

# Short-term Effects of Multiple Ozone Metrics on Outpatient Visits For Urticaria in Lanzhou

Jing Zhang

LZU: Lanzhou University

Yuan He

Lanzhou University Second Hospital

chunrui Shi (✉ [shichr@lzu.edu.cn](mailto:shichr@lzu.edu.cn))

Lanzhou University First Affiliated Hospital

---

## Research Article

**Keywords:** Ambient Ozone, Urticaria, Excess Risk, Metric

**Posted Date:** July 29th, 2021

**DOI:** <https://doi.org/10.21203/rs.3.rs-693728/v1>

**License:** © ⓘ This work is licensed under a Creative Commons Attribution 4.0 International License.

[Read Full License](#)

---

1 Title: Short-term effects of multiple ozone metrics on outpatient visits for urticaria in Lanzhou

2

3 Author: Jing Zhang<sup>1</sup>, Yuan He<sup>3</sup>, Chunrui Shi<sup>2</sup>

4

5 1. The First Clinical Medical College of Lanzhou University, Lanzhou 730000, PR China

6 2. Department of Dermatology, the First Hospital of Lanzhou University, Lanzhou 730000, PR China

7 3. Department of Dermatology, Lanzhou University Second Hospital, Lanzhou 730000, PR China

8

9 Corresponding author: Chunrui Shi

10 Address: No.1, Donggangxi Rd, Chengguan District, Lanzhou City, Gansu Pro., China 730000

11 Telephone: (+86) 0931-8356092; E-mail: shichr@lzu.edu.cn

12

13

14

15

16

17

18

19

20

21

22

23

24

25 **Abstract**

26 **Background** Although mounting evidence suggests that short-term exposure to ozone increases the risk

27 of respiratory disease, cardiovascular disease and mortality, there are few studies comparing the effects  
28 of ozone in relation to urticaria in China.

29 **Objective** To evaluate the risks for urticaria exacerbations related to ambient ozone measured as 1-h  
30 maximum ( $O_3$ -1 h max), maximum 8-h ( $O_3$ -8 h max) and 24-h average ( $O_3$ -24 h avg) concentrations.

31 **Methods** We calculated three metrics of ozone, 1-h maximum, maximum 8-h and 24-h average based  
32 on the hourly data. Generalized additive models with poisson regression incorporating natural spline  
33 functions were used to investigate short-term effects on urticaria associated with ambient ozone  
34 pollution in Lanzhou, China, using 5 years of daily data (2013-2017). We also examined the association  
35 by sex, age and season.

36 **Results** In all-year analyses, a  $10 \mu\text{g}/\text{m}^3$  increase in daily average,  $O_3$ -1 h max,  $O_3$ -8 h max and  $O_3$ -24 h  
37 avg at lag2 corresponded to an increase of 0.58%(95%CI: 0.26%~0.90%), 0.82% (95%CI:  
38 0.47%~1.16%) and 2.17% (95%CI: 1.17%~2.79%), respectively. The elderly populations and females  
39 were susceptible to  $O_3$ , and the associations between ozone and urticaria appeared to be more evident  
40 during warm season than in the cold season.

41

42 **Conclusion** In conclusion, these results indicated that ozone, as a widespread pollutant, affects  
43 outpatient visits for urticaria in Lanzhou.

44 **Key word:** Ambient Ozone; Urticaria; Excess Risk; Metric

45

46

47

48

49

50

51

52

53 **Funding** None.

54

55 **Conflict of interest** None declared.

56

57 **Ethics approval** This study was approved by the Ethics Committee of the First Hospital of Lanzhou  
58 University (the Ethics number: LDYYLL2019-35).

59

60 Availability of data and materials Weather data are available from: <http://data.cma.cn/site/index.html>.

61 Outpatient data were obtained from the three major tertiary hospitals through the Health Information  
62 System, which has not deposited in publicly available repositories. Therefore, it is available from the  
63 corresponding author on reasonable request. Air quality data are obtained from 4 monitoring stations  
64 interspersed in study areas.

65

66

67

68

69

70

71

72

73

74

## 75 **1. Introduction**

76 In recent years, with the development of the economy and science and technology, the  
77 concentration of atmospheric particulate matter in China has been gradually controlled, and the  
78 type of air pollution has begun to change from traditional soot pollution to photochemical smog  
79 pollution. Among them, O<sub>3</sub> is the most toxic component of photochemical smog pollution, and the  
80 concentration of O<sub>3</sub> in the near-surface atmosphere is increasing year by year (Huang et al. 2018).

81 It has been reported that an increase in O<sub>3</sub> concentration can cause a series of acute health hazards,  
82 such as increased risk of skin disease, upper respiratory tract infections and cardiovascular disease  
83 ( Gryparis et al. 2004; Bell et al. 2006; Jerrett et al. 2009; Zhang et al. 2006).

84 In the human body, the skin is the part that comes into most frequent and extensive contact  
85 with the external environment, and it serves as a defense barrier, constantly exposed to air  
86 pollutants and sensing changes in the environment. Urticaria is a common allergic disease with  
87 recurring symptoms and pronounced itching, and the prevalence of urticaria has been increasing  
88 worldwide over the past decade (Zuberbier et al. 2010; Lapi et al. 2016), placing a heavy burden on  
89 patients' quality of life, workability, mental health and socio-economic well-being (Maurer et al.  
90 2016; Broder et al. 2015; Graham et al. 2016). The etiology of urticaria is unknown, but there are a  
91 number of risk factors, both endogenous (e.g. genetic) and exogenous (e.g. environmental),  
92 including indoor environmental conditions, ambient air pollutants and other meteorological factors.  
93 Particulate matters (PM), which are representative of ambient air pollutants, have been shown to  
94 have an effect on urticaria (Kousha et al. 2015). As the second most important air pollutant in  
95 China, O<sub>3</sub> has been little studied in relation to urticaria disease. Only Xu Feng (2011) et al. in  
96 Shanghai found a correlation between the maximum daily 8-h concentration of O<sub>3</sub> and the number  
97 of emergency room visits for urticaria. Since the mechanism of O<sub>3</sub> production is influenced by  
98 solar radiation, temperature and humidity (Wang et al. 2017), and ozone concentrations vary  
99 widely over a 24-hour period, three commonly used concentration monitoring measures are used.  
100 According to the methods of US EPA (Darrow et al. 2011), which included O<sub>3</sub>-1 h max (the highest  
101 hourly value recorded in a given day), O<sub>3</sub>-8 h max (the maximum running or moving 8-hr values  
102 among all 24 h in a day) and O<sub>3</sub>-24 h avg (the mean of 24 individual hourly concentrations measured  
103 from midnight to midnight). Therefore, when studying the health effects of O<sub>3</sub>, it is difficult to  
104 obtain accurate and comprehensive results on the health effects of O<sub>3</sub> in a population if the  
105 concentration of a single measure is used without considering the variation of concentrations in  
106 other measures (Li et al. 2015).

107 To date, epidemiological studies of O<sub>3</sub> using various indicators have only focused on  
108 developed cities in China, such as Beijing (Li et al. 2015), Shanghai and Guangzhou (Sun et al.  
109 2017) etc, and it is difficult to compare the results from these regions. Lanzhou is a landlocked  
110 city in China, with a unique geographical location and meteorological conditions, and

111 photochemical pollution has been one of the problems affecting air pollution in the city  
112 (Wiwatanadate et al. 2014). Therefore, we consider the Lanzhou region as an ideal area to evaluate  
113 the effect of O<sub>3</sub> on urticaria disease.

114 In this study, a Generalised Additive Model (GAM) based on Poisson distribution is proposed  
115 to quantitatively evaluate the near-surface O<sub>3</sub> monitoring data of municipality Lanzhou from 2013  
116 to 2017, using various measures of O<sub>3</sub> concentration, including O<sub>3</sub>-1 h max, O<sub>3</sub>-8 h max and  
117 O<sub>3</sub>-24 h avg to explore the acute health effects of different O<sub>3</sub> concentrations on the urticaria  
118 population. The results of this study may provide clues or evidence for further research on the  
119 effects of environmental ozone on dermatological diseases, and ultimately provide useful  
120 information for our environmental regulatory policies.

121

## 122 **2. Materials and Methods**

### 123 *2.1. Data collection*

124 The clinical data of urticaria were obtained from the medical records of three Grade-A general hospitals  
125 in Lanzhou city. The data were selected from January 1, 2013 to December 31, 2017, and were  
126 exported by the hospital information system (HIS). According to the first four area codes and the ID  
127 number of the home address, the local residents in Lanzhou were screened out. The cases were coded  
128 according to the International Classification of Disease, tenth revision (ICD-10) for urticaria (code:  
129 L50) by the clinicians in a division of Dermatology.

130 Hourly concentrations of ambient air pollutants including Ozone (O<sub>3</sub>), Nitrogen dioxide (NO<sub>2</sub>), Sulfur  
131 dioxide (SO<sub>2</sub>), Particulate Matter (with an aerodynamic diameter  $\leq 2.5 \mu\text{m}$ , PM<sub>2.5</sub>) and Carbon monoxide  
132 (CO) were obtained from Lanzhou Municipal Environmental Monitoring Center. There were four  
133 fixed-site monitoring stations in our study, and the location of the four monitoring stations and the three  
134 tertiary hospitals were shown in the Supplemental Material (Fig. S.1). Hourly meteorological data  
135 (including temperature and relative humidity) measured by national environmental monitoring stations

136 were collected from China Meteorological Administration. We calculated the following three O<sub>3</sub>  
137 exposure metrics for daily O<sub>3</sub> concentrations on an hourly basis including O<sub>3</sub>-1 h max, O<sub>3</sub>-8 h max  
138 and O<sub>3</sub>-24 h avg.

139

## 140 2.2. Statistical Methods

141 Spearman's correlation coefficients were used to analyze the relationships between air pollutants and  
142 meteorological factors. As the number of urticaria outpatient visits approximately obeyed an  
143 overdispersion Poisson distribution, a generalized additive model (GAM) with Poisson chain as the  
144 core analysis was used to estimate the relationship between ambient air pollution and the number of  
145 daily visits to the urticaria outpatient clinic. In this model, we incorporated a days of week (DOW)  
146 categorical variable to control for variation in weekly outpatient visits. We also used a penalized  
147 smoothing spline function incorporating calendar time, daily temperature and relative humidity to  
148 adjust for the effects of seasonal patterns, long-term trends, and meteorological factors and to exclude  
149 their potential nonlinear confounding effects. The A cubic smoothing function is used to control for the  
150 confounding effects of long-term trends, Sunday numbers, and weather factors (temperature and  
151 relative humidity). The model is as following:

152 GAM is shown below (Equation (1)):

$$153 \log E(y_t) = \beta Z_t + ps(time, df = 7) + ps(temp, df = 3) + ps(rhum, df = 3) +$$
$$154 \quad factor(dow) + \alpha \tag{1}$$

155 Where,  $E(y_t)$  defined as the expected number of outpatients visits for urticaria at day  $t$ ;  $\beta$  represents the  
156 relative log rate of urticaria outpatients related to a unit increase of particulate matter;  $Z_t$  indicates the  
157 concentrations of O<sub>3</sub> (O<sub>3</sub>-1 h max, O<sub>3</sub>-8 h max and O<sub>3</sub>-24 h avg) at day  $t$ ;  $ps$  indicates penalized

158 spline function for nonlinear variables, which is adapted to control for potential confounding of  
159 long-term trends and seasonality in daily outpatients visits;  $df$  is the degree of freedom;  $temp$  and  $rhum$   
160 are the daily temperatures ( $^{\circ}\text{C}$ ) and relative humidity (%) at day  $t$ ;  $time$  represents the calendar time;  
161  $dow$  is the dummy variable for the date of the week;  $\alpha$  is the intercept for the model (Li et al. 2015; Sun  
162 et al. 2017).

163 When the core model was established, we introduced air pollution variables into the single pollutant  
164 model to test for possible interactions between urticaria outpatients and air pollutants. Secondly,  
165 because of the delayed health effects of air pollutants, we also considered a moving average lag effect  
166 for the study (e.g., from the current day, lag0 to previous 7 days, lag7) and cumulative day lags (e.g.,  
167 from the current day and the previous 1, lag01 to the previous 8 days, lag07). Third, we also assessed  
168 the robustness of the results in terms of the  $df$  values for calendar time (5, 6 and 8  $df$  per year), and  
169 temperature (4-6), and by adjusting for other pollutants, including  $\text{SO}_2$ ,  $\text{NO}_2$ ,  $\text{PM}_{2.5}$  and  $\text{CO}$ .  
170 Additionally, the associations stratified by sex (female and male), age ( $\leq 18$  years, 19-64 years and  $\geq$   
171 65 years) and season (spring, summer, autumn and winter) were examined. The statistical significance  
172 of subgroup differences was tested through Z-test:  $(\hat{Q}_1 - \hat{Q}_2) \pm 1.96 \sqrt{SE_1^2 + SE_2^2}$  (where  $\hat{Q}_1$  and  $\hat{Q}_2$   
173 were the estimated for age and sex group, and  $SE_1^2$  and  $SE_2^2$  were respective standard errors.)  
174 All analyses were conducted in R Programming Language (V.3.2.2, R Development Core Team) using  
175 the “mgcv” packages. The results of the statistical test were two-sided with  $p$ -values  $< 0.05$  considered  
176 statistically significant. The estimated effect is expressed as the percentage excess risk (ER%) and its  
177 95% confidence interval (95% CI) for urticaria outpatients per 10  $\mu\text{g}/\text{m}^3$   $\text{O}_3$  component increase.

178

### 179 **3. Results**

180 The results of the descriptive analysis of O<sub>3</sub> concentration and meteorological factors and  
181 daily outpatient visits for urticaria are shown in Table 1 From 18 January 2013 to 31 December  
182 2017, the average daily outpatient visits for urticaria in Lanzhou were 40.65±17.34, with slightly  
183 more male than female visits (23.07±10.58). The O<sub>3</sub>-1 h max, O<sub>3</sub>-8 h max and O<sub>3</sub>-24 h avg  
184 values were (80.43±48.75) µg/m<sup>3</sup>, (82.97±45.37) µg/m<sup>3</sup> and (40.14±26.82) µg/m<sup>3</sup> respectively.  
185 The average daily temperature and relative humidity are (11.20 ± 9.93) °C and (49.95 ± 15.57) %,  
186 respectively. Daily mean concentrations of SO<sub>2</sub>, NO<sub>2</sub>, PM<sub>2.5</sub> and CO were (23.41±15.73) µg/m<sup>3</sup>,  
187 (48.09±22.29) µg/m<sup>3</sup>, (56.54±31.86) µg/m<sup>3</sup> and (1.23±0.74) µg/m<sup>3</sup>, respectively. The temporal  
188 trends of O<sub>3</sub> from urticaria outpatient visits and the different measures of O<sub>3</sub> showed a clear  
189 seasonal trend for O<sub>3</sub>-1 h max, O<sub>3</sub>-8 h max and O<sub>3</sub>-24 h avg, with a peak in summer and autumn  
190 and a decrease in winter and spring. The trends in the pattern of urticaria outpatient attendances  
191 were generally similar in the Supplemental Material (Fig. S.2).

192 The Spearman correlation coefficients between air pollutants and meteorological factors are  
193 described in Table 2 The results show that O<sub>3</sub>-1 h max, O<sub>3</sub>-8 h max and O<sub>3</sub>-24 h avg are highly  
194 positively correlated ( $p<0.01$ ), with most of the different indicators for O<sub>3</sub> being positively  
195 correlated with temperature and relative humidity and only O<sub>3</sub>-24 h avg being negatively  
196 correlated with relative humidity,  $r=-0.15$  ( $p<0.01$ ). Three indicators for O<sub>3</sub> are negatively  
197 correlated with SO<sub>2</sub> and CO and positively correlated with PM<sub>2.5</sub>.

198 Table 3 Using the GAM model to control the effects of meteorological factors such as  
199 long-term and seasonal trends, temperature and relative humidity, the associations of O<sub>3</sub>-1 h max,  
200 O<sub>3</sub>-8 h max and O<sub>3</sub>-24 h avg with urticaria outpatient visits were most significant on lag day 2  
201 (lag2), with an increase in ER of 0.58%(95%CI: 0.26%~0.90%), 0.82% (95%CI: 0.47%~1.16%)  
202 and 2.17% (95%CI: 1.17%~2.79%) for each 10 µg/m<sup>3</sup> increase in O<sub>3</sub>-1 h max, O<sub>3</sub>-8 h max and  
203 O<sub>3</sub>-24 h avg, respectively. Therefore, the effect values for lag2 were chosen as the best lag times  
204 for O<sub>3</sub>-1 h max, O<sub>3</sub>-8 h max and O<sub>3</sub>-24 h avg in this study. We also found that O<sub>3</sub>-24 h avg  
205 always had the highest effect value when the effect values were compared in terms of increasing  
206 concentrations of the same unit (10 µg/m<sup>3</sup> or IQR) (see table 4).

207 Fig.1 shows the effect of lag2 on urticaria outpatient attendances for every 10 µg/m<sup>3</sup> increase  
208 in the O<sub>3</sub> exposure index for the different measurement methods. When stratified by gender, the  
209 effect of O<sub>3</sub> was greater in women than in men, with an increase of 0.52%, 0.63% and 1.25% for

210 every 10  $\mu\text{g}/\text{m}^3$  increase in  $\text{O}_3$ -1 h max,  $\text{O}_3$ -8 h max and  $\text{O}_3$ -24 h avg, respectively, which were  
211 not statistically significant. When stratified by age, the effect of  $\text{O}_3$  was more pronounced in those  
212 aged  $\geq 65$  years, but was not statistically significant.

213 In Fig.2, it is shown whether the relationship between  $\text{O}_3$  and the number of urticaria  
214 outpatient visits varies with season for the different measures. We found that the effect values for  
215  $\text{O}_3$  were slightly higher in summer and autumn than in the other two seasons, but were not  
216 statistically significant.

217 As shown in Table 5, when  $\text{SO}_2$  and CO are introduced into the model, the excess risk of  
218 different indicators  $\text{O}_3$  are reduced but still have statistical significance ( $p < 0.05$ ). When  $\text{PM}_{2.5}$  was  
219 introduced, the excess risk increased and it was also statistically significant. However, when  $\text{NO}_2$   
220 was introduced into the model, the results were significantly different. Among them, the effect  
221 value of  $\text{O}_3$ -1 h max increased to 1.26% (95%CI: 0.73%~1.79%) after adding pollutant  $\text{NO}_2$ ,  
222 while the excess risk of  $\text{O}_3$ -8 h max and  $\text{O}_3$ -24 h avg showed a downward trend, respectively 1.15%  
223 and 2.63%, both of which were statistically significant.

224 By changing the *df* of time ( $df=5-8$ ) and *df* of temperature trend ( $df=3-7$ ) in the model, a new  
225 model fitting, compared with the results of the original model change is not obvious, is proved in  
226 our study to establish the model of daily average temperature, time, the two parameters had no  
227 obvious effect on results, the model established in this study is stable. (Table 6.)

228

## 229 4. Discussion

230 In this study, a statistical modelling approach was used and, after controlling for multiple  
231 confounding factors, it was found that elevated levels of  $\text{O}_3$  pollution in Lanzhou had a significant  
232 effect on the increase in urticaria outpatient visits in the whole population. This result is similar to  
233 that of Xu Feng (2011) and Szyszkowicz et al (2012). In the exposure-response relationship  
234 between  $\text{O}_3$  and urticaria outpatient visits (The figures were displayed in the Supplemental Material,  
235 Figure S3), a positive relationship was found between  $\text{O}_3$ -1 h max and  $\text{O}_3$ -8 h max at  
236 concentrations below 100  $\mu\text{g}/\text{m}^3$  and  $\text{O}_3$ -24 h avg below 40  $\mu\text{g}/\text{m}^3$ , confirming that  $\text{O}_3$  can cause  
237 damage to urticaria patients at low concentrations. The Chiang Mai study (Wiwatanadate et al.  
238 2014), however, concluded that there was a negative correlation between  $\text{O}_3$  concentration and the

239 occurrence of rash. This inconsistency may be related to factors such as the climatic characteristics  
240 of the study area, socio-economic conditions, and the age distribution of the study population.

241 In this study, three measurement methods of O<sub>3</sub> exposure index increased by 10µg/m<sup>3</sup>, and  
242 the number of urticaria outpatients increased to different degrees, which meant O<sub>3</sub>-1 h max, O<sub>3</sub>-8  
243 h max and O<sub>3</sub>-24 h avg increased by 0.58%, 0.82% and 2.17% for every 10µg/m<sup>3</sup> increase on lag2  
244 day, suggesting that O<sub>3</sub> has a significant short-term impact on urticaria outpatient visits. By  
245 comparing the results of three different O<sub>3</sub> exposure indicators, it can be seen that O<sub>3</sub>-24 h avg has  
246 the largest effect value, and O<sub>3</sub>-1 h max has the smallest effect value. Therefore, although the  
247 O<sub>3</sub>-8 h max concentration index is currently used as an index that can better represent the acute  
248 health effects of O<sub>3</sub> (Yang et al. 2012). However, in this study, the significance of the other two  
249 exposure indicators cannot be ignored.

250 In a stratified study by sex and age, it was found that outpatient visits for female urticaria  
251 were more affected by O<sub>3</sub> concentration than men. The increase in outpatient visits for people over  
252 65 was related to O<sub>3</sub> concentration. Because there are few similar documents, the accuracy of the  
253 results has not been confirmed.

254 This study found in the seasonal stratification that the relationship between O<sub>3</sub> and urticaria  
255 outpatient visits is more prominent in summer and autumn. There is no statistical significance, but  
256 the possible reasons can be: 1. Ozone has prominent seasonal variability, with average maximum  
257 concentrations ranging from late May to around October, usually 119.1 µg/m<sup>3</sup>, and declining  
258 significantly in winter (Wang et al. 2017). Changes in ozone concentrations also vary throughout  
259 the day and are usually characterised by low levels of concentration between 0:00 and 8: 00,  
260 slowly increasing to the peak between 8:00 and 16: 00, then maintaining high levels of  
261 concentration until 18:00, after which concentrations drop sharply, reaching low levels at 23:00 on  
262 the same day (Zhou et al. 2018). The change pattern of O<sub>3</sub> in a day or a year may be due to the  
263 high temperature in summer and strong solar radiation intensity, which are conditions conducive  
264 to the generation of O<sub>3</sub>. 2. The influence of topographical factors in the valley of Lanzhou.  
265 Lanzhou is surrounded by mountains, with quiet wind all year round, and the transport and  
266 ventilation effects of pollutants are not significant (Li et al. 2020). 3. People often open windows  
267 for ventilation in warm seasons, resulting in increased penetration of outdoor air into the indoor  
268 space. In the cold season, people rarely go out, especially the elderly and the weak. These factors

269 may increase exposure to ambient air pollutants in warm seasons.

270 In this study, it was found that ozone has a strong correlation with particulate matter and that  
271 the effect value of O<sub>3</sub> increases and is statistically significant after the introduction of PM<sub>2.5</sub> into  
272 the model. It has been shown that certain conditions are favourable for the rapid production of  
273 ozone and secondary particulate matter, leading to a complex superposition of ozone and fine  
274 particulate pollutants in the atmosphere. As particulate matter is a complex pollutant and its  
275 composition varies considerably depending on its regional origin, it is not possible to confirm  
276 whether particulate matter is a confounding factor for ozone or a modifying effect, and more  
277 research evidence is needed in the future.

278 It has long been recognized that environmental exposure to ozone affects skin health. By  
279 evaluating the skin of hairless mice exposed to ozone, Thiele et al (1997) found a decrease in  
280 vitamins C and E and the formation of malondialdehyde, a marker of lipid peroxidation, in the  
281 epidermis of hairless mice exposed to varying levels of ozone at increasing and decreasing doses.  
282 McCarthy et al (2013) exposed normal human epidermal keratinocytes (NHEKs) to  
283 environmentally-related ozone concentrations for 30 minutes, and observed that exposure to  
284 0.8ppm ozone resulted in increased DNA damage and depletion of ATP and Sirtuin 3 levels.  
285 However, ozone is also used for its antibacterial and antioxidant effects in the treatment of various  
286 skin diseases such as chronic inflammation, allergic diseases and pruritic skin diseases etc (Bocci  
287 et al. 2015; Abeck et al. 2008). Therefore, the harmful or protective effects of ozone on the skin  
288 need to be explored in further studies.

289

## 290 **5. Conclusions**

291 In this study, it was found that an increase in atmospheric ozone concentrations in Lanzhou  
292 increased the number of urticaria outpatient visits, with a maximum lag time of lag1 for O<sub>3</sub>-1 h  
293 maximum concentration, lag2 for O<sub>3</sub>-8 h maximum concentration and lag2 for O<sub>3</sub>-24 h mean  
294 concentration. There was significant differences in the three ozone exposure indicators for  
295 urticaria outpatient visits, and it could be seen that O<sub>3</sub>-24 h avg had the largest effect value.  
296 Based on gender, age and seasonal analyses at different exposure levels, it was also found that  
297 there was a higher increase in urticaria outpatient visits when O<sub>3</sub> was applied to females, elder

298 population, summer and autumn compared to the rest of the population. The results of the study  
299 provide richer evidence for a comprehensive assessment of the acute effects of ozone on  
300 dermatological diseases and can provide more technical support for policy makers to revise their  
301 public health policies accordingly.

302

## 303 **References**

304 Abeck, D., Plotz, S., 2008. Colloidal silver and ozonized olive oil for atopic dermatitis? *Med*  
305 *Monatsschr Pharm* 31(7), 265-266. <https://pubmed.ncbi.nlm.nih.gov/18808075/>

306• Bell, M.L., Peng, R.D., Dominici, F., 2006. The exposure response curve for ozone and risk of  
307 mortality and the adequacy of current ozone regulations. *Environ health Perspect* 114(4), 532-5  
308 36. <https://doi.org/10.1289/ehp.8816>

309• Bocci, V., Zanardia, I., Valacchi, G., Borrelli, E., Travagli, V., 2015. Validity of oxygen-ozone  
310 therapy as integrated medication form in chronic inflammatory diseases. *Cardiovasc Hematol*  
311 *Disord Drug Targets* 15(2), 127-138. <https://doi.org/10.2174/1871529x1502151209114642>

312• Broder, M.S., Raimundo, K., Antonova, E., Chang, E., 2015. Resource use and costs in an ins  
313 ured population of patients with chronic idiopathic/spontaneous urticaria. *Am J Clin Dermatol* 1  
314 6(4), 313-321. <https://doi.org/10.1007/s40257-015-0134-8>

315• Darrow, L.A., Klein, M., Sarnat, J.A., Mulholland, J.A., Strickland, M.J., Sarnat, S.E., et al., 2  
316 011. The use of alternative pollutant metrics in time-series studies of ambient air pollution and  
317 respiratory emergency department visits. *J Expo Sci Environ Epidemiol* 21(1), 10-19. <https://doi.org/10.1038/jes.2009.49>

319• Graham, J., McBride, D., Stull, D., Halliday, A., Alexopoulos, S.T., Balp, M.M., et al., 2016.  
320 Cost utility of omalizumab compared with standard of care for the treatment of chronic sponta  
321 neous urticaria. *Pharmacoeconomics* 34(8), 815-827. <https://doi.org/10.1007/s40273-016-0412-1>

- 322• Gryparis, A., Forsberg, B., Katsouyanni, K., Analitis, A., Touloumi, G., Schwartz, J., et al., 20  
323 04. Acute effects of ozone on mortality from the “air pollution and health: a European appra  
324 ch” project. *Am J Respir Crit Care Med* 170(10), 1080-1087. [https://doi.org/10.1164/rccm.20040](https://doi.org/10.1164/rccm.200403-333OC)  
325 [3-333OC](https://doi.org/10.1164/rccm.200403-333OC)
- 326 Huang, J., Pan, X., Guo, X., Li, G., 2018. Health impact of China’s Air Pollution Prevention  
327 and Control Action Plan: an analysis of national air quality monitoring and mortality data. *Lan*  
328 *cet Planet Health* 2(7), 313-323. [https://doi.org/10.1016/S2542-5196\(18\)30141-4](https://doi.org/10.1016/S2542-5196(18)30141-4)
- 329• McCarthy, J.T., Pelle, E., Dong, K., Brahmhatt, K., Yarosh, D., Pernodet, N., 2013. Effects of  
330 ozone in normal human epidermal keratinocytes. *Exp Dermatol* 22(5), 360-361. [https://doi.org/](https://doi.org/10.1111/exd.12125)  
331 [10.1111/exd.12125](https://doi.org/10.1111/exd.12125)
- 332• Jens, J.T., Maret, G.T., Maurizio, P., Kenneth, T., Carroll, E.C., Lester, P., 1997. Ozone deplete  
333 s tocopherols and to cotrienols topically applied to murine skin. *FEBS Lett* 401(2), 167-170. [ht](https://doi.org/10.1016/s0014-5793(96)01463-9)  
334 [tps://doi.org/10.1016/s0014-5793\(96\)01463-9](https://doi.org/10.1016/s0014-5793(96)01463-9)
- 335 Jerrett, M., Burnett, R.T., Pope, C.A. 3rd., Ito, K., Thurston, G., Krewski, D., et al., 2009. Lo  
336 ng-term ozone exposure and mortality. *N Engl J Med* 360(11), 1085-1095. [https://doi.org/10.105](https://doi.org/10.1056/NEJMoa0803894)  
337 [6/NEJMoa0803894](https://doi.org/10.1056/NEJMoa0803894)
- 338 Kousha, T., Valacchi, G., 2015. The air quality health index and emergency department visits f  
339 or urticaria in Windsor, Canada. *J Toxicol Environ Health A* 78(8), 524-533. [https://doi.org/10.](https://doi.org/10.1080/15287394.2014.991053)  
340 [1080/15287394.2014.991053](https://doi.org/10.1080/15287394.2014.991053)
- 341 Lapi, F., Cassano, N., Pegoraro, V., Cataldo, N., Heiman, F., Cricelli, I., et al., 2016. Epidemio  
342 logy of chronic spontaneous urticaria: results from a nationwide, population-based study in Italy.  
343 *Br J Dermatol* 174(5), 996-1004. <https://doi.org/10.1111/bjd.14470>
- 344• Li, T., Yan, M., Ma, W., Ban, J., Liu, T., Lin, H., et al., 2015. Short -term effects of multipl  
345 e ozone metrics on daily mortality in a megacity of China. *Environ Sci Pollut Res Int* 22(11)  
346 8738-8746. <https://doi.org/10.1007/s11356-014-4055-5>

- 347• Li, J., Wang, Z., Chen, L., Lian, L., Li, Y., Zhao, L., et al., 2020. WRF-Chem simulations of  
348 ozone pollution and control strategy in petrochemical industrialized and heavily polluted Lanzh  
349 ou City, Northwestern China. *Sci Total Environ* 737:139835. <https://doi.org/10.1016/j.scitotenv.20>  
350 [20.139835](https://doi.org/10.1016/j.scitotenv.2020.139835)
- 351 Maurer, M., Staubach, P., Raap, U., Richter-Huhn, G., Baier-Ebert, M., Chapman-Rothe, N., 2  
352 016. ATTENTUS, a German online survey of patients with chronic urticaria highlighting the b  
353 urden of disease, unmet needs and real-life clinical practice. *Br J Dermatol* 174(4), 892-894. [h](https://doi.org/10.1111/bjd.14203)  
354 [tps://doi.org/10.1111/bjd.14203](https://doi.org/10.1111/bjd.14203)
- 355• Maurer, M., Abuzakouk, M., Bérard, F., Canonica, W., Oude, E.H., Giménez-Arnau, A., et al.,  
356 2017. The burden of chronic spontaneous urticaria is substantial: real-world evidence from AS  
357 SURE-CSU. *Allergy* 72(12), 2005-2016. <https://doi.org/10.1111/all.13209>
- 358 Mieczyslaw, S., Eugeniusz, P., Gordon, S., Brian, H. Rowe., 2012. Ambient ozone and emerge  
359 ncy department visits for skin conditions. *Air Qual. Atmos. Health* 5(3), 303-309. [https://doi.or](https://doi.org/10.1007/s11869-010-0092-5)  
360 [g/10.1007/s11869-010-0092-5](https://doi.org/10.1007/s11869-010-0092-5).
- 361• Sun, Q., Wang, W., Chen, C., Ban, J., Xu, D., Zhu, P., et al., 2017. Acute effect of multiple  
362 ozone metrics on mortality by season in 34 Chinese counties in 2013–2015. *J Intern Med* 283  
363 (5), 481-488. <https://doi.org/10.1111/joim.12724>.
- 364 Wang, WN., Cheng, TH., Gu, XF., Chen, H., Guo, H., Wang, Y., et al., 2017. Assessing Spati  
365 al and Temporal Patterns of Observed Ground-level Ozone in China. *Sci Rep* 7(1), 3615. [https:](https://doi.org/10.1038/s41598-017-03929-w)  
366 [//doi.org/10.1038/s41598-017-03929-w](https://doi.org/10.1038/s41598-017-03929-w)
- 367 Wiwatanadate P., 2014. Acute air pollution-related symptoms among residents in Chiang Mai, T  
368 hailand. *Journal of Environmental Health* 76(6), 76-84.
- 369• Xu, F., Yan, S., Wu, M., Li, F., Xu, X., Song, W., et al., 2011. Ambient ozone pollution as a  
370 risk factor for skin disorders. *Br J Dermatol* 165(1), 224-225. <https://doi.org/10.1111/j.1365-213>  
371 [3.2011.10349.x](https://doi.org/10.1111/j.1365-2133.2011.10349.x)

- 372• Yang, C., Yang, H., Guo, S., Wang, Z., Xu, X., Duan, X., et al., 2012. Alternative ozone metr  
373 ics and daily mortality in Suzhou: the China Air Pollution and Health Effects Study (CAPES).  
374 Sci Total Environ 426: 83-89. <https://doi.org/10.1016/j.scitotenv.2012.03.036>
- 375 Zhang, Y., Huang, W., London, SJ., Song, G., Chen, G., Jiang, L., et al., 2006. Ozone and dai  
376 ly mortality in Shanghai, China. Environ health Perspect 114(8), 1227-1232. [https://doi.org/10.1](https://doi.org/10.1289/ehp.9014)  
377 [289/ehp.9014](https://doi.org/10.1289/ehp.9014)
- 378• Zhou, X., Zhang, T., Li, Z., Tao, Y., Wang, F., Zhang, X., et al., 2018. Particulate and gaseou  
379 s pollutants in a petrochemical industrialized valley city, Western China during 2013-2016. Envi  
380 ron Sci Pollut Res Int 25:15174-15190. <https://doi.org/10.1007/s11356-018-1670-6>
- 381• Zuberbier, T., Balke, M., Worm, M., Edenharter, G., Maurer, M., 2010. Epidemiology of urticar  
382 ia: a representative cross-sectional population survey. Clin Exp Dermatol 35(8), 869-873. [https://](https://doi.org/10.1111/j.1365-2230.2010.03840.x)  
383 [doi.org/10.1111/j.1365-2230.2010.03840.x](https://doi.org/10.1111/j.1365-2230.2010.03840.x)

# Figures

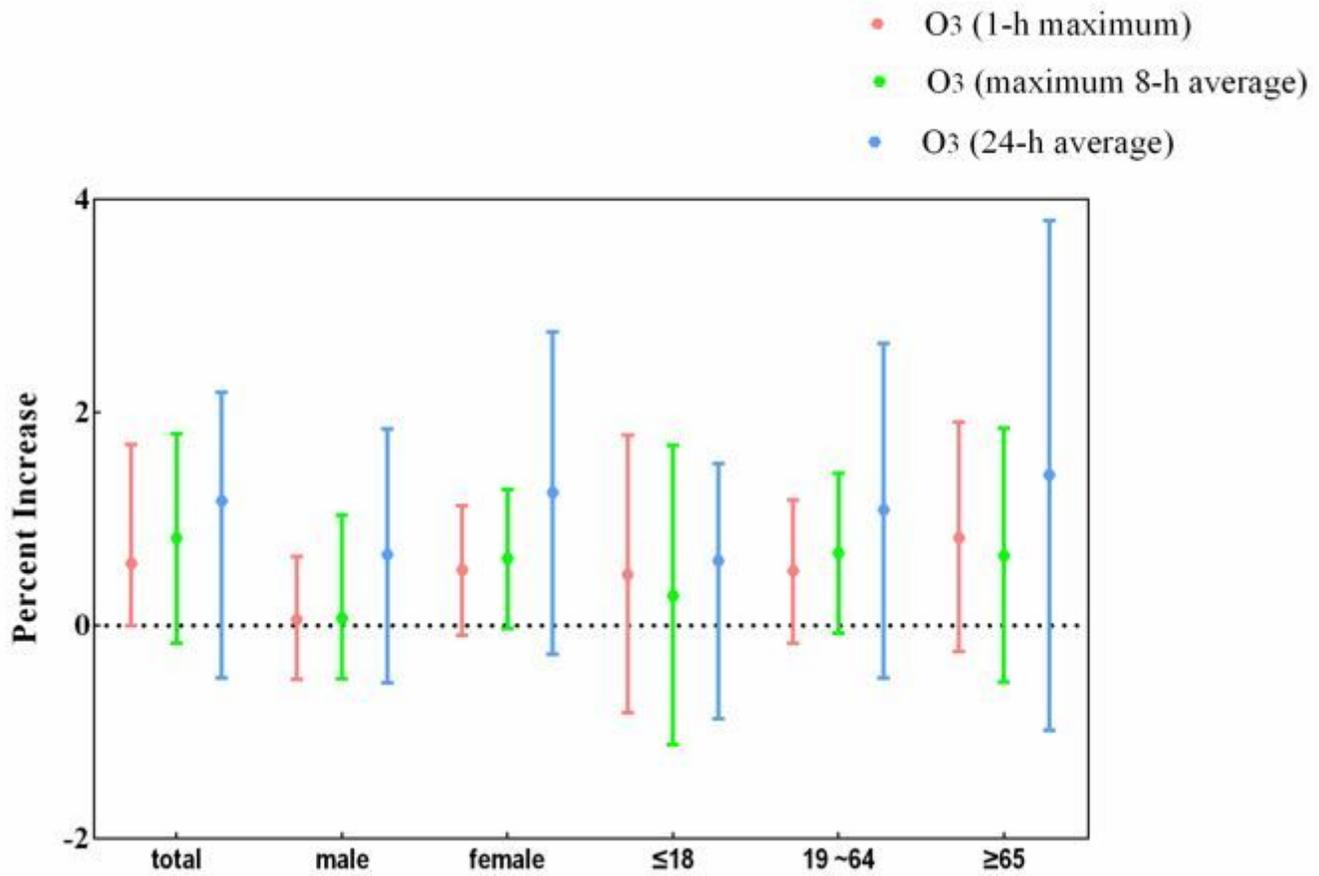


Figure 1

Results of grouped analyses by gender and age on the association between a 10µg/m<sup>3</sup> increases in the average concentration of O<sub>3</sub>-8h and daily outpatient visits for urticaria

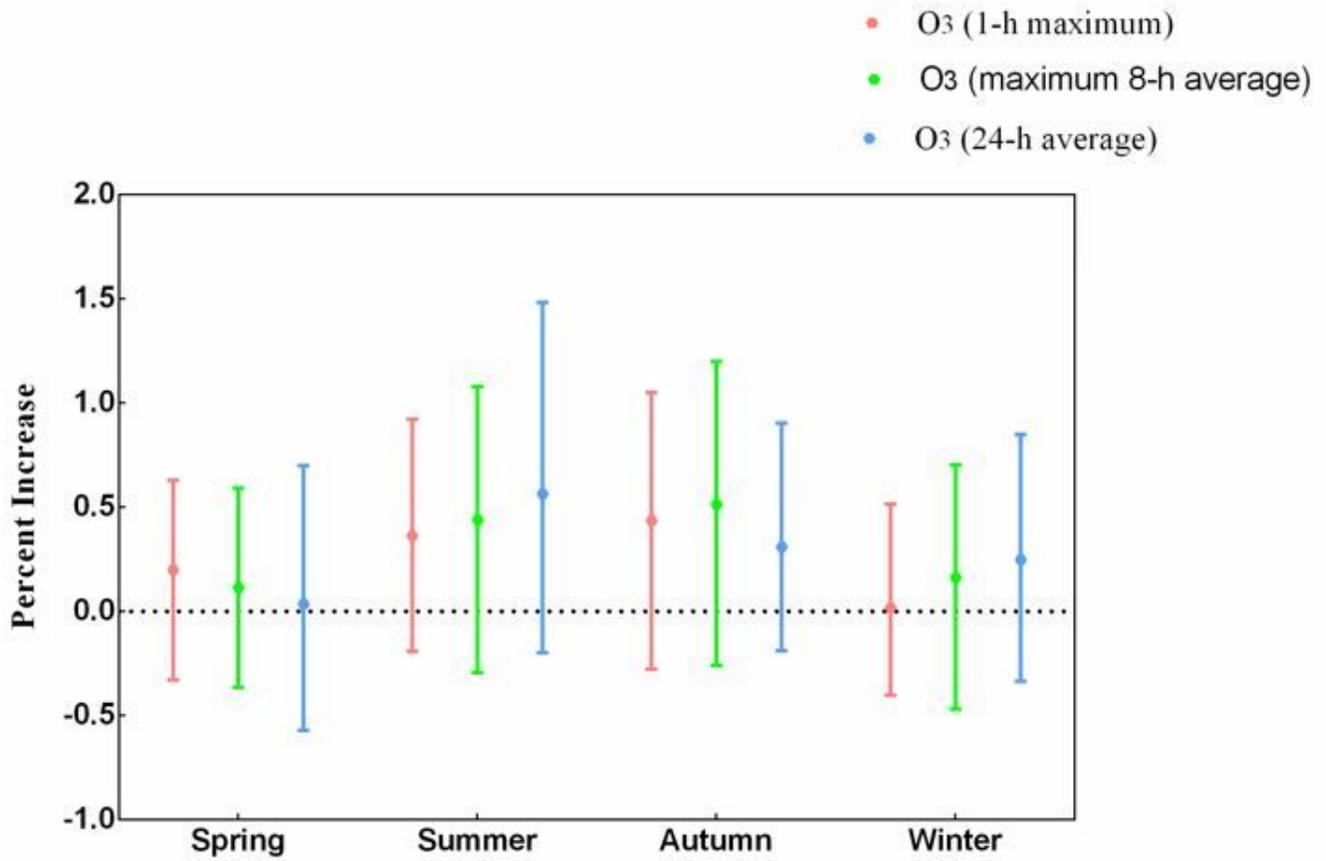


Figure 2

Shows the season-stratified results from a single-day lag and a multi-day cumulative lag.