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1 Black carbon sanitary burden in the Indo-Gangetic 2 plain: exposures, risks and mitigation

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11 ABSTRACT

A large discrepancy between simulated and observed black carbon (BC) surface concentrations over the densely populated Indo-Gangetic plain (IGP) has so far limited our ability to assess the magnitude of BC sanitary impacts in terms of population exposure, morbidity, and mortality. We evaluate these impacts using an integrated modeling framework, including a successfully predicted BC surface concentration in a high-resolution chemical transport model CHIMERE with observationally-constrained BC emissions, combined with consistent health functions for BC. Population exposure to BC is noteworthy, with more than 60 million people identified living over hotspots of BC concentration (wintertime mean $> 20 \mu\text{g m}^{-3}$). A fraction of 62% of the total cardiovascular diseases mortality (CVM) burden for the megacity is found attributable to wintertime BC exposure. The semi-urban area has 50% of the CVM burden attributable to BC exposure in the total population over the IGP. More than 400 thousand lives can potentially be saved from CVM annually, by implementing prioritized emission reduction from the combustion of domestic biofuel in the semi-urban area, and diesel oil in transportation and coal in thermal power plant and brick kiln industries in megacities.

BC aerosols are mainly emitted from incomplete combustion processes, whether it is from combustion engines in the automobile sector, residential burning of wood and coal, industrial power stations using heavy oil or coal, field burning of agricultural wastes, or forest and vegetation fires¹. Recent studies of the health effect of airborne particles provide evidence that BC is more robustly associated with health risks than bulk (non-speciated) particulate matter mass (PM_{2.5} or PM₁₀), thus suggesting that BC concentration is a better indicator to consider²⁻⁴. BC may not be itself a toxic component of fine particulate matter (PM), but it is known to be co-emitted with toxic compounds which are byproducts of incomplete combustion of fuel and can adhere to the BC fractal morphology. BC may thus operate as a universal carrier of a wide variety of chemicals of varying toxicity into the lungs. These chemical can then come in contact with the body's major defence cells and the systemic blood circulation³. Recent studies also indicate health risks (per unit aerosol mass) are as much as 7–8 times larger for BC than for undifferentiated PM^{2,5,6}. There is thus value in targeting specifically BC emission abatement.

Furthermore the spatial and temporal distribution of BC concentration is very heterogeneous in comparison to undifferentiated PM mass. This is attributable to a large number of local combustion sources of BC, while PM also include other species such as dust which comes in part from long-range transport over most part of the Indian subcontinent during summer season and western India during winter season⁷. This heterogeneity may lead to contrast in population exposure to BC compared to that to total PM mass. The World Health Organisation³ suggested that BC should serve as an additional indicator to PM_{2.5} to quantify human exposure to airborne pollution and assess health effects to such exposure, with consequences on how to evaluate the local or regional mitigation measures aimed to reduce population exposed to combustion aerosols.

The IGP region over the Indian subcontinent is known as one of the global hotspots of atmospheric pollutants with very large concentrations of BC aerosols in wintertime⁸. Over India, this region also has the largest population density, with areatypes, e.g., from rural to megacity (refer to Table s2). Epidemiological studies provide evidence of an association between short-term exposure to elevated ambient air pollution (such as that during wintertime) and a higher risk of acute cardiovascular diseases, with systemic oxidative stress induced by air pollution as a potential underlying mechanism^{4,5,9}.

To estimate the magnitude of BC attributable sanitary impacts concerning population exposure, morbidity, and mortality and targeting BC emission abatement for health benefits, a conceptual plan using an integrated modeling framework as implemented in the present study is shown in Figure 1. We obtain a spatially and temporally fine resolved gridded distribution

38 of surface BC concentration for the wintertime over the IGP through BC transport simulation in a chemical transport model
 39 (CHIMERE). Simulating atmospheric BC concentration is indeed necessary to provide information on BC aerosols over large
 40 areas where atmospheric measurements are sparse, particularly in rural and remote locations comprising of a low population
 41 density and experiencing relatively less anthropogenic pollution than urban or megacity sites.

42 Evaluation of BC simulations in CHIMERE over the IGP

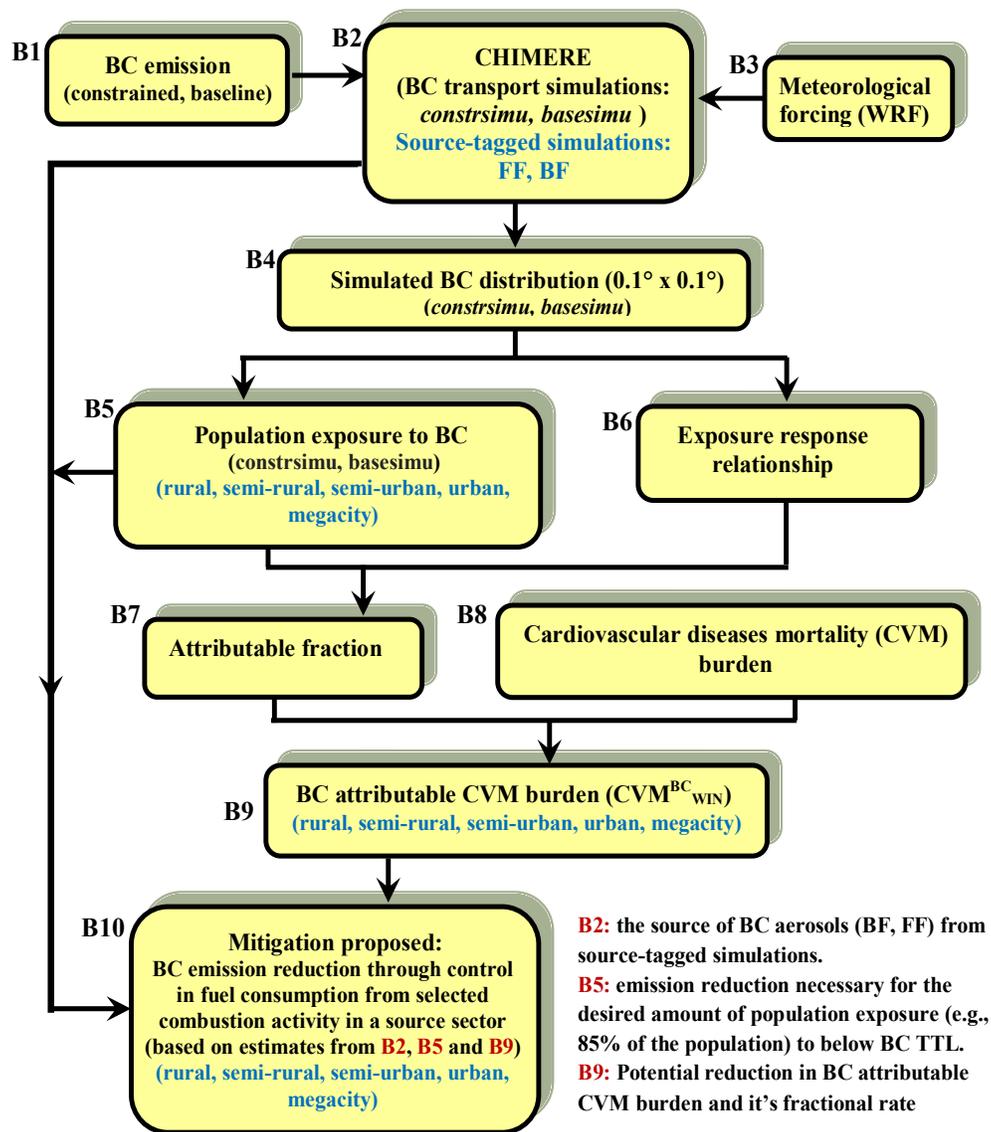


Figure 1. Flowsheet of conceptual plan to estimate the magnitude of BC attributable sanitary impacts and targeting sustainable BC emission reduction for health-benefits. The various blocks in the flowsheet are enumerated as B1, B2, B3...B10.

43 Two simulation experiments, *basesimu* and *constrsimu* are carried out by implementing, respectively, in CHIMERE the
 44 two emission datasets, namely, baseline BC emission and observationally-constrained BC emission (refer to ‘Methods’) to
 45 estimate population exposures under the scenario of simulated atmospheric BC concentration with low baseline BC emissions
 46 and that representing the present-day ambient values (close to observed) with observationally-constrained BC emissions. Spa-
 47 tial distribution ($0.1^\circ \times 0.1^\circ$) of BC surface concentration for all-day and daytime BC concentration averaged over the winter
 48 months (November-December) from the two simulation experiments (*constrsimu* and *basesimu*) and *constrsimu* are presented
 49 in Figures 2a–b and Figure 2c, respectively. In contrast to *basesimu* (Figure 2a), *constrsimu* (Figures 2b–c) is observed simu-
 50 lating adequately the BC concentration over the IGP. Patch of large wintertime BC pollution mostly comprising of wintertime

51 all-day (daytime) BC surface concentration in the range 14–25 (6–8) $\mu\text{g m}^{-3}$ is obtained from *constrsimu* estimates (Fig-
52 ures 2b–c) over the IGP. The feature of a large pollutant load for BC concentration over the IGP, as obtained from *constrsimu*
53 in the present study, is found to be consistent with that exhibited for aerosol optical depth as seen from satellite-based ob-
54 servations (Moderate Resolution Imaging Spectroradiometer, MODIS, e.g., refer to Figure 2e-h of¹⁰). A comparison of the
55 wintertime BC surface concentration (Figure 2f-g) from *constrsimu* estimates shows a good agreement with observations at
56 stations (refer to Figure 2a–c for stations understudy) over the IGP, with the ratio of simulated to measured winter averaged
57 all-day (daytime) concentration being near to one. The comparison from *basesimu* is also shown (Figure 2f-g) to examine the
58 amount of reduction in atmospheric BC concentration desired compared to the measured, thereby targeting abatement in BC
59 pollution (as discussed later) with the low baseline BC emissions. In this regard, BC surface concentration from *basesimu*
60 is accounted for 2 to 10 times lower than its measured counterparts. The normalised mean error (NME) in *constrsimu* is
61 16% (14)% for both the all-day (daytime) winter averaged BC concentration (refer to Table s1 as a supplement), which lies
62 within the uncertainty range in BC measurements (10%–20%). Unlike, the previous modeling studies which showed a large
63 discrepancy in the simulated BC distribution in global or regional chemical transport models, with the magnitude of simulated
64 surface BC concentrations being 4 to 9 times lower than observations, especially during winter months over the IGP^{7,11–13}.
65 Low values for the normalised mean error and root mean square error functions from *constrsimu* indicate that the combination
66 of the CHIMERE model and observationally-constrained BC emissions could simulate BC distribution over the IGP more
67 efficiently than reported in previous studies.

68 Population exposures to ambient BC concentration

69 In order to understand population exposures to BC (expressed as population number or relative percentage of population
70 corresponding to area-types in the IGP exposed to a given BC concentration), the spatial distribution of population density (at
71 a spatial resolution of $10 \times 10 \text{ km}^2$) with the overlay of contours of wintertime BC concentration for all-day and daytime mean
72 is presented in Figures 3a and 3b, respectively. The population over the IGP is mostly spread over the semi-urban area¹⁴ (refer
73 to Table s2), with more than 300 million people over the IGP including, respectively, more than 150, 80, 20, and 35 million
74 over the semi-urban, urban, megacities, and semirural areas being exposed to the wintertime all-day (daytime) mean BC
75 concentration of magnitude mostly higher than 10 (5) $\mu\text{g m}^{-3}$, which is twice (just equivalent to) the BC theoretical threshold
76 limit (TTL, refer to ‘Methods’). Even more than 30 (4) million people over the rural area are also found exposed to magnitude
77 of all-day (daytime) mean BC concentration just above the TTL. More than 60 million people live over hotspots of BC
78 concentration (white contours in Figures 3a and 3b) zone with a typically high wintertime all-day (daytime) mean BC surface
79 concentration larger than 20 (8) $\mu\text{g m}^{-3}$, i.e. comprising of values greater than 4 (1.6)-times the TTL, as identified prominently
80 over urban areas in the eastern and northern IGP, in addition, to that over semi-urban and a few semi-rural locations. Thereby,
81 indicating the alarming feature of BC pollutants present even for the rural/ semi-rural population belts, though, with a low
82 population density and that with a low BC emission rate (15-times and 2–4 times lower than, respectively, megacity and semi-
83 urban/urban belts, refer to Table s2). About 12 million people over the two megacities of the IGP (population density of $>$
84 5000 people per km^{-2}) are exposed to the extreme levels of all-day (daytime) wintertime mean of BC surface concentration (as
85 large as 40 $\mu\text{g m}^{-3}$ (12 $\mu\text{g m}^{-3}$)). The population-weighted averages (refer to Table s2) of the wintertime all-day (daytime)
86 BC surface concentrations for the urban belt and the megacity are larger by about 3-times (1.5–1.8) and 8-times (3-times),
87 respectively, than the rural belt.

88 BC concentrations from *constrsimu* and *basesimu* against the cumulative percentage of population distribution are pre-
89 sented in Figures 3c-h. These figures are shown considering both the overall population of the IGP (Figure 3c) and the pop-
90 ulations corresponding to area-types over the IGP (Figures 3d-h). The cumulative distribution pattern is generally observed
91 to exhibit an S-shaped curvature for all the cases examined, but a reverse curvature for the semi-urban, urban, and megacity
92 area-types with *constrsimu* all-day mean of BC concentration. More than 85% of the population over the IGP (Figure 3c) is
93 exposed to all-day mean BC concentration above the TTL, with about 36% being exposed to more than triple the TTL.

94 Analysis of the cumulative percentage population exposure for each of the area-types (Figures 3d-h) shows that more
95 than 90% ($>70\%$) of the semi-urban, urban population and megacity populations are exposed to all-day (daytime) mean BC
96 concentration above the TTL. The exposure is really alarming as this corresponds to 70% of semi-urban and urban population
97 being exposed to all-day mean BC concentration of value greater than 2–4 times the TTL, and specifically, that to 94% of the
98 megacity being greater than 4-times the TTL. The BC exposure level is distressing for the semi-rural and rural population as
99 well, with about 79% and 43%, respectively, of their population being exposed to all-day mean of BC surface concentration
100 above the TTL. The BC exposure for the population corresponding to these area-types, though, appears to be better considering
101 the daytime mean than the all-day mean, with about 90% of the rural and 70% of the semi-rural population being exposed
102 to daytime mean BC concentration of value within the TTL. The population exposure for all area types, except megacities,
103 is found to be relatively acceptable with the *basesimu* experiment, for which more than 85% of population is exposed to BC
104 concentration below the TTL.

105 Thus, the above analyses indicate the extent of reduction (by a factor of 8 or by 87%) in the present-day BC emissions (as
106 perceived from observationally-constrained BC emission data), which is required to make it equivalent to the low emission
107 scenario from the *basesimu*. This reduction is necessary to bring about 85% of the population exposure to ambient BC levels
108 of value below the TTL.

109 BC attributable cardiovascular diseases mortality (CVM) burden and potential mitigation measures

110 We further evaluate the health functions over the IGP as estimated in the present study (using formulation and assumption
111 given in ‘Methods’) for the cardiovascular diseases mortality (CVM) attributable to wintertime BC exposure. The spatial
112 distribution of relative risk factor (RR) (Figure 4a) indicates this being greater than one over the entire IGP, thereby indicating
113 wintertime BC exposures being a potential risk factor over the entire IGP, including semi-rural and rural areas. Notably, RR is
114 less than one over the rural area covering the parts of central India. The RR value is notably large (1.5 to 4) over the identified
115 BC hotspot locations, with this being the largest for the megacities, Delhi and Kolkata (3 to 4), thereby, indicating the highest
116 potential of BC exposure as a risk factor for the megacities. Spatial distribution of CVM attributed to the wintertime BC
117 exposure (CVM_{BC}^{WIN}) for each gridmesh ($10 \times 10 \text{ km}^2$ resolution), is presented in Figure 4b. Patches of a large burden of
118 CVM_{BC}^{WIN} (> 100 inhabitants mortality per gridmesh) is found spatially distributed over the BC hotspot locations, with the
119 highest burden (> 500 to 1000 inhabitants mortality per gridmesh) observed in and around megacities (Kolkata and Delhi).

120 The CVM_{BC}^{WIN} is estimated comprising of 400 thousand deaths annually over the entire IGP. Of the total CVM_{BC}^{WIN} over the
121 IGP, the largest fraction (49%) exists for the semi-urban area (as it has the largest population exposures to BC concentration
122 above the TTL, refer to Table s2), followed by the urban (28%) and megacity (14%), with the lowest being for the rural (2%),
123 followed by semi-rural (8%). Further, it is noted that CVM_{BC}^{WIN} is 22% of the total annual CVM over the IGP. In other words,
124 a fraction of 22% of the annual mortality due to cardiovascular diseases in the total population over the IGP is attributable
125 to wintertime BC exposure. This fraction when estimated corresponding to the classified domains or area-types (refer to
126 Table s2) is obtained as 62% for the megacity, which is significantly high compared to that as 30% for the urban, 20% for the
127 semi-urban, and 3% for the rural. The above analysis thus indicates an urgent requirement of the domain-wise sustainable
128 mitigation plan towards the reduction in the combustion PM (BC) to provide health benefits to inhabitants from BC pollution
129 exposure. The domain-wise targeted mitigation plan is suggested because the potential emission source of atmospheric BC
130 aerosols are distinctive for the domains (e.g., for the semi-urban and urban/megacity), as discussed later in the text.

131 The uncertainty in CVM_{BC}^{WIN} is estimated to be about 48% (based on uncertainty in β , simulated BC concentration and
132 mortality rate (y_0) as mentioned in ‘Methods’), which is mainly due to the lack of information on β value for the Indian
133 region. Epidemiological studies, including population exposed to various ambient levels of BC concentration, is required on a
134 regional basis to assess more accurately the CVM_{BC}^{WIN} . Nevertheless, the relative fraction (%) of CVM_{BC}^{WIN} and that its relative
135 distribution across the area-types over the IGP as estimated in the present study, is believed to be reasonably consistent.

136 In order to evaluate the domain-wise mitigation measures, we examine the source of BC aerosols over the IGP through
137 an analysis of source-tagged simulation (refer to Figures 2d-e). Wintertime BC concentration is primarily (50%–70%) from
138 emissions of biofuel combustion for cooking and heating in residential household sector over the eastern IGP, and that almost
139 equally (40–50% each) from biofuel combustion in residential and fossil fuel combustion in industrial and transportation
140 sectors over the central part of the IGP. On the other hand, wintertime BC concentration is mainly (50%–70%) from BC
141 emissions of FF combustion over the northern IGP (in and around Delhi and the hotspots) and that over the urban area and
142 hotspots of the eastern India (in and around Kolkata). The origin of ambient BC aerosols primarily from the FF combustion
143 over Delhi-National Capital Region as observed in the present study, is the reason for the failure to attain the desirable level
144 of ambient PM concentration in a recent study for highly polluted areas (e.g., Delhi) even when considering the hypothesized
145 sensitivity experiment using the complete abatement of residential BF combustion¹⁵. This further justifies the usage of BC
146 as a better indicator towards local or regional mitigation measures aimed at reducing population exposure to combustion PM,
147 specifically over the Indian subcontinent.

148 In principle, 100% or so mitigation of combustion activities contributing to BC aerosols is strongly recommended. How-
149 ever, for sustainable attainment of health benefits, we suggest stagewise mitigation measures, to achieve 87% reduction in BC
150 emissions, foremostly, which is required to bring more than 85% of the population exposed to the BC concentration below the
151 TTL (based on our analysis of population exposure as discussed in ‘Population exposures to ambient BC concentration’). The
152 strategic mitigation measures, conceptualised domain-wise, e.g., for the semi-urban, urban, and megacity (refer to Figure 3i),
153 based on prevalent emission sources of BC aerosol, and estimates of population exposures over the domain, in conjunction
154 with that on total number and fraction of population affected from CVM attributable to wintertime BC exposure (also refer to
155 Figure 1), is believed to be leading to sustainable achievement of potential health benefits. These measures, e.g. (i) mitiga-
156 tion of residential BF combustion activity over the entire IGP, preferentially, identifying the population using fuel wood and
157 cropwaste (based on emission composition of BF combustion source, refer to ‘Methods’) would lead to decreasing the CVM
158 burden to a wider extent, saving about 220 thousand lives from CVM_{BC}^{WIN} (benefitting a large number of semi-urban inhabitants

159 spread over a wide area of the IGP, including rural and semi-rural). The CVM burden would decrease at a large fractional rate
160 (saving 50 thousand lives from CVM_{BC}^{WIN} out of 80 thousand lives under CVM burden for megacity) and that at a wider extent,
161 saving 160 thousand lives from CVM_{BC}^{WIN} (out of total of 400 thousand deaths corresponding to CVM_{BC}^{WIN} over the entire IGP),
162 through, specifically targeting urban/megacity for the (ii) reduction of FF combustion sources, preferentially of diesel oil com-
163 bustion in transportation, and coal combustion in industries. As per the above suggested measures, to achieve 87% reduction in
164 the latest BC emissions (constrained) from FF combustion source for urban/megacity, amount of BC emissions required to be
165 reduced from diesel combustion in heavy duty vehicles, and that from coal combustion in thermal power plant (TPP) and brick
166 kiln industries is shown in Figure 3i. This reduction corresponds to a 7-fold decrease in the diesel consumption per km^2 for
167 urban and megacity, planned either by reducing the number of vehicles or curbing the amount of diesel consumed per vehicle
168 (e.g., through switching to hybrid technology measures) by 85% (i.e., $530 \text{ kg vehicle}^{-1} \text{ yr}^{-1}$ to $80 \text{ kg vehicle}^{-1} \text{ yr}^{-1}$). Further,
169 the above reduction corresponds to a 10-fold decrease in the total coal consumption in TPP sector, or coal combustion per
170 unit MW of electricity generated in TPP by 90% (i.e., $160 \text{ tonne MW}^{-1} \text{ yr}^{-1}$ to $16 \text{ tonne MW}^{-1} \text{ yr}^{-1}$). The above reduction
171 also corresponds to a 17- and 10-fold decrease in the total coal consumption in brick kilns sector, respectively, for megacity
172 and urban regions or that to 90% decrease in coal consumption per 1000 bricks produced in the brick kiln (i.e., 300 g to 33 g
173 of coal per 1000 bricks produced). To achieve 87% reduction in the present-day BC emissions (observationally-constrained)
174 from BF combustion source in the semi-urban area, amount of BC emissions required to be reduced from the combustion of
175 fuel wood, crop waste and dung cake in the residential household fuel use is shown in Figure 3i. This reduction corresponds
176 to a 10-fold decrease in the consumption of the BF per km^2 over the IGP.

177 The recently launched national mission, such as Ujjawal Bharat mission of the Government of India, which promotes and
178 aids providing clean energy combustion source preventing the residential coal and biomass fuels combustion for residential
179 cooking, is believed would be effective in reducing the mortality risks associated with the combustion PM over the IGP.
180 Regulatory measures are also required for curbing BC emissions from FF combustion source over the megacity and urban
181 areas as per the above suggested measures. We suggest promoting domain-wise mitigation plans based on estimating the
182 source-wise BC emission reduction credits and the exchange of credits among the potential source sectors which contribute to
183 BC emissions on a regional to national to global scale. Evaluation of these plans over the Indian region is under progress and
184 will be presented in a future study.

185 Conclusion

186 In the present study, we assess the sanitary impacts attributable to BC exposure in terms of population exposures and disease
187 burden (cardiovascular disease mortality, CVM) for the South Asia Indo-Gangetic Plain (IGP) using an integrated modeling
188 framework. The assessment includes using a simulated spatial distribution of BC surface concentration obtained from a high-
189 resolution chemical transport model (CHIMERE) and comparing favorably with observations. The efficacy of CHIMERE
190 to simulate the observed very large concentrations of BC aerosol during the wintertime with low values for the normalized
191 mean error and root mean square error functions, is due to the implementation of an observationally-constrained BC emission
192 dataset for the Indian region.

193 Population exposure to BC is remarkably high for more than three hundred million persons living in the IGP including 94%
194 of the megacity and 70% of semi-urban/urban inhabitants being exposed to an all-day mean BC concentration of magnitude
195 twice larger than the defined theoretical threshold limit (TTL) based on available epidemiological information ($5 \mu\text{g m}^{-3}$).
196 These are also inferred being distressing for rural and semi-rural populations, with about 79% and 43%, respectively, of these
197 populations estimated to be exposed to all-day wintertime mean of BC surface concentration above the TTL. The population-
198 weighted averages of the simulated wintertime BC concentration are larger by factors 4 and 6-8, respectively, for the urban
199 areas and megacities than the rural areas.

200 The spatial mapping of the relative risk factor and CVM burden attributable to BC exposure (CVM_{BC}^{WIN}) implies that all
201 populations living in the IGP, including those in rural areas, are under the risk of BC-attributable health effects. The CVM_{BC}^{WIN}
202 per $10 \times 10 \text{ km}^2$, including more than a hundred deaths over the BC hotspot locations and as large as thousand deaths for the
203 megacities of Kolkata and Delhi is identified in the IGP. The fraction of CVM_{BC}^{WIN} is 22% of the CVM in the total population
204 over the IGP. Among the different area types, this fraction is the largest for the two megacities (62%), or three times that for
205 semi-urban areas.

206 Based on our estimates of population exposures, an 87% reduction in present-day BC emissions is required to avoid
207 the adversity of BC-attributable sanitary impacts and save about 400 thousand lives annually from CVM over the IGP. Health
208 benefits from the sanitary impacts of BC are potentially attainable reducing the CVM burden by a wider extent, thereby, saving
209 about 220 thousand lives of semi-urban including rural population from CVM_{BC}^{WIN} through the preferential mitigation of the
210 residential fuelwood and crop-waste combustion. But, the CVM burden is estimated decreasing by a large fractional rate,
211 saving about 50 thousand lives of megacity population (out of the CVM burden of 80 thousand lives), and by a wider extent
212 saving about 160 thousand lives of urban/ megacity population, through the preferential reduction of diesel oil combustion

213 in transportation, and coal combustion in industries, specifically thermal power plants (TPP) and brick kilns. The domain-
214 wise mitigation plan comprising of the amount of mitigation of emissions required through the corresponding reduction
215 in fuel consumption for energy from identified combustion activity, including the total coal consumption in the industry
216 or consumption per unit MW of electricity generated in TPP, or per 1000 of bricks produced in brick kiln industry, the
217 diesel consumption per vehicle in the transportation, biofuel use per km² in the residential household sector is suggested and
218 presented in the study.

219 **Methods**

220 Available studies indicated that the most likely reasons for the discrepancy between model and observations include an un-
221 derestimation of BC emissions used as input, inadequate meteorology and representation of aerosol treatment, and too coarse
222 spatial resolution in the model^{7,11,16–19}. In a recent study²⁰, an improvement in model performance relative to observations
223 over the East and South Asia regions was shown with a high-resolution chemical transport model, revealing an overlap of
224 high BC exposure concentration and population density. Thereby, indicating that the inadequate BC distribution would lead to
225 the dilution of the population exposure and the reduction of population-weighted concentration. A better agreement between
226 model simulated and observed magnitude of atmospheric aerosol species was, however, delivered over the Indian region in a
227 recent study using the constrained aerosol simulation approach^{7,16}. Further, the constrained BC emissions were estimated over
228 the Indian region constraining the simulated BC concentration in a general circulation model (Laboratoire de Météorologie
229 Dynamique atmospheric General Circulation Model (LMDZT-GCM)) with the observed BC (observationally-constrained BC
230 emissions) by combining forward and receptor modelling approaches¹².

231 In the present study, we implement (Figure 1) the recently estimated spatially and temporally resolved observationally-
232 constrained BC emissions over the Indian region in a chemical transport model (CHIMERE). The efficacy of the high-
233 resolution (0.1°×0.1°) CHIMERE model is evaluated based on its adequacy to simulate the observed wintertime BC dis-
234 tribution, comparing model estimated value with observations at stations over the IGP. The validated model BC distributions
235 are further utilised to assess its implication to sanitary impacts of BC pollution estimating population exposures to ambient
236 BC aerosol, and disease burden (cardiovascular diseases mortality, CVM) attributed to wintertime BC exposure. Further, it is
237 also interesting to examine the prominent emission source of BC aerosols from source-tagged BC simulation and the extent
238 of reasonably sustainable mitigation of BC emissions required for potential health benefits to the population living in different
239 area-types over the IGP, e.g., from rural locations to megacities.

240 **Simulation of BC surface concentration**

241 Simulation of surface concentration of black carbon (BC) aerosol is carried out using two state-of-the-art regional deterministic
242 models: WRF v3 for the meteorology and CHIMERE v2014b for the chemistry-transport²¹. CHIMERE being in offline
243 mode, it is forced by WRF meteorological fields. The two models are running at the same fine grid horizontal resolution
244 of 10 × 10 km² and over the same domain of the IGP (20°N-31°N 75°E-89.4°E, refer to Figure 2a–c). The simulation is
245 performed for the year of 2015 with a spin-up time of 10-days and analysed in the present study averaged for the winter
246 months of November and December.

247 **The WRF meteorological model**

248 The WRF model is a state-of-the-art numerical weather forecast and atmospheric simulation system designed for both research
249 and operational applications (<http://www.mmm.ucar.edu/wrf/users/downloads.html>). The initial and boundary meteorological
250 conditions for WRF simulation are obtained from Global Forecast System (GFS)²² National Center for Environmental Pre-
251 diction - FINAL (NCEP-FNL) operational global analysis data at a spatial resolution of 1° × 1°. Meteorological fields are
252 extracted from WRF at the temporal resolution of one-hour and the horizontal resolution of 10 × 10 km² (same as that for
253 CHIMERE simulation). The characteristics of the land surface such as soil type, vegetation index, albedo or surface topogra-
254 phy are provided by the National Center for Atmospheric Research (NCAR) and available at <http://dss.ucar.edu/datasets/ds083.2/>.
255 The meteorological boundary conditions, including sea surface temperature, are updated every 6 hours. The physical scheme
256 option used for WRF simulation include the scheme for the microphysics²³, subgrid convection by the Grell-3D scheme²⁴, the
257 NOAA land surface module²⁵, the Yonsei University planetary boundary layer scheme²⁶, and the Rapid Radiative Transfer
258 Model (RRTMG) for radiative transfer²⁷.

259 **The CHIMERE chemistry-transport model**

260 CHIMERE is a regional chemistry-transport model dedicated to the simulation of tens of gaseous and aerosols species,²¹. In
261 this study the model version 2014b is used. The aerosol module in CHIMERE provides hourly concentrations of ten chemical
262 species: sulphates, nitrates, ammonium, primary organic and black carbon (OC and BC), secondary organic aerosols (SOA),
263 sea salt, natural and anthropogenic dust and water²⁸. Aerosols are represented with ten bins, with mean mass median diameters
264 ranging from 0.05 to 40 μm. Radiation and photolysis are calculated online using the FastJX module,²⁹. Horizontal transport

265 is solved with the van Leer scheme,³⁰ and with a time-step of 10 minutes. Vertically, mixing is diagnosed from vertical velocity
266 and diffusion follows a classical K_z approach²¹. Boundary layer height is diagnosed using the³¹ scheme and deep convection
267 fluxes are calculated using the³² scheme. The deposition module of the model includes both wet and dry deposition processes
268 for aerosols^{33,34}.

269 Boundary conditions are determined by monthly climatologies of global chemistry transport model LMDZT-GCM coupled
270 to the Interactions between Aerosols and Chemistry (INCA) model³⁵. Meteorological fields required by CHIMERE as input
271 (e.g., 3-D wind, air temperature, relative humidity) are provided from those simulated at an hourly rate in the WRF model.

272 **Anthropogenic emissions and simulation experiments**

273 In the present study, two simulation experiments are carried out with CHIMERE. These are done by implementing in
274 CHIMERE the two emission data sets (at a spatial resolution $0.25^\circ \times 0.25^\circ$ over India): (i) recently estimated constrained BC
275 emission (annual BC emission rate of 2534 Gg yr^{-1}) over the Indian region¹², this experiment is termed as “*constrsimu*”, (ii)
276 extracted baseline BC emission (annual BC emission rate of 388 Gg yr^{-1}) from those implemented for forward BC simulation
277 in the Laboratoire de Météorologie Dynamique atmospheric General Circulation Model (LMDZT-GCM) (Verma et al., 2011)
278 and which were obtained from the bottom-up BC emission data (of year 2000)^{36,37}, termed as “*basesimu*” experiment.

279 Besides the above-mentioned simulation experiments, a source-tagged BC simulation experiment is also carried out us-
280 ing constrained BC emissions in CHIMERE. This is done to examine the relative contribution of BC emission source to
281 atmospheric BC concentration and evaluate the source-based mitigation of emissions for area-types (e.g., megacity, urban,
282 semi-urban) over the IGP through the corresponding reduction in fuel consumption for energy from identified selected com-
283 bustion activity in a source sector (Figure 1). In the source-tagged BC simulation, the BC aerosol transport and atmospheric
284 processes are simulated, implementing BC emissions in CHIMERE for each of the source sectors, including residential biofuel
285 (BF) and fossil fuel (FF) combustion, with emissions outside that source sector being switched off.

286 For the BF and FF source sectors, on estimating the relative composition (%) of BC emission from various fuel combustion
287 activities for energy, it is seen that this composition in BF source sector comprises of BC emissions primarily from the
288 combustion of residential fuelwood (50%) followed by that of cropwaste (20%), dung cake and forest biomass (10% each)
289 for cooking and heating in the residential household sector. That for the FF source sector, this is mainly from the combustion
290 of diesel oil (59%) in transportation, followed by coal combustion (40%, comprising of 23% from brick kiln industry and
291 17% from the thermal power plant). The above estimation includes the merging of the information on emission composition
292 of source sector with the total constrained BC emissions, information on activity-wise BC emission factor as that used in
293 bottom-up emissions in *basesimu*, and that on fuel consumption for combustion activity for energy from the survey data of the
294 Government of India (GoI)^{38,39}, including the report from the Department of Agriculture and Cooperation, MoA, and Land
295 records and survey report, Government of West Bengal.

296 **Surface observations**

297 To evaluate the efficacy of the high-resolution CHIMERE model with the implemented observationally-constrained BC emis-
298 sions, BC surface concentrations from the *constrsimu* is compared with their observed counterparts obtained from available
299 studies at sites over the IGP (refer to Table s1). The comparison includes estimating the ratio between model and observed
300 BC (refer to Figures 2a–c)) and the statistical error, including normalized mean error (NME), and root mean square error
301 (RMSE) for wintertime mean values (refer to Table s1). The comparison is made using daytime, and all-day values averaged
302 over the wintertime. This is because the daytime mean BC concentration exhibits a low hourly variability and corresponds to
303 the well-mixed layer of atmosphere^{40,41}. Evaluation of model estimates for both daytime and all-day averages thus provides a
304 systematic approach to identify the model discrepancy, if primarily due to emissions or that due to physical/chemical processes
305 (which govern the atmospheric residence time of aerosols).

306 **Population exposure and BC attributable disease burden over the IGP**

307 To assess the BC attributable sanitary impacts, we estimate the population exposure to ambient BC aerosols and the burden
308 of disease (cardiovascular diseases mortality, CVM) attributed to wintertime BC exposure. For the spatial mapping of the
309 population exposure to ambient BC, the gridded distribution of BC surface concentration is overlaid on a population density
310 distribution map for the IGP region. The population density distribution is available at a spatial resolution of $10 \times 10 \text{ km}^2$
311 (same as that of the simulated BC concentration) from¹⁴. The spatial distribution of the wintertime mean of BC surface
312 concentration estimated over the IGP from the *constrsimu* is further used to calculate gridded distributions of the health
313 functions like the relative risk factor (RR), the attributable fraction (AF) and burden of disease attributed to wintertime BC
314 exposure. Relative risk (RR) factor is the ratio of the risk of an outcome such as disease or mortality in one group (the exposed
315 group) to that in any other group (generally the control group) (<http://www.medicalbiostatistics.com/rr-or-etc.pdf>). When the
316 RR associated with an agent or a circumstance (here BC exposure) is larger than 1, then the agent is called a risk factor. When
317 the relative risk associated with the agent is less than 1, then the agent is called a protective factor. If the relative risk equals

318 1, then agent is not associated with the outcome. Relative risk factor corresponding to BC exposure is calculated in each grid
 319 according to the following equation^{2,42}:

$$RR = e^{\beta \Delta x} \quad (1)$$

320 where β is the concentration response factor (CRF), which measures the increase in the health risk per unit increase in BC
 321 concentration (Δx in $\mu\text{g m}^{-3}$) beyond the theoretical threshold limit (TTL, defined later). In the present study, health functions
 322 are estimated for the cardiovascular diseases mortality (CVM, i.e. the number of people in a population who die as a result
 323 of cardiovascular diseases) attributable to wintertime BC exposure. This is because the value of β is estimated based on
 324 information from epidemiological studies providing evidence of an association between daily mortality (from cardiovascular
 325 disease) and BC particles for cooler months⁵. This value is obtained as 0.0214 per $\mu\text{g m}^{-3}$ of BC with an uncertainty of 43%
 326 which is nearly equivalent to that provided in². The focus to consider health functions associated to CVM is also consistent
 327 with the presence of a high burden of chronic respiratory diseases in India and that with the highest risk attributable to air
 328 pollution⁴³. It may be noted that β is based on the information of BC concentration exposure and response for CVM from
 329 epidemiological studies performed in the USA. Due to the lack of epidemiological information over the Indian region, and
 330 taking into account that the causes of CVM would be more comparable globally than the all-cause mortality^{44,45}, using β for
 331 CVM from the USA-based study is a useful compromise. The health risk estimates, as obtained from equation 1 for large
 332 BC exposures is, however, found to be consistent. As it is also noted that estimates of RR due to a given BC exposure over
 333 the IGP, using the logarithmic model, equation 1, conform well, when compared to estimates of RR (for chronic obstructive
 334 pulmonary disease, COPD) due to PM exposure using integrated exposure-response (IER) model⁴⁶, the formulation of which
 335 takes into account high PM exposures in ambient environments. The comparison with IER estimates is made corresponding
 336 to 10-times of respective BC concentration, considering health risk estimates (per unit mass) are 7–8 times larger for BC than
 337 PM.

338 In equation 1, Δx (spatially resolved at $10 \times 10 \text{ km}^2$) is estimated as:

$$\Delta x = C - C_0 \quad (2)$$

339 where C_0 is the theoretical threshold limit (TTL), that is the value below which there is no noticeable adverse health effects,
 340 assumed to be $5 \mu\text{g m}^{-3}$ for BC exposure, and C is the wintertime mean (daytime or all-day values averaged over the winter
 341 months) BC concentration to which a given population is exposed to, at the $10 \times 10 \text{ km}^2$ resolution and also expressed in
 342 $\mu\text{g m}^{-3}$. It may be noted that the air quality standards for PM are based on 24-hour mean mass concentration of PM measured
 343 as PM_{10} or $\text{PM}_{2.5}$ (PM with aerodynamic diameters $\leq 10 \mu\text{m}$ or $2.5 \mu\text{m}$), which are respectively 50 and $25 \mu\text{g m}^{-3}$ ³³. There
 344 is no air quality standards for BC concentration, but available studies consider health effect estimates corresponding to BC
 345 as for $1 \mu\text{g m}^{-3}$ of ambient BC concentration per $10 \mu\text{g m}^{-3}$ of PM_{10} ². The TTL for BC assumed is consistent with the
 346 above consideration (i.e. one-tenth of the air quality standard for 24-hour mean mass concentration of PM_{10}). Also, the TTL
 347 considered is consistent with the average ambient BC surface concentration during clean environmental conditions (southwest
 348 monsoon season, rainy season) from observations of BC measurement in an urban location (e.g. Kolkata), which is obtained
 349 as $5 \mu\text{g m}^{-3}$ ⁴¹. The TTL is also considered to be reasonable from the perspective that it is found to lie within the range of
 350 all-day mean BC concentration ($2\text{--}8 \mu\text{g m}^{-3}$) estimated over relatively less polluted rural area (refer to Figure 2b) during
 351 wintertime.

352 Further, the Attributable fraction (AF), defined as the fraction of excess CVM factor attributed to wintertime BC exposure,
 353 is estimated. The AF is calculated for each grid at the same resolution according to the following equation:

$$AF = \frac{RR - 1}{RR} \quad (3)$$

354 The CVM burden attributed to the wintertime BC exposure ($\text{CVM}_{\text{BC}}^{\text{WIN}}$) is then estimated as:

$$\text{CVM}_{\text{BC}}^{\text{WIN}} = y_0 \times AF \times PD \times a \quad (4)$$

355 where y_0 is the age-standardised mortality rate (deaths per 100,000 people in a year) due to cardiovascular diseases, AF is the
 356 Attributable fraction due to wintertime BC exposure, PD is the gridded population density in unit of 100,000 people per km^2
 357 and a is the area of the gridmesh (in km^2). The y_0 parameter is obtained from the global burden of cardiovascular diseases
 358 for the rural/semi-rural, urban/semi-urban, and megacity area-types over India as respectively 241, 360, and 414 deaths per

359 100,000 people in a year (2010) based on information from available studies^{47,48}. The y_0 used correspond to the mortality rate
360 due to cardiovascular diseases for rural type (Andhra Pradesh), urban/ semi-urban (Kerala), megacity (Mumbai) as presented
361 in above mentioned refereed studies. This estimation is shown on Figure 4b and also integrated for each of the classified
362 area-types over the IGP (Table s2). The mortality rate due to cardiovascular diseases for the entire India was inferred as 272
363 deaths per 100,000 people with an uncertainty of 5%^{47,48}.

364 References

- 365 1. Bond, T. C. *et al.* Bounding the role of black carbon in the climate system: A scientific assessment. *J. Geophys. Res.* **118**,
366 5380–5552 (2013).
- 367 2. Janssen, N. A. *et al.* Black carbon as an additional indicator of the adverse health effects of airborne particles compared
368 with pm10 and pm2.5. *Environ. Heal. Perspectives* **119**, 1691–1699 (2011).
- 369 3. Janssen, N. A. *et al.* *Health effects of black carbon* (WHO Regional Office for Europe, Denmark, 2012).
- 370 4. Li, W. *et al.* Short-term exposure to air pollution and biomarkers of oxidative stress: the Framingham Heart Study. *J. Am.*
371 *Heal. Assoc.* **5**, e002742 (2016).
- 372 5. Ostro, B., Feng, W.-Y., Broadwin, R., Green, S. & Lipsett, M. The effects of components of fine particulate air pollution
373 on mortality in California: results from CalFINE. *Environ. health perspectives* **115**, 13–19 (2007).
- 374 6. Smith, K. R. *et al.* Public health benefits of strategies to reduce greenhouse-gas emissions: health implications of short-
375 lived greenhouse pollutants. *The Lancet* **374**, 2091–2103 (2009).
- 376 7. Kumar, D. B., Verma, S., Boucher, O. & Wang, R. Constrained simulation of aerosol species
377 and sources during pre-monsoon season over the Indian subcontinent. *Atmos. Res.* **214**, 91–108, DOI:
378 <https://doi.org/10.1016/j.atmosres.2018.07.001> (2018).
- 379 8. Ramanathan, V. & Carmichael, G. Global and regional climate changes due to black carbon. *Nat. Geosci.* **1**, 221–227,
380 DOI: [10.1038/ngeo156](https://doi.org/10.1038/ngeo156) (2008).
- 381 9. Shah, A. S. *et al.* Short term exposure to air pollution and stroke: systematic review and meta-analysis. *bmj* **350**, h1295
382 (2015).
- 383 10. Pani, S. K. & Verma, S. Variability of winter and summertime aerosols over eastern India urban environment. *Atmos. Res.*
384 **137**, 112–124 (2014).
- 385 11. Pan, X. *et al.* A multi-model evaluation of aerosols over South Asia: common problems and possible causes. *Atmos.*
386 *Chem. Phys.* **15**, 5903–5928, DOI: [10.5194/acp-15-5903](https://doi.org/10.5194/acp-15-5903) (2015).
- 387 12. Verma, S., Reddy, D. M., Ghosh, S., Kumar, D. B. & Chowdhury, A. K. Estimates of spatially and temporally resolved
388 constrained black carbon emission over the Indian region using a strategic integrated modelling approach. *Atmospheric*
389 *Res.* **195**, 9–19, DOI: [10.1016/j.atmosres.2017.05.007](https://doi.org/10.1016/j.atmosres.2017.05.007) (2017).
- 390 13. Ajay, P., Pathak, B., Solmon, F., Bhuyan, P. K. & Giorgi, F. Obtaining best parameterization scheme of RegCM 4.4 for
391 aerosols and chemistry simulations over the CORDEX South Asia. *Clim. Dyn.* DOI: [10.1007/s00382-018-4587-3](https://doi.org/10.1007/s00382-018-4587-3) (2019).
- 392 14. Census. *Primary Census Abstracts* (Registrar General of India, Ministry of Home Affairs, Government of India, New
393 Delhi-110011, 2011).
- 394 15. Chowdhury, S. *et al.* Indian annual ambient air quality standard is achievable by completely mitigating emissions from
395 household sources. *PNAS* 1–6, DOI: [10.1073/pnas.1900888116](https://doi.org/10.1073/pnas.1900888116) (2019).
- 396 16. Santra, S. *et al.* Simulations of black carbon (BC) aerosol impact over Hindu Kush Himalayan sites: validation, sources,
397 and implications on glacier runoff. *Atmos. Chem. Phys.* **19**, 2441–2460, DOI: <https://doi.org/10.5194/acp-19-2441-2019>
398 (2019).
- 399 17. Wang, R. *et al.* Estimation of global black carbon direct radiative forcing and its uncertainty constrained by observations.
400 *J. Geophys. Res. Atmos.* **121**, 5948–5971, DOI: [10.1002/2015JD024326](https://doi.org/10.1002/2015JD024326) (2016).
- 401 18. Verma, S., Venkataraman, C. & Boucher, O. Attribution of aerosol radiative forcing over India during the win-
402 ter monsoon to emissions from source categories and geographical regions. *Atmos. Environ.* **45**, 4398–4407, DOI:
403 [10.1016/j.atmosenv.2011.05.048](https://doi.org/10.1016/j.atmosenv.2011.05.048) (2011).
- 404 19. Reddy, M. S. *et al.* GCM estimates of aerosol transport and radiative forcing during INDOEX. *J. Geophys. Res.* **109**,
405 DOI: [10.1029/2004JD004557](https://doi.org/10.1029/2004JD004557) (2004).

- 406 **20.** Wang, R. *et al.* Exposure to ambient black carbon derived from a unique inventory and high-resolution mode. *PNAS* **111**,
407 2459–2463 (2014).
- 408 **21.** Menut, L. *et al.* Chimere 2013: a model for regional atmospheric composition modelling. *Geosci. Model. Dev.* **6**, 981–
409 1028 (2013).
- 410 **22.** Kalnay, E., Lord, S. J. & McPherson, R. D. Maturity of operational numerical weather prediction: Medium range. *Bull.*
411 *Am. Meteorol. Soc.* **79**, 2753–2770 (1998).
- 412 **23.** Lin, Y. L., Farley, R. D. & Orville, H. D. Bulk parameterization of the snow field in a cloud model. *J. Climate Appl. Met.*
413 **22**, 1065–1092 (1983).
- 414 **24.** Grell, G. & Devenyi, A. D. A generalized approach to parameterizing convection combining ensemble and data assimila-
415 tion techniques. *Geophys. Res. Lett.* **29**, 1693, DOI: [10.1029/2002GL015311](https://doi.org/10.1029/2002GL015311) (2002).
- 416 **25.** Chen, F. & Dudhia, J. Coupling an Advanced Land-Surface Hydrology Model with the Penn State-NCAR MM5 Modeling
417 System. Part I: Model Implementation and Sensitivity. *Mon. Weather. Rev.* **129**, 569–585 (2001).
- 418 **26.** Hong, S. Y., Noh, Y. & Dudhia, J. A new vertical diffusion package with an explicit treatment of entrainment processes.
419 *Mon. Weather. Rev.* **134**, 2318–2341 (2006).
- 420 **27.** Mlawer, E. J., Taubman, S. J., Brown, P. D., Iacono, M. J. & Clough, S. A. Radiative transfer for inhomogeneous
421 atmospheres: RRTM, a validated correlated-k model for the longwave. *J. Geophys. Res.* **102**, 16663–16682 (1997).
- 422 **28.** Bessagnet, B. *et al.* Aerosol modeling with CHIMERE-preliminary evaluation at the continental scale. *Atmos. Environ.*
423 **38**, 2803–2817, DOI: [10.1016/j.atmosenv.2004.02.034](https://doi.org/10.1016/j.atmosenv.2004.02.034) (2004).
- 424 **29.** Wild, O., Zhu, X. & Prather, M. J. Fast-j: Accurate simulation of in- and below-cloud photolysis in tropospheric chemical
425 models. *J. Atmos. Chem.* **37**, 245–282 (2000).
- 426 **30.** van Leer, B. Towards the ultimate conservative difference scheme: IV. A new approach to numerical convection. *J.*
427 *Comput. Phys.* **23**, 276–299 (1979).
- 428 **31.** Troen, I. & Mahrt, L. A simple model of the atmospheric boundary layer: Sensitivity to surface evaporation. *Boundary-*
429 *Layer Meteorol.* **37**, 129–148 (1986).
- 430 **32.** Tiedtke, M. A comprehensive mass flux scheme for cumulus parameterization in large-scale models. *Mon. Weath. Rev.*
431 **107**, 1779–1800 (1989).
- 432 **33.** Wesely, M. Parameterization of Surface Resistances to Gaseous Dry Deposition in Regional-Scale Numerical Models.
433 *Atmos. Environ.* **23**, 1293–1304 (1989).
- 434 **34.** Zhang, L., Gong, S., Padro, J. & Barrie, L. A size-segregated particle dry deposition scheme for an atmospheric aerosol
435 module. *Atmospheric Environment* **35**(3), 549–560 (2001).
- 436 **35.** Szopa, S., Foret, G., Menut, L. & Cozic, A. Impact of large scale circulation on european summer surface ozone and
437 consequences for modelling forecast. *Atmospheric Environ.* **43**, 1189–1195 (2009).
- 438 **36.** Reddy, M. S. & Venkataraman, C. Inventory of aerosol and sulphur dioxide emissions from India: I - Fossil fuels
439 combustion. *Atmos. Environ.* **36**, 677–697 (2002a).
- 440 **37.** Reddy, M. S. & Venkataraman, C. Inventory of aerosol and sulphur dioxide emissions from India: II - Biomass combus-
441 tion. *Atmos. Environ.* **36**, 699–712 (2002b).
- 442 **38.** MoSPI. *Statistical Year Book* (Central Statistics Office, Ministry of Statistics and Programme Implementation, Govern-
443 ment of India, New Delhi, 2018).
- 444 **39.** MoSPI. *Energy statistics* (Central Statistics Office, Ministry of Statistics and Programme Implementation, Government
445 of India, New Delhi, 2018).
- 446 **40.** Verma, S., Pani, S. & Bhanja, S. Sources and radiative effects of wintertime black carbon aerosols in an urban atmosphere
447 in east india. *Chemosphere* **90**, 260–269, DOI: [10.1016/j.chemosphere.2012.06.063](https://doi.org/10.1016/j.chemosphere.2012.06.063) (2013).
- 448 **41.** Pani, S. K. & Verma, S. Variability of winter and summertime aerosols over eastern India urban environment. *Atmos. Res.*
449 **137**, 112–124 (2014).
- 450 **42.** Krewski, D. *et al.* Extended follow-up and spatial analysis of the american cancer society study linking particulate air
451 pollution and mortality. *Res. Rep. Heal. Eff. Inst.* **140** (2009).
- 452 **43.** Salvi, S. *et al.* The burden of chronic respiratory diseases and their heterogeneity across the states of India: the Global
453 Burden of Disease Study 1990–2016. *Lancet Glob. Heal.* **6**, e1363–e1374 (2018).

- 454 **44.** Anenberg, S. C. *et al.* Impacts of global and regional and sectoral black carbon emission reductions on surface air quality
455 and human mortality. *Atmos. Chem. Phys.* **11**, 7253–7267 (2011).
- 456 **45.** Lelieveld, J., Evans, J. S., Fnais, M., Giannadaki, D. & Pozzer, A. The contribution of outdoor air pollution sources to
457 premature mortality on a global scale. *Nature* **525**, 367 (2015).
- 458 **46.** Burnett, R. T. *et al.* An integrated risk function for estimating the global burden of disease attributable to ambient fine
459 particulate matter exposure. *Environ. Heal. Perspectives* **122**, 397–403, DOI: <https://doi.org/10.1289/ehp.1307049> (2014).
- 460 **47.** Prabhakaran, D., Jeemon, P. & Roy, A. Cardiovascular diseases in India: current epidemiology and future directions.
461 *Circulation* **133**, 1605–1620 (2016).
- 462 **48.** Prabhakaran, D. *et al.* The changing patterns of cardiovascular diseases and their risk factors in the states of India: the
463 Global Burden of Disease Study 1990–2016. *The Lancet Glob. Heal.* **6**, e1339–e1351 (2018).

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468
469 *Data availability.* The data in this study are available from the corresponding author upon request (shubha@iitkgp.ac.in).
470

471 *Author contributions.* SV planned and coordinated the study and wrote the paper. SG conducted the BC transport simu-
472 lations, performed calculations of health functions, emission reduction, and produced figures. OB advised throughout the
473 paper’s preparation, analysis of results, and contributed to writing and completion of this paper. RW contributed to the writing
474 of the paper. LM advised and checked for the technicality of the CHIMERE model configuration and contributed to the paper.
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476 paper. All authors contributed to the manuscript.

477
478 *Competing interests.* The authors declare that they have no conflict of interest.

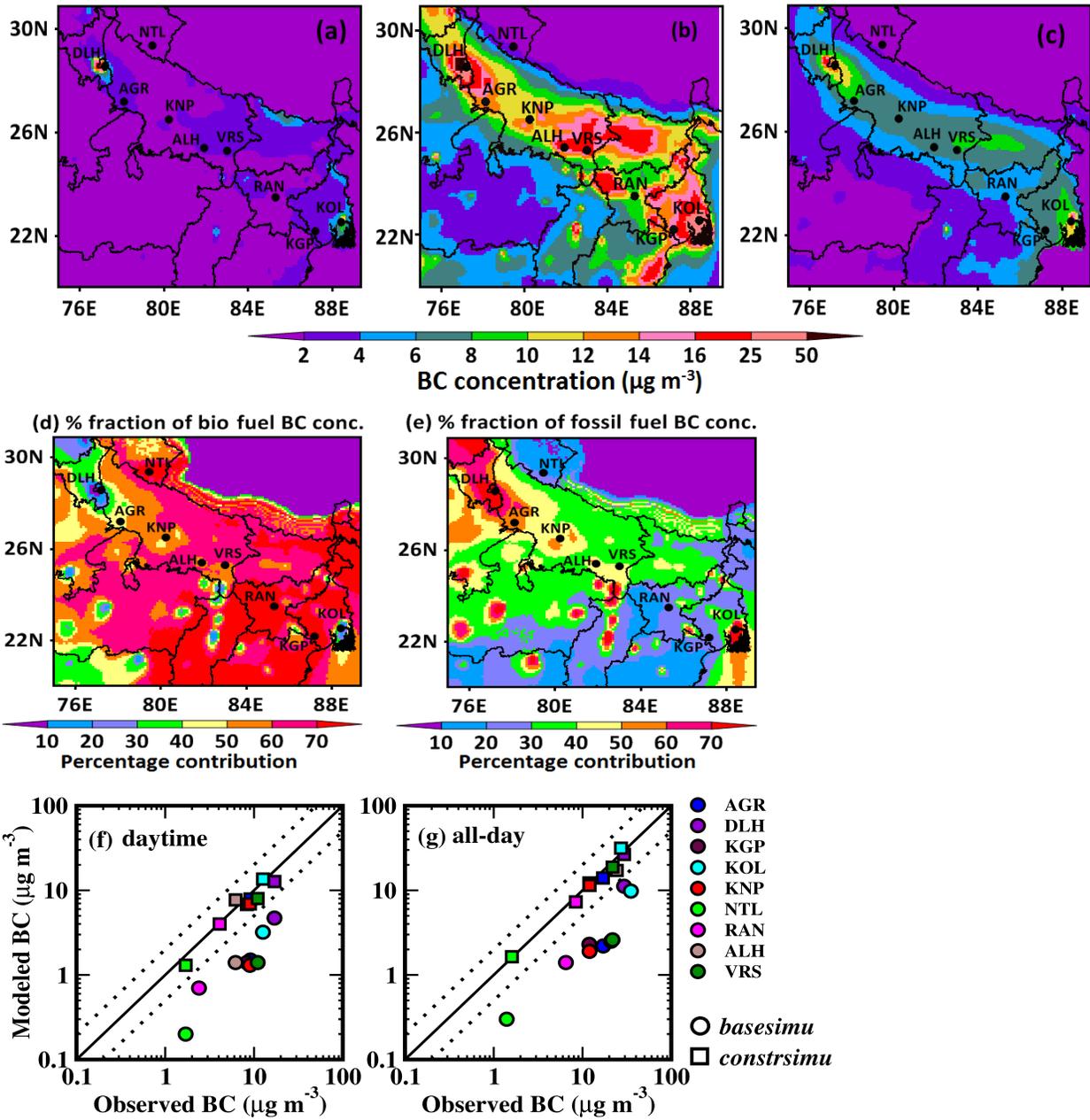


Figure 2. Spatial distribution of the simulated BC surface concentration ($\mu\text{g m}^{-3}$) averaged over the wintertime for (a) the *basesimu* experiment using baseline BC emission for all-day values and (b, c) the *constrsimu* experiment using constrained BC emission for (b) all-day and (c) daytime values. Fraction (%) of simulated BC concentration for *constrsimu* from (d) biofuel and (e) fossil-fuel emissions; (f–g) comparison of (f) daytime and (g) all-day wintertime mean of surface BC concentrations ($\mu\text{g m}^{-3}$) from simulations with measurements at stations over the IGP from available studies (refer to Table s1): Agra (Agr), Delhi (DEL), Kharagpur (KGP), Kolkata (KOL), Kanpur (KNP), Nainital (NTL), Ranchi (RAN), Allahabad/Prayagraj (ALH), Varanasi (VRS).

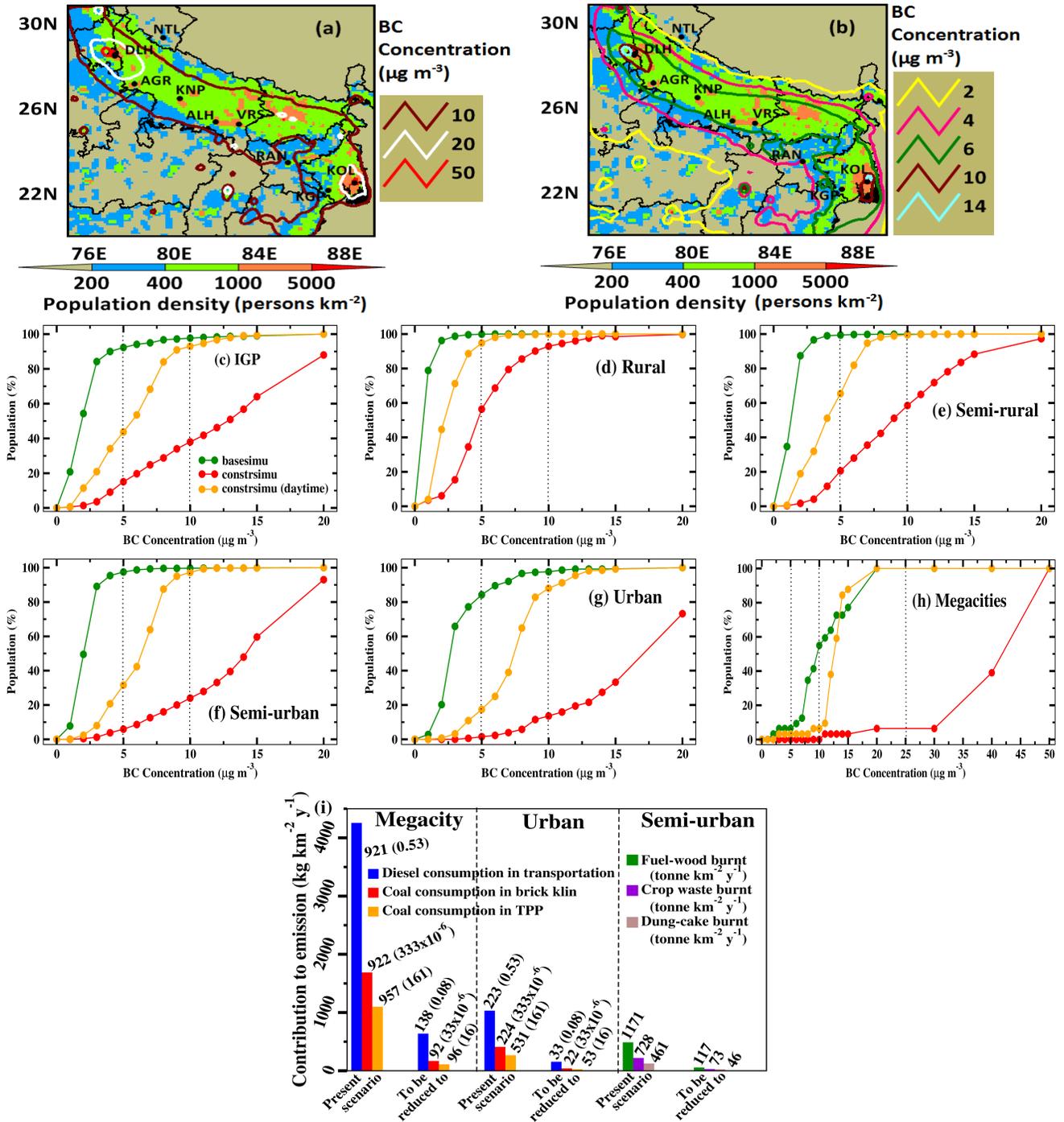


Figure 3. Spatial distribution of population density (number of inhabitants per km^2) overlaid with contours of BC concentration for wintertime (a) all-day mean of BC concentration and (b) daytime mean of BC concentration. Cumulative population exposure (%) to BC concentration from *constrsimu* and *basesimu* for the (c) entire IGP, and area-types of (d) rural (e) semi-rural, (f) semi-urban (g) urban and (h) megacity. (i) Domain-wise (megacity, urban, semi-urban) sustainable mitigation plan comprising of the amount of mitigation of emissions required (x – axis) from the present scenario (*constrsimu*) to the desired level (*basesimu*) for health-benefits from BC sanitary impacts through the corresponding reduction in fuel consumption for energy from identified combustion activity in a source sector. The y – axis shows BC emission (with unit as $\text{kg km}^{-2} \text{yr}^{-1}$) arising from fuel consumption for energy (given as number on bars with unit as $\text{tonne km}^{-2} \text{yr}^{-1}$). The number inside the bracket shown on bars represent tonne yr^{-1} per heavy vehicle for diesel consumption, tonne yr^{-1} per 1000 bricks produced for coal consumption in brick kiln industry, tonne yr^{-1} per unit MW of electricity for coal consumption in TPP). The fuel consumption shown corresponds to the one which needs reduction on priority for health benefits to the population corresponding to an area-type.

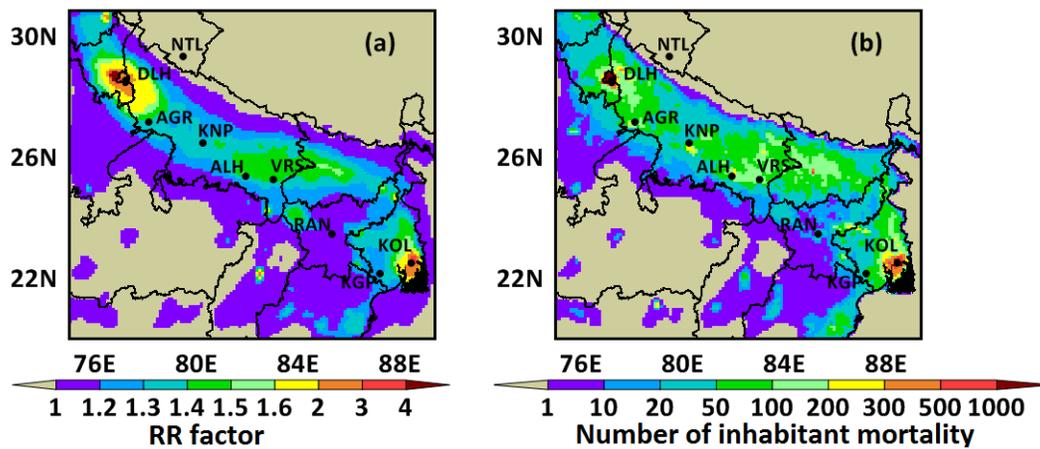


Figure 4. Spatial distribution ($10 \times 10 \text{ km}^2$ resolution) of health impact functions to assess BC attributable sanitary impacts for (a) Relative Risk factor, and (b) cardiovascular diseases mortality attributed to the wintertime BC exposure ($\text{CVM}_{\text{BC}}^{\text{WIN}}$, shown as number of inhabitants mortality per gridmesh on color scale).

Figures

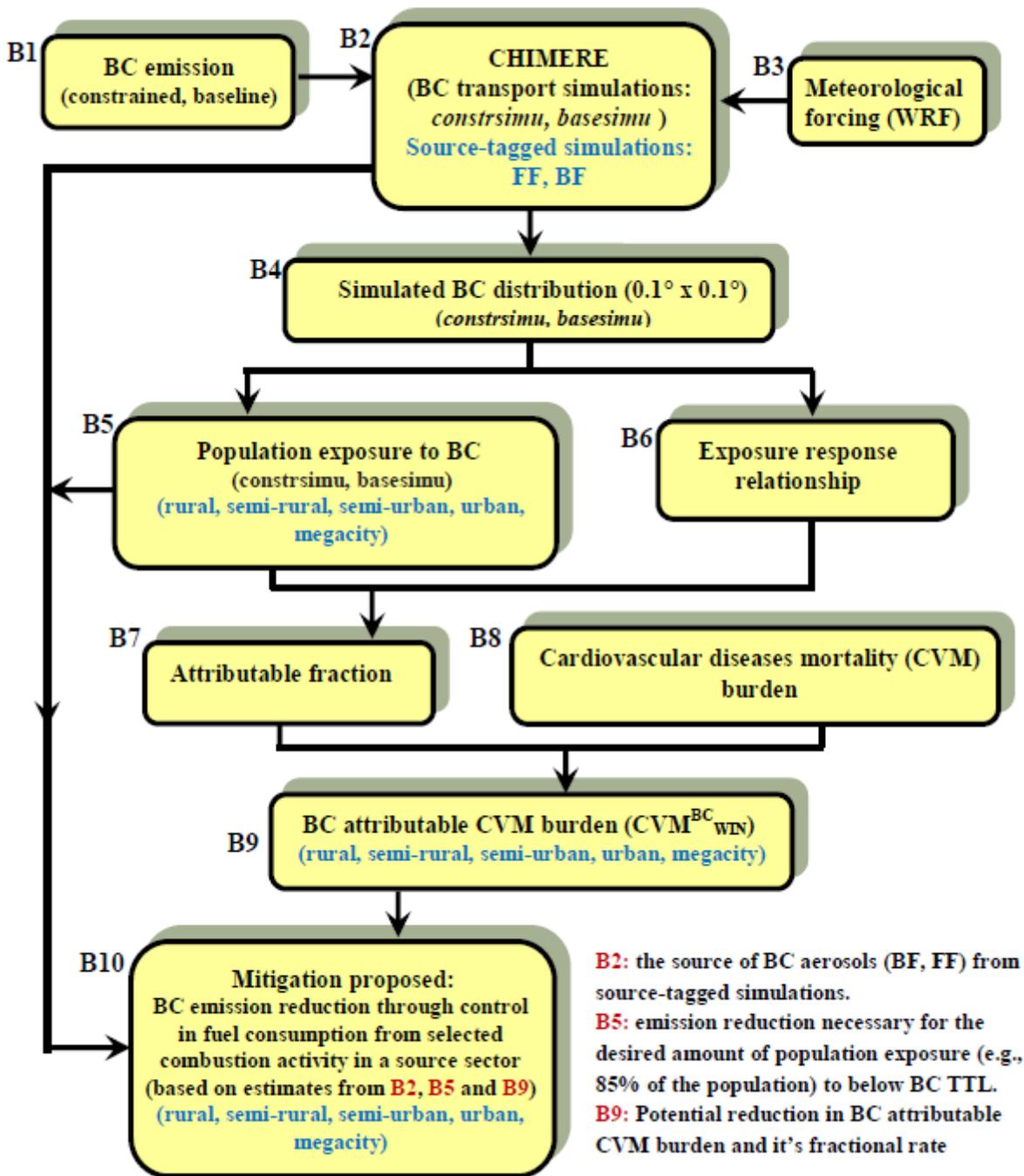


Figure 1

Flowsheet of conceptual plan to estimate the magnitude of BC attributable sanitary impacts and targeting sustainable BC emission reduction for health-benefits. The various blocks in the flowsheet are enumerated as B1, B2, B3...B10.

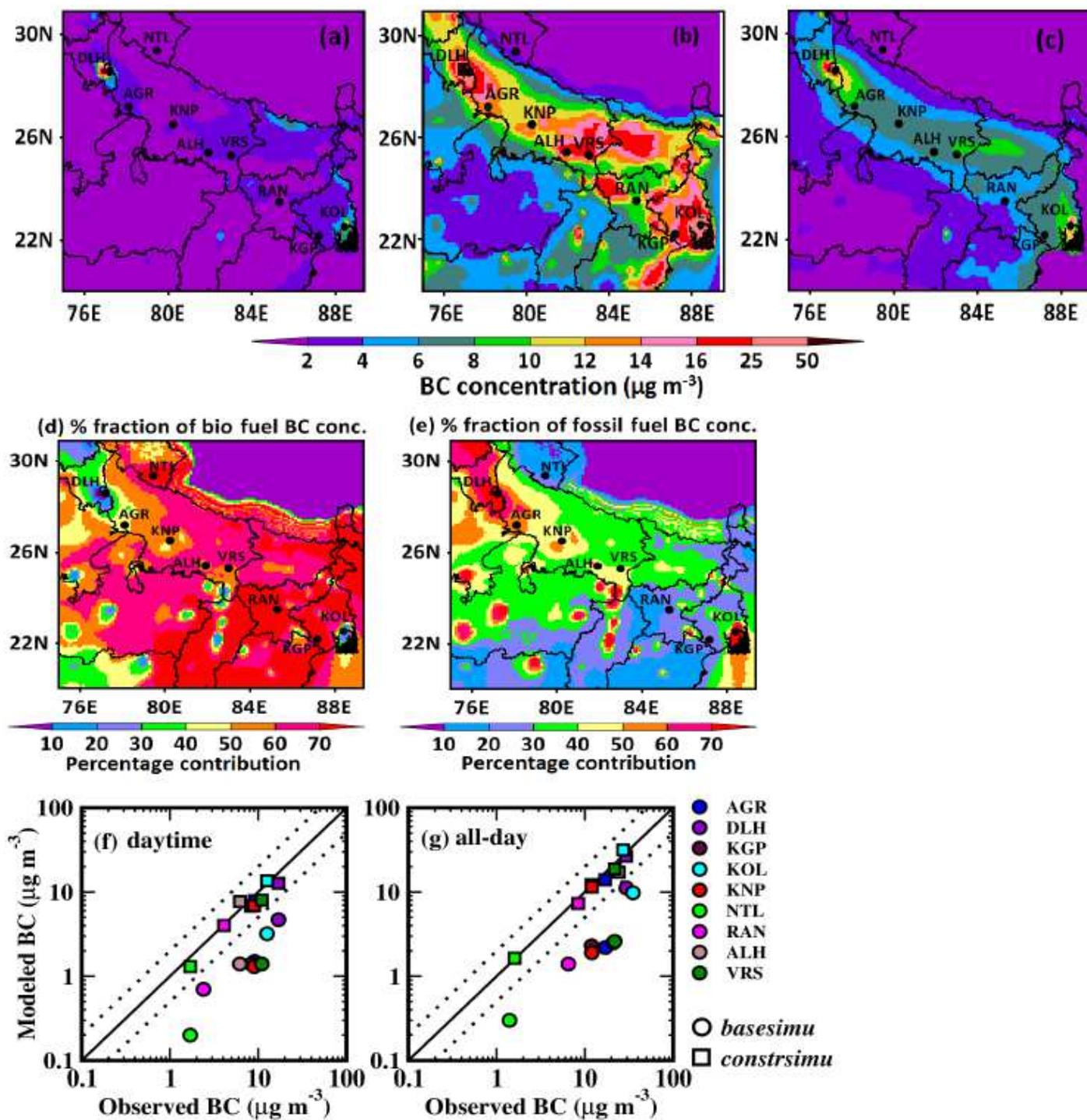


Figure 2

Spatial distribution of the simulated BC surface concentration ($\mu\text{g m}^{-3}$) averaged over the wintertime for (a) the *basesimu* experiment using baseline BC emission for all-day values and (b, c) the *constrsimu* experiment using constrained BC emission for (b) all-day and (c) daytime values. Fraction (%) of simulated BC concentration for *constrsimu* from (d) biofuel and (e) fossil-fuel emissions; (f-g) comparison of (f) daytime and (g) all-day wintertime mean of surface BC concentrations ($\mu\text{g m}^{-3}$) from simulations with measurements at stations over the IGP from available studies (refer to Table s1): Agra

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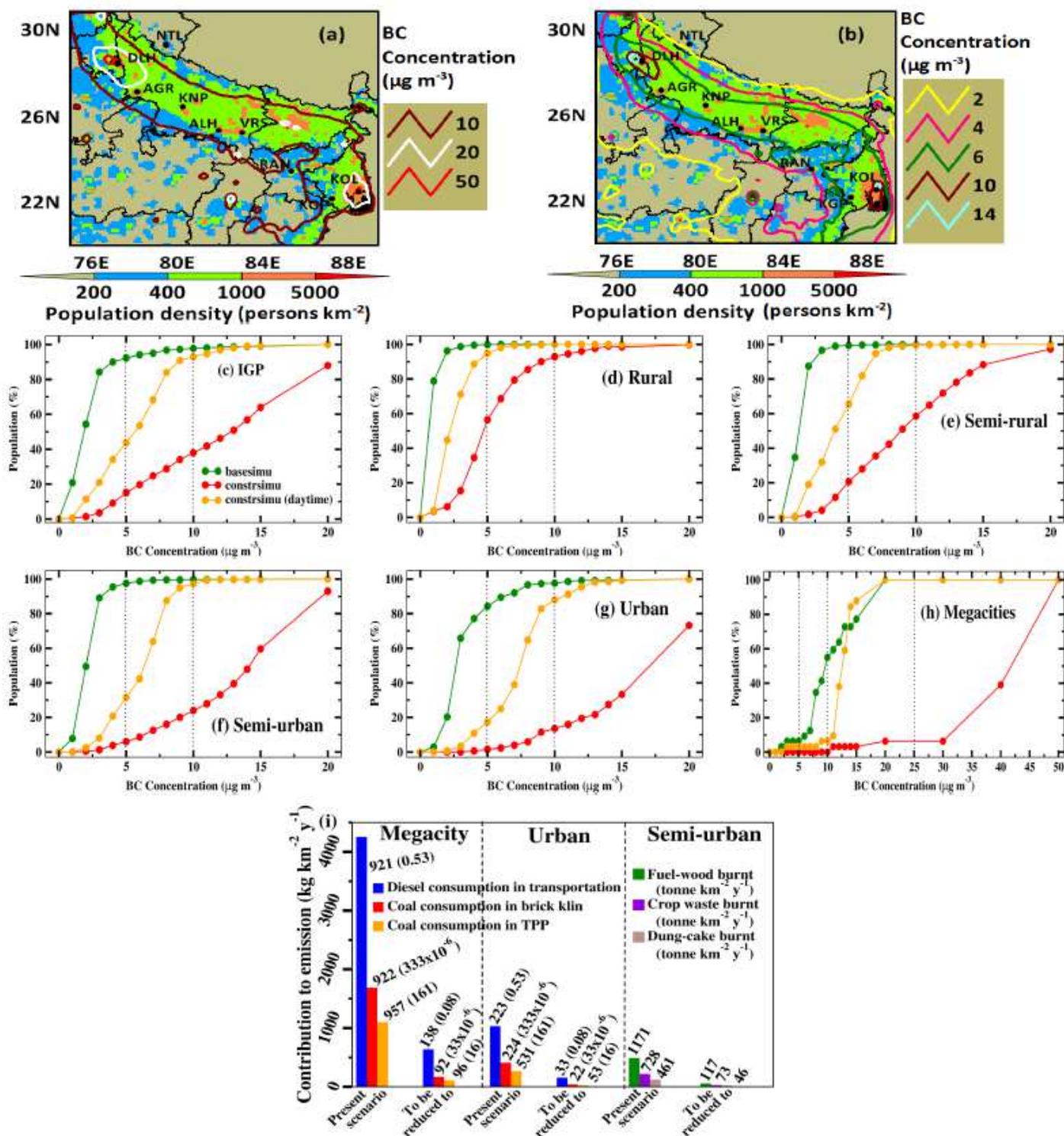


Figure 3

Spatial distribution of population density (number of inhabitants per km²) overlaid with contours of BC concentration for wintertime (a) all-day mean of BC concentration and (b) daytime mean of BC concentration. Cumulative population exposure (%) to BC concentration from constrsimu and basesimu

for the (c) entire IGP, and area-types of (d) rural (e) semi-rural, (f) semi-urban (g) urban and (h) megacity. (i) Domain-wise (megacity, urban, semi-urban) sustainable mitigation plan comprising of the amount of mitigation of emissions required (x-axis) from the present scenario (constrsimu) to the desired level (basesimu) for health-benefits from BC sanitary impacts through the corresponding reduction in fuel consumption for energy from identified combustion activity in a source sector. The y-axis shows BC emission (with unit as $\text{kg km}^{-2} \text{ yr}^{-1}$) arising from fuel consumption for energy (given as number on bars with unit as $\text{tonne km}^{-2} \text{ yr}^{-1}$). The number inside the bracket shown on bars represent tonne yr^{-1} per heavy vehicle for diesel consumption, tonne yr^{-1} per 1000 bricks produced for coal consumption in brick kiln industry, tonne yr^{-1} per unit MW of electricity for coal consumption in TPP). The fuel consumption shown corresponds to the one which needs reduction on priority for health benefits to the population corresponding to an area-type.

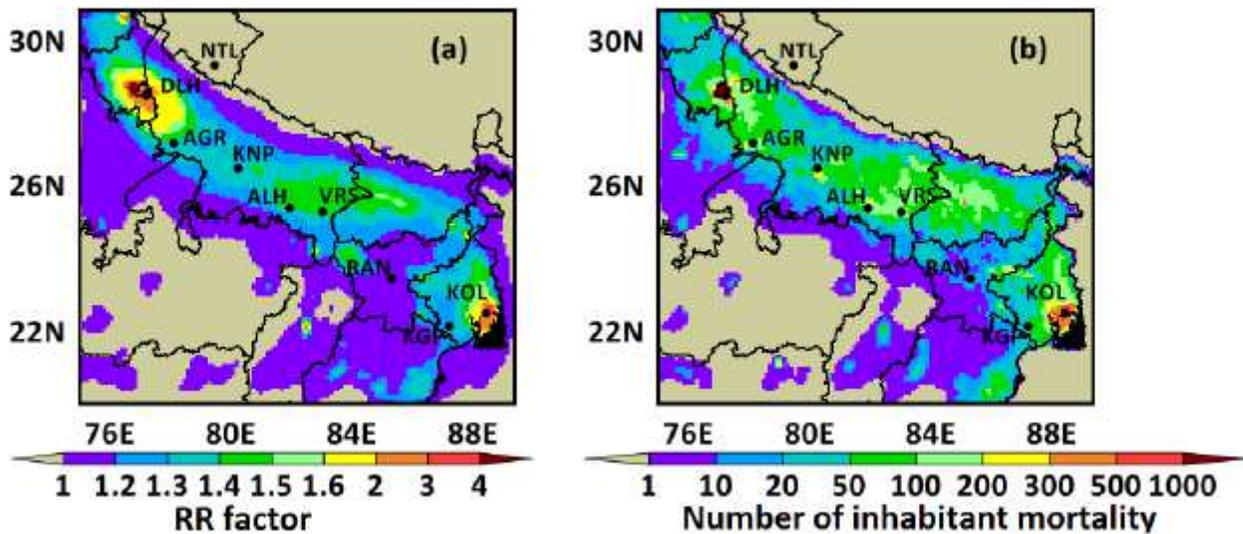


Figure 4

Spatial distribution ($10 \times 10 \text{ km}^2$ resolution) of health impact functions to assess BC attributable sanitary impacts for (a) Relative Risk factor, and (b) cardiovascular diseases mortality attributed to the wintertime BC exposure (CVMWIN BC, shown as number of inhabitants mortality per gridmesh on color scale).

Supplementary Files

This is a list of supplementary files associated with this preprint. Click to download.

- [supplementBCHealthimpact.pdf](#)