hunting	Xavante ritual hunting drives in Brazil <sup>64</sup> .	65
	HF5. Kill, injure, or tire animals when hunting	Teke-Alima people in Gabon burning to kill grasshoppers for gathering <sup>65</sup> .	12
Gathering	G1. Enhance productivity of foraged resources	Western Mono tribes in California burning to promote black oak acorns and reduce oak pests <sup>66</sup> .	134
	G2. Ease the collection of a foraged resource by improving visibility or access	Burning to clear the ground of leaf litter to facilitate collection of mahua flowers in Orissa, India <sup>67</sup> .	41
	G3. Drive wild bees away from hives for honey collection	Dayak and Malay people smoking wild bees away from bee trees in Indonesia <sup>68</sup> .	48
Charcoal and firewood production	C1. Produce charcoal	Production of charcoal by Maasai women in Tanzania <sup>69</sup> .	16
	C2. Produce fuelwood for gathering, or enable gathering of fuelwood	Khanyayo people in South Africa burning woodlands to speed up the drying process in wood that can be collected for fuel <sup>70</sup> .	20
Movement	M1. Maintain and open trails and waterways for general access	Bayei people burning to clear vegetation blocking river channels in the Okavango delta <sup>71</sup> .	50
Human health and wellbeing	HW1. Reduce animals that are dangerous to or unwanted by humans	Chiquitano people in Bolivia burning <i>pampas</i> to reduce hiding places for snakes <sup>72</sup> .	62
	HW2. Reduce fuel loads to reduce risk of wildfires at a landscape scale	Fires set by Bambara and Malinke people in Mali fragment the fuel landscape to prevent large fires later in the dry season <sup>19</sup> .	87
	HW3. Create firebreak using fire to protect e.g., resources, farms, sacred sites	Krahô people in Brazil burning to create firebreaks to protect fruiting trees <sup>73</sup> .	44
	HW4. Suppress a wildfire (using backing fire to fight fire with fire)	Wapichan people in Guyana using backing fire to suppress wildfires <sup>74</sup> .	6
	HW5. Produce a more aesthetically pleasing landscape, or for enjoyment	Métis people in Canada burning off dead grass in fields and around houses so new grass grows back thicker and greener, for aesthetic reasons <sup>75</sup> .	23
Social signals	S1. Communicate about current activity	Khwe people in Namibia burning to signal to distant family members to notify them of location and success of a hunt <sup>76</sup> .	17
	S2. Show disapproval or protest (arson)	Burning by Galician peasants in resistance to state forestry <sup>77</sup> .	51
	S3. For ritual or ceremonies	Burning in Spanish chestnut forests in the cultural feast of the <i>Calbote</i> , marking summer's end and the start or end of the chestnut harvest <sup>78</sup> .	16
	S4. Assert or maintain cultural identity	Banbai people in Australia burning to reinstate cultural practices and care for country <sup>79</sup> .	112

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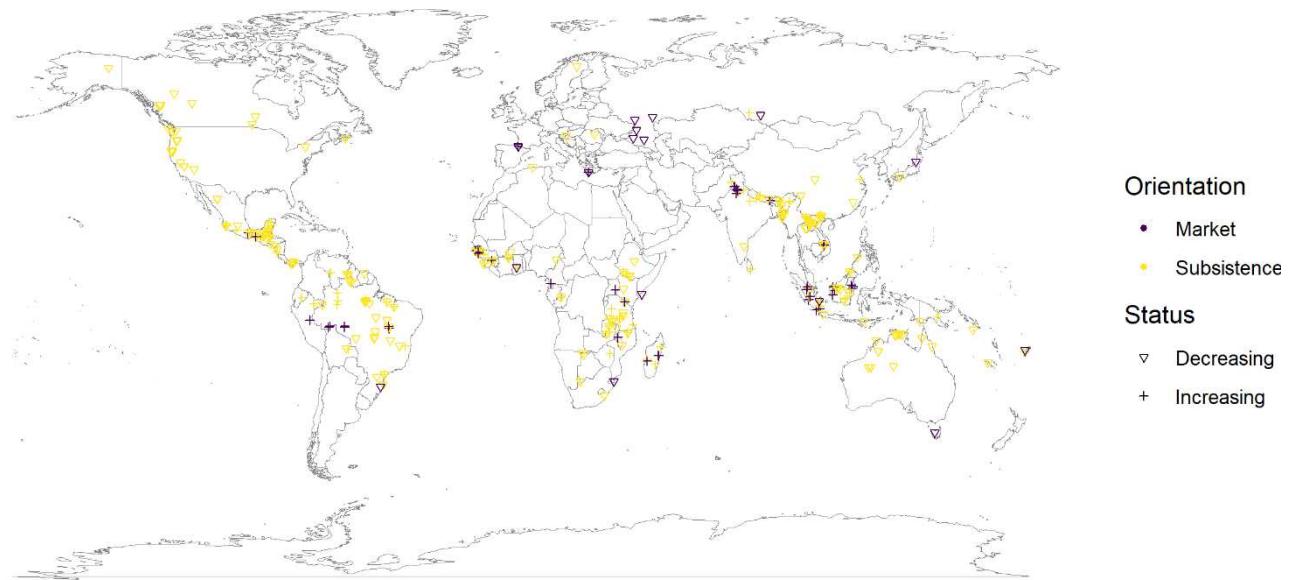
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**Table 2.** Trends in fire practices (codes for fire use purposes are given in table 1). '+' and blue colouration indicates a significant increasing trend ( $P<.05$  in two-tailed binomial test). '-' and red colouration indicates a significant decreasing trend ( $P<.05$  in two-tailed binomial test). 'NA' and white colouration indicates an insignificant trend ( $P>.05$  in two-tailed binomial test).

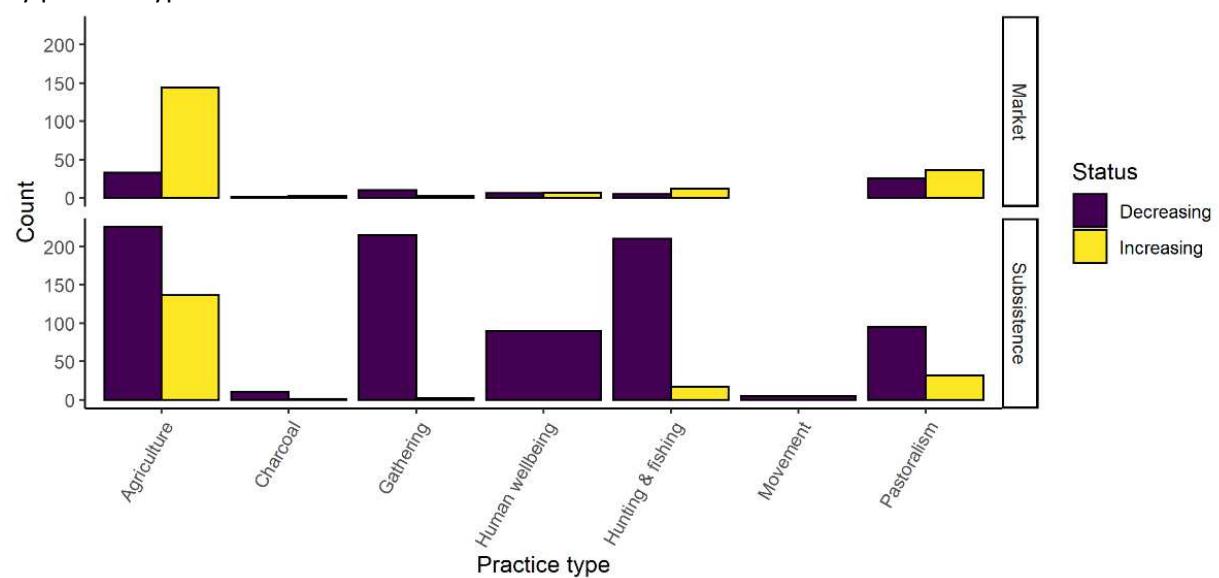
Fire practice type	Proportion of population practicing	Proportion of landscape burned per year	Frequency of fire in each patch of landscape
All market oriented	+ ( $n=120, P<.001$ )	+ ( $n=145, P<.001$ )	+ ( $n=17, P=.01$ )
All subsistence oriented	- ( $n=328, P<.001$ )	- ( $n=378, P<.001$ )	- ( $n=219, P<.001$ )
A1	- ( $n=95, P<.001$ )	- ( $n=137, P<.001$ )	+ ( $n=95, P<.001$ )
A2	+ ( $n=33, P<.001$ )	+ ( $n=38, P<.001$ )	NA ( $n=2, P=.50$ )
A3	NA ( $n=11, P=1.00$ )	NA ( $n=12, P=1.00$ )	NA ( $n=0$ )
A4	NA ( $n=62, P=.25$ )	NA ( $n=62, P=.10$ )	NA ( $n=9, P=1.00$ )
A5	NA ( $n=10, P=.75$ )	NA ( $n=9, P=1.00$ )	NA ( $n=3, P=1.00$ )
P1	NA ( $n=7, P=.45$ )	NA ( $n=10, P=.11$ )	NA ( $n=1, P=1.00$ )
P2	- ( $n=83, P<.001$ )	- ( $n=96, P<.001$ )	- ( $n=39, P<.001$ )
P3	NA ( $n=8, P=.29$ )	NA ( $n=7, P=.45$ )	NA ( $n=5, P=1.00$ )
P4	- ( $n=15, P<.001$ )	- ( $n=16, P<.001$ )	NA ( $n=8, P=.07$ )
HF1	- ( $n=17, P=.01$ )	- ( $n=17, P=.01$ )	- ( $n=11, P<.001$ )
HF2	- ( $n=31, P<.001$ )	- ( $n=31, P<.001$ )	NA ( $n=14, P=.06$ )
HF3	- ( $n=37, P<.001$ )	- ( $n=34, P<.001$ )	- ( $n=16, P<.001$ )
HF4	- ( $n=37, P<.001$ )	- ( $n=39, P<.001$ )	- ( $n=19, P<.001$ )
HF5	- ( $n=7, P=.02$ )	- ( $n=7, P=.02$ )	NA ( $n=3, P=.25$ )
G1	- ( $n=78, P<.001$ )	- ( $n=83, P<.001$ )	- ( $n=61, P<.001$ )
G2	- ( $n=22, P<.001$ )	- ( $n=25, P<.001$ )	- ( $n=17, P<.001$ )
G3	- ( $n=9, P=.04$ )	- ( $n=10, P=.02$ )	NA ( $n=4, P=.13$ )
C1	NA ( $n=2, P=1.00$ )	NA ( $n=2, P=1.00$ )	NA ( $n=0$ )
C2	- ( $n=7, P=.02$ )	NA ( $n=4, P=.13$ )	NA ( $n=3, P=1.00$ )
M1	- ( $n=19, P<.001$ )	- ( $n=21, P<.001$ )	- ( $n=12, P=.006$ )
HW1	- ( $n=25, P<.001$ )	- ( $n=27, P<.001$ )	- ( $n=11, P=.01$ )
HW2	- ( $n=37, P<.001$ )	- ( $n=42, P<.001$ )	- ( $n=15, P<.001$ )
HW3	- ( $n=18, P<.001$ )	- ( $n=15, P<.001$ )	- ( $n=11, P<.001$ )
HW4	NA ( $n=1, P=1.00$ )	NA ( $n=1, P=1.00$ )	NA ( $n=0$ )
HW5	- ( $n=17, P<.001$ )	- ( $n=19, P<.001$ )	NA ( $n=4, P=1.00$ )
S1	- ( $n=11, P<.001$ )	- ( $n=11, P<.001$ )	NA ( $n=5, P=.38$ )
S2	+ ( $n=7, P=.02$ )	+ ( $n=9, P=.004$ )	NA ( $n=1, P=1.00$ )
S3	NA ( $n=5, P=.06$ )	- ( $n=6, P=.03$ )	NA ( $n=3, P=.25$ )
S4	- ( $n=45, P<.001$ )	- ( $n=36, P<.001$ )	NA ( $n=13, P=.09$ )

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413 **Fig. 2**  
414 **a.** Map showing status (increasing or decreasing) of market- and subsistence-oriented practices in  
415 the LIFE database.

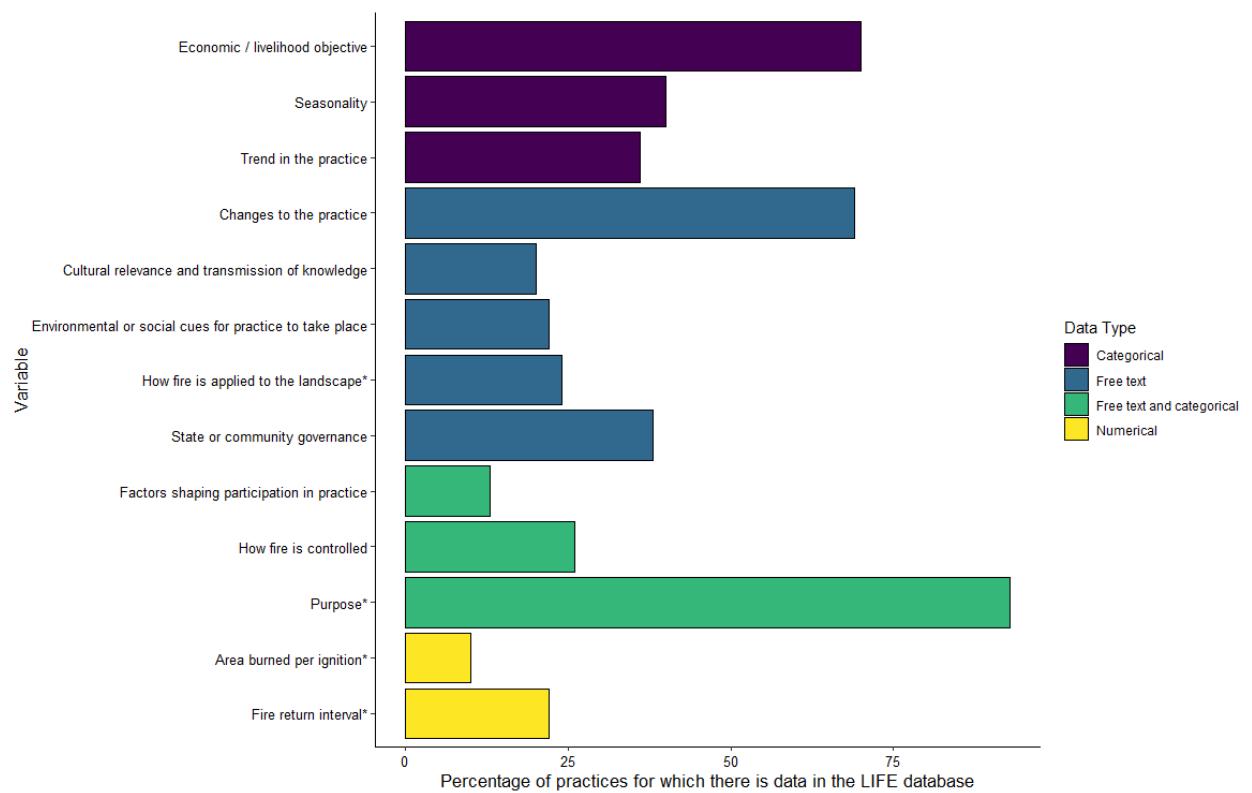


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417 **b.** Status (decreasing/increasing) of market- and subsistence-oriented practices in the LIFE database,  
418 by practice type.



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428 **Fig. 3** Completeness of LIFE database for selected variables. \*Variable applicable only to fire use  
429 practices.  
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## References

1. Bowman, D. M. et al. The human dimension of fire regimes on Earth. *Journal of biogeography* 38(12), 2223-2236 (2011).
2. Huffman, M. R. The many elements of traditional fire knowledge: synthesis, classification, and aids to cross-cultural problem solving in fire-dependent systems around the world. *Ecology and Society* 18(4) (2013).
3. Bird, R. B., Bird, D. W., Codding, B. F., Parker, C. H. & Jones, J. H. The “fire stick farming” hypothesis: Australian Aboriginal foraging strategies, biodiversity, and anthropogenic fire mosaics. *Proceedings of the National Academy of Sciences* 105(39), 14796-14801 (2008).
4. Trauernicht, C., Brook, B. W., Murphy, B. P., Williamson, G. J. & Bowman, D. M. Local and global pyrogeographic evidence that indigenous fire management creates pyrodiversity. *Ecology and Evolution* 5(9), 1908-1918 (2015).

5. Doerr, S. H. & Santín, C. Global trends in wildfire and its impacts: perceptions versus realities in a changing world. *Philosophical Transactions of the Royal Society B: Biological Sciences* 371(1696), 20150345 (2016).
6. Andela, N. *et al.* A human-driven decline in global burned area. *Science*, 356(6345), 1356-1362 (2017).
7. Bilbao, B. *et al.* in *Adaptation to climate change risks in Ibero-American countries — RIOCCADAPT report* (eds Bilbao, B *et al.*) 435-496 (McGraw Hill, Madrid, Spain, 2020).
8. Goss, M. *et al.* Climate change is increasing the likelihood of extreme autumn wildfire conditions across California. *Environmental Research Letters* 15(9), 094016 (2020).
9. Flannigan, M. *et al.* Global wildland fire season severity in the 21st century. *Forest Ecology and Management*, 294 54-61 (2013).
10. Andela, N. *et al.* The Global Fire Atlas of individual fire size, duration, speed, and direction. *Earth System Science Data* 11(2), 529-552 (2019).
11. Carmenta, R., Parry, L., Blackburn, A., Vermeylen, S. & Barlow, J. Understanding human-fire interactions in tropical forest regions: a case for interdisciplinary research across the natural and social sciences. *Ecology and society* 16(1) (2011).
12. Scherjon, F., Bakels, C., MacDonald, K. & Roebroeks, W. Burning the land: an ethnographic study of off-site fire use by current and historically documented foragers and implications for the interpretation of past fire practices in the landscape. *Current Anthropology* 56(3), 299-326 (2015).
13. Coughlan, M. R., Magi, B. I. & Derr, K. M. A global analysis of hunter-gatherers, broadcast fire use, and lightning-fire-prone landscapes. *Fire* 1(3), 41 (2018).
14. Nikolakis, W. & Roberts, E. Indigenous fire management: a conceptual model from literature. *Ecology and Society* 25(4) (2020).
15. Coughlan, M. R. & Petty, A. M. Linking humans and fire: a proposal for a transdisciplinary fire ecology. *International Journal of Wildland Fire* 21(5), 477-487 (2012).
16. Teckentrup, L. *et al.* Response of simulated burned area to historical changes in environmental and anthropogenic factors: a comparison of seven fire models. *Biogeosciences*, 16(19), 3883-3910 (2019).
17. Benali, A. *et al.* Bimodal fire regimes unveil a global-scale anthropogenic fingerprint. *Global Ecology and Biogeography* 26(7), 799-811 (2017).
18. Lauk, C. & Erb, K-H. in *Social ecology: society-nature relations across time and space* (eds Haberl, H., Fischer-Kowalski, M., Krausmann, F. & Winiwarter, V.) 335–348 (Springer, 2016).

19. Laris, P. Burning the seasonal mosaic: preventative burning strategies in the wooded savanna of southern Mali. *Human Ecology* 30(2), 155-186 (2002).
20. Bliege Bird, R. & Bird, D.W. *Current Anthropology* 56(3), 314-315 (2015).
21. Pyne, S. J. *Vestal fire: an environmental history, told through fire, of Europe and Europe's encounter with the world* (University of Washington Press, 1997).
22. Kull, C. A. *Isle of fire: the political ecology of landscape burning in Madagascar* (University of Chicago Press, 2004).
23. Moura, L. C., Scariot, A. O., Schmidt, I. B., Beatty, R. & Russell-Smith, J. The legacy of colonial fire management policies on traditional livelihoods and ecological sustainability in savannas: Impacts, consequences, new directions. *Journal of environmental management* 232, 600-606 (2019).
24. Welch, J. R. & Coimbra Jr, C. E. Indigenous fire ecologies, restoration, and territorial sovereignty in the Brazilian Cerrado: the case of two Xavante reserves. *Land Use Policy*, 104055 (2019).
25. Temudo, M. P., Oom, D. and Pereira, J. M. Bio-cultural fire regions of Guinea-Bissau: Analysis combining social research and satellite remote sensing. *Applied Geography* 118, 102203 (2020).
26. Johnston, F. H., Jacups, S. P., Vickery, A. J. & Bowman, D. M. Ecohealth and Aboriginal testimony of the nexus between human health and place. *EcoHealth* 4(4), 489-499 (2007).
27. Peller, H.A. Soil fertility, agroecology, and social change in southern Belize. (The Ohio State University, 2021).
28. Angassa, A. & Oba, G. Herder perceptions on impacts of range enclosures, crop farming, fire ban and bush encroachment on the rangelands of Borana, Southern Ethiopia. *Human ecology* 36(2), 201-215 (2008).
29. Rai, N. D., Benjaminsen, T. A., Krishnan, S. & Madegowda, C. Political ecology of tiger conservation in India: Adverse effects of banning customary practices in a protected area. *Singapore Journal of Tropical Geography* 40(1), 124-139 (2019).
30. Lestrelin, G., Vigiak, O., Pelletreau, A., Keohavong, B. & Valentin, C. Challenging established narratives on soil erosion and shifting cultivation in Laos. *Natural Resources Forum* 36(2012), 63–75 (2012).
31. Mathews, A. S. Power/knowledge, power/ignorance: forest fires and the state in Mexico. *Human Ecology* 33(6), 795-820 (2005).
32. Dressler, W. H., Smith, W., Kull, C. A., Carmenta, R. & Pulhin, J. M. Recalibrating burdens of blame: anti-swidden politics and green governance in the Philippine Uplands. *Geoforum* (2020).

33. Bremer, L. L., Farley, K. A., Lopez-Carr, D. & Romero, J. Conservation and livelihood outcomes of payment for ecosystem services in the Ecuadorian Andes: what is the potential for ‘win-win’? *Ecosystem Services* 8, 148-165 (2014).
34. Dobler-Morales, C., Roy Chowdhury, R. & Schmook, B. Governing intensification: the influence of state institutions on smallholder farming strategies in Calakmul, Mexico. *Journal of Land Use Science* 15(2-3), 108-126 (2020).
35. Rodríguez, I. Pemon perspectives of fire management in Canaima National Park, southeastern Venezuela. *Human Ecology* 35(3), 331-343 (2007).
36. Dressler, W. & Pulhin, J. The shifting ground of swidden agriculture on Palawan Island, the Philippines. *Agriculture and Human Values* 27(4), 445-459 (2010).
37. Codding, B. F., Bird, R. B., Kauhanen, P. G. & Bird, D. W. Conservation or co-evolution? Intermediate levels of aboriginal burning and hunting have positive effects on kangaroo populations in Western Australia. *Human Ecology* 42(5), 659-669 (2014).
38. Yaro, J. A. & Tsikata, D. Savannah fires and local resistance to transnational land deals: the case of organic mango farming in Dipale, northern Ghana. *African Geographical Review* 32(1), 72-87 (2013).
39. Chokkalingam, U., Kurniawan, I. & Ruchiat, Y. Fire, livelihoods, and environmental change in the middle Mahakam peatlands, East Kalimantan. *Ecology and Society* 10(1), (2005).
40. Erenstein, O. Cropping systems and crop residue management in the Trans-Gangetic Plains: Issues and challenges for conservation agriculture from village surveys. *Agricultural Systems* 104(1), 54-62 (2011).
41. Scott, J. C. *Weapons of the weak: everyday forms of peasant resistance* (Yale University Press, 1985).
42. Dennis, R. A. *et al.* Fire, people and pixels: linking social science and remote sensing to understand underlying causes and impacts of fires in Indonesia. *Human Ecology* 33(4), 465-504 (2005).
43. Wang, F., Wang, M. & Yin, H. Can campaign-style enforcement work: when and how? Evidence from straw burning control in China. *Governance* 2021, 1–20 (2021).
44. Van Vliet, N. *et al.* Trends, drivers and impacts of changes in swidden cultivation in tropical forest-agriculture frontiers: a global assessment. *Global Environmental Change* 22(2), 418-429 (2012).

45. Carmenta, R., Coudel, E. & Steward, A. M. Forbidden fire: does criminalising fire hinder conservation efforts in swidden landscapes of the Brazilian Amazon?. *The Geographical Journal* 185(1), 23-37 (2019).
46. Fule, P. Z., Ramos-Gómez, M., Cortés-Montaño, C. & Miller, A. M. Fire regime in a Mexican forest under indigenous resource management. *Ecological Applications* 21(3), 764-775 (2011).
47. McGregor, S. *et al.* Indigenous wetland burning: conserving natural and cultural resources in Australia's World Heritage-listed Kakadu National Park. *Human Ecology* 38(6), 721-729 (2010).
48. Altman, J., Ansell, J. & Yibarbuk, D. No ordinary company: Arnhem Land Fire Abatement (Northern Territory) Limited. *Postcolonial Studies* 23(4), 552-574 (2020).
49. Petty, A. M., deKoninck, V. & Orlove, B. Cleaning, protecting, or abating? Making Indigenous fire management "work" in Northern Australia. *Journal of Ethnobiology* 35(1), 140-162 (2015).
50. Laris, P. On the problems and promises of savanna fire regime change. *Nature Communications* (In Press).
51. Morton, J. F. The impact of climate change on smallholder and subsistence agriculture. *Proceedings of the national academy of sciences* 104(50), 19680-19685 (2007).
52. Khalil, C. A., Conforti, P., Ergin, I. & Gennari, P. *Defining small-scale food producers to monitor target 2.3. of the 2030 agenda for sustainable development* (Food and Agriculture Organisation of the United Nations (FAO), 2017); <http://www.fao.org/3/i6858e/i6858e.pdf>
53. Perkins, O. & Millington, J. DAFl: a global database of anthropogenic fire. figshare <https://doi.org/10.6084/m9.figshare.c.5290792> (2021).
54. Fire Research Institute. (2021). At <<http://fireresearchinstitute.org/>>
55. Diemont, S. A. Ecosystem management and restoration as practiced by the indigenous Lacandon Maya of Chiapas, Mexico. (The Ohio State University, 2006).
56. Hecht, S. B. in *Amazonian dark earths: Wim Sombroek's vision* (eds Woods, W. I. *et al.*) 143-162 (Springer, 2009).
57. Toussaint, M. L'épreuve du feu: politiques de la nature, savoirs, feux de brousse et décolonisation en Nouvelle-Calédonie. (Paris Sciences et Lettres, 2018).
58. Sorrensen, C. Contributions of fire use study to land use/cover change frameworks: understanding landscape change in agricultural frontiers. *Human Ecology* 32(4), 395-420 (2004).
59. Coughlan, M. R. Errakina: pastoral fire use and landscape memory in the Basque region of the French Western Pyrenees. *Journal of Ethnobiology* 33(1), 86-104 (2013).
60. Davydov, V. Coming back to the same places: the ethnography of human-reindeer relations in the Northern Baikal region. *Journal of Ethnology and Folkloristics* 8(2), 7-32 (2014).

61. Johansson, M. U., Fetene, M., Malmer, A. & Granström, A. Tending for cattle: traditional fire management in Ethiopian montane heathlands. *Ecology and Society* 17(3) (2012).
62. Vaarzon-Morel, P. & Gabrys, K. Fire on the horizon: contemporary Aboriginal burning issues in the Tanami Desert, central Australia. *GeoJournal* 74(5), 465 (2009).
- 63 Sletto, B. Conservation planning, boundary-making and border terrains: The desire for forest and order in the Gran Sabana, Venezuela. *Geoforum* 42(2), 197-210 (2011).
- 64 Welch, J. R. Xavante ritual hunting: anthropogenic fire, reciprocity, and collective landscape management in the Brazilian cerrado. *Human ecology* 42(1), 47-59 (2014).
- 65 Walters, G. The Land Chief's embers: ethnobotany of Batéké fire regimes, savanna vegetation and resource use in Gabon. (University College London, 2010).
- 66 Long, J. W., Goode, R. W., Gutteriez, R. J., Lackey, J. J. & Anderson, M. K. Managing California black oak for tribal ecocultural restoration. *Journal of Forestry* 115(5), 426-434 (2017).
- 67 Nanda, P. K. & Sutar, P. C. in *Community-based fire management: case studies from China, The Gambia, Honduras, India, the Lao People's Democratic Republic and Turkey*, 80-96 (Food and Agriculture Organisation of the United Nations (FAO), 2003); <https://coin.fao.org/coin-static/cms/media/9/13171073526480/ad348e00.pdf>
- 68 Mulder, V., Heri, V. & Wickham, T. Traditional honey and wax collection with *Apis dorsata* in the upper Kapuas Lake Region, West Kalimantan. *Borneo Research Bulletin* 31, 246-261 (2000).
- 69 Butz, R. J. Changing land management: a case study of charcoal production among a group of pastoral women in northern Tanzania. *Energy for Sustainable Development* 17(2), 138-145 (2013).
- 70 Kepe, T. Grasslands ablaze: vegetation burning by rural people in Pondoland, South Africa. *South African Geographical Journal* 87(1), 10-17 (2005).
- 71 Bernard, T. & Moetapele, N. Desiccation of the Gomoti River: biophysical process and Indigenous resource management in Northern Botswana. *Journal of Arid Environments* 63(1), 256-283 (2005).
- 72 McDaniel, J., Kennard, D. & Fuentes, A. Smokey the tapir: traditional fire knowledge and fire prevention campaigns in lowland Bolivia. *Society and Natural Resources* 18(10), 921-931 (2005).
- 73 Mistry, J. et al. Indigenous fire management in the cerrado of Brazil: the case of the Krahô of Tocantins. *Human ecology* 33(3), 365-386 (2005).
74. Rodríguez, I., Albert, P., La Rose, C. & Sharpe, C. *A study of the use of fire by Amerindian communities in South Rupununi, Guyana, with recommendations for sustainable land management* (Forest Peoples Project, 2011).
75. Christianson, A. N. Wildfire risk perception and mitigation at Peavine Métis Settlement. (The University of Alberta, 2011).

76. Humphrey, G. The role of humans, climate, and vegetation in the complex fire regimes of north-east Namibia. (The University of Cape Town, 2018).
77. Seijo, F. The politics of fire: Spanish forest policy and ritual resistance in Galicia, Spain. *Environmental Politics* 14(3), 380-402 (2005).
78. Seijo, F. *et al.* Forgetting fire: traditional fire knowledge in two chestnut forest ecosystems of the Iberian Peninsula and its implications for European fire management policy. *Land Use Policy* 47, 130-144 (2015).
79. McKemey, M. B. *et al.* Cross-cultural monitoring of a cultural keystone species informs revival of Indigenous burning of country in South-Eastern Australia. *Human Ecology* 47(6), 893-904 (2019).

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