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Mycelium chemistry differs markedly between ectomycorrhizal and arbuscular mycorrhizal fungi

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Abstract

1 The chemical quality of soil carbon (C) inputs is a major factor controlling litter
2 decomposition and soil C dynamics. Mycorrhizal fungi constitute one of the dominant
3 pools of soil microbial C, while their litter quality is understood poorly, leading to the
4 major uncertainties in estimating soil C dynamics. We examined chemical recalcitrance of
5 arbuscular mycorrhizal (AM) and ectomycorrhizal (EM) fungal species using fungal
6 samples obtained from *in vitro* cultivation. We show that the chemical composition of AM
7 and EM fungal mycelium differs significantly: EM fungi have higher concentrations of
8 labile (water-soluble, ethanol-soluble) and recalcitrant (non-extractable) chemical
9 components, while AM fungi have higher concentrations of acid-hydrolysable components.
10 Our results imply that differences in chemical decomposability traits among mycorrhizal
11 fungal guilds represent a critically important driver of the soil C cycle, which could be as
12 vital as is recognized for differences among aboveground plant litter.

Introduction

13 The soil carbon (C) cycle is among the least understood terrestrial biogeochemical cycles,
14 while it is critically important for ecosystem functioning and mitigation of climate
15 change^{1,2}. A major knowledge gap in this field is the lack of data on belowground influxes
16 of C, and their fate in terms of contribution to stable C pools^{3,4}. A particularly poorly
17 understood aspect is the magnitude of C input into the soil pool of potentially
18 decomposable C components provided by belowground organisms, and decomposability
19 patterns of these organisms⁵. The chemical quality of the C components comprising input
20 of C into the soil is among the key factors that influence the soil C turnover process⁶⁻⁸, and
21 arguably mediate the ultimate fate of soil C, i.e. to be sequestered or respired⁹. Due to the
22 large uncertainty about the contribution of belowground organisms and their chemical

23 quality, until now, the largest known source of variability in the quality of C inputs into the
24 soil has been associated with differences among plant species in terms of aboveground
25 litter quality¹⁰⁻¹². However, that might be an underestimate of the true variability in the
26 quality of C inputs.

27 Our knowledge about the factors that control the decomposability of C compounds entering
28 into the soil pool through residues of microorganisms, especially so from widespread soil-
29 borne fungi, is extremely limited. In soil ecosystems, mycorrhizal fungi living in symbiosis
30 with plant roots are among the key soil microorganisms controlling the exchange of C and
31 nutrients between soil and plants^{13,14}. The living and dead biomass of these microorganisms
32 constitute one of the most dominant pools of soil microbial C^{15,16}. Depending on the soil
33 ecosystem environment and mycorrhizal type, mycorrhizal hyphal biomass can constitute
34 up to half of the standing mycelial biomass¹⁷ and one-third of total microbial biomass¹⁵.
35 Mycorrhizal fungi are important C sinks of net primary production (NPP)^{18,19}, and
36 depending on the mycorrhizal guild, the annual mycelial accumulation can reach around
37 175-200 g C m⁻² ^{20,21}. This is particularly evident in some forest ecosystems, where the
38 allocation of photosynthesized C into fungi can represent up to 30% of the NPP^{20,22}. Yet,
39 the magnitude of the potential contribution of mycorrhizal fungal pools to long-term soil
40 C storage is unknown. Hence, a quantitative assessment of the chemical composition of
41 microorganisms relevant for assessing the decomposability of microbial necromass is
42 critically needed to narrow down the uncertainties in estimating belowground contributions
43 to soil C pools^{23,24}.

44 Among the four principal types of mycorrhiza, the two globally dominant ones are
45 arbuscular mycorrhiza (AM) and ectomycorrhiza (EM)²⁵. These soil fungi associate with
46 the roots of most terrestrial plants²⁵, and are predominant across the majority of the
47 terrestrial vegetated areas²⁶. Ecophysiological traits of these two main guilds of
48 mycorrhizal fungi differ in many aspects (e.g. in the ability of enzymatic degradation of
49 organic matter)^{27,28}. Also, the microscopic structure of AM and EM fungal hyphae differs⁵.
50 EM fungal hyphae have thicker walls, pigmentation, and septa between cells, and are
51 generally believed to have a longer life span than AM fungal hyphae^{16,29}. These differences
52 in morphology could potentially determine the chemistry quality of EM and AM fungi, and
53 have raised the hypothesis that dead EM hyphae are likely more recalcitrant to
54 decomposition than AM hyphae. However, our knowledge about the chemical differences
55 among mycorrhizal fungal guilds, particularly on the chemical components that contribute
56 differently to necromass decomposition, is remarkably limited.

57 Thus far, studies of the impacts of decomposition of fungal mycelium on soil organic
58 matter (SOM) have focused primarily on EM and ericoid mycorrhizal fungi³⁰, and have
59 examined mostly the abundance of individual chemical components in the fungal
60 mycelium, such as concentrations of nitrogen, chitin and melanin. The latter is known to
61 be negatively correlated to necromass decomposition of mycorrhizal fungal biomass^{9,31}.
62 While the outcomes of these analyses shed new light on ecophysiological traits of
63 mycorrhizal fungi at the individual level, they (1) do not provide an integrative assessment

64 of the potential fate of fungal biomass in the process of organic matter decomposition, and
65 (2) neglect the most ancient and widespread mycorrhizal fungal guild, currently associated
66 with the largest part of Earth's terrestrial vegetation – the AM fungi^{25,26}.

67 The objective of this study was to fill a main knowledge gap in the soil C cycle by
68 examining inherent differences between EM and AM fungi in terms of their ultimate
69 decomposability potential. Similar to plant residues that have a variety of components that
70 differ in recalcitrance^{32,33}, soil fungi also consist of components of distinct
71 decomposability²⁴. Fungi contain relatively recalcitrant components, such as melanin, that
72 require costly oxidative enzymes for further decomposition³⁴, as well as relatively labile
73 components (e.g. chitin) that are utilized as a source of C and N for the soil microbial
74 community³⁵. Upon release, these components enrol in principally different types of
75 physical and chemical interactions with mineral surfaces and soil aggregates^{3,36}. Yet the
76 integrative understanding of principal differences among EM and AM fungi in terms of
77 their decomposability is lacking.

78 An important reason underpinning this knowledge gap is the need for samples of *in vitro*
79 pure biomass of mycorrhizal fungi to examine their chemical composition. For the AM
80 fungi, this constitutes a particular challenge due to their obligate symbiotic lifestyle, which
81 requires a suitable host root established on a poor medium to avoid any contamination by
82 unwanted microbes. Using unique methods of cultivation of mycorrhizal fungi^{37–39} well
83 established in the laboratory of mycology of the UCLouvain (Belgium), we cultivated
84 multiple species of AM fungi under *in vitro* culture conditions and obtained amounts of
85 fungal mycelia sufficient to examine their chemical compositions. To assess the differences
86 in chemical traits between AM and EM fungal mycelium, we also cultivated EM fungi *in*
87 *vitro* following standard laboratory techniques⁴⁰. We subsequently assessed the chemical
88 recalcitrance of AM and EM fungal mycelium. With this dataset, we tested two hypotheses
89 crucial to understanding the contribution of major mycorrhizal fungal guilds to the soil C
90 cycle:

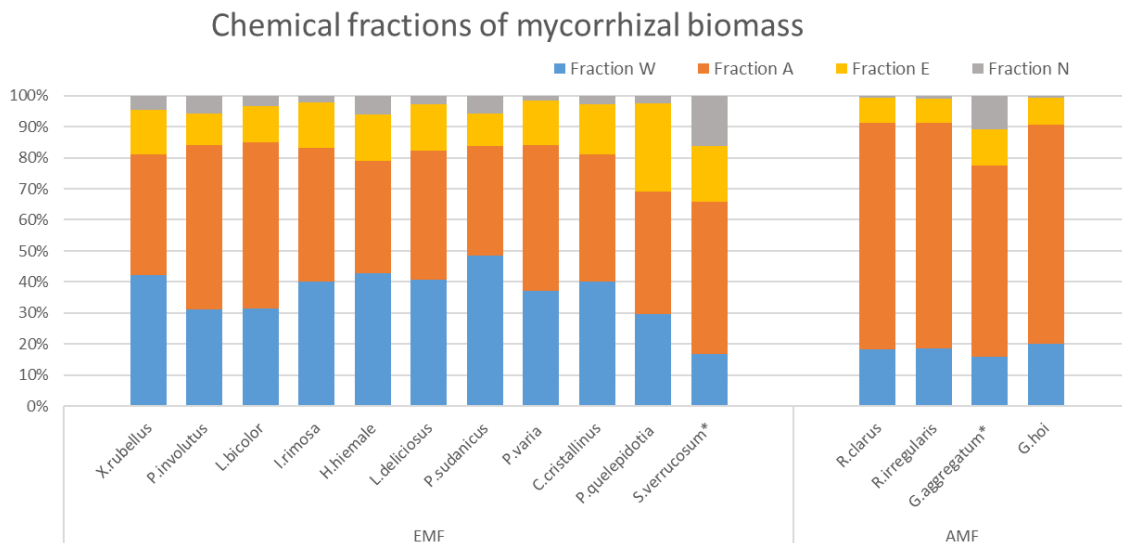
- 91 (1) AM and EM fungal guilds differ principally in their chemical composition traits
92 relevant for decomposability.
- 93 (2) Differences of decomposability between AM and EM fungal guilds are larger than
94 the differences among litters from distinct plant functional types.

95 **Results and discussion**

96 ***Distinct chemical composition of AM and EM fungi***

97 We used 11 species of EM fungi and 4 species of AM fungi from *in vitro* cultures
98 (CBS/MUL number see Table.S1): EM fungal species are *Xerocomus rubellus*, *Paxillus*
99 *involutus*, *Laccaria bicolor*, *Inocybe rimosa*, *Hebeloma hiemale*, *Lactarius deliciosus*,
100 *Phaeogyroporus sudanicus*, *Peziza varia*, *Cortinarius cristallinus*, *Peziza quelepidotia*,
101 and *Scleroderma verrucosum*; AM fungal species are *Rhizophagus clarus*, *Rhizophagus*
102 *irregularis*, *Glomus aggregatum* and *Glomus hoi*. Samples of dried fungal biomass were
103 examined for water-soluble, acid-hydrolyzable, ethanol-soluble and non-extractable

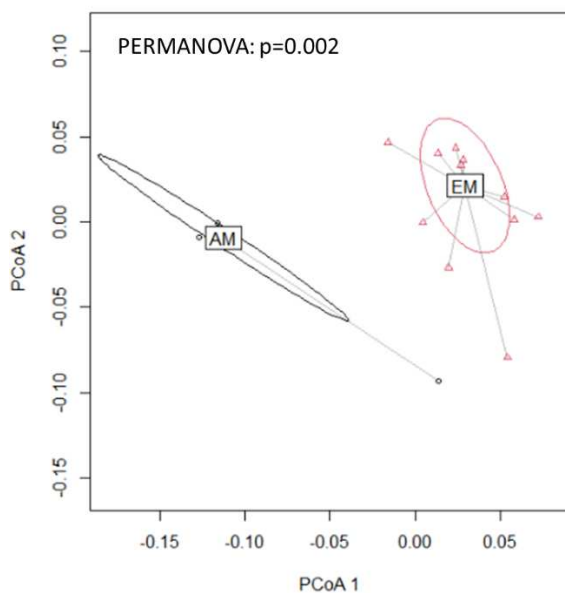
104 components (hereafter W, A, E, N components, respectively), allowing direct comparison
 105 of AM and EM fungi for the entire suite of recalcitrance traits (Fig.1).



106

107 Fig.1 Relative abundance of water-soluble (W), acid-hydrolyzable (A), ethanol-soluble (E) and non-extractable (N)
 108 components in AM and EM fungi. Mycelia of *S.verrucosum** and *G.aggregatum** were assessed in a mixture with plant
 109 litter (details see Methods section).

110 We found that the biomass of AM fungi exhibits a distinct set of decomposability-related
 111 traits compared to that of EM fungi (outcomes of a perMANOVA test on the WAEN
 112 components: $p=0.002$, $df=1$; data for individual fungal species was treated as replicates
 113 within AM and EM fungal guilds). The unambiguous difference between centroids of AM
 114 and EM fungi in the multidimensional space of WAEN components (Fig.2) suggests that
 115 these two groups of fungi are likely to contribute to different pathways of soil C
 116 transformations as being direct sources of soil C compounds.

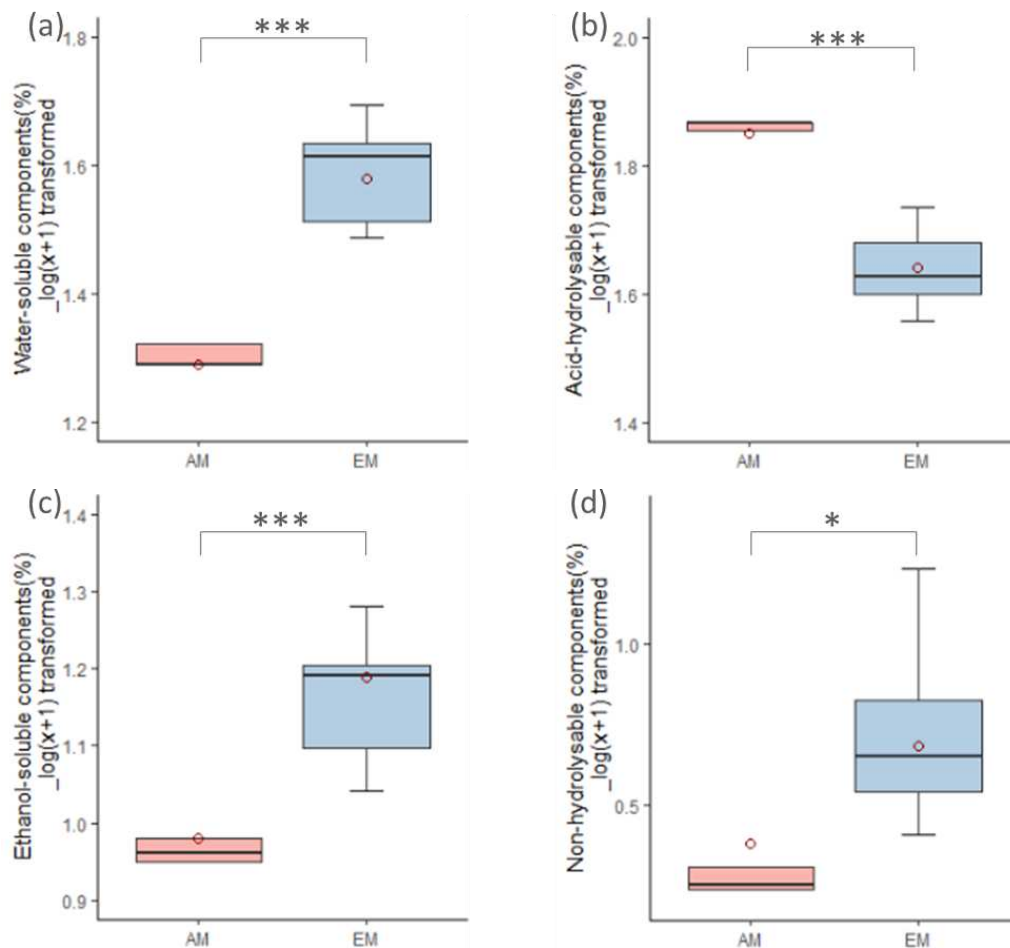


117

118 Fig.2 Clustering and centroids of AM and EM mycorrhizal fungal species in a multidimensional space of WAEN
 119 components.

120 **Analyses of individual chemical compositions**

121 Subsequently, we examined which individual WAEN components differ among AM and
122 EM fungi. We found that concentrations of the most easily decomposable component (W),
123 the ethanol-soluble fraction (E) and the most recalcitrant component (N) were significantly
124 higher in EM fungi (Mann-Whitney tests with data for individual fungi within a fungal
125 guild treated as replicates; $p < 0.001$, $p < 0.001$ and $p < 0.05$, respectively; Fig.3a,c, and d). In
126 contrast, AM fungi had a significantly higher A fraction (Mann-Whitney test conducted in
127 the same manner as for W, A, and N components: $p < 0.001$; Fig.3b). While both A and E
128 components are of intermediate recalcitrance, acid-hydrolysable components have been
129 shown to have higher decomposability than ethanol-soluble components⁴¹. The higher
130 relative amount of E and N components in EM fungi compared to that of AM fungi
131 constitutes novel empirical evidence in support of previous suggestions^{23,42,43} that per
132 fungal biomass units of EM fungi provide an important contribution to the soil pool of
133 intact or partially oxidized mycorrhizal fungal biomass. At the same time, the high
134 abundance of most easily decomposable W components supports empirical evidence of the
135 high rate decomposability of EM fungal mycelium during the initial stages of
136 decomposition^{44,45}.



137 Fig.3 Boxplots (a) Water-soluble components, $P < 0.001$, (b) Acid- hydrolysable components, $P < 0.001$, (c) Ethanol-
138 soluble components, $P < 0.001$, (d) Non-extractable components, $P < 0.05$. All P values refer to the Mann-Whitney test,
139 AM (n=4) and EM (n=11). Upper and lower limits of the box- quartiles around the weighted-mean, horizontal lines

140 *within boxes- weighted-median values within each mycorrhiza group, and red circles - weighted-mean of each*
141 *mycorrhiza group.*

142 We compared the magnitude of differences in recalcitrance of mycorrhizal fungal litter to
143 the magnitude of differences in recalcitrance of plant leaf litter (see Methods section for
144 details) which for decades has been recognized as the major factor controlling soil C
145 dynamics¹⁰⁻¹². Hereto we compared 4 plant functional types: evergreen trees, deciduous
146 trees, evergreen shrubs, herbaceous plants. The differences between AM and EM
147 mycorrhizal fungal species in chemical recalcitrance of litter, measured as the relative
148 abundance of WAEN components, was nearly twice as large as the differences in chemical
149 recalcitrance of litter between plant species of distinct functional types. For the particular
150 case of the water-soluble components, it was even three times higher (Table.1). The effect
151 sizes (η^2) of the difference between mycorrhizal fungal guilds for each individual WAEN
152 component were higher. This was particularly evident for the major components of W and
153 A (see Fig.1) in the mycelium, which comprised the majority of the variation in chemical
154 components. In contrast, a similar analysis conducted for plant species grouped into
155 functional types (for details see Methods section) showed that the effect size of chemical
156 differences of plant functional types was much smaller, and only acid-hydrolysable
157 components contributed to the major variation in the group. Taken together, this suggests
158 that the potential contribution to distinct pathways of C transformations differs markedly
159 between mycorrhizal fungal guilds, and that differences in the decomposability pathways
160 of mycorrhizal fungal material are even more striking than the differences observed in leaf
161 litter among plant functional types, till now considered as one of the most important factors
162 determining soil C circulation. Thus, chemical differences between mycorrhizal fungi
163 types might be essential underestimated sources of (variation in) below-ground soil C
164 dynamics.

165 *Table.1 Effect sizes (η^2 , one-way ANOVA) of chemical composition differences in leaf litter within plant functional*
166 *types and mycelial biomass within mycorrhizal fungal guilds*

	Water-soluble	Acid-extractable	Ethanol-soluble	Non-extractable
Mycorrhizal fungal guilds	0.68	0.76	0.46	0.26
Plant functional types	0.20	0.41	0.28	0.19

167 Our test of chemical recalcitrance of mycorrhizal mycelium biomass of multiple EM and
168 AM fungal species provides the first empirical evidence of the inherent difference between
169 AM and EM fungi in terms of their chemical composition related to the decomposition
170 pathway. Differences between these fungal guilds in decomposability support and
171 mechanistically underpin previous speculations that EM fungi might contain a higher ratio
172 of components recalcitrant to decomposition than AM fungi⁵. While microbiologists seek
173 to specify the fungus-specific macromolecular compounds and basic chemical elements,
174 these characters are difficult to link to soil C cycle mechanisms. Instead of analysing
175 individual chemical components or complex chemical compounds of fungal biomass that
176 are possibly a proxy for decomposability^{5,9,31}, we opted to characterize fungal biomass
177 through general traits of litter quality known to drive soil C cycling^{41,46,47}. Recently, it has

178 been suggested that labile and recalcitrant C compounds originating from decomposing
179 organic matter might follow distinct pathways of stabilization depending on the abundance
180 of soil saprotrophic organisms^{48,49}. This suggests that C components originating from
181 mycorrhizal fungi of distinct guilds are likely involved in distinct pathways of C
182 transformations in soil. Moreover, through the differential release of labile and more
183 recalcitrant C components, the temporal dynamics of contributions of different mycorrhizal
184 fungal guilds to distinct soil C transformation pathways will also differ among EM and
185 AM fungi.

186 For decades, foliage litter and its variability among species or plant functional types has
187 been considered as one of the main factor controlling soil C cycle process¹⁰. Our analysis
188 shows that the magnitude of differences in chemical decomposability traits between fungi
189 of distinct mycorrhizal guilds is much higher than that of the leaf litter of plant species
190 belonging to distinct functional groups. This suggests that the quality of mycorrhizal fungal
191 biomass is a critically important factor for pathways of soil C transformation processes.
192 Such pathways have been previously hypothesized^{5,24} but in practice neglected or
193 underestimated due to the high uncertainty associated with this phenomenon. Given that
194 plants allocate a significant part (up to 30%) of NPP to mycorrhizal fungal biomass^{20,22}, an
195 amount comparable to the allocation into plant leaves in some ecosystems^{50,51}, the
196 differential contributions of mycorrhizal fungal guilds to the processes of soil C turnover
197 should be considered as a critical SOM formation factor. As mycorrhizal fungal necromass
198 is among the most important sources of below ground soil C input, our results provide
199 chemical quality information of soil C inputs which is essential in narrowing down major
200 uncertainties in estimating soil C fluxes dynamics.

201 **Methods**

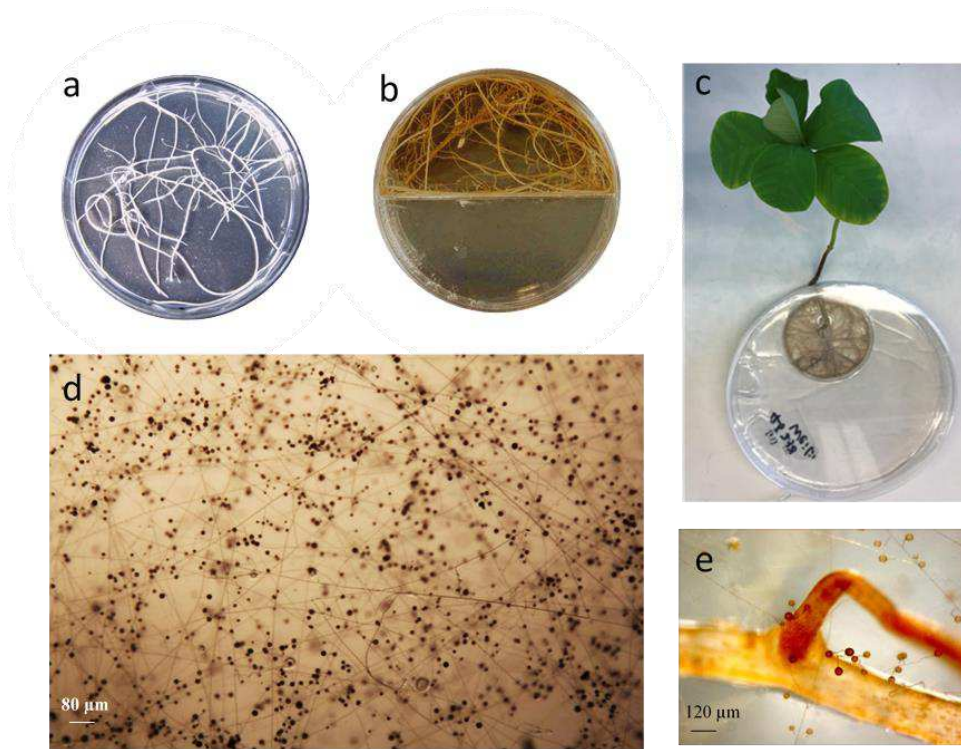
202 ***Cultivation of AM and EM fungi***

203 We selected available AM fungal species strains from the Glomeraceae family, as this
204 family is globally the most dominant family of AM fungi⁵², while they can be grown *in*
205 *vitro* producing reasonably large amounts of fungal biomass. We selected EM fungal
206 species to cover relatively abundant strains of various families. In addition, we opted to use
207 a higher number of EM fungi species compared to AM fungi, because (1) EM fungi consist
208 of ca. 20,000-25,000 species^{53,54} which entail high diversity of chemical traits, while AM
209 fungi have been known to exhibit lower diversity with ca. 300 species identified within this
210 fungal phylum⁵⁵⁻⁵⁷, (2) mass-production of AM fungi to reach the amounts of biomass
211 necessary for the recalcitrance assessments is complicated, necessitating hundreds of Petri
212 plates. Through cultivation, all manipulations were conducted under sterile conditions to
213 prevent contamination of fungal material, by using a laminar flow hood, and with sterile or
214 sterilized laboratory material.

215 ***EM fungi cultivation and sample preparation***

216 Original cultures of EM fungal species were obtained from Westerdijk Fungal Biodiversity
217 Institute (the Netherlands), which also provided standard laboratory instructions for EMF

218 cultivation (except strain of *Scleroderma verrucosum*, which was obtained from the
 219 collection of GINCO). Each species was inoculated in 30-80 Petri plates (90 mm,
 220 diameter), containing species-specific medium (Table.S1) solidified with bacteriological
 221 agar, then sealed with film and incubated in climate rooms (temperature 21-27°C according
 222 to the preference of each strain, in the dark) for 4-5 weeks (Fig.S1). Harvested fresh
 223 mycelium of EM fungi was washed with distilled water for 10 seconds, collected by
 224 filtration, and stored at -20°C. The frozen fungi biomass samples were dried using a freeze
 225 dryer or oven under 55°C for at least 12 hours (weighed after another 4 h until the weight
 226 is stable, drying methods see Table.S1), then stored at -20°C before chemical recalcitrance
 227 assessments.



228

229 *Fig.4 In vitro cultivation of mycorrhizal fungi arbuscular mycorrhizal fungi. Cultivation system in (a) mono-*
 230 *compartmented or (b) bi-compartmented Petri dish on excised Ri T-DNA transformed root organs of carrot. (c)*
 231 *Cultivation with the whole plant of Crotalaria Spectabilis in a bi-compartment Petri dish system; (d) Mycelium and spore*
 232 *production of the AM fungus (Rhizophagus irregularis, MUCL 41833) grown in a bi-compartmented Petri dish and (e) a*
 233 *mono-compartment Petri dish.*

234 **AM fungi cultivation and sample preparation**

235 All AM fungal strains were obtained from the Glomeromycota *in vitro* Collection (GINCO,
 236 Belgium). The cultivation protocol of AM fungi followed the methods well-established in
 237 the laboratory of mycology of UCLouvain (Belgium). As AM fungi are relatively slow-
 238 growing, and there was no prior knowledge on biomass output among *in vitro* cultivation
 239 approaches, each strain was cultivated using four different systems (Fig.4): autotrophic
 240 whole plants system either with a Petri (S1) or a mesh (S2) root compartment (RC),
 241 transformed root organ culture (ROC) system in bi- (S3) or mono-compartmented (S4)
 242 Petri plates. In the end, we established over 600 AM systems, all biomass produced by each

243 strain in the four different *in vitro* systems described below was needed to fulfil the
244 standard amount required for the chemical analysis.

245 System S1 consisted of a lid of a small 50-mm-diameter Petri dish placed inside a large
246 145-mm-diameter Petri dish, to create an RC inside a mycelial compartment (MC) (Fig.
247 4c). System S2 was similar to S1 with the difference that a 55-mm-diameter cap made with
248 40- μm nylon mesh and filled with cotton was used as RC instead (Fig.S2). The RC
249 contained roots of mycorrhizal plants to sustain fungal growth into the MC. Both
250 compartments were filled with modified Strullu–Romand (MSR⁵⁸) medium without
251 sucrose and vitamins. The large plates were covered with black plastic foil to minimize
252 light exposure. In each large plate, the plant shoot grew outside through a 2-mm-diameter
253 lateral opening sealed with sterile silicon grease as described⁵⁹. The systems were kept in
254 a growth chamber with a 16-hour photoperiod, 130 $\mu\text{mol m}^{-2} \text{s}^{-1}$ light intensity, 27 °C
255 temperature, and 80% relative humidity. Each RC was refilled with medium every 2-3
256 weeks.

257 The system S3 (Fig.4b) consisted of 94-mm-diameter bi-compartmented Petri plates with
258 RC and MC. The RC contained mycorrhizal Ri T-DNA transformed roots clone DC2 of
259 *Daucus carota* growing in MSR medium to sustain the fungal growth into the MC. The
260 MC was filled with MSR medium without sucrose and vitamins. These bi-compartmented
261 plates were incubated inverted in the dark at 27°C for 6 months; The system S4 (Fig.4a)
262 consisted of a 145-mm-diameter mono-compartmented Petri plate with mycorrhizal Ri T-
263 DNA transformed roots of *D.carota* clone DC2 growing in MSR medium⁵⁸. The plates
264 were incubated inverted in the dark at 27°C for 4-5 months.

265 For systems S1, S2, and S3, roots were trimmed before invading the MC to keep the MC
266 root-free. Once the MC was full with mycelium, the medium was harvested to extract the
267 mycelium as described below, and the MC was re-filled with medium to allow fungal re-
268 growth. The harvesting procedure was repeated for each plate every 4 to 6 weeks until
269 another 6-10 months according to the productivity of each plate; For system S4, only the
270 sections of the medium without any roots were harvested once to exclude roots and root
271 exudates after incubation.

272 The absence of roots in the harvested medium was carefully evaluated and confirmed using
273 a stereomicroscope. For all different systems and strains, the harvested medium which only
274 contained mycelium was immediately liquefied inside a beaker in a water bath at 70 °C for
275 2 hours - this procedure also killed the mycelium. The mycelium was then collected using
276 a 38- μm filter, washed with demineralised water for 10 seconds to remove any remnants
277 of medium and root exudates (only possibly exist in the harvested medium from S4), and
278 stored at -20°C until further use. Prior to chemical analyses, all mycelia was dried using
279 the same procedure as for the EM fungi described above (Table.S1).

280 ***Chemical recalcitrance analysis***

281 Mycelia were examined for the chemical composition of different recalcitrance in the
282 laboratory of Natural Resources Institute (Finland). Litter quality of fungal dead material

283 was determined by examining the samples for different chemical decomposable
284 substances: water-soluble (W), ethanol- or dichloromethane-soluble (E), acid-hydrolysable
285 (A) and non-extractable (N) components. These components are key to determining the
286 dynamics of litter decomposition and soil C cycling in soil C modelling^{41,46,60}. The
287 decomposability of each component was determined based on the Yasso soil C model^{41,46}
288 and followed the order W-A-E-N. The amounts of extractable substances were determined
289 gravimetrically by incubating samples with a solvent and weighing the samples after
290 filtration and drying. Mass loss during each extraction was considered to be equal to the
291 amount of a compound being extracted. For details of the protocols, see Ryan et al. (1990)
292 and Wieder and Starr (1998). The raw measurement results corresponding to Fig.1 are
293 provided in Table.S2. Samples that did not reach 0.5g were measured with a mixture of
294 plant litter (with a known content of WAEN components) to reach the necessary quantity
295 for analysis, and the chemical composition was calculated based on the proportion of
296 fungal biomass in the sample (details see Note.S1).

297 *Data process and statistics*

298 We assessed the significance of the overall differences in recalcitrance between AM and
299 EM fungi, with the permutational analysis of variance – perMANOVA⁶³, performed with
300 999 permutations in the Vegan package in R. WAEN values were log-transformed to meet
301 model assumptions. Dispersions of beta diversity (the distance from an individual measure
302 to the group's centroid) were calculated by each beta diversity metric within AM and EM
303 fungal groups for estimating within-group variation across individuals. Significant
304 differences in beta diversity variation⁶⁴ were tested using permutational statistical tests for
305 the homogeneity of group dispersions with 999 permutations in Vegan. We used Principal
306 Coordinates Analysis (PCoA) for visualization of the data present in the beta diversity
307 distance matrix (Fig.3).

308 We tested the hypothesis that AM fungi exhibit higher amounts of easily soluble, and acid-
309 hydrolysable compounds, while EM fungi have higher amounts of compounds that are
310 neither soluble nor hydrolysable, by a non-parametric Mann-Whitney U test to determine
311 if there were statistically significant differences in each chemical component between the
312 two mycorrhizal groups. To account for the fact that WAEN of two fungal strains were
313 assessed in a mixture with plant litter, all statistical analyses of fungal WAEN were
314 conducted as weighted analyses according to the accuracy assessment for the results of
315 sample from a mixture (details see Notes.S1).

316 We examined the magnitude of the difference between the recalcitrance of mycorrhizal
317 fungal types vs. the recalcitrance of plant material (Hypothesis 2) comparing the effect
318 size- Eta square (η^2)⁶⁵ of ANOVAs on WAEN values of AM vs EM fungi to the effect
319 sizes of ANOVAs on WAEN of plant functional types (other effect size indices are
320 provided in supplements, Notes.S2). To meet the normality assumptions, WAEN values
321 were log-transformed.

322 The data on plant functional types used for this analysis were obtained as follows: We
323 gathered plant leaf WAEN chemical composition data for 59 species from CIDET³² and

324 LIDET³³ datasets (details see Notes.S2). Those data were grouped into the evergreen tree,
325 deciduous tree, evergreen shrub and herb, based on plant growth form information from
326 the TRY database⁶⁶. Species with multiple form definitions were defined according to the
327 highest occurrence frequency⁶⁷.

328

329 Data availability

330 All data generated or analysed during this study are included in this article (and its
331 supplementary information files).

332

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340

341 Author contribution statements

342 W.H., P.B. and N.S. conceived the original idea and plan the project. S.D. and J.L. were
343 involved in planning the project. W.H. carried out the cultivation with assistance from S.D.
344 and M.C. on AM fungi. J.H. provided the sample measurements and processed the
345 experimental data. W.H. performed the numerical calculations and analysed the data. P.B,
346 N.S, J.H. and T.V. aided in interpreting the results. W.H. and N.S. wrote the manuscript
347 in consultation with P.B, S.D., M.C., J.H., T.V. and J.L.. All authors discussed the results
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349

350 Competing interests

351 The authors declare no competing interests.

352 Supplementary Information

353 Table.S1 Samples of AM and EM fungi with corresponding cultivation and drying details

354 Table.S2 Chemical compositions of AM and EM fungi samples

355 Fig.S1 Cultivations for EM fungi

356 Fig.S2 Schematic representation of two physically different autotrophic whole plant
357 systems (S1&S2)

358 Notes.S1 Accuracy assessment for results of litters measured with the mixture

359 Notes.S2 Plant litter data and statistics

360

361 References

- 362 1. Melillo, J. M. *et al.* Soil warming and carbon-cycle feedbacks to the climate system.
363 *Science (80-.)*. **298**, 2173–2176 (2002).
- 364 2. Stockmann, U. *et al.* The knowns, known unknowns and unknowns of sequestration of
365 soil organic carbon. *Agric. Ecosyst. Environ.* **164**, 80–99 (2013).

- 366 3. Sokol, N. W., Sanderman, J. & Bradford, M. A. Pathways of mineral-associated soil
367 organic matter formation: Integrating the role of plant carbon source, chemistry, and
368 point of entry. *Glob. Chang. Biol.* **25**, 12–24 (2019).
- 369 4. Krull, E. S., Baldock, J. A. & Skjemstad, J. O. Importance of mechanisms and processes of
370 the stabilisation of soil organic matter for modelling carbon turnover. *Funct. Plant Biol.*
371 **30**, 207–222 (2003).
- 372 5. Langley, J. A. & Hungate, B. A. Mycorrhizal controls on belowground litter quality.
373 *Ecology* **84**, 2302–2312 (2003).
- 374 6. Strickland, M. S., Osburn, E., Lauber, C., Fierer, N. & Bradford, M. A. Litter quality is in the
375 eye of the beholder: Initial decomposition rates as a function of inoculum characteristics.
376 *Funct. Ecol.* **23**, 627–636 (2009).
- 377 7. Couêteaux, M. M., Bottner, P. & Berg, B. Litter decomposition, climate and litter quality.
378 *Trends Ecol. Evol.* **10**, 63–66 (1995).
- 379 8. Prescott, C. E. Litter decomposition: What controls it and how can we alter it to
380 sequester more carbon in forest soils? *Biogeochemistry* **101**, 133–149 (2010).
- 381 9. Fernandez, C. W., Heckman, K., Kolka, R. & Kennedy, P. G. Melanin mitigates the
382 accelerated decay of mycorrhizal necromass with peatland warming. *Ecol. Lett.* **22**, 498–
383 505 (2019).
- 384 10. Brovkin, V. *et al.* Plant-driven variation in decomposition rates improves projections of
385 global litter stock distribution. *Biogeosciences* **9**, 565–576 (2012).
- 386 11. Aponte, C., García, L. V. & Marañón, T. Tree species effect on litter decomposition and
387 nutrient release in mediterranean oak forests changes over time. *Ecosystems* **15**, 1204–
388 1218 (2012).
- 389 12. Hättenschwiler, S. & Jørgensen, H. B. Carbon quality rather than stoichiometry controls
390 litter decomposition in a tropical rain forest. *J. Ecol.* **98**, 754–763 (2010).
- 391 13. Heijden, M. G. a Van Der *et al.* Mycorrhizal ecology and evolution: the past, the present,
392 and the future. *New Phytol.* **205**, 1406–1423 (2015).
- 393 14. Lin, G., McCormack, M. L., Ma, C. & Guo, D. Similar below-ground carbon cycling
394 dynamics but contrasting modes of nitrogen cycling between arbuscular mycorrhizal and
395 ectomycorrhizal forests. *New Phytol.* **213**, 1440–1451 (2017).
- 396 15. Högberg, M. N. & Högberg, P. Extramatrical ectomycorrhizal mycelium contributes one-
397 third of microbial biomass and produces, together with associated roots, half the
398 dissolved organic carbon in a forest soil. *New Phytol.* **154**, 791–795 (2002).
- 399 16. Leake, J. *et al.* Networks of power and influence: the role of mycorrhizal mycelium in
400 controlling plant communities and agroecosystem functioning. *Can. J. Bot.* **82**, 1016–
401 1045 (2004).
- 402 17. Bååth, E., Nilsson, L. O., Göransson, H. & Wallander, H. Can the extent of degradation of
403 soil fungal mycelium during soil incubation be used to estimate ectomycorrhizal biomass
404 in soil? *Soil Biol. Biochem.* **36**, 2105–2109 (2004).

- 405 18. Kaiser, C. *et al.* Exploring the transfer of recent plant photosynthates to soil microbes:
406 Mycorrhizal pathway vs direct root exudation. *New Phytol.* **205**, 1537–1551 (2015).
- 407 19. Konvalinková, T., Püschel, D., Řezáčová, V., Gryndlerová, H. & Jansa, J. Carbon flow from
408 plant to arbuscular mycorrhizal fungi is reduced under phosphorus fertilization. *Plant Soil*
409 **419**, 319–333 (2017).
- 410 20. Ouimette, A. P. *et al.* Accounting for carbon flux to mycorrhizal fungi may resolve
411 discrepancies in forest carbon budgets. *Ecosystems* **23**, 715–729 (2019).
- 412 21. Wallander, H., Nilsson, L. O., Hagerberg, D. & Bååth, E. Estimation of the biomass and
413 seasonal growth of external mycelium of ectomycorrhizal fungi in the field. *New Phytol.*
414 **151**, 753–760 (2001).
- 415 22. Allen, M. F. & Kitajima, K. Net primary production of ectomycorrhizas in a California
416 forest. *Fungal Ecol.* **10**, 81–90 (2014).
- 417 23. Godbold, D. L. *et al.* Mycorrhizal hyphal turnover as a dominant process for carbon input
418 into soil organic matter. *Plant Soil* **281**, 15–24 (2006).
- 419 24. Frey, S. D. Mycorrhizal fungi as mediators of soil organic matter dynamics. *Annu. Rev.*
420 *Ecol. Evol. Syst.* **50**, 237–259 (2019).
- 421 25. Brundrett, M. C. & Tedersoo, L. Evolutionary history of mycorrhizal symbioses and global
422 host plant diversity. *New Phytol.* **220**, 1108–1115 (2018).
- 423 26. Soudzilovskaia, N. A. *et al.* Global mycorrhizal plant distribution linked to terrestrial
424 carbon stocks. *Nat. Commun.* **10**, 1–10 (2019).
- 425 27. Phillips, R. P., Brzostek, E. & Midgley, M. G. The mycorrhizal-associated nutrient
426 economy : a new framework for predicting carbon – nutrient couplings in temperate
427 forests. *New Phytol.* 41–51 (2013).
- 428 28. Miyauchi, S. *et al.* Large-scale genome sequencing of mycorrhizal fungi provides insights
429 into the early evolution of symbiotic traits. *Nat. Commun.* **11**, 5125 (2020).
- 430 29. Harley, J. L. Fungi in Ecosystems. *J. Ecol.* **59**, 653 (1971).
- 431 30. Fernandez, C. W., Langley, J. A., Chapman, S., McCormack, M. L. & Koide, R. T. The
432 decomposition of ectomycorrhizal fungal necromass. *Soil Biology and Biochemistry* **93**,
433 38–49 (2016).
- 434 31. Fernandez, C. W. & Koide, R. T. Initial melanin and nitrogen concentrations control the
435 decomposition of ectomycorrhizal fungal litter. *Soil Biol. Biochem.* **77**, 150–157 (2014).
- 436 32. Trofymow, J. A. *The Canadian Institute Decomposition Experiment (CIDET) : project and*
437 *site establishment report / J.A. Trofymow and the CIDET Working Group.* (1998).
- 438 33. Gholz, H. L., Wedin, D. A., Smitherman, S. M., Harmon, M. E. & Parton, W. J. Long-term
439 dynamics of pine and hardwood litter in contrasting environments: Toward a global
440 model of decomposition. *Glob. Chang. Biol.* **6**, 751–765 (2000).
- 441 34. Kögel-Knabner, I. The macromolecular organic composition of plant and microbial
442 residues as inputs to soil organic matter. *Soil Biol. Biochem.* **34**, 139–162 (2002).

- 443 35. Zeglin, L. H. & Myrold, D. D. Fate of Decomposed Fungal Cell Wall Material in Organic
444 Horizons of Old-Growth Douglas-fir Forest Soils. *Soil Sci. Soc. Am. J.* **77**, 489–500 (2013).
- 445 36. Kleber, M. *et al.* Mineral-organic associations: formation, properties, and relevance in soil
446 environments. in *Advances in Agronomy* **130**, 1–140 (2015).
- 447 37. Fortin, J. A. *et al.* Arbuscular mycorrhiza on root-organ cultures. *Can. J. Bot.* **80**, 1–20
448 (2002).
- 449 38. Declerck, S., Séguin, S. & Dalpé, Y. The monoxenic culture of Arbuscular Mycorrhizal fungi
450 as a tool for germplasm collections. in *In Vitro Culture of Mycorrhizas* 17–30 (Springer-
451 Verlag, 2005).
- 452 39. Lalaymia, I. & Declerck, S. *The Mycorrhizal Donor Plant (MDP) in vitro culture system for*
453 *the efficient colonization of whole plants.* **2146**, (Springer US, 2020).
- 454 40. Crous, P. W., Verkley, G. J. M., Groenewald, J. Z. & Houbbraken, J. *Westerdijk Laboratory*
455 *Manual Series 1: Fungal Biodiversity.* (2019).
- 456 41. Tuomi, M. *et al.* Leaf litter decomposition-Estimates of global variability based on
457 Yasso07 model. *Ecol. Modell.* **220**, 3362–3371 (2009).
- 458 42. Clemmensen, K. E. *et al.* Carbon sequestration is related to mycorrhizal fungal
459 community shifts during long-term succession in boreal forests. *New Phytol.* **205**, 1525–
460 1536 (2015).
- 461 43. Averill, C., Turner, B. L. & Finzi, A. C. Mycorrhiza-mediated competition between plants
462 and decomposers drives soil carbon storage. *Nature* **505**, 543–5 (2014).
- 463 44. Staddon, P. L., Ramsey, C. B., Ostle, N., Ineson, P. & Fitter, A. H. Rapid turnover of hyphae
464 of mycorrhizal fungi determined by AMS microanalysis of ¹⁴C. *Science (80-.).* **300**, 1138–
465 1140 (2003).
- 466 45. Adamczyk, B., Sietiö, O., Biasi, C. & Heinonsalo, J. Interaction between tannins and fungal
467 necromass stabilizes fungal residues in boreal forest soils. *New Phytol.* **223**, 16–21 (2019).
- 468 46. Liski, J., Palosuo, T., Peltoniemi, M. & Sievänen, R. Carbon and decomposition model
469 Yasso for forest soils. *Ecol. Modell.* **189**, 168–182 (2005).
- 470 47. Guendehou, G. H. S. *et al.* Decomposition and changes in chemical composition of leaf
471 litter of five dominant tree species in a West African tropical forest. *Trop. Ecol.* **55**, 207–
472 220 (2014).
- 473 48. Paterson, E. *et al.* Labile and recalcitrant plant fractions are utilised by distinct microbial
474 communities in soil: Independent of the presence of roots and mycorrhizal fungi. *Soil*
475 *Biol. Biochem.* **40**, 1103–1113 (2008).
- 476 49. Cotrufo, M. F., Wallenstein, M. D., Boot, C. M., Deneff, K. & Paul, E. The Microbial
477 Efficiency-Matrix Stabilization (MEMS) framework integrates plant inputs form stable soil
478 organic matter? *Glob. Chang. Biol.* **19**, 988–995 (2013).
- 479 50. Xia, J. *et al.* Global patterns in Net Primary Production allocation regulated by
480 environmental conditions and forest stand age: a model-data comparison. *J. Geophys.*
481 *Res. Biogeosciences* **124**, 2039–2059 (2019).

- 482 51. Malhi, Y., Doughty, C. & Galbraith, D. The allocation of ecosystem net primary
483 productivity in tropical forests. *Philos. Trans. R. Soc. B Biol. Sci.* **366**, 3225–3245 (2011).
- 484 52. Davison, J. *et al.* Plant functional groups associate with distinct arbuscular mycorrhizal
485 fungal communities. *New Phytol.* **226**, 1117–1128 (2020).
- 486 53. Tedersoo, L., May, T. W. & Smith, M. E. Ectomycorrhizal lifestyle in fungi: global diversity,
487 distribution, and evolution of phylogenetic lineages. *Mycorrhiza* **20**, 217–263 (2010).
- 488 54. Rinaldi, A. C. ., Comandini, O. . & Kuyper, T. W. Ectomycorrhizal fungal diversity:
489 separating the wheat from the chaff. *Fungal Divers.* **33**, 1–45 (2008).
- 490 55. Krüger, M., Krüger, C., Walker, C., Stockinger, H. & Schüßler, A. Phylogenetic reference
491 data for systematics and phylotaxonomy of arbuscular mycorrhizal fungi from phylum to
492 species level. *New Phytol.* **193**, 970–984 (2012).
- 493 56. Lee, E.-H., Eo, J.-K., Ka, K.-H. & Eom, A.-H. Diversity of arbuscular mycorrhizal fungi and
494 their roles in ecosystems. *Mycobiology* **41**, 121–125 (2013).
- 495 57. Schüßler, A., Schwarzott, D. & Walker, C. A new fungal phylum, the Glomeromycota:
496 phylogeny and evolution. *Mycol. Res.* **105**, 1413–1421 (2001).
- 497 58. Declerck, S., Strullu, D. G. & Plenchette, C. Monoxenic culture of the intraradical forms of
498 *Glomus* sp. isolated from a tropical ecosystem: a proposed methodology for germplasm
499 collection. *Mycologia* **90**, 579 (1998).
- 500 59. Voets, L. *et al.* Extraradical mycelium network of arbuscular mycorrhizal fungi allows fast
501 colonization of seedlings under in vitro conditions. *Mycorrhiza* **19**, 347–356 (2009).
- 502 60. Viskari, T. *et al.* Soil carbon estimates by Yasso15 model improved with state data
503 assimilation. 1–22 (2020). doi:10.5194/gmd-2020-141
- 504 61. Ryan, M. G., Melillo, J. M. & Ricca, A. A comparison of methods for determining
505 proximate carbon fractions of forest litter. *Can. J. For. Res.* **20**, 166–171 (1990).
- 506 62. Wieder, R. K. & Starr, S. T. Quantitative determination of organic fractions in highly
507 organic, Sphagnum peat soils. *Commun. Soil Sci. Plant Anal.* **29**, 847–857 (1998).
- 508 63. Anderson, M. J. Permutational Multivariate Analysis of Variance (PERMANOVA). *Wiley*
509 *StatsRef Stat. Ref. Online* 1–15 (2017).
- 510 64. Anderson, M. J., Ellingsen, K. E. & McArdle, B. H. Multivariate dispersion as a measure of
511 beta diversity. *Ecol. Lett.* **9**, 683–693 (2006).
- 512 65. Tomczak, M. & Tomczak, E. The need to report effect size estimates revisited. An
513 overview of some recommended measures of effect size. *Trends Sport Sci.* **1**, 19–25
514 (2014).
- 515 66. Kattge, J. *et al.* TRY - a global database of plant traits. *Glob. Chang. Biol.* **17**, 2905–2935
516 (2011).
- 517 67. Engemann, K. *et al.* A plant growth form dataset for the New World. *Ecology* **97**, 3243
518 (2016).

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